

Incidence of Class 1 Integrons in a Quaternary Ammonium Compound-Polluted Environment

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Samples of effluent and soil were collected from a reed bed system used to remediate liquid waste from a wool finishing mill with a high use of quaternary ammonium compounds (QACs) and were compared with samples of agricultural soils. Resistance quotients of aerobic gram-negative and gram-positive bacteria to ditallowdimethylammonium chloride (DTDMAC) and cetyltrimethylammonium bromide (CTAB) were established by plating onto nutrient agar containing 5 µg/ml or 50 µg/ml DTDMAC or CTAB. Approximately 500 isolates were obtained and screened for the presence of the *intI1* (class 1 integrase), *qacE* (multidrug efflux), and *qacEΔ1* (attenuated *qacE*) genes. QAC resistance was higher in isolates from reed bed samples, and class 1 integron incidence was significantly higher for populations that were preexposed to QACs. This is the first study to demonstrate that QAC selection in the natural environment has the potential to coselect for antibiotic resistance, as class 1 integrons are well-established vectors for cassette genes encoding antibiotic resistance.

Integrons are recombination and expression systems that capture genes as part of a genetic element known as a gene cassette (33). Most cassettes with known functions confer antibiotic or quaternary ammonium compound (QAC) resistance.

QAC resistance genes fall into two families. The *qacA/B* genes belong to the major facilitator superfamily and are only found in staphylococci on multiresistance plasmids (30). Other QAC resistance genes belong to the small multidrug resistance family and include *qacC/D*, now known as *smr*, *qacE*, *qacEΔ1*, *qacF*, *qacG*, *qacH*, and *qacJ* (3, 7, 15–17, 21, 23, 31, 32). *qacE*, *qacEΔ1*, *qacF*, and *qacG* have been identified on integrons, and the remaining genes have been identified on multiresistance plasmids in staphylococci (*qacG* is unusual in that it has been reported to be carried by staphylococcal multiresistance plasmids and class 1 integrons in gram-negative bacteria).

Class 1 integrons consist of a 5' conserved region consisting of an integrase gene encoding a site-specific recombinase (42), an *attI* site (28) where cassettes are integrated, and a promoter, P_{ant} , that regulates the expression of gene cassettes (10). Gene cassettes contain a protein coding region and a recombination site known as a 59-be site which is responsible for the orientation of integration (9). The variable regions of class 1 integrons contain the cassette genes, to the right of which lies the 3' conserved region, which may have one of three different backbone structures (27). The first backbone type consists of a Tn402 (In16)-like arrangement consisting of a *tni* module containing three transposition genes and a resolvase gene. The second, In5 type consists of *qacEΔ1*, *sulI*, orf5, orf6, and a partial *tni* module, *tniΔ*, consisting of two transposition genes. The third, In4 type carries just *qacEΔ1*, *sulI*, orf5, and orf6.

Integrons carrying the complete *tni* module are able to undergo self-transposition, and it is thought that the In5 and In4 types may also be able to move if the *tni* gene products are supplied in *trans* (27).

The role of class 1 integrons in conferring antibiotic resistance to clinical isolates of many bacterial strains is well documented (5, 14, 22, 38, 44). Studies of the incidence of class 1 integrons in bacterial pathogens associated with agriculture and fish farming, such as *Escherichia coli* and *Aeromonas salmonicida*, have also shown a link between integrons and antibiotic resistance (2, 41). Studies of the incidence of integrons in environmental bacteria have been undertaken by Rosser and Young (34), who studied the incidence of class 1 integrons in isolates from the Tay estuary, where 3.6% of isolates carried class 1 integrons. Nield and coworkers (25) identified three new integron classes in total community DNAs extracted from Australian soils by using primers for a conserved region of the integrase gene and the 59-be site. A similar approach using degenerate primers targeting the conserved regions of 59-be sites identified 164 gene cassettes, with the majority showing no relationship to known sequences.

Antibiotic and QAC resistance genes, e.g., *qacE*, are both carried on class 1 integrons, so selection for QAC resistance may cause coselection for antibiotic resistance. This provides a potential reservoir of antibiotic-resistant bacteria in QAC-polluted environments. In staphylococci, the *qacA/B* genes carried on multiresistance plasmids confer a low-level resistance to chlorhexidine and QACs, and it has been suggested that the introduction of chlorhexidine into clinical environments has resulted in the selection of staphylococci containing *qacA* carried on multiresistance plasmids (35).

For this study, resistance to ditallowdimethylammonium chloride (DTDMAC) and cetyltrimethylammonium bromide (CTAB) was assessed in environmental bacteria. Bacteria were isolated from a reed bed system (Fig. 1) used to remediate effluent from a textile mill with high QAC usage (wool finishing

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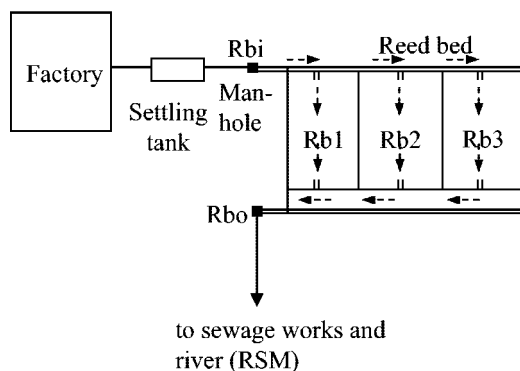


FIG. 1. Diagram of mill and reed bed system. Rbi, reed bed inlet; RB1 to RB3, reed beds 1 to 3; Rbo, reed bed outlet.

effluent only) and also from control sites with low QAC exposure. The incidence of class 1 integrons was also investigated to test the hypothesis that QAC exposure coselects for antibiotic resistance by selecting for class 1 integrons.

MATERIALS AND METHODS

Sample collection. Samples were taken from a textile mill in Yorkshire, including effluent leaving the mill, core samples from reed beds used for remediation, reed bed outflow, and river sediment downstream of the sewage farm outfall to which the effluent was discharged (Fig. 1). The workload at the mill was highly variable, as was the type of cloth treated; consequently, the nature of the effluent was highly variable. A detailed breakdown of the compounds used is not available, although a variety of QACs were used as scouring agents, antistatic agents (used at 0.1 to 0.2% of the fabric weight), dyeing retardants, softeners, finishers, moth-proofers, etc. Control soil samples were taken from agricultural land in Warwick, the Cotswolds, and Droitwich.

Isolation. Viable counts were carried out in triplicate on nutrient agar (NA) containing 5 $\mu\text{g/ml}$ or 50 $\mu\text{g/ml}$ CTAB or DTDMAC or with no selection; all plates contained 50 $\mu\text{g/ml}$ cycloheximide to reduce fungal growth. Plates were incubated at 26°C. Forty to 50 colonies encompassing the range of colony morphologies observed were picked for each sample and streaked onto NA plates.

PCR. Isolates were grown overnight in nutrient broth at 26°C, and phenol-chloroform extraction was performed. DNA concentrations were determined by measuring the absorbance of the sample at 260 nm on a spectrophotometer and were standardized to 100 ng/ μl . *intA*- and *intB*-specific primers (34) were used to amplify the class 1 integrase gene, and specific primers were used to amplify the *qacE* and *qacE Δ 1* genes (19). Integron-positive isolates and approximately 20 additional isolates from both contaminated and control sites were identified by PCR amplification and sequencing of 16S rRNAs by the use of pA-pH primers (11). Standard PCR conditions were used, as described by Rosser and Young (34) and Kazama et al. (19). Confirmation of the presence of *qacE* on a class 1 integron was achieved by using CASS1/CASS2 primers (34), which amplify the variable region of integrons priming the *intI1* and *qacE* or *qacE Δ 1* genes. PCR products were detected by electrophoresis in a 1.5% agarose gel, sequenced by the use of Big Dye Terminator, version 3.1, chemistry (Applied Biosystems), and run on a 3100 genetic analyzer. To aid in the identification of some isolates, we used an API 20NE test kit (bioMérieux) according to the manufacturer's instructions, incubating the bacteria at 30°C. The identification of strains was carried out according to the instructions in the API 20NE identification manual.

Statistical analysis. Plate counts were expressed as resistance quotients (RQs), for which the counts with selection were expressed as percentages of the counts with no selection. Differences in class 1 integrase gene frequencies between the polluted and control isolates were tested for significance by use of a chi-square test for comparisons of two proportions (from independent samples).

RESULTS

Plate counts. Viable plate counts and RQs for all samples inoculated onto QAC-selective plates at 5 μg and 50 $\mu\text{g/ml}$ are shown in Fig. 2 and 3, respectively. Most culturable bacteria appeared resistant to 5 $\mu\text{g/ml}$ DTDMAC, with RQs varying

from 62% for Cotswold soil sample 1 (CW1) to 147% for reed bed sample 1 (RB1). RQs were marginally higher for the mill samples overall, with three reed bed samples yielding RQs of >100%. For selection with 50 $\mu\text{g/ml}$ DTDMAC, bacteria from the control samples were much more sensitive than those selected with 5 $\mu\text{g/ml}$ DTDMAC, with RQs between 2.5 and 11%, and the mill samples had RQs of 49 to 117%, which were slightly lower than those at 5 $\mu\text{g/ml}$ DTDMAC. The RQ for the river sediment sample (RSM) taken downstream of the sewage outfall into which the mill effluent discharges was intermediate between those of contaminated and control samples. Isolates from the river sample showed less resistance than mill samples but more resistance than isolates from agricultural soils at 50 $\mu\text{g/ml}$ DTDMAC.

Resistance to CTAB at 5 $\mu\text{g/ml}$ was low, with high RQs limited to the reed bed samples, ranging from 67% to 99%. Control samples all had RQs of <3%. At the higher concentration for CTAB selection, virtually all culturable bacteria were sensitive, with the exception of reed bed samples, which had RQs of 1.3 to 23%.

PCR screening. The results of screening for *intI1*, the integrase gene carried on class 1 integrons, are shown in Table 1. For isolates from contaminated samples (including river sediment), 7.98% were integron positive, in contrast to 0% of control samples, which was a statistically significant difference, with a chi-square value of 19.604 and a *P* value of <0.0001 when a comparison of proportions test was applied. When restricted to the reed bed samples, the incidence increased to 14.9%, which was also significantly different from that of the control samples.

The 19 class 1 integron-positive isolates were screened by the use of specific primers for the *qacE* and *qacE Δ 1* genes, with 18 isolates testing positive for *qacE* and 0 testing positive for *qacE Δ 1* (Table 1). Putative *qacE* PCR products were sequenced and illustrated 100% similarity to known *qacE* genes. Amplification of the integron variable region by use of the CASS1/CASS2 primers produced an 1,800-bp product, indicating that 18/19 integrons carried the *qacE* gene. 16S rRNA PCRs and subsequent sequencing of a 700-bp fragment revealed that the 18 isolates carrying *qacE* showed the highest similarities to *Pseudomonas* sp. (3 isolates), *Serratia* sp. (2 isolates), and *Aeromonas hydrophila* (13 isolates) and that the isolate lacking *qacE* showed the highest similarity to *Enterobacteriaceae*. A full list of the species identified is given in Table 2. In many cases, the 16S rRNA sequences showed high similarities to those of *Aeromonas hydrophila*, *Aeromonas salmonicida*, and *Vibrio parahaemolyticus*. The taxonomy of the genus *Aeromonas* is not straightforward, with the so-called "*A. hydrophila* complex" also including the closely related *A. salmonicida* and *Aeromonas bestiarum* (40). *A. hydrophila* is motile, whereas *A. salmonicida* strains are generally nonmotile. The API 20NE test kit gave an identification of *A. hydrophila* with a 99.3% certainty. Microscopic observation revealed the cells to be highly motile rods, strengthening the identification of *A. hydrophila*.

DISCUSSION

The majority of isolates were resistant to DTDMAC at 5 $\mu\text{g/ml}$, suggesting that this concentration only affects the growth of a few sensitive strains. Among the reed bed samples,

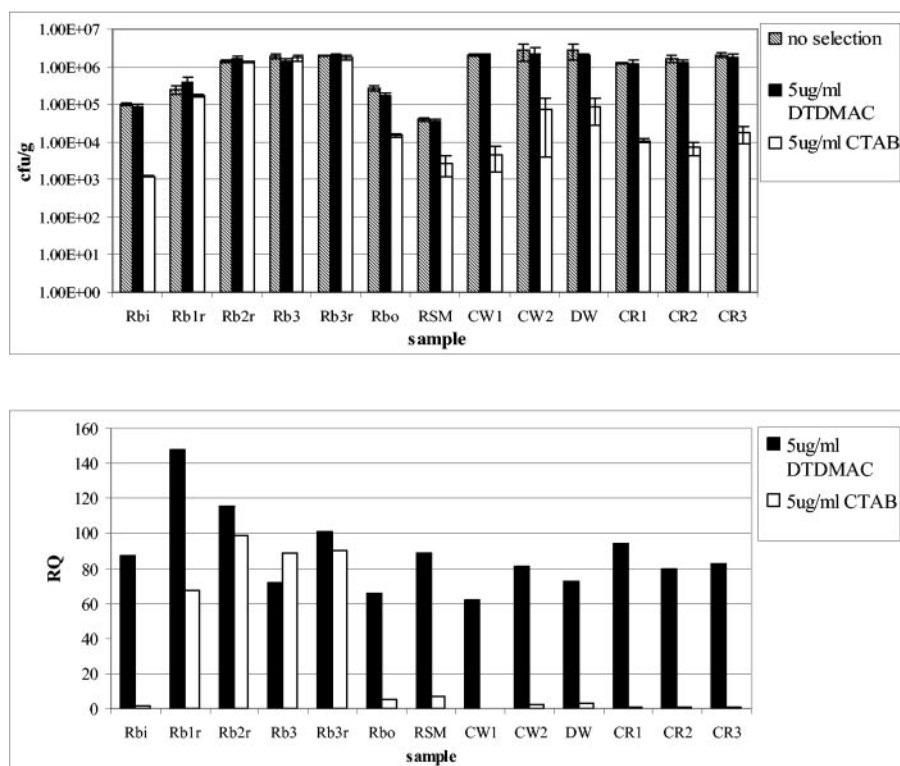


FIG. 2. Relative resistance to DTDMAC and CTAB at 5 µg/ml. (Top) Viable plate counts, with error bars giving standard deviations. (Bottom) RQs showing percentages of resistant bacteria in each sample. Rbi, reed bed inlet; RB1 to RB3, reed beds 1 to 3; r, root-associated sample material; Rbo, reed bed outlet; RSM, river sediment downstream of outfall; CW1 and CW2, Cotswold farmland soil; DW, Droitwich farmland soil; CR1 to CR3, Warwickshire farmland soil.

three of four had RQs of >100% due to the fact that DTDMAC may act as a carbon source. *Pseudomonas fluorescens* has been reported to utilize a dialkyl-QAC, didecyldimethylammonium chloride, as a sole carbon source (26). *Pseudomonads* have also been reported to utilize dodecyldimethylamine (20) and hexadecyltrimethylammonium chloride (43). In addition, *Acinetobacter* sp., *Serratia marcescens*, and *Aeromonas hydrophila* have also been reported to degrade QACs (1, 29, 36). It is probable that the *Pseudomonas* sp., *Serratia* sp., and *Aeromonas hydrophila* isolates identified in this study as carrying class 1 integrons are involved in QAC degradation. With 50 µg/ml DTDMAC, there was a much larger distinction between the RQs observed for the polluted and control sites, illustrating that culturable bacteria were much more highly resistant in populations that were preexposed to QACs. This differential in resistance was likely the result of QAC selective pressure within the factory effluent and reed bed system and correlates with a significantly higher class 1 integron frequency in QAC-polluted than control samples. RQs of >100% were still apparent for two reed bed samples, suggesting that some strains are still able to metabolize DTDMAC at 50 µg/ml.

Resistance to CTAB at both concentrations showed a clear linkage with the reed bed habitat. Isolates from the mill effluent flowing into the reed bed and from the effluent discharged from the reed bed illustrated much less resistance than other isolates, for the former because there was little time for selection to occur upstream of the reed beds and for the latter because resistant bacteria may be associated with root systems

or biofilms. Bacteria are known to be less susceptible to QACs when growing in biofilms (6), and the low levels of resistant bacteria in the reed bed outlet effluent confirmed that resistant strains reside within the reed bed sediment. It is also possible that low counts in the reed bed outlet effluent can be accounted for by imperfect operation of the system, i.e., some effluent flowed over the surface of the reed bed, diluting that passing through the beds. Very little resistance was seen for control samples at 5 µg/ml CTAB, and almost none was seen at 50 µg/ml. Again, the integron frequency was generally higher for samples with populations with high RQs. Integron-positive isolates were screened for *qacE* and *qacEΔ1*, and 18/19 isolates from the reed bed samples carried *qacE*, a fully functional multidrug efflux gene carried on a class 1 integron that has been shown to confer increased resistance to a variety of QACs, including CTAB (31). The integron frequency was also highest in RB1, which had a higher effluent flow rate than RB2 and RB3. Rosser and Young (34) reported a 3.6% incidence of class 1 integrons in environmental coliforms, i.e., *Pseudomonas* spp. and *Vibrio* spp. collected from the River Tay estuary water, which was similar to the overall incidence of 3.8% seen for this study. However, it is clear that class 1 integrons were not evenly distributed among samples but were only found in reed bed samples exposed to QACs (14.9% of bacterial isolates from reed bed samples) and not in isolates from agricultural soil. The high incidence of *qacE* observed (approximately 95% of integrons) is higher than those observed in other studies of environmental bacteria, which showed incidences of 46% (34)

TABLE 2. Identification of isolates

Sample	Isolate (% similarity to 16S sequences, no. of isolates) ^a
Reed bed inlet	<i>Bacillus licheniformis</i> (99, 1) <i>Brevundimonas vesicularis</i> (99, 1) <i>Rhodococcus erythropolis</i> (99, 1)
Reed beds	<i>Acinetobacter venetianus</i> (98, 1) <i>Acinetobacter</i> sp. (99, 1) <i>Acinetobacter</i> sp. (92, 1) <i>Bacillus weihenstephanensis</i> (99, 1) <i>Enterobacter</i> sp. (98, 1) (pathogen) <i>Pseudomonas</i> sp. (99, 1) <i>Pseudomonas aurantiaca</i> (97, 1) <i>Stenotrophomonas</i> sp. (98, 2) <i>Aeromonas hydrophila</i> (99, 13) ^{*b} <i>Enterobacteriaceae</i> (97, 1)* <i>Pseudomonas fluorescens</i> (98, 1)* <i>Pseudomonas</i> sp. (97, 2)* <i>Serratia proteamaculans</i> (98, 1)* <i>Serratia</i> sp. (95, 1)*
Reed bed outlet	<i>Citrobacter</i> sp. (99, 2) <i>Rhodococcus erythropolis</i> (99, 1)
River sediment	<i>Acinetobacter</i> sp. (98, 2) <i>Bacillus</i> sp. (98, 2) <i>Exiguobacterium</i> sp. (98, 2) <i>Pseudomonas</i> sp. (98, 3)
Cryfield soil	<i>Bacillus weihenstephanensis</i> (99, 3) <i>Paenibacillus</i> sp. (96, 2) <i>Pseudomonas</i> sp. (99, 3) <i>Pseudomonas veronii</i> (99, 2)
Cotswold soil	<i>Bacillus</i> sp. (98, 2) <i>Paenibacillus</i> sp. (99, 1) <i>Pseudomonas</i> sp. (98, 1)
Droitwich soil	<i>Bacillus</i> sp. (99, 2) <i>Paenibacillus</i> sp. (97, 1) <i>Pseudomonas</i> sp. (98, 1)

^a 16S sequences were compared to those in the National Center for Biotechnology Information database. *, isolates carrying class 1 integrons.

^b 99.3% by the API NE20 test.

integrons on R plasmids (37). Pseudomonads were common at both sample sites, and as shown in this study, commonly carry class 1 integrons in environments that select for biocide resistance. It is also interesting that many of the genera identified from the contaminated site contain pathogenic species.

It is evident that there are a variety of mechanisms to explain the observed resistance of cultured isolates from the QAC-contaminated reed bed system, but this study is unique in linking increased class 1 integron frequencies with increased QAC resistance. Ninety-five percent of the class 1 integrons carried *qacE*, reinforcing the hypothesis that QAC pollution selects for integrons. Since class 1 integrons constitute a well-known mechanism for the horizontal transfer of antibiotic resistance genes, it is clear that QAC pollution significantly increases the chances for coselection of antibiotic resistance within environmental bacteria. Recent studies concentrating on investigations of household biocide use and antibacterial resistance (8) failed to establish a link due to the fact that samples were taken from areas where lethal exposures to biocides occur. A comprehensive review of the literature (12) also concluded that the risk of reduced susceptibility to antimicrobial agents associated with exposure to sublethal concentrations of biocides was small. However, these last authors speculated that although biocides at the point of use are rapidly bactericidal and do not therefore pose a risk of increasing

antibiotic resistance, downstream of their application a gradient may occur by which sublethal concentrations may provide selective pressure. This hypothesis appears to concur with the data presented in this paper showing that at certain concentrations downstream of application, such as occurs in the reed bed system described here, selection can and does occur for mobile genetic elements carrying biocide resistance genes that are also capable of carrying antibiotic resistance determinants.

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