

1 **Deep autoencoder-based behavioral pattern recognition outperforms standard statistical**
2 **methods in high-dimensional zebrafish studies**

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48 **Abstract**

49 Zebrafish have become an essential tool in screening for developmental neurotoxic chemicals
50 and their molecular targets. The success of zebrafish as a screening model is partially due to their
51 physical characteristics including their relatively simple nervous system, rapid development,
52 experimental tractability, and genetic diversity combined with technical advantages that allow
53 for the generation of large amounts of high-dimensional behavioral data. These data are complex
54 and require advanced machine learning and statistical techniques to comprehensively analyze
55 and capture spatiotemporal responses. To accomplish this goal, we have trained semi-supervised
56 deep autoencoders using behavior data from unexposed larval zebrafish to extract quintessential
57 “normal” behavior. Following training, our network was evaluated using data from larvae shown
58 to have significant changes in behavior (using a traditional statistical framework) following
59 exposure to toxicants that include nanomaterials, aromatics, per- and polyfluoroalkyl substances
60 (PFAS), and other environmental contaminants. Further, our model identified new chemicals
61 (Perfluoro-n-octadecanoic acid, 8-Chloroperfluorooctylphosphonic acid, and
62 Nonafluoropentanamide) as capable of inducing abnormal behavior at multiple chemical-
63 concentrations pairs not captured using distance moved alone. Leveraging this deep learning
64 model will allow for better characterization of the different exposure-induced behavioral
65 phenotypes, facilitate improved genetic and neurobehavioral analysis in mechanistic
66 determination studies and provide a robust framework for analyzing complex behaviors found in
67 higher-order model systems.

68

69 **Author Summary**

70 We demonstrate that a deep autoencoder using raw behavioral tracking data from zebrafish
71 toxicity screens outperforms conventional statistical methods, resulting in a comprehensive
72 evaluation of behavioral data. Our models can accurately distinguish between normal and
73 abnormal behavior with near-complete overlap with existing statistical approaches, with many
74 chemicals detectable at lower concentrations than with conventional statistical tests; this is a
75 crucial finding for the protection of public health. Our deep learning models enable the
76 identification of new substances capable of inducing aberrant behavior, and we generated new
77 data to demonstrate the reproducibility of these results. Thus, neurodevelopmentally active
78 chemicals identified by our deep autoencoder models may represent previously undetectable
79 signals of subtle individual response differences. Our method elegantly accounts for the high
80 degree of behavioral variability associated with the genetic diversity found in a highly outbred
81 population, as is typical for zebrafish research, thereby making it applicable to multiple
82 laboratories. Utilizing the vast quantities of control data generated during high-throughput
83 screening is one of the most innovative aspects of this study and to our knowledge is the first
84 study to explicitly develop a deep autoencoder model for anomaly detection in large-scale
85 toxicological behavior studies.

86 Introduction

87 Significant progress continues to be made in our understanding of neurodevelopmental disorders
88 such as autism spectrum disorder, attention deficit hyperactivity disorder (ADHD), developmental
89 delay, learning disabilities, and other neurodevelopmental problems. As incidences continue to
90 rise globally and affect 10-15% of all births, more work must be done to improve our
91 understanding of these disorders (Boyle et al., 2011; *Neurodevelopmental Diseases*, 2021; US
92 EPA, 2015b). Meta-analyses suggest strong and consistent epidemiological evidence that the
93 developing nervous system is particularly vulnerable to low-level exposure to widespread
94 environmental contaminants, as the anatomical and functional architecture of the human brain is
95 mainly determined by developmental transcriptional processes during the prenatal period
96 (Grandjean & Landrigan, 2014; Green & Planchart, 2018; Miller et al., 2014; Rock & Patisaul,
97 2018; US EPA, 2015b). Therefore, identifying associations between developmental exposures and
98 neurological effects is a core objective to improve public health by informing disease and disability
99 prevention (*A Blueprint for Brain Development*, 2014; *Neurodevelopmental Diseases*, 2021).

100 As the number of environmental contaminants grows to nearly one million, comprehensive data
101 on the neurodevelopmental toxicity of these contaminants remain sparse or nonexistent (Krewski
102 et al., 2020; US EPA, 2015a, 2015b; Wambaugh et al., 2013). In response, high-throughput
103 screening (HTS) assays have been developed to expedite chemical toxicity testing using *in vitro*
104 and *in vivo* systems (Judson et al., 2010; Richard et al., 2016; Truong et al., 2014). However, *in*
105 *vitro* cell and cell-free assays cannot fully capture systemic organismal responses in terms of
106 anatomy, physiology, or behavior (Thomas et al., 2012). Zebrafish (*Danio rerio*) have emerged
107 as an ideal model for studying low-level chemical exposure because of their high fecundity,
108 rapid development, genetic tractability, and amenability to high-throughput data generation
109 (Bugel et al., 2014; Planchart et al., 2018; Truong et al., 2014). The zebrafish brain's structural
110 organization, cellular morphology, and neurotransmitter systems are very similar to other
111 vertebrates, including chickens, rats, and humans (Horzmann & Freeman, 2016; Kalueff et al.,
112 2014; Lowery & Sive, 2004; Tropepe & Sive, 2003). Furthermore, zebrafish have behavioral
113 patterns highly similar to mammals, and genetic homologs for 70% of human genes and 82% of
114 human disease genes, making them a powerful tool for revealing the neuronal developmental
115 pathways underlying behavior (Basnet et al., 2019; Howe et al., 2013; Postlethwait et al., 1998).

116 Zebrafish larvae show mature swimming patterns following swim bladder development at four to
117 five days post-fertilization (dpf), which can be assessed using various locomotor behavioral
118 assays (Hernandez et al., 2018; Tegelenbosch et al., 2012). One of these assays, the larval
119 photomotor response (LPR), utilizes a sudden transition from light to dark, eliciting a
120 stereotyped large-angle O-bend, followed by several minutes of increased movement, which
121 gradually reduces (Burgess & Granato, 2007c; Emran et al., 2008). Exposure to toxicants has
122 been shown to alter this stereotypical behavioral response (Basnet et al., 2019; Truong et al.,
123 2016). Current HTS for behavioral neurotoxicity focuses heavily on analyzing locomotor
124 behavior using distance moved and population-based statistical methods (Basnet et al., 2019; G.
125 Zhang et al., 2017). However, while the behavior repertoire of larval zebrafish is less
126 sophisticated when compared to that of adult zebrafish and other higher-order vertebrates, they
127 are capable of numerous distinct behaviors (Basnet et al., 2019; Kalueff et al., 2013; Mirat et al.,
128 2013). These behaviors, such as thigmotaxis, and light avoidance cannot always be captured
129 when using distance moved as a sole indicator of neurobehavioral toxicity in analyses of this
130 data. Moreover, as most laboratory zebrafish populations feature significant genetic
131 heterogeneity, individual responses to exotic toxicants cannot be expected to be homogeneous
132 for simplistic measures such as distance moved (Balik-Meisner et al., 2018).

133 Improved accessibility to computing resources and application interfaces, together with recent
134 advances in deep-learning makes it possible to analyze complex behavioral data in novel ways
135 and predict neurodevelopmental toxicity (Arifoglu & Bouchachia, 2017; Pereira et al., 2020; Xia
136 et al., 2018). The volume and diversity of data generated during HTS experiments, combined
137 with the variety in toxicological response within populations, present an opportunity that is well-
138 suited for machine learning (ML). In particular, analysis of zebrafish HTS data from five dpf
139 larvae exposed to 1,060 unique chemicals reveals that only 8% of chemical-concentration pairs
140 (a unique combination of chemical and concentration, e.g. 6.4 μ M Nicotine) exhibit changes in
141 distance moved (G. Zhang et al., 2017), which is alarmingly low given the known toxicity
142 profiles of the chemical set. This challenge provides an opportunity to apply methods developed
143 for anomaly detection from areas such as financial fraud (Awoyemi et al., 2017), medical
144 application faults (Pachauri & Sharma, 2015), security systems intrusion (Sargolzaei et al.,
145 2016), system faults (Warriach & Tei, 2013), and others (Fazai et al., 2019; Jaiswal & Ruskin,

146 2019). In anomaly detection, we learn the pattern of a normal process, and anything that does not
147 follow this pattern is classified as an anomaly. This learning model is particularly applicable, as
148 many HTS data sets have large amounts of control data to analyze (G. Zhang et al., 2017). One
149 intriguing approach to achieving this is by applying an autoencoder (Feng et al., 2021; Frassek et
150 al., 2021; Goodfellow et al., 2016; Le Borgne et al., 2022; Nicholaus et al., 2021; Ranjan et al.,
151 2019). An autoencoder is a neural network of two modules, an encoder and a decoder
152 (Goodfellow et al., 2016; Gupta & Singh, 2019). The encoder learns the underlying features of a
153 process, and these features are typically in a reduced dimension. The decoder then uses this
154 reduced dimension to recreate the original data from these underlying features.

155 In the present study, we trained deep autoencoder models to recognize the pattern of
156 quintessential larval zebrafish behavior and identify abnormal behavior following developmental
157 chemical exposure. The performance of our deep autoencoders was compared against traditional
158 statistical methodologies, the gold standard for behavioral assessment. In addition to model
159 development, we assessed the features driving performance through feature permutation and
160 generated new confirmatory data to assess model reproducibility and confirm novel findings.

161 **Results**

162 **Statistical classification of behavior**

163 After classifying each of the 96-well plates by differences in the movement of controls into
164 hyperactive, normal, or hypoactive, we compared treated vs control behavioral response to
165 light/dark cycling in zebrafish larvae at five dpf. We identified 39 chemical-concentration
166 combinations from ten chemicals capable of inducing a significantly different ($p < 0.05$)
167 behavioral response (Supp. Table 2). Using the 30th and 70th percentiles, we defined 227
168 individual larvae as abnormal (Fig 1a). These 227 larvae formed the validation set used to test
169 the performance of our models.

170 **Training performance**

171 Autoencoder models were trained using only control data for each of the activity states
172 (hypoactive, normal, and hyperactive) per phase of the second light cycle. This resulted in six
173 trained models (Supp Fig 1 the training loss plots for the models). Table 1 shows the results for
174 the six deep autoencoder models trained using control data and validated using data from
175 zebrafish defined as abnormal using the K-S test. All the models performed well with values
176 ranging from 0.615 – 0.867 and 0.740 – 0.922 for the Kappa and AUROC, respectively. As
177 expected, the models consistently produced high specificity (SP) levels as this value indicated
178 how well the models classify control data. There was greater variability in the sensitivity (SE)
179 with the dark phase models matching or outperforming the light phase models for each activity
180 state. Further, we observed a noteworthy trend across all models producing high positive
181 predictive value (PPV). Overall, these results show that deep autoencoders trained using control
182 data is capable of distinguishing between normal and abnormal larval zebrafish behavior with a
183 high degree of accuracy.

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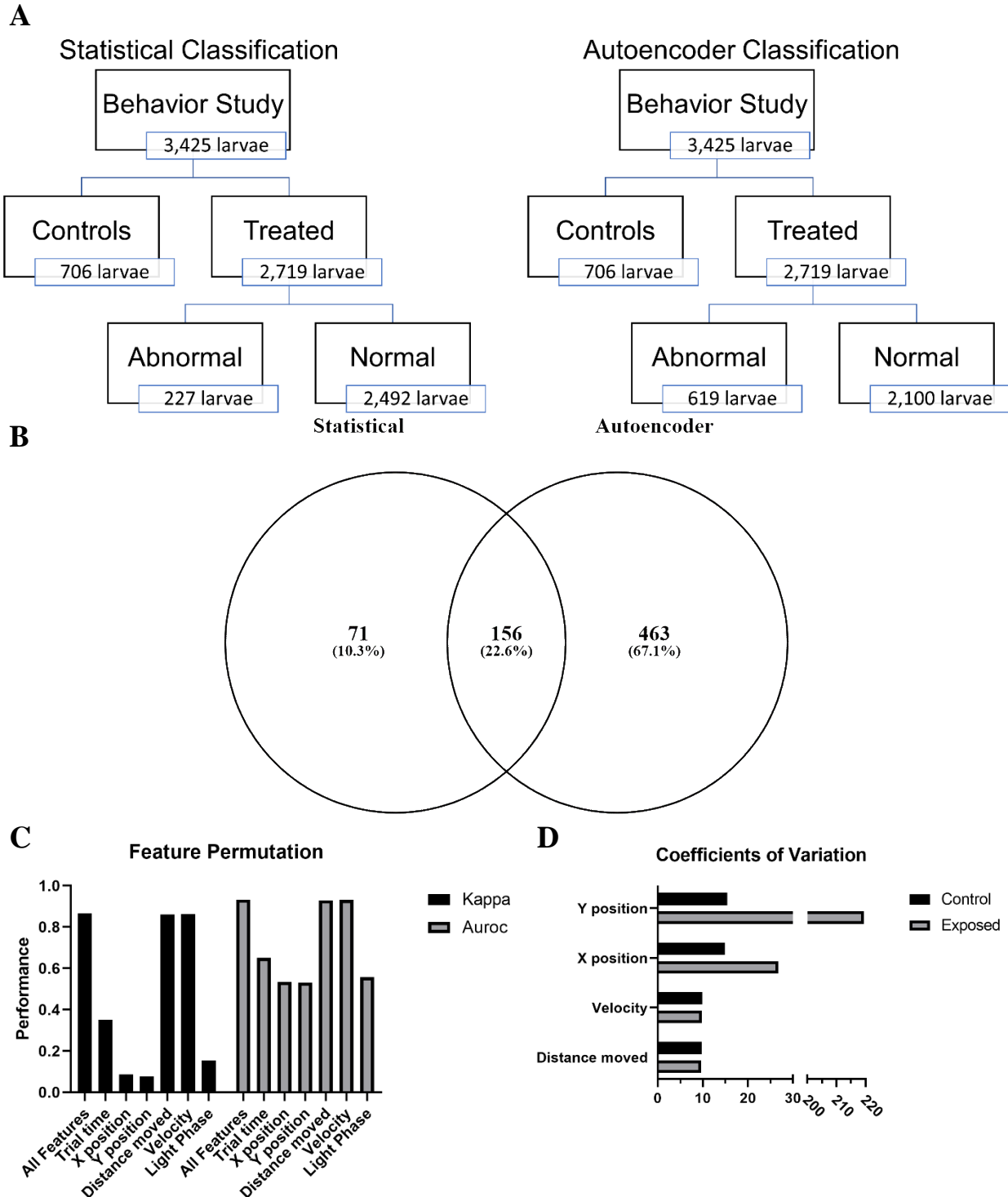


Figure 1: Assessment autoencoder performance. (A) Schematic representation of the differences in statistical and autoencoder based classification of behavioral response in larval zebrafish. (B) Venn diagram showing overlap between statistical and autoencoder classified abnormal zebrafish. (C) Evaluating the change in model performance when the values of a single feature are randomly shuffled. Kappa – Cohen’s Kappa statistic, AUROC - area under the receiver operating characteristic. Figure depicts means \pm SEM. (D) Coefficients of variation for each of the main numerical features.

186 Table 1. Deep autoencoder model performance in behavioral classification. Table showing
187 performance of model trained using different activity states of the control data in both light and
188 dark phases.

Model		Performance Metrics				
Baseline Control Activity Level	Light Phase	SE	SP	PPV	Kappa	AUROC
Hypoactive	Light	78.5	100	99.7	0.867	0.892
	Dark	78.3	98.0	88.4	0.800	0.882
Normal	Light	48.3	99.7	93.1	0.615	0.740
	Dark	73.3	94.8	77.6	0.695	0.840
Hyperactive	Light	79.2	97.5	85.5	0.790	0.883
	Dark	86.9	97.5	90.2	0.855	0.922

189 Evaluation of unknowns

190 Using the six trained models, we evaluated the 2,719 treated zebrafish larvae (Fig 1). The
191 autoencoders correctly classified 156 of the 227 larvae that fell below or above the 30th and 70th
192 percentiles, respectively. In addition, our deep autoencoders identified 463 larvae as abnormal
193 from the 2,492 larvae defined as normal using the K-S test (Fig 1b). The majority (422) of these
194 619 larvae were from one of 66 chemical-concentration combinations from 13 chemicals (Table
195 2). The deep autoencoders successfully identified nine of the ten statistically abnormal chemicals
196 and identified these chemicals at or below the lowest concentration shown to be statistically
197 significant. While the deep autoencoders did not identify Perfluorodecylphosphonic acid as
198 capable of inducing abnormal behavior, but they did identify 3-Perfluoropentyl propanoic acid
199 (5:3), Perfluoro-n-octadecanoic acid, 8-Chloroperfluorooctylphosphonic acid, and
200 Nonafluoropentanamide, which were missed in the statistical testing framework. These results,
201 summarized in fig 2, show that deep autoencoders can match the performance of the K-S test and
202 are more sensitive at detecting abnormal behavior.

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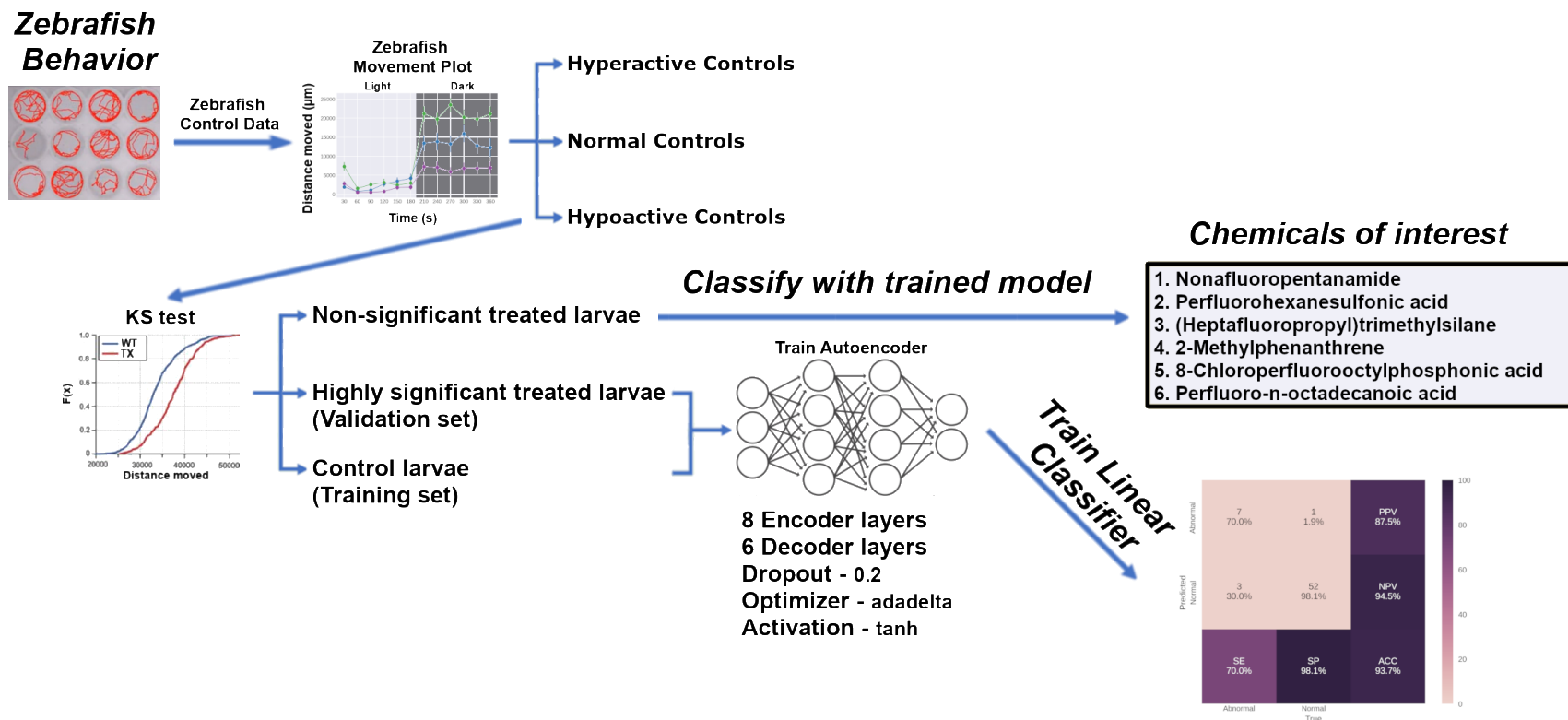


Figure 2: Summary of behavioral analysis pipeline and results. Utilizing our analysis pipeline produced six deep autoencoder models (three for the light phase and three for the dark phase) capable of classifying larval zebrafish behavior with high Kappa and AUROC values. The trained models were then used to classify the non-significant exposed larvae and identified Nonafluoropentanamide, Perfluorohexanesulfonic acid, (Heptafluoropropyl)trimethylsilane, 2-Methylphenanthrene, 8-Chloroperfluorooctylphosphonic acid, Perfluoro-n-octadecanoic acid, and others as capable of inducing abnormal behavior.

205 Table 2. Autoencoders identified chemicals. Table showing chemicals and concentrations
206 flagged for displaying abnormal behavioral when evaluated using Autoencoder. Compounds that
207 were picked up by Autoencoder, but not KS test are highlighted in red.

CASRN	Chemical Name	Concentration (μ M)
71751-41-2	Abamectin	0.1, 0.2, 0.4, 0.6
308068-56-6	Multi-Walled Carbon Nanotube	10, 23.2, 50, 75, 100
2531-84-2	2-Methylphenanthrene	1, 2.54, 6.45, 16.4, 35, 74.8, 100
832-69-9	1-Methylphenanthrene	1, 2.54, 6.45, 16.4, 35, 74.8, 100
914637-49-3	3-Perfluoropentyl propanoic acid (5:3)	0.25
192-51-8	Dibenzo[e-l]pyrene	0.01, 0.025, 0.065, 0.164, 0.35, 0.75, 1, 2.54, 16.4, 35, 100
16517-11-6	Perfluoro-n-octadecanoic acid	0.25
355-46-4	Perfluorohexanesulfonic acid	0.015, 0.14, 0.41, 3.7, 11.1, 33.3, 66.5, 100
3834-42-2	(Heptafluoropropyl)trimethylsilane	0.015, 0.046, 0.41, 1.23, 11.1, 33.3
	8-Chloroperfluorooctylphosphonic acid	0.167
31253-34-6	2-Aminohexafluoropropan-2-ol	0.015, 0.046, 0.41, 1.23, 3.7, 11.1, 33.3, 66.5, 100
13485-61-5	Nonafluoropentanamide	0.41, 3.7, 11.1
439-14-5	Diazepam	1, 3, 5, 8, 12

208

209 Features driving improved autoencoder performance

210 To determine the features in the model that were most important in driving classification
211 performance, we employed permutation feature importance. This technique is a model agnostic
212 inspection technique used for any fitted estimator to determine the importance of each feature in
213 the model. Larger the decrease in model performance (Kappa or AUROC) when a single feature
214 value is randomly shuffled, the more important the feature. Our results, shown in fig 1c, indicate
215 that phase, trial time, x position, and y position are the largest drivers of model performance,
216 while distance moved and velocity contribute very little. Coefficients of variation show greater
217 variability in the x and y positional data between control and exposed groups compared to either
218 velocity or distance moved (fig 1d). This trend is consistent irrespective of the larval activity
219 state (hypoactive, normal activity, or hyperactive) relative to their respective controls (Fig 3).

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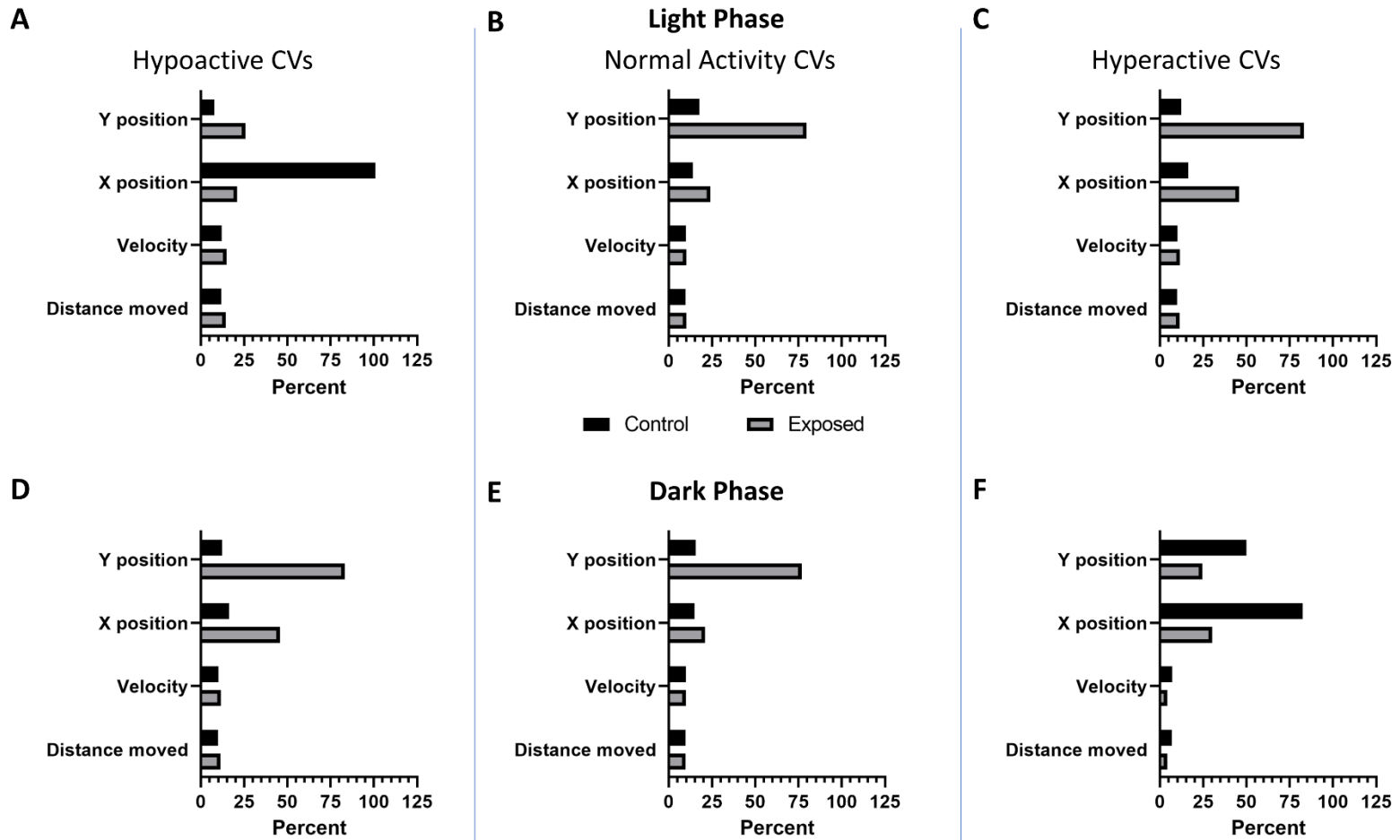


Figure 3: Coefficients of variation per larval activity state. Coefficients of variation (CVs) for each of the main numerical features (A – C) in the light (D – F) and in the dark. Columns show CVs of larval zebrafish significantly ($p < 0.05$) (A, D) hypoactive, (B, E) normal activity, or (C, F) hyperactive relative to their respective controls.

223 Experimental confirmation of autoencoder findings

224 To provide an unbiased evaluation of the final model fits, we generated new data using 2-
225 Methylphenanthrene, and Nonafluoropentanamide. The data collected confirmed that our models
226 accurately classified all controls as normal while detecting similar levels of abnormal behavior
227 response across the concentration range (Fig 4). These results show that the trained model is
228 capable of producing similar results across experimental replicates.

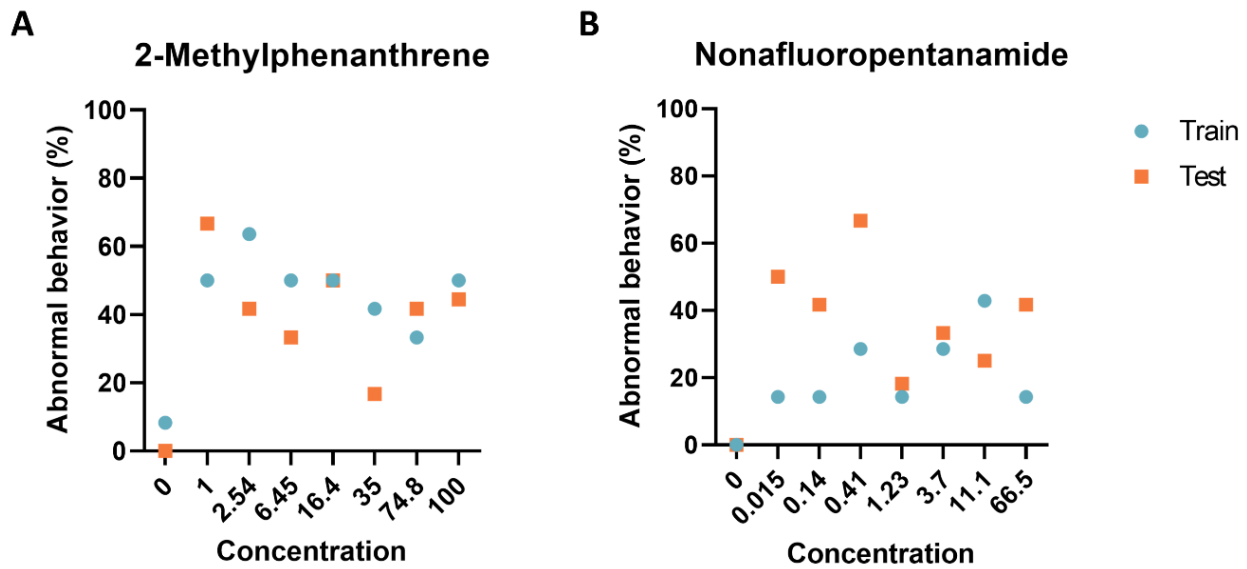


Figure 4: Experimental model evaluation. Comparison of the performance of deep autoencoder models between the training set and two chemicals identified by the models to elicit abnormal larval zebrafish behavior. Percent of larval zebrafish classified as abnormal based on their behavioral response to developmental exposure to (A) 2-Methylphenanthrene and (B) Nonafluoropentanamide

229

230 Discussion

231 Statistical analysis identified 39 chemical-concentration combinations from ten chemicals
232 capable of inducing a significantly different ($p < 0.05$) behavioral response. Utilizing the 227
233 abnormal individuals identified by the statistical test as our validation set, we trained six deep
234 autoencoder models using control data for each of the activity states (hypoactive, normal, and
235 hyperactive). All of the resulting models performed well with values ranging from 0.615 – 0.867
236 and 0.740 – 0.922 for the Kappa and AUROC, respectively. All models achieved SP values
237 above 94.8% and PPV values above 77.6% while SE values for all dark phase models
238 outperformed the light phase models for each activity state (Table 1). Assessment of permutation
239 feature importance indicates that phase, trial time, x-position, and y-position are the largest
240 drivers of model performance (fig 1c). The calculated coefficients of variation shed some light
241 on this surprising finding (fig 1d). They show that variation in the x and y positional data is
242 greater than observed for velocity or distance moved between control and exposed groups. These
243 differences in variation likely make it easier for the models to distinguish between treated and
244 exposed groups.

245 When we examined exposed larvae defined as normal using the K-S test (Fig 1), our deep
246 autoencoders identified 66 chemical-concentration combinations from 12 chemicals (Table 2)
247 with Perfluoro-n-octadecanoic acid, 8-Chloroperfluorooctylphosphonic acid, and
248 Nonafluoropentanamide only identified by our autoencoders. These results show that a deep
249 autoencoder-based model can classify larval zebrafish behavior as normal or abnormal with very
250 good efficacy and often identified abnormal behaviors at lower concentrations than current
251 statistical methods. Further, the models identified three novel chemicals, Perfluoro-n-
252 octadecanoic acid, 8-Chloroperfluorooctylphosphonic acid, and Nonafluoropentanamide as
253 capable of inducing abnormal behavior (Fig 3).

254 Recognition and categorization of swimming patterns in larvae is a challenging task and a
255 number of approaches have been used. These can range from subjective analysis based on
256 experienced observations (Fero et al., 2011; Kalueff et al., 2013, p. 0) or more recently through
257 the application of unsupervised ML (Budick & O'Malley, 2000; Burgess & Granato, 2007a,
258 2007b, 2007c; Kimmel et al., 1974; Mirat et al., 2013; H. Zhang et al., 2013). These studies have
259 focused on the analysis and categorization of behavioral patterns in wild-type strains (Burgess &

260 Granato, 2007c; H. Zhang et al., 2013), mutant strains (Burgess & Granato, 2007b; Mirat et al.,
261 2013), or larvae exposed to neuroactive chemicals (Mirat et al., 2013) but do not classify
262 behavior as normal or abnormal. In addition, these unsupervised approaches have utilized
263 highspeed camera systems which are medium to low throughput and have limited potential in the
264 screening of tens of thousands of chemicals for behavioral effects. As introduced above,
265 classification of behavior is one of the primary goals of toxicological screening and tends to
266 result in highly imbalanced datasets and lend themselves to anomaly detection methodologies.
267 While these methods are common in manufacturing (Fan et al., 2018; Fazai et al., 2019; Jaiswal
268 & Ruskin, 2019; Nicholaus et al., 2021), information systems (Pachauri & Sharma, 2015;
269 Warriach & Tei, 2013), security systems (Feng et al., 2021; Sargolzaei et al., 2016), and financial
270 fraud (Awoyemi et al., 2017) they have only very recently been applied to biological data
271 (Frassek et al., 2021; Homayouni et al., 2021; Nwokedi et al., 2021). To the best of our
272 knowledge, this is the first study to explicitly develop a deep autoencoder model for anomaly
273 detection in toxicological behavior studies.

274 Overall, our results show that a deep autoencoder utilizing raw behavioral tracking data from
275 five dpf zebrafish larvae can accurately distinguish between normal and abnormal behavior. We
276 show that these results are reproducible and allow for the identification of new compounds
277 capable of eliciting abnormal behavior. Further, our models were able to identify abnormal
278 behavior following chemical exposure at lower concentrations than with traditional statistical
279 tests. Our approach accounts for the high degree of behavioral variability associated with the
280 genetic diversity found within a highly outbred population typical of zebrafish studies, thereby
281 making it extensible to use across labs. Looking to the future, neurodevelopmentally active
282 chemicals identified using our deep autoencoder models may represent heretofore undetectable
283 signals of subtle differences in individual responses, suggesting chemicals that should be
284 investigated further as eliciting differential population responses (i.e. interindividual
285 susceptibility differences).

286 These findings will facilitate the application of behavioral characterization methods discussed
287 above, such as Zebrazoom (Mirat et al., 2013), using highspeed cameras to identify the
288 behavioral traits most perturbed by the chemical exposure and allow for more mechanistic
289 discovery. One of the key innovations presented in this study is leveraging vast amounts of

290 control data generated as part of any high-throughput screening (HTS) – setting the stage for
291 predictive toxicological applications and safety assessments for the enormous backlog of as-yet
292 untested chemicals.

293

294 **Materials and methods**

295 This section describes the autoencoder models utilizing a semi-supervised ML algorithm and
296 logistic regression (LR) to discriminate between normal and abnormal behavior in chemically
297 exposed five dpf zebrafish. An overview of our approach is shown in Fig 3. Briefly, we created
298 and trained six autoencoder models for each phase of the assay; namely, hyperactive, normal,
299 and hypoactive depending on the control movement in the light or dark phases of the assay.
300 Finally, treated plates were tested on one of these, depending on which category, its controls fell
301 under. We used experimental data collected on a large and diverse compound set of 30
302 chemicals including an insecticide, nanomaterial, perfluorinated chemicals, and aromatic
303 pollutants at a range of concentrations (133 chemical-concentration pairs) to assess the
304 neurotoxic effects of these chemicals following developmental exposure (Supp. Table 1).

305 **Zebrafish husbandry**

306 Tropical 5D wild-type zebrafish were housed at Oregon State University's Sinnhuber Aquatic
307 Research Laboratory (SARL, Corvallis, OR) in densities of 1000 fish per 100-gallon tank
308 according to the Institutional Animal Care and Use Committee protocols (Barton et al., 2016).
309 Fish were maintained at 28 °C on a 14:10 h light/dark cycle in recirculating filtered water,
310 supplemented with Instant Ocean salts. Adult, larval and juvenile fish were fed with size-
311 appropriate GEMMA Micro food 2–3 times a day (Skretting). Spawning funnels were placed in
312 the tanks the night prior, and the following morning, embryos were collected and staged
313 (Kimmel et al., 1995; Westerfield, 2007). Embryos were maintained in embryo medium (EM) in
314 an incubator at 28 °C until further processing. EM consisted of 15 mM NaCl, 0.5 mM KCl,
315 1 mM MgSO₄, 0.15 mM KH₂PO₄, 0.05 mM Na₂HPO₄, and 0.7 mM NaHCO₃ (Westerfield,
316 2007).

317 **Developmental chemical exposure**

318 The empirical data used to develop our model were gathered as described in Truong et al. and
319 Noyes et al. (Noyes et al., 2015; Truong et al., 2014, 2022). The experimental design consisted of
320 the 30 unique chemicals tested (Supp Table 1) with at least 7 replicates (an individual embryo in
321 singular wells of a 96-well plate) at each concentration for each chemical.

322 **Developmental toxicity assessments**

323 *Mortality and morphology*

324 At 24 hours post-fertilization (hpf), embryos were screened for mortality and 2 developmental
325 endpoints. At 120 hpf, mortality and incidence of abnormality in 9 morphology endpoints were
326 evaluated as binary outcomes. Any individuals identified with a physical abnormality were
327 excluded from any behavioral analysis as these abnormalities might confound the results.

328 *Photomotor responses*

329 The larval photomotor response (LPR) assay was conducted at 120 hpf when the 96-well plates
330 of larvae were placed into a ZebraBox (Viewpoint LifeSciences) and larval movement was
331 recorded. The recorded videos were then tracked with Ethovision XT v.11 analysis software for
332 24 min across 3 cycles of 3 min light: 3 min dark. The trial time(s), x-position, y-position,
333 distance moved (μm), and velocity (mm/s) by each larva in the 2nd light/dark cycle were the
334 features used for behavioral assessment (Supp Fig 2). The 2nd light/dark cycle was chosen as it
335 exhibited less noise than the 1st cycle and was less influenced by any learning that might have
336 occurred in the 3rd cycle. For all assessments, data were collected from embryos exposed to
337 nominal concentrations of chemical and uploaded under a unique well-plate identifier into a
338 custom LIMS (Zebrafish Acquisition and Analysis Program [ZAAP]) – a MySQL database and
339 analyzed using custom R scripts that were executed in the LIMS background (Truong et al.,
340 2016).

341 **Data preprocessing and statistical analysis pipeline**

342 *Preprocessing*

343 All data processing, statistical analysis and ML were implemented in Python using the open
344 source libraries Tensorflow (Martín Abadi et al., 2015), Keras (U.S. Environmental Protection
345 Agency, 2021), Scikit-learn (Pedregosa et al., 2011), Pandas (McKinney, 2010), and Numpy
346 (Harris et al., 2020) within a purpose build Singularity container environment (Sylabs.io, 2019).
347 The x-position and y-position data was standardized relative to the center of each well and

348 forward filled if datapoints were missing. Outliers were normalized to the maximum likely
349 distance a zebrafish larva could move in $1/25^{\text{th}}$ of a second. Considering that the average length
350 of a 5 dpf larval zebrafish is 3.9 mm and can move about 2.5 times its body length during a
351 startle response (120 frames at 1000 frames/second) the threshold for distance moved in our
352 system was set at 3.25 mm per frame (Burgess & Granato, 2007b; *ZFIN Zebrafish*
353 *Developmental Stages*, n.d.). This resulted in 5,445 of the 30,825,000 frames being normalized.

354 Statistical analysis

355 Interexperimental zebrafish larval response to light/dark cycling is highly variable (Supp Fig 2).
356 Therefore, a two sample Kolmogorov–Smirnov test (K-S test) was used to compare mean of
357 controls from individual 96-well plates to mean control movement across all plates. The K-S test
358 is a non-parametric two-sided test and no adjustments were made for normality or multiple
359 comparisons. Controls from individual plates with statistically significant ($p < 0.01$) differences
360 in movement compared to the average of all controls were grouped together as hyperactive,
361 normal, or hypoactive. Following grouping the K-S test was used to compare each chemical-
362 concentration combination with their respective same plate control ($p < 0.05$). Individuals in the
363 30th and 70th percentiles of each chemical-concentration combination were defined as abnormal.

364 **Autoencoder architecture**

365 Deep autoencoders were developed using zebrafish control data to distinguish between normal
366 and abnormal zebrafish behavior. The model was trained on a Dell R740 containing two Intel
367 Xeon processors with 18 cores per processor, 512 GB RAM, and a Tesla-V100-PCIE (31.7 GB).
368 The autoencoders consisted of an input and output layer of fixed-size based on the size of a
369 single phase (25 frames per 180s) of the second light cycle (4500 frames by 5 features). The
370 encoder network was composed of eight fully connected hidden layers using a normal kernel
371 initialization, tanh activation, a dropout value of 0.2, L1 and L2 regularization values of $1e^{-05}$,
372 and an adadelta optimizer. The size of each hidden layer was reduced by increasing multiples of
373 15 and resulted in a compressed representation (bottleneck) size of 250. The decoder network
374 was composed of six fully connected hidden layers using tanh activation, and a dropout value of
375 0.2. All hidden layers used an adadelta optimizer (learning_rate=0.001, rho=0.95, and

376 epsilon=1e-07) and mean squared error for the loss function (He et al., 2015; Osl et al., 2012;
377 Ramachandran et al., 2017). For each model, we optimized the hyperparameters (i.e., the number
378 of hidden layers, the number of nodes in the layers, loss functions, optimizers, regularization
379 rates, and dropout rates) by grid search technique trained on all control data over 500 epochs
380 using Cohens Kappa statistic as the objective metric. The final encoder models were trained over
381 the course of 125000 epochs. The resulting compressed representation was used as input into a
382 logistic regression layer trained using a 100 fold cross-validation with each fold consisting of
383 4000 epochs using a limited-memory BFGS solver. The code and sample training data that
384 implements the models are available at GitHub [[https://github.com/Tanguay-
385 Lab/Manuscripts/tree/main/Green_et_al_\(2023\)_Manuscript](https://github.com/Tanguay-Lab/Manuscripts/tree/main/Green_et_al_(2023)_Manuscript)]. A complete dataset is available
386 upon request.

387

388 **Network performance and evaluation**

389 The data showed strong normal vs abnormal class imbalance (Fig 1). Classifiers may be biased
390 towards the major class (normal) and therefore, show poor performance accuracy for the minor
391 class (abnormal) (Lemaître et al., 2017). Normal vs abnormal classification accuracy was
392 evaluated using a confusion matrix, Cohen's Kappa statistic, and area under the receiver
393 operating characteristic (AUROC) as Kappa and AUROC measure model accuracy, while
394 compensating for simple chance (Ben-David, 2008). The primary metrics we used from the
395 confusion matrix included sensitivity (SE), specificity (SP), and positive predictive value (PPV)
396 as these parameters give us the true positive rate, true negative rate, and the proportion of true
397 positives amongst all positive calls (Parikh et al., 2008; Pearson, 1904; Townsend, 1971).
398 Chemical-concentration combinations were defined as abnormal if the autoencoders identified
399 more individual as abnormal in the exposed than their respective controls and at least 25% of the
400 individuals were abnormal. Permutation feature importance was used to evaluate which features
401 are the most important for model performance. In brief, one feature (variable) is shuffled
402 randomly and all features are fed into the model the resulting Kappa and AUROC values are
403 calculated. This is repeated 1000 times per feature and average Kappa and AUROC are
404 calculated across each shuffle (Breiman, 2001). To determine why one feature might be more

405 important than another a coefficient of variation was calculated for each of the features in the
406 control and exposed groups (variation() in the Scipy package).

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412

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