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Pretreatment methods for monitoring microplastics in soil and freshwater sediment samples: A comprehensive review

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Abstract

This paper reviews the currently used pretreatment methods for microplastics (MPs) analysis in soil and freshwater sediments, primarily sample processing, pretreatment, and characterization methods for MPs analysis. In addition, analytical tools (e.g., lab instruments), MPs characteristics, and MPs quantity, are included in this review. Prior to pretreatment, soil and sediment samples are typically processed using sieving and drying methods, and a sample quantity of <50 g was mostly used for the pretreatment. Density separation was commonly performed before organic matter removal. Sodium chloride (NaCl) and zinc chloride (ZnCl₂) were most often used for density separation, and hydrogen peroxide (H₂O₂) oxidation was most frequently used to remove organic matter. Although advantages of each pretreatment method have been investigated, it is still challenging to determine a universal pretreatment method due to sample variability (e.g., sample

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Appendix A. Supplementary data

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characteristics). Furthermore, it is highly required to establish standard pretreatment methods that can be used for various environmental matrices, including air, water, and wastes as well as soil and sediment.

Graphical Abstract



Keywords

Environment; Microplastics; Pretreatment methods; Soil; Freshwater sediment

1. Introduction

Plastic is one of the most widely used and frequently detected materials in the environment worldwide, owing to its versatile applications and complex structure. Eriksen et al. (2014) reported that the world's oceans are polluted with >250,000 tons of plastics, comprising about 5 trillion plastic particles (Eriksen et al., 2014). According to a report by Plasticsthe facts 2021, in 2020, over 360 million tons of plastics, including 55 million tons in Europe alone, were produced worldwide and about 23% of the plastics collected in Europe were landfilled, indicating that a significant amount of the used plastics have entered the environment through various routes (Plastics Europe and EPRO, 2021). Additionally, according to a Statista report (2021), the United States is the world's largest producer of plastic waste, where out of the 35.7 million tons plastic waste produced, only 3.1 million tons of plastic waste were recycled, with plastic waste being discharged into marine litter, which can cause marine pollution (Statista, 2021). In general, when plastics enter the environment, they are broken down into small pieces due to environmental weathering (i.e., environmental aging), such as photo-oxidation by UV and chemical oxidation by reactive oxygen species generated in the environment (Han et al., 2019a; Han et al., 2018a; Han et al., 2018b). Plastic particles smaller than 5 mm, which are referred to as Microplastics (MPs), pose significant environmental concerns (Verschoor, 2015). Plastics are persistent in the environment due to their complex chemical structures and large molecular weights, making them less prone to environmental degradation (Liu et al., 2022). Moreover, plastics contain additives and toxic chemicals (e.g., brominated flame retardants, phthalate, nanomaterials, and harmful impurities) (Lithner et al., 2011). As plastics environmentally age and decompose into smaller pieces, their surface area dramatically increases, and the additives are more rapidly released (Rillig et al., 2021).

Therefore, plastic and MPs have significant adverse effects on human health (Barboza et al., 2018; Karbalaei et al., 2018; Prata et al., 2020), animals (Franzellitti et al., 2019; Guzzetti et al., 2018; Prata et al., 2021; Rillig et al., 2019), and the ecosystem (Ivleva et al., 2017; Rillig and Lehmann, 2020). In particular, human exposure to plastic debris and the released additives, can cause lung and intestine damage (Wright and Kelly, 2017). It is easier for smaller sized additives and plastic debris to penetrate organs such as cell membranes (Hale et al., 2020), the blood-brain barrier, and the human placenta (Ragusa et al., 2021), resulting in damage. Therefore, it is of great importance to determine and monitor MPs in the environment. Currently, extensive research effort is devoted to monitoring MPs in the environment as they have been ubiquitously detected in various environmental media including water (Luo et al., 2019), air (Gasperi et al., 2018), soil (Harms et al., 2021), sediment (Uddin et al., 2021) and even organisms (Rillig et al., 2017; Zhang et al., 2021). In addition, to detect and analyze MPs in the environment, several different techniques have been established (Imhof et al., 2012), including microscopy (Stereo), spectroscopy (Fourier transform infrared (FTIR) and Raman spectroscopy (Raman)), and thermal analyses (Li et al., 2018; Shim et al., 2017). Even though analytical techniques were already established for MPs monitoring, sample pretreatment is essential to remove any organic or other interfering compounds prior to MPs analysis in environmental samples. In particular, soil and sediment are heterogeneous materials containing many components such as garbage, dead leaves, natural organic matter, silt, and sand (Rillig, 2012). Therefore, separating and extracting MPs from soil and sediment samples is more challenging than air and water matrices. Many studies of MPs focused on the pollution assessment of MPs (Cole et al., 2011) as well as their behavior and role in the environment (Andrady, 2017). Most studies have focused on MPs analysis in ocean and marine sediment (Besley et al., 2017; Coppock et al., 2017; Hidalgo-Ruz et al., 2012; Matsuguma et al., 2017), whereas research on MPs in soil and freshwater sediment is still very limited. Moreover, there are few studies related to pretreatment methods and measurement criteria for extracting MPs from soil and freshwater sediment and thus, their standards have not been well established. Therefore, this paper focuses mainly on pretreatment methods through a review of recent papers analyzing MPs in soil and freshwater sediment (Adomat and Grischek, 2021; Prata et al., 2019; Stock et al., 2019). The outcome of this study will provide guidelines for understanding suitable pretreatment procedures for monitoring MPs in the environment, particularly for soil and freshwater sediment samples. To our knowledge, this study is the first one focusing on the pretreatment of soil and sediment samples for MPs detection.

2. Methodology of literature selection

Relevant literature were searched using Google Scholar databases with papers published from 2017 to 2022. Used keywords were 'Microplastics', 'Soil', 'Sediment', 'Freshwater', 'Pretreatment' and their combinations. Forty studies met the criteria (20 publications for soil and freshwater sediment, respectively) and were critically reviewed. We focused on sample pretreatment methods for MPs analysis in the papers, targeting only soil and freshwater sediment as the environmental matrix. Therefore, marine sediment papers were excluded. If there were more than one environmental matrix (e.g. water or sludge and sediment) in a study, contents of soil and sediment in the study were only used. Different processes for

MPs sample pretreatment in each paper were divided into four representative categories: i) sampling sites and lab sample processing before main pretreatment processes (Section 3), ii) main pretreatment methods for monitoring MPs: density separation and organic removal (Section 4), iii) analytical instruments and characteristics for MPs detection in samples after pretreatment (Section 5) and iv) recovery test of pretreatment methods (Section 6).

3. Sampling sites and lab sample processing before main pretreatment

processes

Sampling is the first step toward MPs monitoring in the soil and sediment. Before MPs analysis in the laboratories, samples are collected, stored, and preprocessed. For safety purposes, generally the MPs samples are collected, stored, and transported in glass, stainless or aluminum containers rather than plastic bottles or containers. In this review, sample processing methods in soil and freshwater sediment are classified into three categories of 1) sampling sites, 2) sample processing methods (sieving/drying), and 3) sample quantity.

3.1. Sampling sites

Tables 1 and 2, and Fig. 1 summarize the information collected from the selected forty papers. Half of these studies were based on samples of freshwater sediment obtained from twelve Asian countries (nine sites in China, one site in Korea, Indonesia, and India), six European countries (two sites in Germany, one site in Portugal, Denmark, Scotland, and Spain), one South American country (Mexico) and one in the Oceania region (New Zealand). The freshwater sediment samples were collected from fourteen rivers, three lakes, one pond, and two estuaries. Among the twenty soil samples, fifteen samples were collected from Asia (including ten samples in China, two in Korea, two in Japan, and one in Pakistan), two samples from Europe (one in Switzerland and Spain), two samples from South America (Chile) and one sample from Africa (Mauritius). Mainly agricultural lands were selected for the soil samples, while flood plains, coastal beaches, and river delta wetlands were selected for sediment samples.

3.2. Sample processing methods

The sample processing methods are typically represented by sieving and drying. A sieving process is a standard method for analyzing MPs in sediment or soil samples and is used to: 1) homogenize the samples, 2) obtain MPs of the desired sizes, and 3) remove unwanted large sized grains or impurities. In soil and sediment studies, sample sieving is usually performed during the sample collection or before pretreatment. Among various sieve pore sizes, 5 mm and 2 mm were mostly used. Other pore sized sieves (10 mm and 1 mm, etc.) have also been used and several other studies performed multi-sieving using two or more sieves with different pore sizes. For example, Ragoobur et al. (Ragoobur et al., 2021) used 10, 6 and 2 mm sieves to remove impurities such as gravel and roots. In a different study, Rafique et al. (Rafique et al., 2020) used a 50 μ m sieve to remove particles of <50 μ m that were not needed for analysis after sample sieving using 5 mm sieve. Multi-sieving is important when selecting different MPs analytical instruments for each section through size fractionation and to investigate the correlation between MPs type and size. As an example, to classify the analyzed particles sizes, Sarkar et al. (Sarkar et al.,

2019) used four sieves with different pore sizes (i.e., 10 mm, 5 mm, $850 \mu\text{m}$, and $63 \mu\text{m}$) to investigate MPs size distributions ranging from 63-850 µm, 850 µm-5 mm, and 5-10 mm (mesoplastics). Similarly, Scherer et al. (Scherer et al., 2020) employed three different sized sieves (20 μ m, 125 μ m, and 1000 μ m) and evaluated three size distributions of MPs $(20-125 \,\mu\text{m}, 125-1000 \,\mu\text{m}, \text{and} > 1000 \,\mu\text{m})$. It is important to note that MPs with particle sizes larger than 1000 µm can be directly and visually sorted. In this study, attenuated total reflectance-Fourier transform infrared spectroscopy (ATR-FTIR), and visual analysis were performed for 125-1000 µm and 1000-5000 µm size MPs, and MPs of 20-125 µm sized were stored for pyrolysis-gas chromatography-mass spectrometry (Pyr-GC-MS) analysis (Scherer et al., 2020). Additionally, soil and sediment have different characteristics, such as moisture, organic matter, and biota, that are a function of sampling sites. For consistency across studies, samples were dried at a constant weight before the experiments. Dry sieving was performed if accurate sieving was difficult due to high viscosity and cohesiveness of the moisture-containing samples. In soil and sediment studies, sample drying was carried out at different temperatures and time durations. Samples were dried in a range of 25-105 °C; however, 50–75 °C was the commonly used temperature in most cases. A freeze-drying temperature was used in some cases as well (Jiang et al., 2018). Since MPs can melt or be damaged at high temperatures, it is recommended that drying is not performed at high temperatures. However, in some cases, high drying temperatures (100–105 $^{\circ}$ C) have been used for determining the standard weight of samples (Alam et al., 2019; Blair et al., 2019; Zhou et al., 2018). Blair et al. (Blair et al., 2019) also reported that MPs did not deform, i.e., melt or decompose, in the above temperature range. The drying time also varied, ranging from 12 h to 7 days, and in most cases, 24–48 h was chosen as the optimal drying time period. It is critical to select a specific drying temperature and sufficient drying time to achieve constant weight, depending on the characteristics of the sample. The details of the drying temperature and drying time frame are summarized in Tables 1 and 2.

3.3. Sample quantity

Sample quantity is critical for MPs analysis in soil and freshwater sediment. For instance, if the sample contains a high concentration of MPs, spectroscopy, including Raman and FTIR analyses, might not be suitable. On the contrary, if samples contain low concentration of MPs, thermal analysis, such as thermogravimetric analysis-Fourier transfer infrared spectroscopy (TGA-FTIR), thermal extraction and desorption-gas chromatography-mass spectrometry (TED-GC-MS), and Pyr-GC-MS, will not accurately detect MPs due to their insufficient detection limit. Therefore, it is important to establish a standardized sample quantity for the MP analysis. Yet, the quantity of samples used in MPs analysis in soil and freshwater sediment could be different. In case of freshwater sediment, 10-500 g of dry weight (DW) sample was mostly used. DW method was mainly used to measure the sample quantity. However, in some cases, the sample was measured using wet weight (WW), followed by drying (Eo et al., 2019; Scherer et al., 2020). In the case of soil MPs analysis, the sample quantity varied significantly, ranging from 1 to 1000 g. The significant difference in the sample quantity resulted from the dramatically different distribution of MPs in the samples. In most soil studies, DW was used to measure the sample quantity, but sometimes, volume was used instead of DW. For example, Dikareva & Simon (Dikareva and Simon, 2019) elutriated 1 L of sediment samples (WW) and dried the residue remaining on the 63

µm sieve. Scheurer & Bigalke (Scheurer and Bigalke, 2018) used 50 mL of dried soil sample and mixed it with sodium chloride (NaCl) solution. The details of the sample quantity used in soil and sediment studies can be found in Tables 1 and 2, and Fig. 2.

4. Main pretreatment methods for monitoring MPs: density separation and organic removal

Extraction and separation are two essential steps before any MPs analysis in the soil and sediment. For effective extraction and separation of MPs in soil and sediment, density separation (DS) and organic removal (OR) are the most commonly used techniques. In addition to density separation, an organic removal step is needed prior to MPs analysis in soil and sediment samples as the presence of organic matter on the surface of MPs may alter FTIR and Raman results. As summarized in Tables 3 and 4, different pretreatment techniques were used during MP analysis in the soil and sediment samples. In general, samples went through pretreatment processes before the density separation and organic removal, followed by filtration and drying, as discussed below.

4.1. Pretreatment procedures

Establishing a pretreatment procedure for MPs analysis is very important as it affects the results of sample analysis. In the reviewed studies, the order of pretreatment processes varied. In the pretreatment process for both freshwater sediment and soil, most studies used both density separation and organic removal, and a few studies performed only density separation without the organic matter removal process. The pretreatment processes for freshwater sediment and soil samples were very similar in different studies. The most used processes were DS and then OR (DS \rightarrow OR) for both samples. The order is as follows: $DS \rightarrow OR (55\% \text{ of the studies}) > DS \text{ only } (25\% \text{ of the studies}) > OR \rightarrow DS (10\% \text{ of})$ the studies) > OR \rightarrow DS \rightarrow OR (5% of the studies) = DS & OR simultaneously (5% of the studies) for freshwater sediment matrixes, and DS \rightarrow OR (50% of the studies) > DS only (20% of the studies) = OR \rightarrow DS (20% of the studies) > OR \rightarrow DS \rightarrow OR (10% of the studies) for soil matrixes as shown in Figs. 3 and 4. Although uncommon, organic oxidation and density separation processes were also used simultaneously for a sediment sample. Also, if necessary, DS processes were repeated several times to increase the efficiency of MPs separation/extraction from samples (Simon-Sánchez et al., 2019). The order of pretreatment can significantly change the method used in each step. If organic matter removal was performed before density separation, a large amount of organic matter removal reagent was used because soil and sediment samples contain a lot of organic matter. For example, in the case of oxidation using H_2O_2 , a significant amount of foam may be produced due to organic oxidation by H_2O_2 , resulting in sample loss. Conversely, if the density separation was performed first, many impurities can be mixed in the supernatant, and then filter clogging may occur when collecting MPs, because of the presence of unremoved organic compounds in the samples. Therefore, it is of great importance to configure the order of pretreatment considering these advantages and disadvantages.

4.2. Density separation methods

Density separation is a key process for MPs analysis in soil and sediment samples due to the complexity of sample matrices. Due to differences in the density of each material, various types of MPs can float (or remain unsettled) in samples when several solutions with different densities are used. Table 5 summarizes different polymers' densities and various density solutions for MPs separation. The density solutions used for freshwater sediment and soil samples were slightly different from each other. For both the sediment and soil samples, 1.20 g/cm³ NaCl was the most commonly used density solution. Similarly, zinc chloride (ZnCl₂) and sodium iodide (NaI) with a density of 1.50–1.80 g/cm³ and 1.60–1.80 g/cm³. respectively, were widely used as density solutions. Moreover, water and other solutions including potassium formate (CHKO₂) of 1.54 g/cm³, lithium metatungstate (LMT) of 1.60 g/cm³, NaI-NaCl mixtures of 1.50 g/cm³, sodium bromide (NaBr) of 1.55 g/cm³, calcium chloride (CaCl₂) of 1.40 g/cm³, and ZnCl₂:CaCl₂ mixtures of 1.55-1.58 g/cm³ were used as density solutions. Li et al. (Li et al., 2019) compared the efficiencies of three different solutions (i.e., NaCl, ZnCl₂, and NaI) and found NaI as the most efficient separating solution. When high-density fiber MPs were negligible, NaCl was recommended as the optimum separation solution with economic consideration (Li et al., 2019). In some cases, multiple density separation solutions with varying densities were used simultaneously with water components. For example, NaCl and NaI (Di and Wang, 2018; Yuan et al., 2019; Zhou et al., 2020; Zhou et al., 2018) or deionized (DI) water and NaI have been frequently used in the literature (Dikareva and Simon, 2019). Other studies have combined water, NaCl, and ZnCl₂ (Corradini et al., 2021; Corradini et al., 2019). Details on the use of density solutions are given in Fig. 5. The density and characteristics of the density separation solution should be determined and used for MPs extraction. Considering the most used density solution, NaCl has a low density and may not be able to extract MPs such as polyvinyl chloride (PVC), polyethylene terephthalate (PET), and polytetrafluorethylene (PTFE). In addition, NaCl is easily crystalized during MPs extraction and thus, MPs were attached on beakers with the crystalized NaCl, resulting in sample loss. High density ZnCl₂ has an advantage of extracting high density polymers, including PVC and PET, but it can form salt and cause toxicity (Franklin et al., 2007), which must be considered. Therefore, for the selection of density solution for MPs extraction, characteristics of density solutions must be carefully checked. The separation solutions were added to the freshwater sediment or soil samples and mixed homogeneously using different techniques, such as aeration, magnetic stirrer, glass rod, sonication, or centrifugation. The resulting mixtures were left for some time until the soil or sediment compounds settled and the suspended MPs were extracted. In freshwater sediment and soil studies, the volume of density solution used to pretreat the same amount of sample was similar. For samples weighing 100 g or less, the majority of studies for both freshwater sediment (Alam et al., 2019; Blair et al., 2019; Jiang et al., 2018; Mani et al., 2019; Shruti et al., 2019; Simon-Sánchez et al., 2019; Wang et al., 2017) and soil (Choi et al., 2021; Corradini et al., 2021; Corradini et al., 2019; Huang et al., 2020; Li et al., 2019; Pérez-Reverón et al., 2022; Rafique et al., 2020; Ragoobur et al., 2021; Zhang and Liu, 2018) used solution volumes under 500 mL. For larger samples (i.e., above 200 g), a larger volume of the solutions was used (e.g., 1 L) regardless of the sample type (i.e., whether soil or freshwater sediment). The amount of the density separation solution should be sufficient enough for separating MPs from the sample, which depends on the characteristics of the

container used in the density separation step. In many studies, during density separation, the samples and solutions were mixed using mechanical mixing. Other studies mixed the samples and solutions and then centrifuged them for the separation of MPs. Centrifugation accelerates the settling step, so the time required for particles to settle would be reduced. The supernatants of the centrifuged mixture were simply collected. In most cases, when centrifugation was applied, only a small amount of sample was used such as amounts aliquoted from samples or residue particles after the elutriation step (e.g., 5–30 g of DW). On the other hand, Ragoobur et al. (Ragoobur et al., 2021) used 150 g of dried sample, but in this case, the sample was used for density separation first, and then its supernatant was used for several steps of centrifugation. A few studies also used an unusual way of collecting the separated MPs, also known as the overflow method, where the same type of solution used in density separation was poured into the settled or separated mixture, and the MPs were collected from the overflowing supernatants.

4.3. Organic removal on the surface of MPs

The supernatants or particles collected from the density separation step were passed through a step of organic removal. While most studies applied the organic removal process right after the density separation process, a few studies performed the sample oxidation for organics removal before the density separation. The organic removal methods used in the freshwater sediment and soil samples are shown in Fig. 6. Different oxidants including hydrogen peroxide and Fenton's reagents, have been widely used for organics removal during MPs analysis. For treating freshwater sediment and soil samples, a wet peroxide oxidation process (WPO), employing hydrogen peroxide solution, was mainly used to decompose organics in the collected particles. In freshwater sediment and soil studies, both room temperature and heat-treated WPO could be used. In several studies, H₂SO₄, NaOH, and butyl alcohol were added when performing WPO and Fenton process in order to improve the efficiency of organic removal (Choi et al., 2021; Li et al., 2019; Liu et al., 2019a; Scherer et al., 2020; Zhang and Liu, 2018). Interestingly, for soil samples, oxidation processes using HNO3 and H2SO4 were also reported (Li et al., 2020; Scheurer and Bigalke, 2018). In most of the literature reviewed, acid digestion with HNO₃ or HCl and alkaline digestion with NaOH were barely used for sample pretreatment because these can cause discoloration of MPs (Roch and Brinker, 2017) or damage MPs (Köhn et al., 2017; Qiu et al., 2016), resulting in inaccurate MPs analysis. In a few cases, enzymes were used for organic removal on the surface of MPs. In these cases, removing the different organic matter is difficult because a specific enzyme could only decompose a specific target substance. On the other hand, Fenton's oxidation and H_2O_2 oxidation were able to efficiently remove organic matter without major problems under appropriate reagent concentrations and temperature conditions. That is why both oxidation processes were widely used due to their convenience of organic removal under different conditions.

4.4. Filtration

Filtration is considered the final step of the pretreatment procedure of the sample, accomplished using various types of microfilters, mostly in the range of $0.2-20 \,\mu\text{m}$. However, in a few cases, filters with a pore size of 50 μm or larger were also used for the filtration. Filters were selected, depending on the particle sizes of the analyzed MPs. In

addition to the pore sizes, different filter materials, including glass fiber, nylon, aluminum oxide, cellulose nitrate, etc., were used. In general, the filter types have no correlations with analytical methods (i.e., microscope, FTIR, and Raman). It may be due to different adsorption patterns of MPs on various filter materials. However, in one study, Mani et al. (Mani et al., 2019) used different filter types, depending on the analytical method. The 0.2 μ m anodisc was used for ATR-FTIR analysis (<500 μ m of MPs) and mixed cellulose ester filter with 0.45 μ m pore size was used for μ -FTIR analysis (500 μ m-2 mm of MPs). Similarly, Corradini et al. (Corradini et al., 2021) also used a 2.5 μ m cellulose filter for optical microscopy analysis and 0.4 μ m polycarbonate membrane for FTIR analysis.

After filtration, the drying process was conducted at 40-60 °C, or room temperature, or even in desiccators below the threshold temperature to avoid MPs deformation, as mentioned in Section 3.2.

5. Analytical instruments and characteristics for MPs detection in

samples after pretreatment

After sample pretreatment with different steps, MPs were analyzed in terms of physical properties, types, and abundance in samples.

5.1. Instrument for MP analysis

To monitor MPs in the freshwater sediment and soil samples, a suite of visualization methods have been used including visual sorting with bare eyes, magnifying glass, optical microscope, a fluorescent stereo microscope, and a scanning electron microscope (SEM). To simplify categorizing visualization methods, visual sorting and a magnifying glass were included in the microscope category. Moreover, spectroscopic and thermal analyses were performed using Raman, FTIR, and Pyr-GC–MS.

Among the analytical methods, the most commonly used instruments were microscopical techniques (i.e., dissect, polarized, fluorescence, scanning electron, and atomic microscopy), followed by FTIR. Generally, FTIR or an optical microscope was solely used, but, mostly, two or more analytical tools were used together to determine MPs in samples. Interestingly, Schere et al. (Scherer et al., 2020) used two different methods, depending on the sizes of MPs in samples. A microscope was used for MPs of 125-5000 µm and Pyr-GC-MS for 20-125 µm MPs. Detailed information about the instrument and analytical size of MPs for each instrument are summarized in Tables S1 and S2 and Fig. 7. Also, Fig. 8 shows the minimal size of detectable MPs correlated to the selected instrument. In general, MPs with a size of 50 µm or less could be determined with a combination of a microscope and FTIR or Raman. Likewise, MPs with the size of 50-100 µm were most detectable in analyses involving spectroscopy (FTIR and Raman). Moreover, Grause et al. (Grause et al., 2022) were able to detect MPs with a size of at least 10 µm using a fluorescence microscope. However, it is very difficult to use an optical microscope to analyze MPs under 100 μ m, given the accuracy and efficiency of the analysis (Shim et al., 2017). Therefore, the instrument must be carefully selected based on the size of MPs for purposes of different studies.

5.2. Characterization of MPs

The MPs in freshwater sediment and soil samples were analyzed and characterized in terms of size, shape, color, and type. The shape of MPs was classified into fiber, film, fragment, pellet, and bead. Most MPs were classified as fibers and fragments. In addition, MPs color varied between red, blue, green, black, and even transparent. Color analysis of MPs is mainly possible with a microscope, but there were still many studies that did not investigate the color using this method. The types of MPs can be identified by using spectroscopy or thermal decomposition, such as FTIR, Raman, and Pry-GC–MS. Various types of MPs, including polypropylene (PP), polyethylene (PE), polystyrene (PS), and PVC, were identified with these instruments. In particular, most of the identified MPs consisted of polyolefin, which includes PE and PP. The sizes of MPs were observed in a wide range of 10–5000 μ m. However, as mentioned in Section 5.1, because each instrument has different detection limits for size measurement, the instrument must be carefully selected before the analysis. To accurately analyze types of small-sized MPs, a spectroscopy or pyrolysis instrument is recommended.

6. Recovery tests of pretreatment methods

A recovery test is needed to validate the pretreatment method for MPs analysis in environmental samples. Among the reviewed papers, only 50% of the studies carried out the recovery test as seen in Table 6. For the recovery tests, MPs types such as PP, PE, PS, PVC, PET, and poly(methyl methacrylate) (PMMA) were mainly used. The recovery was calculated by comparing the number of spiked MPs in samples before and after pretreatment. More specifically, the fluorescence of MPs was measured when MPs were fluorescently tagged. A recovery for 63–150 µm sized PET was 81% (Grause et al., 2022) and 55% for 62–125 µm sized PMMA (Mani et al., 2019). In most studies, when largersized MPs were used, a higher recovery rate was obtained. When an insufficient amount of fluorescence was tagged on MPs, it was very challenging to determine the exact recovery rate. Unfortunately, most of the studies did not specify the shape of MPs used for the recovery tests. However, in the studies that performed recovery tests based on the type and shape of MPs, fragments of low density polyethylene (LDPE) showed high recovery rates of up to 98% and fiber of PVC up to 90% (Corradini et al., 2019). In addition, the recovery rates according to the shape (fiber, film, and particle) of PE was measured to be 85-98.3% in the order of film<fiber<particle when NaBr solution was used for density separation (Liu et al., 2019b). In addition, Blair et al. (Blair et al., 2019) performed recovery tests using secondary MPs and obtained recovery rates of 49-58%.

7. Conclusions and future perspective

This paper thoroughly reviewed recent articles published from 2017 to 2022 about the pretreatment methods for MP analysis in freshwater sediment and soil samples. As known, soil and sediment sample pretreatment is very difficult due to the complicated compositions of samples. For sample processing, sieving is required to remove large sized impurities and MPs separation in sizes. Drying was essential to determine accurate sample amounts for MPs analysis without any MPs damages. The quantity of sample used for pretreatment was

mostly <100 g after drying. Density separation and organic removal are the main processes of pretreatment and in general, density separation was performed before organic removal because the required amounts of chemicals for the pretreatment were significantly reduced. As density separation solutions, NaCl, ZnCl₂, and NaI are widely used, and their selection is dependent on the polymer types analyzed in different studies. Moreover, advantages and disadvantages of different density solutions must be considered to choose the best reagent. In a few cases, acids, alkalies, and enzymes could be used to remove organic matter. However, they may damage MPs, and in particular, an enzyme has its own specific target organic and thus, its use is very limited along with its expensiveness. H_2O_2 oxidation or Fenton reactions are the most effective organic removal methods. If MPs size distribution must be investigated, sample filtration using sieves with different pore sizes could be used. In addition, sample pretreatment could be dependent on different analytical equipment. Also, sample pretreatment could significantly influence the results of MP analysis. Therefore, we emphasize the importance of sample pretreatment of soil and sediment.

Unfortunately, so far, there is no standardized pretreatment method for MPs monitoring in soil and sediment and it is still challenging to suggest a representative method for sample pretreatment due to varying factors, parameters and methods that are study specific. Moreover, the order of pretreatment steps could be effectively changed, depending on the sizes and types of MPs for purposes of different studies. To develop an appropriate pretreatment method, it is vital to establish a standard or representative method by referring to optimal conditions tested in multiple studies. Finally, investigations on pretreatment methods for various environmental media (e.g., air, water, and wastes) with different characteristics should be expanded for further determination of MPs in the environment.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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Data availability

No data was used for the research described in the article.

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HIGHLIGHTS

- Soil & sediment sample pretreatment is a significant process for MP analysis.
- A standardized method is still not available for soil & sediment sample pretreatment.
- Current pretreatment methods were carefully reviewed for soil & sediment samples.
- A dominant procedure for sample pretreatment in current studies was determined.
- However, a standardized pretreatment method is highly required in near future.



Fig. 1.

The geographical location of the sites where the studies were conducted; (a) freshwater sediment, (b) soil and (c) number of studies in different countries.



Fig. 2.

Sample quantity for selected weight ranges used for pretreatment in freshwater sediment and soil studies.



Fig. 3.

Five flowcharts of pretreatment procedures: (a) density separation only, (b) density separation followed by organic removal, (c) organic removal followed by density separation, (d) density separation followed by organic removal after removal of organic in advance and (e) simultaneously perform density separation and organic removal.









Density solution types used in freshwater sediment and soil studies (When using 2 or 3 solutions, selected from DI water, NaCl, NaI, ZnCl₂ is used).





Organic removal methods in freshwater sediment and soil studies; H_2O_2 (9.8–11.4 M) used in WPO.







Fig. 8. Minimum size of MPs detected using analytical instruments.

No.	Sampling site	Sample processing			Sample amounts	Ref.
		Sieving	Drying			
			Temperature (°C)	Time		
-	Beijing river, China	N/A	50	48 h	30 g (DW)	(Wang et al., 2017)
7	Qinghai lake, China	2 mm	N/A	N/A	N/A	(Xiong et al., 2018)
3	Yangtze river, China	N/A	N/A	N/A	500 g (WW)	(Di and Wang, 2018)
4	Dongting lake, China	N/A	Freeze-drying	N/A	50 g (DW)	(Jiang et al., 2018)
5	Huangpu river, China	N/A	70	N/A	100 g (DW)	(Peng et al., 2018)
9	Antuã river, Portugal	N/A	06	48 h	N/A	(Rodrigues et al., 2018)
٢	Atoyac river, Mexico	N/A	<40	N/A	30 g (DW)	(Shruti et al., 2019)
8	Retention ponds, Denmark	2 mm (wet-sieving), settle for a week, gather	50 (After pre-	N/A	50 g (WW; pre-oxidation)	(Liu et al., 2019a)
		particles in the supernatant using a 10 µm s.t.s.inter, sonicate STS filter for 15 min, transfer the filtered particles back to the settled sediment	OXIGATIOIL)		200 g (DW; density separation)	
6	Wei river, China	N/A	70	24 h	100 g (DW)	(Ding et al., 2019)
10	Kelvin river, Scotland	2.8 mm, 2.0 mm, 1.4 mm, 1.0 mm, 0.71 mm, 0.5 mm, 0.355 mm, 0.255 mm, 0.185 mm, 0.09 mm and 0.063 mm, use an automatic shaker for 10 min	100	24 h	25 g (DW)	(Blair et al., 2019)
1	Rhine river, German	<2 mm, 2–5 mm, >5 mm	60	4 days	60 g (DW)	(Mani et al., 2019)
12	Poyang lake, China	N/A	50	48 h	500 g (DW)	(Yuan et al., 2019)
13	Ciwalengke river, Indonesia	N/A	100	48 h	100 g (DW)	(Alam et al., 2019)
14	Nakdong river, Korea	N/A	60	N/A	Determine by drying	(Eo et al., 2019)
					100 g (WW)	
15	Ganga river, India	63–850 μm, 850 μm–5 mm, 5–10 mm (mesoplastics)	65	36 h	N/A	(Sarkar et al., 2019)
16	Auckland streams, New Zealand		N/A	N/A	Determine by drying the residue remaining after elutriation step of 1 L (WW)	(Dikareva and Simon, 2019)
17	Ebro river, Spain	N/A	N/A	N/A	10–20 g (DW)	(Simon-Sánchez et al., 2019)
18	Elbe river, German	Wet-sieving, 20–125 µm (dry at 45–55 °C for 5–7 days), 125–1000 µm (perform density separation), 1000–5000 µm (store for analysis)	45	5–7 days	Determine by drying 200 g (WW)	(Scherer et al., 2020)

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Table 1

No.	Sampling site	Sample processing			Sample amounts	Ref.
	-	Sieving	Drying			
			Temperature (°C)	Time		
19	Changjiang estuary, China	N/A	70	24 h	100 g (DW)	(Peng et al., 2017)
20	Bohai bay river estuary, China	0.9 mm	60	N/A	200 g (DW)	(Han et al., 2019b)
Note: I	JW (Dried Weight) and WW (Wet	Weight).				

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No.	Sampling site	Sample processing			Sample amounts	Ref.
		Sieving	Drying			
			Temperature (°C)	Time		
-	Shanghai farmlands, China	N/A	70	24 h	N/A	(Liu et al., 2018)
7	Floodplain, Swiss	1 mm	65	N/A	50 mL of dried and sieved soil sample	(Scheurer and Bigalke, 2018)
ю	Chai river valley cropped area, China	10 mm	25	N/A	30 g (DW)	(Zhang and Liu, 2018)
4	Shandong province coastal beaches, China	N/A	105	>12	N/A	(Zhou et al., 2018)
5	Nanjing, Wuxi and Jiangsu province agriculture lands, China	5 mm, 1 mm	Air dry	N/A	50 g	(Li et al., 2019)
9	Shanghai farmlands, China	N/A	70	24 h	N/A	(Lv et al., 2019)
٢	Melipilla agricultural field, Chile	2 mm	40 ± 2	N/A	$5 \pm 0.01 \text{ g}$	(Corradini et al., 2019)
×	Shanghai farmlands soil, yellow-brown soil, paddy soil and floodplain soil, China	5 mm	70	24 h	50-200 g (DW)	(Liu et al., 2019b)
6	Shihezi cotton fields, China	5 mm	Air dry	N/A	20 g (DW)	(Huang et al., 2020)
10	Paotai town, Shihezi, Xinjiang agricultural soils, China	2 mm, 0.9 mm, 0.28 mm, 0.15 mm	N/A	N/A	1 kg (DW)	(Li et al., 2020)
11	Hangzhou bay coastal plain farmlands, China	5 mm	25	N/A	50 g (DW)	(Zhou et al., 2020)
12	Yellow river delta wetland in Shangdong, China	2 mm	N/A	N/A	1 kg (DW)	(Duan et al., 2020)
13	Eight different land use types in Lahore, Pakistan	5 mm, 50 µm (wet sieving)	N/A	N/A	200 g (DW)	(Rafique et al., 2020)
14	Arable soils, pastures, rangelands and natural grasslands in metropolitan, Chile	2 mm	N/A	N/A	5 g (for visual identification), 1 g (for FTIR)	(Corradini et al., 2021)
15	Tedori river alluvial fan, Japan	N/A	60	48 h	1 kg (DW)	(Katsumi et al., 2021)
16	Forest, suburban and agricultural land in Yeoju, Korea	5 mm	60	48 h	50 g (DW)	(Choi et al., 2021)
17	Agricultural soils in Yongin, Korea	5 mm	N/A	N/A	200 g (DW)	(Kim et al., 2021)
18	Agricultural fields, Mauritius	10 mm, 6 mm, 2 mm	60	72 h	150 g	(Ragoobur et al., 2021)
19	Agricultural fields in Miyagi, Japan	2 mm	60	72 h	10 g	(Grause et al., 2022)
20	Irrigated soils in Fuerteventura, Spain	N/A	N/A	N/A	10 g	(Pérez-Reverón et al., 2022)
Note:	DW (Dried Weight) and WW (Wet Weight).					

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Table 2

Sample pretreatment before density separation and organic removal for MPs analysis in soil.

Dell	suy separauon an	ווו טו צמוועי וייוווי					
N0.	Order	Density separatic	on (DS)	Organic removal (OR)	Filtration		Ref.
		Density solution	Procedure		Filter type	Drying conditions	
-	DS only	NaCl (1.20 g/ cm ³)	1. Add 200 mL of NaCl solution Stir for 2 min Settle for 2 h	N/A	1.0 µm glass microfiber filter	50 °C for 48 h	(Wang et al., 2017)
			2. Sonicate for 5 min Settle overnight				
7	$\mathrm{DS} ightarrow \mathrm{OR}$	$\frac{CHKO_2 (1.54 \text{ g})}{\text{cm}^3}$	To overflow supernatant into stainless steel tray add solution and settle overnight	Using H ₂ O ₂ (30%) at 60 °C overnight	1.2 µm glass microfiber filter	Dry in a desiccator	(Xiong et al., 2018)
ς	$\mathrm{DS} ightarrow \mathrm{OR}$	1. NaCl (saturated) 2. Nal (60%)	Perform a two step density separation 1. Add 1 L of NaCl solution Stir for 2 min Settle for 10 min Sieve using a 48 µm stainless steel sieve Repeat three times	WPO (30% H ₂ O ₂) for 12 h	0.45 µm glass microfiber filter	50 °C	(Di and Wang, 2018)
			 Add 375 mL of Nal solution Shake 200 ppm on a shaker for 2 min Settle for 10 min 				
4	$\mathrm{DS}\to\mathrm{OR}$	$ZnCl_2 (1.50 \text{ g/} \text{cm}^3)$	Add 400 mL of ZnCl ₂ solution Stir for 30 min Settle for 24 h	WPO (30% H ₂ O ₂)	0.22 µm glass microfiber filter	60 °C	(Jiang et al., 2018)
ŝ	DS only	NaCl (1.20 g/ cm ³)	Add NaCl solution Stir for 2 min Settle for 24 h	N/A	1.0 µm glass microfiber filter	Dry to constant weight	(Peng et al., 2018)
9	DS → OR	ZnCl ₂ (1.60 g/ cm ³)	Disaggregate with 400 mL of sodium polyphosphate (5.5 g/L) Stir for 1 h at high rpm Sieve sediments using a 5 mm and 0.055 mm sieve Add 300 mL of ZnCl ₂ solution and stir Settle for 1 h and transfer supernatant to 0.055 mm sieve and rinse with Milli-Q water	WPO (30% H ₂ O ₂ with 0.05 M Fe ²⁺ catalyst) at 75 °C for 5/10 min until it starts boiling and afterwards react at room temperature for 15 h	0.45 µm membrane filter	40 °C for 3–5 days	(Rodrigues et al., 2018)
Г	$OR \rightarrow DS$	ZnCl ₂ (1.58 g/ cm ³)	Add 36 mL of ZnCl ₂ solution Settle for 24 h	WPO (30 mL of 30% H ₂ O ₂) for overnight	1.2 µm nitrocellulose filter paper	A ir-dry	(Shruti et al., 2019)
∞	$\mathrm{OR} \to \mathrm{DS} \to \mathrm{OR}$	ZnCl ₂ (1.97 g/ cm ³)	Perform a 2 step density separation 1. Add 1.5 L of ZnCl ₂ solution Aerate for 1 h Settle for 5 h Drain off solution using a funnel 2. Aerate for 30 min	 Pre-oxidation: g of WW subsample was filled with 200 mL Milli-Q water before adding 30 mL of 30% H₂O₂ (achieving final concentration of 4%) Heat at 50 °C and stir 	500 µm sieve, 10 µm stainless steel filter	 Dry particles on a 500 µm sieve at 50 °C Collect the particles on a 10 µm filter into 	(Liu et al., 2019a)
				Add 30 mL 30% H ₂ O ₂ every 24 h		by ultrasonication	

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Table 3

N0.	Order	Density separatic	on (DS)	Organic removal (OR)	Filtration		Ref.
		Density solution	Procedure		Filter type	Drying conditions	
			Settle overnight Drain off solution	(until no foarning) Dry at 50 °C Repeat processing for all of the 3 L sediments		and evaporate by nitrogen (N5.0) until dry	
				2. Post-oxidation: Fenton reaction by adding 146 mL of 50% H ₂ O ₂ , 63 mL of 0.1 M FeSO ₄ and 65 mL of 0.1 M NaOH at 15–19 °C Leave for 2 days at room temperature			
6	$\mathrm{DS} ightarrow \mathrm{OR}$	NaCl (saturated)	Add NaCl solution Stir for 2 min Settle for 24 h	WPO (30 mL of 30% H_2O_2) at 100 rpm and 65 °C for 24 h	0.45 µm membrane filter paper	$50 \circ C$ for 24 h	(Ding et al., 2019)
10	DS only	NaCl (1.20 g/ cm ³)	Add 40–68 mL of NaCl solution Stir for 1 min Settle for 24 h	N/A	11 µm cellulose filter	Room temperature (18-21 °C)	(Blair et al., 2019)
Ξ	DS → OR	ZnCl ₂ (1.70 g/ cm ³)	Leave a 1.5 cm margin, fill ZnCl ₂ solution into a 100 mL flask Sonicate at 160 W/35 kHz for 15 min Stir at 250 rpm for 1 h Settle for 30 min Add 20 mL of ZnCl ₂ , allow the excess fluid to overflow Store supernatants at 4 °C	 Sieve using 500 µm sieve Filter using 10 µm stainless steel disk -500 µm particles: oxidation by adding Fenton reagent (10 mL of FeSO₄ (7,2 mM, pH 5)) in a 20 °C water bath Add 20 mL 30% H₂O₂ after 15 min and sonicate at 215 W/35 kHz for 5 min 	 0.2 µm anodisc (<500 µm) 0.45 µm mixed 0.45 µm mixed unur2 	50°C for least 2 days	(Mani et al., 2019)
				500 µm particles: retain for subsequent analysis			
12	$\mathrm{DS} ightarrow \mathrm{OR}$	1. NaCl (saturated) 2. Nal (60%)	 Add 1 L of NaCl solution Settle for 2 h Sieve using 50 µm stainless steel sieve and rinse with distilled water Repeat three times 	WPO (30 mL of 30% H ₂ O ₂)	0.7 µm glass microfiber filter	N/A	(Yuan et al., 2019)
			2. Add 500 mL of Nal solution and perform the same way and repeat three times				
13	DS only	NaCl (30%)	Add 400 mL of NaCl solution Stir for 2 min using spoon Leave until no more visible material floats	N/A	1.2 µm glass microfiber filter	105 °C for 30 min	(Alam et al., 2019)
14	$\mathrm{DS} ightarrow \mathrm{OR}$	LMT (1.60 g/ cm ³)	Shake with 300 mL of LMT solution for 1 min Settle for 10 min Transfer supematant Repeat twice with 200 mL of LMT	WPO (20 mL of 35% H ₂ O ₂ with Fe ²⁺ solution) at 75 °C at 180 rpm for 30 min Add H ₂ O ₂ every 30 min	5 µm polycarbonate filter paper	Room temperature	(Eo et al., 2019)

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	ï	Density separatio	n (DS)	Organic removal (OR)	Filtration		Ref.
		Density solution	Procedure	ſ	Filter type	Drying conditions	
15 DS –	→ OR	ZnCl ₂ (1.80 g/ cm ³)	Mix with ZnCl ₂ (1:10(w/v)) Stir for 15–20 min Settle overnight Collect on filter paper (0.7 μm glass microfiber filter)	WPO (30% H ₂ O ₂) for 3 h	0.7 µm glass microfiber filter	Dry in a desiccator for 36 h	(Sarkar et al., 2019)
16 DS 01	nly	1. DI water (1.0 g/cm ³) 2. Nal (1.60 g/ cm ³)	 Elutriation step with water for 20 min Mix sediments by air stone Collect overflow materials on a 63 µm sieve by adding water (250 L/h flow rate) Transfer into a 50 mL centrifuge tube 	N/A	1.2 µm glass microfiber filter	60 °C	(Dikareva and Simon, 2019)
			 Add 30 mL of 3.3 M Nal solution Centrifuge for 5 min at 3500 G Settle for 1 h Repeat once again 				
17 DS, C simul	JR Itaneously	NaCl (1.20 g/ cm ³)	Use 200 mL of NaCl +20 mL of 30% H ₂ O ₂ solut Heat at 50 °C Stir for 20 min at 200 pm Settle at 50 °C for 1 h and afterwards settle at roc Transfer foam from reaction and mix with 50 mL heat, mix, settle, and filter supernatant	ion om temperature for 1 h , of NaCl + 20 mL of H ₂ O ₂ and then	0.7 µm glass microfiber filter	40 °C overnight	(Simon- Sánchez et al., 2019)
			Meantime, add 20 mL of H_2O_2 to WPO solution the mixture again and then collect and filter super	and sediments, heat, mix and settle rnatant			
18 DS –	→ OR	ZnCl ₂ (1.60– 1.80 g/cm ³)	Fill with ZnCl ₂ solution Settle for 24 h	WPO (30% H_2O_2 and 10% $H_2SO_4)$ at 55 $^\circ C$ for 5 days	2.7 µm glass microfiber filter	55 °C for 7 days (For pyr-GC–MS analvsis)	(Scherer et al., 2020)
				After filtration, repeat the washing process at least twice (until the washing water pH is 7) Finally, transfer the suspended powder on the filter to the new filter			
19 OR –	→ DS	NaCl (1.20 g/ cm ³)	Add NaCI solution Stir for 2 min Settle for 24 h	WPO (30% H ₂ O ₂)	1.0 µm glass microfiber filter	40 °C for 24 h	(Peng et al., 2017)
20 DS –	→ OR	NaI - NaCl (1.50 g/cm ³)	Add 250 mL of NaI - NaCl solution Air floatation for 40 s Settle for 5 min	WPO (30 mL of 35% H_2O_2) at room temperature for 7	0.45 µm membrane filter	N/A	(Han et al., 2019b)

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Dens	sity separat	tion and organic rei	noval for MPs analysis in soil.				
No.	Order	Extraction (density se	paration; DS)	Organic removal (OR)	Filtration		Ref.
		Solution	Procedure	1	Filter type	Drying conditions	
-	DS → OR	NaCl (1.19 g/cm ³)	Add NaCl solution Ultrasonication for 2 min Stir for 30 min Settle for 24 h Repeat three times	WPO (30% H ₂ O ₂) at 50 °C for 72 h	20 µm nylon net filter	Room temperature	(Liu et al., 2018)
0	$DS \rightarrow OR$	NaCl (1.20 g/cm ³)	 Add 160 mL of 27% NaCl solution Stir for 10 min Centrifuge at 3450 G for 30 min Filter supernatant using 0.45 µm membrane filter 	Wash particles from the filter and then treat them with 40–80 mL of 65% HNO ₃ at 90 °C for 48 h	0.2 µm infrared transparent anodisc filter	N/A	(Scheurer and Bigalke, 2018)
			2. Remove particles stuck to beaker walls by 27% NaCl and settle for 24 h				
			Mix sample mixture at 800 rpm for 10 min Centrifuge at 3450 G for 30 min				
σ	OR ↓ ↓	Nal (1.80 g/cm ³)	Add 150 mL of Nal solution Centrifuge again Wash with Nal solution	1. Pre-oxidation: WPO (10 mL of 35% H ₂ O ₂ + 1 mL of 10% FeSO ₄) at $50 ^{\circ}$ C Add a few drops of butyl acoohol (if frothing was excessive) Add 1 mL of 10% FeSO ₄ (to decompose the H ₂ O ₃) Add 30 mL of 0.5 M NaOH Store for 24 h Adjust to 150 mL using DI water Adjust to 150 mL using DI water Agitate for 20 min using ultrasonic Centrifuge at 2300 rpm for 10 min	1 mm, 0.25 mm, 0.05 mm sieve	0 °C	(Zhang and Liu, 2018)
				2. Post-oxidation: WPO (2 mL of 35% H ₂ O ₂)			
4	DS only	1. NaCl (1.20 g/cm ³) 2. NaI (1.60 g/cm ³)	 Use a continuous flow and floating separation apparatus Float using NaCl solution Clean residues in the sieve (300 mesh) were cleaned using water 	N/A	N/A	Air dry	(Zhou et al., 2018)
			 Residues were floated using NaI solution Filter if it contained a large number of solid particles 				

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Table 4

N0.	Order	Extraction (density sel	paration; DS)	Organic removal (OR)	Filtration		Ref.
	-	Solution	Procedure		Filter type	Drying conditions	
Ś	$\begin{array}{c} 1. \text{ OR} \\ \text{DS} \\ \text{OR} \\ \text{OR} \end{array}$ $\begin{array}{c} 2. \text{ DS} \\ \text{OR} \end{array}$	1. NaCl (1.20 g/cm ³), NaI (1.80 g/cm ³), ZnCl ₂ (1.50 g/cm ³) 2. Only using NaCl (1.20 g/cm ³)	 Test three types of floatation solutions Add 200 mL of a solution Stir at 200 rpm for 1 h using a shaker Settle for 48 h and siphon out 100 mL of supernatant Add 100 mL of the same floatation solution Stir for 30 min and settle for 48 h Combine supernatants Repeat steps but by only using NaCl solution 	 Pre-oxidation: WPO (200 mL of 30% H₂O₂) If temperature remained stable, heating at 70 °C Post-oxidation: WPO (30% H₂O₂ + H₂SO₄ (3:1, v/v)) 	20 µm nylon filter	Air dıy	(Li et al., 2019)
9	DS → OR	NaCl (1.24 \pm 0.05 g/ cm ³)	Add NaCl solution Stir for 30 min Settle for 24 h Repeat three times Filter with 20 →m	Wash filter WPO (100 mL of 30% H_2O_2) at 65 °C at 80 rpm for 72 h	20 µm nylon net filter	Room temperature	(Lv et al., 2019)
7	DS only	1. DI water (1.0 g/ cm ³) 2. NaCl (1.20 g/cm ³) 3. ZnCl ₂ (1.55 g/ cm ³)	 Add 20 mL of DI water stir at 21,000 rpm for 30 s Centrifuge at 2000 rpm for 15 min Filter supernatant using by 2–3 µm filter paper Add 20 mL of NaCl Stir, centrifuge and filter the same way Add 20 mL of ZnCl₂ Add 20 mL of ZnCl₂ Add 20 mL of ZnCl₂ Add 200 rpm for 30 s Centritize 2000 rpm for 15 min 	NA	2.5 µm cellulose filter paper	V/√	(Corradini et al., 2019)
8	$DS \rightarrow OR$	NaBr (1.55 g/cm³)	Filter the same way Add NaBr solution Stir for 5 min Settle for 2 h Use automatic cycling device for 30 min Filter using a 20 um filter	WPO (30% H ₂ O ₂) at 60 °C for 3 days	20 µm membrane filter	N/A	(Liu et al., 2019b)
6	$DS \rightarrow OR$	NaI	Add 200 mL of Nal solution Stir for 30 min Settle for 12 h Filter using a 7 µm glass fiber filter	WPO (100 mL of 30% H ₂ O ₂) for 12 h	7 μm glass fiber filter	Room temperature	(Huang et al., 2020)
10	$DS \rightarrow OR$	DI water (1.0 g/cm ³)	Leach soil specimens using pressurized water Add distilled water Shake for 30 s Precipitate for 5 min Repeat three times	Oxidation using 10 mL of 98% H ₂ SO ₄ Sonicate for 20 min Place in ice bath Dilute with distilled water	Membrane filter paper	°C s0−60°°C	(Li et al., 2020)
11	DS ↓ OR	1. NaCl (1.20 g/cm ³) 2. Nal (1.60 g/cm ³)	Disperse dried soil with (NaPO ₃) ₆ (0.5 mol/L) Add NaCl solution into beaker (flow rate 1.0 L/min) with air blowing at the bottom Collect the over-flow materials on a vibrating sieve (pore 50 µm) Settle residues on the sieve for 48 h in NaI solution	WPO (30% H ₂ O ₂) at 70 °C for 72 h	20 µm glass fiber filter	N/A	(Zhou et al., 2020)

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N0.	Order	Extraction (density se	paration; DS)	Organic removal (OR)	Filtration		Ref.
		Solution	Procedure	1	Filter type	Drying conditions	
			Filter using a 5 µm cellulose nitrate filter Repeat three times				
12	$\underset{OR}{DS} \rightarrow$	ZnCl ₂ (1.60 g/cm ³)	Add ZnCl ₂ solution Settle for 6 h at room temperature Repeat three times	WPO (30% H ₂ O ₂) at 70 °C in water bath	N/A	Air dry	(Duan et al., 2020)
13	$\underset{DS}{OR} \rightarrow$	NaCl (30%; w/v)	Add 600 mL of NaCl solution Settle for 12 h	WPO (50 mL of mixture (35% $H_2O_2 + 0.5$ M of ferrous sulfate)) at room temperature for $24-72$ h	300, 150, 50 µm sieve	N/A	(Rafique et al., 2020)
14	DS only	1. DI water $(1.0 \text{ g}/\text{cm}^3)$,	 For visual identification, centrifuge g of soil and 20 mL of DI water at 2000 rpm for 15 min 	N/A	1. 2.5 µm cellulose filter	Dried at 40 °C for 12 h (for	(Corradini et al., 2021)
		2. NaCl (1.20 g/cm ³),	Filter supematant through a 2.5 µm filter paper		paper (Jor Visual identification by	F LIK allalysis)	
		3. ZnCl ₂ (1.55 g/ cm ³)	Fill remaining sediments with 20 mL of NaCl Centrifuge at 21000 rpm for 30 s and filter supernatant a another time		2. 0.4 µm polycarbonate		
			Fill remaining sediments with 20 mL of $ZnCl_2$ Centrifuge last time and filter supernatant through the same filter used the previous two times		membrane filter (for FTIR analysis)		
			 For FTIR microscope analysis, sonicate g soil and 10 mL of ZnCl₂ for 10 min Agitate at 2000 rpm for 15 s Shake for 20 min at 180 oscillations per min Before filtration, centrifuge for 10 min at 2500 rpm 				
			Filter supernatant through a 0.4 µm polycarbonate membrane filter Re-fill with ZnCl ₂ and underwent same step once again				
15	DS only	Tap water	Add tap water Stir and settle overnight	N/A	1 mm sieve (collect >1 mm size)	Dry at 60 °C for 24 h	(Katsumi et al., 2021)
16	DS → OR	ZnCl ₂ (1.70 g/cm ³)	Add 500 mL of ZnCl ₂ solution Stir at 300 rpm for 5 min Settle for 24 h Repeat three times	WPO (20 mL of 0.05 M FeSO ₄ and 3 mL of conc. H ₂ SO ₄ and 20 mL of 30% H ₂ O ₂) at 75 °C for 24 h Rinse with water and sieve using a 1 mm sieve	0.45 µm membrane filter	Room temperature	(Choi et al., 2021)
17	$\underset{DS}{OR} \rightarrow$	ZnCl ₂ :CaCl ₂ (2:1.4; 1.55-1.58 g/cm ³)	 Filter digested supernatant through 20 µm sieve, afterward, visually sort and analyze the >1 mm particles 	WPO (200 mL of $30\% H_2O_2$)	0.7 µm glass fiber filter	N/A	(Kim et al., 2021)
			 Fill digested remnant with density solution Filter supernatant through 20 μm sieve Sort >1 mm Repeat this process two more times 				

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N0.	Order	Extraction (density s	eparation; DS)	Organic removal (OR)	Filtration		Ref.
		Solution	Procedure		Filter type	Drying conditions	
			3. Perform density separation of the particles of <1 mm obtained from the first process and >20 µm obtained from the second process and filter supernatant finally				
18	DS → OR	NaCl (1.19 g/cm ³)	Shake soil samples at 150 rpm for 30 min with 300 mL of distilled water using an orbital shaker Settle for 24 h Repeat 3-4 times until no more observed floating extract Centrifuge the collected supernatant solution for 29 min at 1500 pm Mix supernatant with 100 mL of NaCl solution for 15 min at 150 rpm and settle overnight Centrifuge supernatant for 10 min at 1500 rpm for 1 s	WPO (100 mL of 30% H_2O_2) for 7 days at 25 °C	0.45 µm cellulose nitrate filter	Room temperature	(Ragoobur et al., 2021)
19	$\underset{DS}{OR}\rightarrow$	CaCl ₂ (1.40 g/cm ³)	Centrifuge with CaCl ₂ solution for 10 min at a speed of 3700 min ⁻¹ Stain samples with Nile red	Fenton reaction by adding 30 mL of 30% H ₂ O ₂ , 30 mL of iron catalyst for 1 h at 0 $^{\circ}$ C	Glass fiber filter	C 09	(Grause et al., 2022)
20	$\underset{DS}{OR}\rightarrow$	NaCl (1.20 g/cm ³)	Add 100 mL of NaCl solution Stir for 1 min Settle for 1.5 h	WPO (40 mL of 33% (w/v) H ₂ O ₂) for 2 hat 60 °C Stir at 300 rpm	50 µm stainless steel filter	N/A	(Pérez- Reverón et al., 2022)

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Polymer		Ref.	Density	solution	Ref.
Type	Density (g/cm ³)		Type	Density (g/cm ³)	
Polypropylene (PP)	0.83-0.85	(Andrady, 2015)	Water	1.00	(Dikareva and Simon, 2019)
Polyethylene (PE)	0.91 - 0.94	(Andrady, 2015)	NaCl	1.20	(Wang et al., 2017)
Polystyrene (PS)	1.05	(Andrady, 2015)	$CaCl_2$	1.40	(Grause et al., 2022)
Polyamide or nylon (PA)	1.13-1.15	(Sundt et al., 2014)	K_2CO_3	1.54	(Gohla et al., 2021; Prepilková et al., 2022)
Polyurethane (PUR)	1.20	(Sundt et al., 2014)	NaBr	1.55	(Liu et al., 2019b)
Polyethylene terephthalate (PET)	1.37	(Andrady, 2015)	ZnBr_2	1.71	(Quinn et al., 2017)
Polyvinyl chloride (PVC)	1.38	(Andrady, 2015)	$ZnCl_2$	1.60 - 1.80	(Scherer et al., 2020)
Polytetrafluorethylene (PTFE)	2.20	(Sundt et al., 2014)	NaI	1.60 - 1.80	(Dikareva and Simon, 2019; Zhang and Liu, 2018)

N0.	Matrix	Recovery test method	Type and size of MPs	Recovery rate	Ref.
-	Sediments	Spike 10 of each size range for all types of MPs Repeat in duplicate	PE, PS, PP, PVC (<1 mm, 1–2 mm, 2–5 mm)	80, 90, 100% (PE, PP; each size group), 70, 90, 100% (PS; each size group), 60, 80, 90% (PVC; each size group)	(Di and Wang, 2018)
5		Spike 300 MPs into 200 g of the muffled sand Filter using 10 mm stainless steel filter Count MPs using a stereomicroscope Repeat in triplicate	Red PS (beads) (100 µm)	66 ± 5.6%	(Liu et al., 2019a)
ω		Spike 10 beads or 15 secondary MPs into 20 g of dried sediment Repeat in triplicate for each MPs type	Beads: PE (0.71–0.85 mm), PP (2.45 mm), PS (4.4 mm) (4.4 mm) Secondary MPs: Nylon toothbrush bristles, PP cleaning brush bristles, rope fragments and PE mesh fruit packaging fragments	100% (microbead) 49 ± 10.2–58 ± 7.7% (secondary MPs)	(Blair et al., 2019)
4		Assess adapted ZnCl ₂ protocol in a pilot spike-recovery test Spike 40 (62–250 µm) and 20 (250–700 µm) MPs into 60 g of sediment Repeat in triplicate	PMMA fragments (1.18 g/cm ³) (62–125 µm, 125–250 µm, 250–500 µm, 500–700 µm)	$\pm 8.7\%$ (62–125 µm), 80.8 $\pm 6.3\%$ (125–250 µm), $\pm 5.0\%$ (220–500 µm), $\pm 0.0\%$ (500–700 µm)	(Mani et al., 2019)
2		Spike 40–80 MPs into 12.3 g of dried sediment Filter using 0.7 µm GF/F and dry at 40 °C overnight	Acrylic, nylon, polyester, PE fibers	70%	(Simon-Sánchez et al., 2019)
9		Spike 125 MPs into 1500 g of quartz sand	PE, PP, PS, PMMA, PVC (25 particles each, sized 125-1000 µm)	$87.2 \pm 4.5\%$	(Scherer et al., 2020)
٢		Spike 10 MPs into 200 g of elutriated soil and sediment Repeat five times	PE, PP, PVC, PET, PS, expandable polystyrene (EPS) (<1 mm)	%06<	(Han et al., 2019b)
∞	Soils	Stain MPs with Nile Red in acetone (500 µg/mL) for 10 min Spike 20 MPs into 50 g of clean soil Repeat in triplicate	PP, PE, PA, PET, PVC, PC, acrylonitrile butadiene styrene (ABS), PMMA, PS (1–5 mm)	90.0 \pm 10.0% (PP, PA), \pm 0.0% (PE, PC, PMMA), 95.0 \pm 5.0% (ABS), 96.7 \pm 2.9% (PS), 0.0 \pm 0.0% (PET, PVC due to the high density)	(Liu et al., 2018)
6		Use 27% NaCl (1.2 g/cm ³) for small MPs Spike 10 MPs into 50 g of pure plastic free sand Test four different ways of density separation: 1. Sedimentation cylinder method, settle for 48 h 2. Self-constructed MP separator with additional air bubbling 3. Stir for 10 min and centrifuge for 30 min at 3450 G 4. Identical with third but with rubber disc inserted after centrifugation Removal of organic matter in a short time using HNO ₃	PP (0.5–1 mm)	93% (3), 97% (1,2), 98% (4)	(Scheurer and Bigalke, 2018)
10		Spike MPs into cleaned sands	PP, PE (200 µm-5 mm)	97%	(Zhou et al., 2018)
11		Spike 10 MPs into 50 g of soil Pre-digestion (30% H ₂ O ₂), density separation and filtration	PE, PP, PS, PA, ABS, PET (<2 mm)	Nal: 97.78 ± 1.57% NaCI: 80.56 ± 2.08% (except PET) ZnCl ₃ : 97.22 ± 2.08%	(Li et al., 2019)

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Table 6

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No.	Matrix	Recovery test method	Type and size of MPs	Recovery rate	Ref.
12		Spike each type of MPs into 5 g of soil sample Repeat in triplicate	Fiber: Acrylic (length: $2.7 \pm 1.4 \text{ mm}$) (width: $0.04 \pm 0.01 \text{ mm}$) (area: $0.12 \pm 0.06 \text{ mm}^2$) Polyester (length: $1.6 \pm 1.1 \text{ mm}$) (width: $0.04 \pm 0.01 \text{ mm}$) (area: $0.07 \pm 0.06 \text{ mm}^2$) Nylon (length: $2.3 \pm 0.8 \text{ mm}$) (width: $0.05 \pm 0.01 \text{ mm}$) (area: $0.98 \pm 0.37 \text{ mm}^2$)	98% (LDPE), 90% (Polyester), 88% (PVC), 77% (Nylon), 49% (Acrylic)	(Corradini et al., 2019)
			LDPE (0.16 \pm 0.1 mm ²) PVC (0.10 \pm 0.08 mm ²)		
13		Stain MPs with Nile Red in acetone for 10 min except for PMMA, PS and ABS Calculate each recovery rate using three density solutions (NaCl, NaBr or CaCl ₂)	Type: PE., PET, POM, PVC, PC, ABS, PMMA, PS (shredding/grinding 3 mm bead and passed through 7–160 mesh) Size: PE (100–500 µm, 500–1000 µm, 1000–3000 µm) Size: PE (particle, fiber, film) (100–500 µm) Shape: PE (particle, fiber, film) (100–500 µm)	 >90% (All type of MPs) 75.0– 96.7% (100–500 µm PE) 100% (>1 mm PE) 65–98.3% (Three shape-different PE; of them, NaBr: 85–98.3%) 	(Liu et al., 2019b)
14		Spike 20 films or particles into 20 g of clean soil	LDPE film (1 × 1 mm) LDPE particle (250 µm)	100% (film), 98% (particle)	(Huang et al., 2020)
15		Spike 30 mg of MPs into 200 g of clean soil sample Stir for 20 min at 25 °C Weigh by electronic balance	PE (>2 mm, 0.9–2 mm, 0.28–0.9 mm, 0.15–0.28 mm)	96% (>2 mm) 85% (0.9–2 mm) 87% (0.28–0.9 mm) 84% (0.15–0.28 mm)	(Li et al., 2020)
16		Spike 30 MPs into 50 g of clean soil Count by stereomicroscope	PVC, PE, PP, PS, Polyamide (500 µm-2 mm)	75.9–112.4%	(Zhou et al., 2020)
17		Spike each 10 of six type MPs with two size range	PE (sheet), PET, PP, PS (fragment), PVC, nylon (fiber) (300–500 μm, 700–1000 μm)	$97.8 \pm 4.8\%$	(Kim et al., 2021)
18		Spike MPs into clean soil	High density polyethylene (HDPE), PP, LDPE, PS, PP fibers (1.2–1.6 mm)	85% (HDPE), 90% (PP), 85% (LDPE), 100% (PS)	(Ragoobur et al., 2021)
19		Soil sample was heated at 500 $^{\circ}$ C to remove MPs and organic materials Spike 100 mg of MPs into 10 g of soil and centrifuge at a speed of 3700 min ⁻¹ for 10 min	LDPE, PP, PS, flexible PVC (63–1000 μm) (recovery test) PET (63–150 μm, 150–212 μm, 212–500 μm, 500–1000 μm, 63–1000 μm) (Tests to identify how size affects MPs recovery)	97–98% (LDPE, PP, PS, FPVC) 81% (PET; 63–150 µm), 94.0 ± 2.1% (PET; 500–1000 µm), 94.9 ± 2.4% (PET; 63–1000 µm)	(Grause et al., 2022)