

Exposure to agricultural pesticides and wheezing among 5–12-year-old children in the Imperial Valley, CA, USA

Yoshira Ornelas Van Horne^{1a,b,*}, Jill E. Johnston^{1b}, Dayane Duenas Barahona^b, Mitiasoa Razafy^{1b}, Elizabeth M. Kamai^{1b}, Brandyn C. Ruiz^b, Sandrah P. Eckel^{1b}, Esther Bejarano^c, Luis Olmedo^c, Shohreh F. Farzan^{1b}

Background: Exposure to pesticides has been linked to adverse respiratory health outcomes in children.

Methods: We leveraged the Children's Assessing Imperial Valley Respiratory Health and the Environment cohort located in the rural community of Imperial Valley near the US–Mexico border. We calculated the kilograms of total pesticides applied within 400 m of children's residential addresses for the years 2016–2020. Estimated pesticide usage near homes was categorized into three groups (none vs. low vs. high [split at the median]). All health variables (i.e., asthma status and wheezing) were derived from a parent-reported questionnaire on respiratory health. We used generalized linear models, controlling for child sex, the language of survey, health insurance, respondents' highest education, and exposure to environmental secondhand smoking, to calculate prevalence differences between none versus low and high exposure to agricultural pesticides.

Results: Approximately 62% of the 708 children (aged 5–12 years) lived within 400 m of at least one pesticide application within 12 months prior to survey administration. Exposure to pesticides within 400 m of children's residences was associated with 12-month prior wheeze. Those in the “high” exposure group had a prevalence of wheezing that was 10 (95% confidence interval: 2%, 17%) percentage points higher than among children not exposed to pesticide applications. Associations for high exposure to specific categories of pesticide applications, sulfur only, all pesticides except sulfur, chlorpyrifos, and glyphosate, also were observed with a higher prevalence of wheezing than among children not exposed to pesticide applications.

Conclusions: We observed associations between living near pesticide applications and more wheeze symptoms among children.

Keywords: Environmental justice; Pesticides; Wheezing; Asthma; Children's health

^aDepartment of Environmental Health Sciences, Mailman School of Public Health, Columbia University, New York City, New York; ^bDepartment of Population and Public Health Sciences, Keck School of Medicine, University of Southern California, Los Angeles, California; and ^cComite Civico Del Valle, Brawley, California

Supported by R01ES029598 and 3R01ES029598-04S1 from the National Institute of Environmental Health Sciences.

Data are confidential but available upon reasonable request. The AIRE study Principal Investigators welcome new collaborations with other investigators and have actively engaged in collaborative data-sharing projects. Interested investigators should contact J.E.J. (jill@usc.edu) and S.F.F. (sffarzan@usc.edu) to obtain additional information about the AIRE study, discuss collaborative opportunities, and request a project proposal form. Code available upon request. Interested investigators should contact Y.O.V.H (yov2000@cumc.columbia.edu).

SDC Supplemental digital content is available through direct URL citations in the HTML and PDF versions of this article (www.enviroepidem.com).

*Corresponding Author. Address: Department of Environmental Health Sciences, Mailman School of Public Health, Columbia University, 722 West 168th Street, 12th Floor, New York, NY 10032. E-mail: yov2000@cumc.columbia.edu (Y.O. Van Horne).

Copyright © 2024 The Authors. Published by Wolters Kluwer Health, Inc. on behalf of The Environmental Epidemiology. All rights reserved. This is an open-access article distributed under the terms of the Creative Commons Attribution-Non Commercial-No Derivatives License 4.0 (CCBY-NC-ND), where it is permissible to download and share the work provided it is properly cited. The work cannot be changed in any way or used commercially without permission from the journal.

Environmental Epidemiology (2024) 8:e325

Received 31 January, 2024; Accepted 12 July, 2024

Published online 19 August 2024

DOI: 10.1097/EE9.0000000000000325

Introduction

The widespread application of agricultural pesticides disproportionately impacts rural communities living in close proximity to fields and may contribute to documented health disparities.¹ The health effects associated with pesticides are wide ranging.² Exposure to pesticides has been associated with preterm delivery,³ an increased risk of autism spectrum disorder,^{4,5} low APGAR scores (a marker of infant health at birth),⁶ and respiratory effects.⁷ Disparities in pesticide exposures by race/ethnicity also persist, with higher exposures concentrated in structurally marginalized communities.^{8,9} In the United States, the agricultural sector accounts for nearly 90% of the total pesticide usage.¹⁰ In California, from 2017 to 2020, total pesticide usage in industrial agriculture was approximately 200 million pounds annually.¹¹ These pesticide applications occur in agricultural fields in close proximity to schools and homes,¹² indicating that the pesticide drift exposure pathway is of particular concern for rural communities.^{13,14}

What this study adds:

We leveraged data from the California Pesticide Use Registry to examine the association of agricultural pesticide use (i.e., sulfur only, all pesticides except sulfur, chlorpyrifos, and glyphosate) to parent-reported respiratory health symptoms in predominantly Latino school-aged children. Exposure to pesticide applications within 400 m of children's residences was associated with 12-month prior wheeze. These findings contribute to our understanding of pesticide exposures and their impact on children's health.

Pesticides can enter respiratory airways as small irritating molecules, aggravating the airways.¹⁵ Children living in agricultural communities are exposed to multiple pesticides, which may vary by season, application, and crop type.¹⁶ Previous studies have found that sulfur, a widely used fumigant, has been associated with an increase in respiratory symptoms and a decrease in pulmonary function in children.¹⁷ The fumigants 1,3-dichloropropene and metam sodium have been reported to cause respiratory distress in adults in two separate case studies.^{18,19} Results from adults in the Agricultural Health Study indicate that exposure to glyphosate is associated with atopic asthma,²⁰ while pendimethalin, trifluralin, and chlorpyrifos have been associated with wheezing in farmers.²¹ The respiratory health of children, such as wheezing, can have long-term consequences for respiratory health into adulthood.^{22–24} Established predictors of wheezing in children include age,²⁵ sex,²² early-life sensitization,²⁶ and environmental exposures such as maternal smoking during pregnancy.^{26–28} Despite prior research identifying adverse health impacts from exposure to pesticides,^{5,6,29} there is a paucity of research focusing on their effects on children's health in rural communities across US populations.

There is a limited but growing number of studies that have identified an association between various types of pesticides and wheezing in children.^{30,31} In the Infants' Environmental Health Study in Costa Rica, recent exposure to pyrethroids was associated with higher wheezing in children at age 5.³² In Limpopo, South Africa, prenatal exposure to pyrethroids was also associated with an increased risk of wheezing in children at age 3.5 and 5.³³ In contrast, in a cohort in New York City, no association was found between prenatal exposure to organophosphate (OP) and pyrethroids and wheezing at age 5.³⁴ Additionally, among children of the Outcomes and Measures of the Environment (HOME) Study in Cincinnati, Ohio, prenatal exposure to OP and pyrethroids were not associated with wheeze at 8 years or from birth to 8 years.³⁵ Overall pooling of these studies in two different reviews suggests that exposure to pesticides increases the odds of developing wheezing among children.^{30,31}

Residents of the Imperial Valley, located in the rural border county of Imperial County of southeastern California, experience a multitude of social and environmental issues that may impact their well-being.³⁶ Residents of this area are primarily Latino, 1 in 3 children live in poverty, and there is a 20% unemployment rate.³⁷ The county faces poor air quality and excess particulate matter levels (PM).³⁸ Further, one in five children is diagnosed with asthma and the rate of asthma-related pediatric emergency room visits and hospitalizations is two times the CA state average.^{37,39} Industrial agriculture in the region results in one of the highest amounts of pesticide applications in the state, with Imperial County ranking in the top 12 (of 58 counties) since 2017^{11,40–44} with approximately 5 million pounds of pesticides applied annually in this community. The purpose of this study was to assess the respiratory health impacts of pesticide usage in a rural, structurally marginalized population of school-aged children in Imperial Valley, CA.

Methods

Study setting and design

The Assessing Imperial Valley Respiratory Health and the Environment (AIRE) study was initiated as a direct response to address concerns about environmental pollution in Imperial Valley. In 2017, through a community-academic partnership with a local community organization, Comite Civico del Valle, a prospective cohort was initiated to understand the impacts of environmental exposures on children's respiratory health. Research protocols were approved by the University of Southern

California Institutional Review Board (HS-17-00204). Written informed consent was obtained from a parent or legal guardian at enrollment.

Participants

From 2017 to 2019, the cohort enrolled 731 elementary school-aged children (aged 5–12 years) from five different schools across five communities in Imperial Valley.⁴⁵ Participation of the schools and study participants was voluntary. To recruit participants, study staff met with teachers, principals, superintendents, and parents from each school to answer questions. Staff visited 1st, 2nd, and 3rd grade classrooms to introduce the study and distribute packets containing materials to be taken home, including study information, a consent form and a child health survey to be completed by a caregiver. A total of 804 children were eligible for the study, of which 731 children enrolled, for a 91% acceptance rate. Parents (caregivers or guardians) of 731 children consented to their child's participation in the AIRE Study, however, only 708 (96%) completed and returned the accompanying questionnaire. The questionnaire included demographic and lifestyle questions that assessed the children's characteristics such as age, sex, race/ethnicity, and parent/guardian information including, educational attainment, insurance coverage, and tobacco usage (Table 1). In the questionnaire, we also asked parents/guardians their residential address and their children's residential history (i.e., if the child had lived in the same residence since birth and if the child had lived in the same residence the 12 months prior).

Wheezing

Parents were asked about the child's respiratory symptoms using questions adapted from the International Study of Asthma and Allergies in Childhood (ISAAC) questionnaire.^{29,46–49} Children were categorized as asthmatic based on a positive response to the question, "has a doctor ever diagnosed your child as having asthma?" We defined wheezing as a binary outcome based on a positive response to the following during the previous 12 months: wheezing or whistling in the chest in the past 12 months.¹⁷

Geographic-based estimates of nearby pesticide applications

Using residential address information reported by the parent/guardian, we geocoded the location of the child's home at baseline. At enrollment, all children reported living at the same address for the previous 12 months. We estimated agricultural pesticide use near the child's residence using data from the California Department of Pesticide Regulation (CDPR).¹¹ Since 1990, California has required mandatory reporting of all agricultural pesticide applications, with data available through the Pesticide Use Report (PUR) Program. We used validated geospatial methods employed by other research teams and edited for outliers to assign exposures.^{50,51} PURs include active ingredient, brand name, date/time applied, amount, and acreage within 1-square mile sections (~1.6 km × 1.6 km) defined by the Public Land Survey System. For each child, we computed the total pesticides applied within a 400-m buffer distance of their home, using a reported address from the questionnaire, for the 12 months before the date of the baseline survey using the following pesticide groups: all pesticides, all pesticides except sulfur, sulfur only, chlorpyrifos only, and glyphosate only. Analysis of all pesticides was chosen because of the concerns surrounding cumulative effects, given that many of the pesticides sprayed in Imperial County are classified as respiratory irritants (i.e., chlorpyrifos, glyphosate, and sulfur) and previous studies have also assessed

general proximity to fields sprayed near fields.^{17,20,21,52,53} The 400 m buffer was selected based both on (1) previous studies identifying an association between a 500-m buffer correlation with pesticides measured in homes located near agricultural fields^{17,54–56} and (2) a 2018 law in California that limits the application of pesticides within 402 m (1/4 mile) of schools. Thus 400 m represents a buffer with both scientific and policy implications.¹² For all pesticides, we weighted use near homes based on the proportion of each square mile that was within each buffer surrounding a residence centroid. Due to the large amount of residence centroids with zero pesticide applications within 400 m for chlorpyrifos, and to maintain consistency across pesticide groups, we categorized exposure based on

three groupings: none, low, and high, with the low and high categories split at the group-specific median. The median used was 67 kg for all pesticide applications, 80 kg for sulfur, 34 kg for all minus sulfur, 2 kg for chlorpyrifos, and 7kg for glyphosate (Table 2).

Data analysis

We estimated respiratory symptoms associated with exposure to pesticides within 400 m by quantifying prevalence differences (rather than odds ratios) using generalized linear models, specifying identity links, Gaussian distribution, and sandwich-based errors.⁵⁷ Based on prior literature and directed acyclic graphs

Table 1.
AIRE cohort study characteristics at baseline, N = 708

| Characteristic | All | | Wheezing | | No wheezing | |
|---|-----------|----|-----------|----|-------------|----|
| | N | % | n | % | n | % |
| Age mean (years) | 8.5 (1.1) | | 8.3 (0.9) | | 8.5 (1.0) | |
| Child sex | | | | | | |
| Female | 381 | 54 | 62 | 48 | 293 | 55 |
| Male | 327 | 46 | 68 | 52 | 236 | 45 |
| Highest household education | | | | | | |
| Less than 12th grade | 137 | 19 | 29 | 22 | 102 | 19 |
| Completed high school | 197 | 28 | 27 | 21 | 167 | 32 |
| Some college | 236 | 33 | 55 | 42 | 174 | 33 |
| Completed 4 years of college | 60 | 8 | 14 | 11 | 31 | 6 |
| Graduate or professional | 50 | 7 | 3 | 2 | 29 | 5 |
| Missing | 28 | 4 | 2 | 2 | 26 | 5 |
| Health insurance | | | | | | |
| Private | 134 | 19 | 3 | 2 | 88 | 17 |
| Public | 494 | 70 | 103 | 79 | 391 | 74 |
| None | 43 | 6 | 20 | 15 | 17 | 3 |
| Missing | 37 | 5 | 4 | 3 | 33 | 6 |
| Secondhand smoke exposure | | | | | | |
| Yes | 47 | 7 | 5 | 4 | 14 | 3 |
| No | 566 | 80 | 112 | 86 | 454 | 86 |
| Missing | 95 | 13 | 12 | 9 | 61 | 12 |
| Maternal smoking during pregnancy | | | | | | |
| Yes | 54 | 8 | 16 | 12 | 21 | 4 |
| No | 625 | 88 | 110 | 85 | 489 | 92 |
| Do not know | 29 | 4 | 4 | 3 | 19 | 4 |
| Language of survey | | | | | | |
| English | 471 | 67 | 97 | 75 | 340 | 64 |
| Spanish | 237 | 33 | 32 | 25 | 189 | 36 |
| Lived in the same residence since birth | | | | | | |
| Yes | 176 | 25 | 34 | 26 | 136 | 26 |
| No | 501 | 71 | 94 | 72 | 381 | 72 |
| Missing | 31 | 4 | 1 | <1 | 12 | 2 |
| Child's history of asthma | | | | | | |
| Yes | 166 | 23 | 90 | 69 | 65 | 12 |
| No | 539 | 76 | 39 | 30 | 463 | 88 |
| Missing | 3 | <1 | 0 | 0 | 1 | <1 |
| Wheezing in past 12 months | | | | | | |
| Yes | 130 | 18 | | | | |
| No | 527 | 74 | | | | |
| Missing | 49 | 7 | | | | |

Table 2.
Distributions of agricultural pesticide applications (kg) within 400 m of child's residence, N = 708

| Pesticide use (kg) | Total | Mean | 25th | 50th | 75th | 90th | Max |
|--------------------|--------|----------|------|------|------|------|-------|
| All | 56,824 | 77 (133) | 0 | 20 | 109 | 209 | 1,379 |
| Sulfur | 28,060 | 38 (80) | 0 | 0 | 49 | 147 | 975 |
| All minus sulfur | 28,763 | 39 (83) | 0 | 14 | 53 | 99 | 404 |
| Chlorpyrifos | 523 | 1 (2) | 0 | 0 | 0 | 1.3 | 18 |
| Glyphosate | 3,319 | 5 (9) | 0 | 0 | 4 | 14 | 80 |

All includes the 150 active ingredients applied within 400 m. Median for analysis refers to the split of low vs. high for those in the exposed category.

Table 3.

Adjusted association for pesticide exposures (kg) within a 400-m radius of a child’s residence and wheezing symptoms at baseline in the AIRE cohort, stratified by history of doctor-diagnosed asthma

| Exposure ^c | n | Wheezing prevalence (95% CI) Min ^a | | Wheezing prevalence (95% CI) Full ^b | | Wheezing prevalence (95% CI) Min | | Wheezing prevalence (95% CI) Full | |
|-----------------------|-----|---|-----|--|-----|----------------------------------|-----|-----------------------------------|--|
| | | Asthmatic | n | Asthmatic | n | Nonasthmatic | n | Nonasthmatic | |
| All | | | | | | | | | |
| None | 40 | | 36 | | 154 | | 143 | | |
| Low (<67 kg) | 59 | 0.12 (−0.09, 0.32) | 56 | 0.13 (−0.09, 0.35) | 176 | 0.03 (−0.03, 0.09) | 166 | 0.03 (−0.03, 0.09) | |
| High (≥67 kg) | 56 | 0.18 (−0.02, 0.38) | 51 | 0.17 (−0.04, 0.39) | 173 | 0.04 (−0.01, 0.10) | 162 | 0.05 (−0.01, 0.11) | |
| Sulfur | | | | | | | | | |
| None | 78 | | 70 | | 319 | | 299 | | |
| Low (<80 kg) | 37 | −0.05 (−0.24, 0.14) | 36 | −0.03 (−0.23, 0.17) | 96 | −0.003 (−0.07, 0.06) | 89 | −0.001 (−0.07, 0.07) | |
| High (≥80 kg) | 40 | 0.18 (−0.01, 0.37) | 37 | 0.14 (−0.06, 0.35) | 88 | 0.05 (−0.01, 0.11) | 83 | 0.05 (−0.01, 0.12) | |
| All minus sulfur | | | | | | | | | |
| None | 41 | | 37 | | 156 | | 145 | | |
| Low (<34 kg) | 49 | 0.09 (−0.10, 0.28) | 47 | 0.09 (−0.12, 0.30) | 183 | 0.02 (−0.04, 0.07) | 174 | 0.02 (−0.04, 0.08) | |
| High (≥34 kg) | 65 | 0.19 (−0.02, 0.40) | 59 | 0.19 (−0.04, 0.42) | 164 | 0.06 (0.002, 0.12) | 152 | 0.06 (0.002, 0.12) | |
| Chlorpyrifos | | | | | | | | | |
| None | 123 | | 112 | | 407 | | 380 | | |
| Low (<2 kg) | 16 | 0.15 (−0.10, 0.41) | 16 | 0.17 (−0.10, 0.43) | 46 | 0.03 (−0.05, 0.11) | 43 | 0.03 (−0.05, 0.11) | |
| High (≥2 kg) | 16 | 0.29 (0.03, 0.55) | 15 | 0.24 (−0.02, 0.51) | 50 | 0.09 (0.007, 0.17) | 48 | 0.10 (0.008, 0.18) | |
| Glyphosate | | | | | | | | | |
| None | 79 | | 69 | | 277 | | 260 | | |
| Low (<7 kg) | 37 | −0.06 (−0.25, 0.13) | 35 | −0.04 (−0.24, 0.16) | 111 | 0.03 (−0.10, 0.41) | 103 | 0.04 (−0.04, 0.09) | |
| High (≥7 kg) | 39 | 0.20 (0.01, 0.39) | 39 | 0.19 (−0.008, 0.38) | 111 | 0.02 (−0.03, 0.08) | 105 | 0.03 (−0.03, 0.10) | |

^aMinimally adjusted for sex and language of survey.

^bFully adjusted for sex, language of survey, health insurance, parent/guardian education, and exposure to environmental smoking.

^cLow and high categories split at the median for those with pesticide exposures greater than 0 within a 400-m radius of the residence. None indicates no pesticide applications were applied within a 400-m radius of the child’s residence. Generalized linear models, with Gaussian distribution, identity link function, and sandwiched errors.

(Figure S1; <http://links.lww.com/EE/A296>), we chose to minimally adjust models for child sex and language of survey (Spanish or English). We chose a priori to additionally evaluate models adjusted for health insurance (none, public, and private), respondents’ highest education, and exposure to environmental secondhand smoke. Child exposure to environmental secondhand smoke at home was defined as living with a current smoker or the presence of any regular visitors who smoke inside the home, regardless of whether the child is also present or maternal smoking during pregnancy. In sensitivity analyses, we evaluated models by stratifying by asthma status (Table 3) and excluding participants with exposure to current environmental secondhand smoking from fully adjusted models (Table S1; <http://links.lww.com/EE/A296>), excluding participants whose biological mom smoked during pregnancy (Tables S2 and S3; <http://links.lww.com/EE/A296>) and further adjusting (Table S4; <http://links.lww.com/EE/A296>) and stratifying by time of residence in the same household (Tables S5 and S6; <http://links.lww.com/EE/A296>).

Results

The mean age of participants at enrollment in the AIRE cohort was 8.5, with a range of 7–9 years of age. A majority of children were female (n = 381, 54%), had a parent-completed survey in English (n = 471, 67%), and had public health insurance (n = 494, 69%) (Table 1). Less than 8% reported current exposure to environmental secondhand smoke. Most parents reported having attended some college (n = 236, 33%) or completing high school (n = 197, 28%). A total of 166 (23%) parents reported their child had a history of asthma at enrollment. There were 130 children (18%) with wheezing in the 12 months prior to the survey.

Pesticide applications

Within 400 m of child residences, there were 150 different pesticides applied over a 12-month period. This amounted to a

total of 56,824 kg of pesticides applied to agricultural fields in our Imperial Valley study area. Applications of sulfur only were the main contributor with a total of 28,060 kg, followed by all pesticides except sulfur (28,763 kg), glyphosate (3319 kg), and chlorpyrifos (523 kg) (Table 2).

Wheezing

Of the 708 participants, 658 had complete information on sex, the language of survey, asthma status, and a complete geocoded address for inclusion in minimally adjusted models. In minimally adjusted models, children in the “high” total pesticide exposure group had a prevalence of 12-month prior wheeze that was 10 percentage points (95% confidence interval [CI]: 2%, 17%) higher than that of children not exposed to any pesticides (Table 4). Similarly, the difference in the prevalence of 12-month prior wheeze in the “high” exposure group to sulfur compared to the “no” exposure group was 12 percentage points higher (95% CI: 5%, 20%). Estimates were similar for the “high” exposure group to all other pesticides excluding sulfur to the unexposed group, with a difference of 9.0 percentage points (95% CI: 1%, 16%). For chlorpyrifos, children in the “high” exposure group had a 15.0 percentage point (95% CI: 5%, 26%) higher prevalence than the unexposed group. Finally, those in the “high” exposure group for glyphosate had a prevalence difference of 9.0 percentage points (95% CI: 1%, 16%) higher than the unexposed group. There was limited evidence of an association in the prevalence of 12-month prior wheeze in the “low” exposure groups to the unexposed groups for any of the pesticide groups, although estimates were suggestive of a positive trend in wheezing prevalence with any pesticide exposure. In fully adjusted models, a total of 614 had full information for health insurance, parent or guardian education, and exposure to secondhand smoking; adjustment for these additional covariates did not materially change estimates (Table 4).

Table 4. Adjusted association for pesticide exposures (kg) within a 400-m radius of a child’s residence and wheezing symptoms at baseline in the AIRE cohort

| Exposure | n | Wheezing prevalence (95% CI) | n | Wheezing prevalence (95% CI) | n | Wheezing prevalence (95% CI) |
|------------------|-----|------------------------------|-----|------------------------------|-----|------------------------------|
| | | Crude | | Min adjusted ^a | | Fully adjusted ^b |
| All | | | | | | |
| None | 194 | Ref. | 194 | Ref. | 179 | Ref. |
| Low (<67 kg) | 235 | 0.07(−0.01, 0.14) | 235 | 0.07 (−0.007, 0.14) | 222 | 0.07 (−0.008, 0.15) |
| High (≥67 kg) | 229 | 0.09 (0.02, 0.17) | 229 | 0.10 (0.02, 0.17) | 213 | 0.10 (0.03, 0.18) |
| Sulfur | | | | | | |
| None | 397 | Ref. | 397 | Ref. | 371 | Ref. |
| Low (<80 kg) | 133 | 0.04 (−0.04, 0.12) | 133 | 0.03 (−0.05, 0.11) | 123 | 0.04 (−0.04, 0.12) |
| High (≥80 kg) | 128 | 0.12 (0.04, 0.20) | 128 | 0.12 (0.05, 0.20) | 120 | 0.13 (0.05, 0.21) |
| All minus Sulfur | | | | | | |
| None | 197 | Ref. | 197 | Ref. | 182 | Ref. |
| Low(<34 kg) | 232 | 0.07(−0.005, 0.15) | 232 | 0.07 (−0.004, 0.15) | 221 | 0.07 (−0.006, 0.15) |
| High (≥34 kg) | 229 | 0.08 (0.007, 0.16) | 229 | 0.09 (0.01, 0.16) | 211 | 0.09 (0.02, 0.17) |
| Chlorpyrifos | | | | | | |
| None | 530 | Ref. | 530 | Ref. | 492 | Ref. |
| Low (<2 kg) | 62 | 0.07(−0.03, 0.17) | 62 | 0.06 (−0.04, 0.16) | 59 | 0.07 (−0.04, 0.17) |
| High (≥2 kg) | 66 | 0.15 (0.04, 0.25) | 66 | 0.15 (0.05, 0.26) | 63 | 0.16 (0.06, 0.27) |
| Glyphosate | | | | | | |
| None | 360 | Ref. | 360 | Ref. | 332 | Ref. |
| Low (<7 kg) | 148 | 0.02 (−0.05, 0.10) | 148 | 0.02 (−0.05, 0.10) | 138 | 0.04 (−0.04, 0.12) |
| High (≥7 kg) | 150 | 0.09 (0.01, 0.17) | 150 | 0.09 (0.01, 0.16) | 144 | 0.10 (0.02, 0.18) |

Low and high categories split at the median (kg) for those with pesticide exposures greater than 0 within a 400-m radius of the residence. None indicates no pesticide applications were applied within a 400-m radius of the child’s residence.

^aMinimally adjusted for sex and language of survey.

^bFully adjusted for sex, language of survey, health insurance, parent/guardian education, and exposure to environmental smoking.

Ref indicates reference group.

Associations between exposure to pesticides and respiratory symptoms appear to be higher in children with asthma (Table 3). Among children with a history of asthma, those in the “high” exposure group had a prevalence of wheezing that was 17% (95% CI: −4%, 39%), 14% (95% CI: −6%, 35%), 19% (95% CI: −4%, 42%), 24% (95% CI: −2%, 51%), and 19% (95% CI: −0.8%, 38%) percentage points higher than that among children not exposed to all pesticides, sulfur only, all pesticides except sulfur, chlorpyrifos, or glyphosate, respectively (Table 3). While a stronger effect is observed among asthmatics there are significant positive associations for nonasthmatics in the high category of chlorpyrifos and all minus sulfur. Among nonasthmatic children, those in the “high” exposure group had a prevalence of wheeze that was 6 (95% CI: 0.2%, 12%) and 10 (95% CI: 0.8%, 18%) percentage points higher than that among unexposed children to all pesticides except sulfur and chlorpyrifos, respectively (Table 3).

In sensitivity analyses that excluded children exposed to environmental secondhand smoking, estimates of association were similar (Table S1; <http://links.lww.com/EE/A296>). Excluding participants whose biological mom smoked during pregnancy did not materially change the estimates (Tables S2 and S3; <http://links.lww.com/EE/A296>). We also investigated adjustment and stratifying by time of residence in the home since birth (lifetime never moved vs. moved), with estimates attenuating for all participants (Tables S4–S6; <http://links.lww.com/EE/A296>).

Discussion

We observed consistent cross-sectional associations between exposure to pesticides applied within 400 m of children’s residences within the past 12 months and reported wheeze during this time. Those in the highest exposure group experienced a higher prevalence of wheeze compared to unexposed children for all pesticides groups we examined: all pesticides, sulfur only, all pesticides except sulfur, chlorpyrifos, and glyphosate,

respectively. There was limited evidence of an association in the prevalence of 12-month prior wheeze between the “Low” exposure group and the unexposed group for any of the pesticide groups; however, we saw a suggestive positive trend across categories. Associations between exposure to pesticides and respiratory symptoms appear to be higher in children with asthma, but positive, albeit weaker trends, also were observed among nonasthmatic children.

Our findings are consistent with previous studies reporting an association between general exposure to pesticides and increased wheezing in children.^{58–61} While our sample size is small, we did find a consistent association between recent exposures (12 months) to chlorpyrifos or glyphosate and wheezing in children. Previous studies in urban environments have not reported an association between prenatal exposure to OPs and wheezing in children at ages 5³⁴ or 8.³⁵ Exposure to glyphosate and wheezing among children has not been previously reported, however, exposure to glyphosate has been linked to wheezing among farmers.^{62,63} Furthermore, a previous cohort study in rural Salinas Valley, CA, USA, the CHAMACOS (the Center for the Health Assessment of Mothers and Children of Salinas) study, utilized the CDPR PUR database to assess pesticide exposures and respiratory health.^{17,64} Similar to our findings, a 10-fold increase in exposure to sulfur within 1000 m of a child’s residence was associated with a decrease in lung function.¹⁷ However, in the same CHAMACOS cohort, no associations were found between exposure to chloropicrin, metam sodium, 1,3-dichloropropene, or methyl bromide within 8000 m and lung function.⁶⁴ In contrast, a different study in Central and Southern California found that ambient concentrations of methyl bromide and 1,3-dichloropropene were each associated with asthma emergency department visits.^{65,66} In our present AIRE cohort, we were limited in finding associations with the same pesticides (i.e., 1,3-dichloropropene, pendimethalin, bensulide, or trifluralin [data not shown]) as their use in Imperial Valley was less widespread.

Mechanism of action and class of pesticide (e.g., herbicide, fumigant, and OPs) are important factors to consider in relation to public health concerns. While the mechanisms of action may differ, the fumigant sulfur, the OP chlorpyrifos, and the herbicide glyphosate are all considered respiratory irritants.^{67–70} Experimental models also suggest that low-level organophosphorus pesticides (OP's) exposure can activate many of the inflammatory mediators involved in asthma hyperactivity.⁷¹ The mechanism of action will vary by class and type of pesticide, for example, OPs (e.g., chlorpyrifos) contribute to airway hyper-reactivity via neurogenic inflammation^{72–75} and oxidate stress.⁷⁶ Lung inflammation, characterized by type 2 (TH2) immune responses, is associated with allergy and plays a key role in asthma development.⁷⁷ However, unlike OPs, sulfur has not been found to induce an inflammatory response in the lungs.⁷⁸

For children residing in rural communities, such as the Imperial Valley, it is important to not only consider mechanisms of action but also cumulative exposures. In 2018, California started limiting the application of pesticides within 402 m (1/4 mile) of schools.¹² However, in the AIRE study, we find that within 400 m of child residences, there were 150 different pesticides applied over a 12-month period. Our findings suggest that reducing pesticide applications near residential homes could protect children's respiratory health. While both the CHAMACOS and AIRE cohorts enrolled primarily children from under-resourced Latino families, the majority of the children in the AIRE cohort are not from a household where a parent or caretaker is a farmworker. This suggests that pesticide drift and inhalation may be of greater concern than the take-home exposure pathway of the CHAMACOS cohort.^{1,13,14,79} Interventions aimed at reducing exposure through the take-home pathway include having the parents/caregivers change into different clothes, recommending wet cloth mopping, and extra mats outside the home; however, these recommendations are unlikely to reduce pesticide exposure through drift.^{1,13,14}

Pesticide exposures have long been identified by researchers and community voices as an important environmental justice issue impacting not only agricultural workers but also nearby residents.^{2,80–82} In the United States alone, the agricultural sector accounts for nearly 90% of the total pesticide usage, making agricultural farmworkers, their families, and nearby residents particularly vulnerable to the effects of pesticides.⁴⁴ While efforts have been made to advance labor and occupational safety laws, these have continuously excluded agricultural workers, who have fewer protections than most other occupations in the United States.⁸³ Over the years, significant gains in monitoring, reporting, and policy changes such as limiting agricultural pesticide use within 0.4 km of schools have been achieved.⁸⁴ However, disparities persist, pesticide use is disproportionately located in communities experiencing the highest levels of poverty, and communities of color.⁸⁵ Rural communities such as Imperial Valley continue to rank among the top communities most burdened by toxic pollution, including pesticides according to CalEnviroScreen 4.0 (Table S7; <http://links.lww.com/EE/A296>).⁸⁵

Our study has several strengths and limitations. We leveraged data from a relatively large, ongoing cohort study of children's respiratory health, with detailed health and covariate information, and available agricultural pesticide application data to address a key gap in the literature. We utilized the validated ISAAC questionnaire to assess respiratory symptoms and had a geographical location of the child's residence, which allowed us to individually link each child to administrative records of nearby pesticide applications through the CDP. Because we did not use direct biological or environmental monitoring, we were unable to estimate body burden, which may have resulted in some exposure misclassification during the 12-month period exposure period. Additionally, while a previous study has estimated that a large amount of the variability in outdoor air for the fumigant methyl bromide can be explained by the use of

proximity to agricultural applications using the PUR data, the fumigant sulfur assessed in this study has not been evaluated in a similar manner.⁸⁶ Furthermore, the sample size for the children in “high” exposure group for chlorpyrifos and glyphosate is small ($n = 63$ and $n = 144$, respectively). We also cannot rule out unmeasured confounding by other sources of pesticide exposure, such as pesticide use in the home, as we do not have reliable data on pesticide use at home for these participants.

Conclusion

We found that children highly exposed to pesticide applications (including sulfur, chlorpyrifos, and glyphosate) within 400 m of their homes had an increased prevalence of 12-month prior wheeze compared to unexposed children in a rural, predominantly Latino community in southeast California. Children living in agricultural communities are often exposed to multiple pesticides¹⁶ and Imperial Valley remains one of the highest pesticide use areas in CA.^{40–44} In this intensively farmed community, fields are planted and harvested several times a year. There are over 100 different types of crops cultivated and over 50 pesticides (differentiated by active ingredient) used in Imperial Valley amounting to over five million pounds of pesticides applied annually. As schools and homes are located in close proximity to agricultural fields, exposure to pesticides is of great concern. The pesticide drift exposure pathway is of particular concern for rural children, as these often come on top of social and economic challenges faced in rural communities. In addition to future studies evaluating the relationships between other classes of pesticides and objective measures of respiratory health, policies aimed at reducing pesticide applications near residential zones should be prioritized.

Conflicts of interest statement

The authors declare that they have no conflicts of interest with regard to the content of this report.

Acknowledgments

We are grateful to all of the children and families in the AIRE cohort for their enthusiasm and participation in this study. This study could not have been completed without the support and guidance of our community partners at Comite Civico del Valle.

References

1. Van Horne YO, Farzan SF, Razafy M, Johnston JE. Respiratory and allergic health effects in children living near agriculture: a review. *Sci Total Environ*. 2022;832:155009.
2. Donley N, Bullard RD, Economos J, et al. Pesticides and environmental injustice in the USA: root causes, current regulatory reinforcement and a path forward. *BMC Public Health*. 2022;22:1–23.
3. Ling C, Liew Z, Von Ehrenstein OS, et al. Prenatal exposure to ambient pesticides and preterm birth and term low birthweight in agricultural regions of California. *Toxics*. 2018;6:41.
4. von Ehrenstein OS, Ling C, Cui X, et al. Prenatal and infant exposure to ambient pesticides and autism spectrum disorder in children: population based case-control study. *bmj*. 2019;364:1962.
5. Shelton JF, Geraghty EM, Tancredi DJ, et al. Neurodevelopmental disorders and prenatal residential proximity to agricultural pesticides: the CHARGE study. *Environ Health Perspect*. 2014;122:1103–1109.
6. Parra KL, Harris RB, Farland LV, Beamer P, Furlong M. Associations of prenatal agricultural farm work with fetal overgrowth and pregnancy complications in State of Arizona birth records. *J Occup Environ Med*. 2023;65:635–642.
7. Buralli RJ, Dultra AF, Ribeiro H. Respiratory and allergic effects in children exposed to pesticides—a systematic review. *Int J Environ Res Public Health*. 2020;17:2740.
8. Cushing L, Faust J, August LM, Cendak R, Wieland W, Alexeff G. Racial/ethnic disparities in cumulative environmental health impacts in

- California: evidence from a statewide environmental justice screening tool (CalEnviroScreen 1.1). *Am J Public Health*. 2015;105:2341–2348.
9. Temkin AM, Uche UI, Evans S, et al. Racial and social disparities in Ventura County, California related to agricultural pesticide applications and toxicity. *Sci Total Environ*. 2022;853:158399.
 10. EPA PIS. *Usage, 2008–2012 Market Estimates*. US Environmental Protection Agency; 2017.
 11. *2021 Pesticide Use Report Highlights*. CDPR; 2021. Available at: https://www.cdpr.ca.gov/docs/pur/pur21rep/pur_highlights_2021.pdf. Accessed 19 July 2023.
 12. *Agricultural Pesticide Use Near Public Schools in California*. California Environmental Health Tracking Program; 2014. Available at: <https://www.phi.org/wp-content/uploads/migration/uploads/application/files/m01vrkqvtqh6897f165fyegso0p8qqqudrto9v13d6uiocq0r.pdf>. Accessed 18 July 2023.
 13. López-Gálvez N, Wagoner R, Quirós-Alcalá L, et al. Systematic literature review of the take-home route of pesticide exposure via biomonitoring and environmental monitoring. *Int J Environ Res Public Health*. 2019;16:2177.
 14. Hyland C, Laribi O. Review of take-home pesticide exposure pathway in children living in agricultural areas. *Environ Res*. 2017;156:559–570.
 15. Hernández AF, Parrón T, Alarcón R. Pesticides and asthma. *Curr Opin Allergy Clin Immunol*. 2011;11:90–96.
 16. Dereumeaux C, Fillol C, Quenel P, Denys S. Pesticide exposures for residents living close to agricultural lands: a review. *Environ Int*. 2020;134:105210.
 17. Raanan R, Gunier RB, Balmes JR, et al. Elemental sulfur use and associations with pediatric lung function and respiratory symptoms in an agricultural community (California, USA). *Environ Health Perspect*. 2017;125:087007.
 18. Cone JE, Wugofski L, Balmes JR, et al. Persistent respiratory health effects after a metam sodium pesticide spill. *Chest*. 1994;106:500–508.
 19. Hernandez A, Martin-Rubi J, Ballesteros J, et al. Clinical and pathological findings in fatal 1, 3-dichloropropene intoxication. *Hum Exp Toxicol*. 1994;13:303–306.
 20. Hoppin JA, Umbach DM, London SJ, et al. Pesticides and atopic and nonatopic asthma among farm women in the agricultural health study. *Am J Respir Crit Care Med*. 2008;177:11–18.
 21. Hoppin JA, Umbach DM, London SJ, Alavanja MC, Sandler DP. Chemical predictors of wheeze among farmer pesticide applicators in the agricultural health study. *Am J Respir Crit Care Med*. 2002;165:683–689.
 22. Goksör E, Åmark M, Alm B, Ekerljung L, Lundbäck B, Wennergren G. High risk of adult asthma following severe wheezing in early life. *Pediatr Pulmonol*. 2015;50:789–797.
 23. Ma H, Li Y, Tang L, et al. Impact of childhood wheezing on lung function in adulthood: a meta-analysis. *PLoS One*. 2018;13:e0192390.
 24. Rubner FJ, Jackson DJ, Evans MD, et al. Early life rhinovirus wheezing, allergic sensitization, and asthma risk at adolescence. *J Allergy Clin Immunol*. 2017;139:501–507.
 25. Hyvärinen MK, Kotaniemi-Syrjänen A, Reijonen TM, Korhonen K, Korppi MO. Teenage asthma after severe early childhood wheezing: an 11-year prospective follow-up. *Pediatr Pulmonol*. 2005;40:316–323.
 26. Rodríguez-Martínez CE, Sossa-Briceño MP, Castro-Rodríguez JA. Factors predicting persistence of early wheezing through childhood and adolescence: a systematic review of the literature. *J Asthma Allergy*. 2017;10:83–98.
 27. Piippo-Savolainen E, Korppi M. Wheezy babies—wheezy adults? Review on long-term outcome until adulthood after early childhood wheezing. *Acta Paediatr*. 2008;97:5–11.
 28. Owora AH, Zhang Y. Childhood wheeze trajectory-specific risk factors: a systematic review and meta-analysis. *Pediatr Allergy Immunol*. 2021;32:34–50.
 29. Raanan R, Harley KG, Balmes JR, Bradman A, Lipsett M, Eskenazi B. Early-life exposure to organophosphate pesticides and pediatric respiratory symptoms in the CHAMACOS cohort. *Environ Health Perspect*. 2015;123:179–185.
 30. Keleb A, Daba C, Asmare L, et al. The association between children's exposure to pesticides and asthma, wheezing, and lower respiratory tract infections: a systematic review and meta-analysis. *Front Public Health*. 2024;12:1402908.
 31. Gilden RC, Harris RL, Friedmann EJ, et al. Systematic review: association of pesticide exposure and child wheeze and asthma. *Curr Pediatr Rev*. 2023;19:169–178.
 32. Islam JY, Hoppin J, Mora AM, et al. Respiratory and allergic outcomes among 5-year-old children exposed to pesticides. *Thorax*. 2023;78:41–49.
 33. Elsiwi B, Eskenazi B, Bornman R, et al. Maternal exposure to pyrethroid insecticides during pregnancy and respiratory allergy symptoms among children participating in the Venda Health Examination of Mothers, Babies and their Environment (VHEMBE). *Environ Res*. 2024;242:117604.
 34. Reardon AM, Perzanowski MS, Whyatt RM, Chew GL, Perera FP, Miller RL. Associations between prenatal pesticide exposure and cough, wheeze, and IgE in early childhood. *J Allergy Clin Immunol*. 2009;124:852–854.
 35. Gilden R, Friedmann E, Holmes K, et al. Gestational pesticide exposure and child respiratory health. *Int J Environ Res Public Health*. 2020;17:7165.
 36. State of the Air. American Lung Association; 2022. Available at: <https://www.lung.org/getmedia/74b3d3d3-88d1-4335-95d8-c4e47d0282c1/sota-2022.pdf>. Accessed 18 July 2023.
 37. Farzan SF, Razafy M, Eckel SP, Olmedo L, Bejarano E, Johnston JE. Assessment of respiratory health symptoms and asthma in children near a drying saline lake. *Int J Environ Res Public Health*. 2019;16:3828.
 38. Johnston JE, Razafy M, Lugo H, Olmedo L, Farzan SF. The disappearing Salton Sea: a critical reflection on the emerging environmental threat of disappearing saline lakes and potential impacts on children's health. *Sci Total Environ*. 2019;663:804–817.
 39. Lipsett M, Smorodinsky S, English P, Copan L. *BASTA Border Asthma & Allergies Study: Final Report*. Richmond; 2009.
 40. *2021 Pesticide Use Report Highlights*. California Department of Pesticide Use registry; 2021. Available at: https://www.cdpr.ca.gov/docs/pur/pur21rep/pur_highlights_2021.pdf. Accessed 19 July 2023.
 41. *2020 Pesticide Use Report Highlights*. California Environmental Protection Agency Department of Pesticide Regulation; 2020. Available at: https://www.cdpr.ca.gov/docs/pur/pur20rep/pur_highlights_2020.pdf. Accessed 19 July 2023.
 42. *2019 Pesticide Use Report Highlights*. California Environmental Protection Agency Department of Pesticide Regulation; 2019. Available at: https://www.cdpr.ca.gov/docs/pur/pur19rep/pur_highlights_2019.pdf. Accessed 19 July 2023.
 43. *2018 Pesticide Use Report Highlights*. California Environmental Protection Agency Department of Pesticide Regulation; 2018. Available at: https://www.cdpr.ca.gov/docs/pur/pur18rep/pur_highlights_2018.pdf. Accessed 19 July 2023.
 44. *2017 Pesticide Use Report Highlights*. California Environmental Protection Agency Department of Pesticide Regulation; 2017. Available at: https://www.cdpr.ca.gov/docs/pur/pur17rep/pur_highlights_2017.pdf. Accessed 19 July 2023.
 45. Farzan SF, Kamai E, Duenas Barahona D, et al. Cohort profile: the Assessing Imperial Valley Respiratory Health and the Environment (AIRE) study. *Paediatr Perinat Epidemiol*. 2024;38:359–369.
 46. Asher M, Keil U, Anderson H, et al. International Study of Asthma and Allergies in Childhood (ISAAC): rationale and methods. *Eur Respir J*. 1995;8:483–491.
 47. Holguin F, Flores S, Ross Z, et al. Traffic-related exposures, airway function, inflammation, and respiratory symptoms in children. *Am J Respir Crit Care Med*. 2007;176:1236–1242.
 48. Kraai S, Verhagen LM, Valladares E, et al. High prevalence of asthma symptoms in Warao Amerindian children in Venezuela is significantly associated with open-fire cooking: a cross-sectional observational study. *Respir Res*. 2013;14:1–10.
 49. Stellman SD, Thomas PA, S. Osahan S, Brackbill RM, Farfel MR. Respiratory health of 985 children exposed to the World Trade Center disaster: report on World Trade Center Health Registry wave 2 follow-up, 2007–2008. *J Asthma*. 2013;50:354–363.
 50. Gatto NM, Rhodes SL, Manthripragada AD, et al. α -Synuclein gene may interact with environmental factors in increasing risk of Parkinson's disease. *Neuroepidemiology*. 2010;35:191–195.
 51. Furlong MA, Paul KC, Cockburn M, et al. Ambient pyrethroid pesticide exposures in adult life and depression in older residents of California's central valley. *Environ Epidemiol*. 2020;4:e123.
 52. Bukalasa JS, Brunekreef B, Brouwer M, et al. Associations of residential exposure to agricultural pesticides with asthma prevalence in adolescence: the PIAMA birth cohort. *Environ Int*. 2018;121:435–442.
 53. Kudagammana ST, Mohotti K. Environmental exposure to agrochemicals and allergic diseases in preschool children in high grown tea plantations of Sri Lanka. *Allergy Asthma Clin Immunol*. 2018;14:1–5.
 54. Gunier RB, Ward MH, Airola M, et al. Determinants of agricultural pesticide concentrations in carpet dust. *Environ Health Perspect*. 2011;119:970–976.
 55. Harnly ME, Bradman A, Nishioka M, et al. Pesticides in dust from homes in an agricultural area. *Environ Sci Technol*. 2009;43:8767–8774.

56. Harnly M, McLaughlin R, Bradman A, Anderson M, Gunier R. Correlating agricultural use of organophosphates with outdoor air concentrations: a particular concern for children. *Environ Health Perspect.* 2005;113:1184–1189.
57. Naimi AI, Whitcomb BW. Estimating risk ratios and risk differences using regression. *Am J Epidemiol.* 2020;189:508–510.
58. Maritano S, Moirano G, Popovic M, et al. Maternal pesticides exposure in pregnancy and the risk of wheezing in infancy: a prospective cohort study. *Environ Int.* 2022;163:107229.
59. Salameh P, Baldi I, Brochard P, Raheison C, Abi Saleh B, Salamon R. Respiratory symptoms in children and exposure to pesticides. *Eur Respir J.* 2003;22:507–512.
60. Xu X, Nembhard WN, Kan H, Becker A, Talbott EO. Residential pesticide use is associated with children's respiratory symptoms. *J Occup Environ Med.* 2012;54:1281–1287.
61. Malaeb D, Hallit S, Sacre H, Hallit R, Salameh P. Factors associated with wheezing among Lebanese children: results of a cross-sectional study. *Allergol Immunopathol (Madr).* 2020;48:523–529.
62. Hoppin JA, Umbach DM, Long S, et al. Pesticides are associated with allergic and non-allergic wheeze among male farmers. *Environ Health Perspect.* 2017;125:535–543.
63. Islam JY, Mohamed A, Umbach DM, et al. Allergic and non-allergic wheeze among farm women in the Agricultural Health Study (2005–2010). *Occup Environ Med.* 2022;79:744–751.
64. Gunier RB, Raanan R, Castorina R, et al. Residential proximity to agricultural fumigant use and respiratory health in 7-year old children. *Environ Res.* 2018;164:93–99.
65. Gharibi H, Entwistle MR, Schweizer D, Tavallali P, Cisneros R. The association between 1, 3-dichloropropene and asthma emergency department visits in California, USA from 2005 to 2011: a bidirectional-symmetric case crossover study. *J Asthma.* 2020;57:601–609.
66. Gharibi H, Entwistle MR, Schweizer D, Tavallali P, Thao C, Cisneros R. Methyl-bromide and asthma emergency department visits in California, USA from 2005 to 2011. *J Asthma.* 2020;57:1227–1236.
67. Lee S, McLaughlin R, Harnly M, Gunier R, Kreutzer R. Community exposures to airborne agricultural pesticides in California: ranking of inhalation risks. *Environ Health Perspect.* 2002;110:1175–1184.
68. Fryer AD, Lein PJ, Howard AS, Yost BL, Beckles RA, Jett DA. Mechanisms of organophosphate insecticide-induced airway hyperreactivity. *Am J Physiol Lung Cell Mol Physiol.* 2004;286:L963–L969.
69. Pope CN. Organophosphorus pesticides: do they all have the same mechanism of toxicity? *J Toxicol Environ Health Part B.* 1999;2:161–181.
70. Banks CN, Lein PJ. A review of experimental evidence linking neurotoxic organophosphorus compounds and inflammation. *Neurotoxicology.* 2012;33:575–584.
71. Shaffo FC, Grodzki AC, Fryer AD, Lein PJ. Mechanisms of organophosphorus pesticide toxicity in the context of airway hyperreactivity and asthma. *Am J Physiol Lung Cell Mol Physiol.* 2018;315:L485–L501.
72. Fukuyama T, Tajima Y, Ueda H, et al. Allergic reaction induced by dermal and/or respiratory exposure to low-dose phenoxyacetic acid, organophosphorus, and carbamate pesticides. *Toxicology.* 2009;261:152–161.
73. Kumar S, Khoudoun M, Kettleson EM, et al. Glyphosate-rich air samples induce IL-33, TSLP and generate IL-13 dependent airway inflammation. *Toxicology.* 2014;325:42–51.
74. Nishino R, Fukuyama T, Watanabe Y, Kurosawa Y, Ueda H, Kosaka T. Effect of mouse strain in a model of chemical-induced respiratory allergy. *Exp Anim.* 2014;63:435–445.
75. Fukuyama T, Tajima Y, Ueda H, et al. A method for measuring mouse respiratory allergic reaction to low-dose chemical exposure to allergens: an environmental chemical of uncertain allergenicity, a typical contact allergen and a non-sensitizing irritant. *Toxicol Lett.* 2010;195:35–43.
76. Saulsbury MD, Heyliger SO, Wang K, Johnson DJ. Chlorpyrifos induces oxidative stress in oligodendrocyte progenitor cells. *Toxicology.* 2009;259:1–9.
77. Fahy JV. Type 2 inflammation in asthma—present in most, absent in many. *Nat Rev Immunol.* 2015;15:57–65.
78. Lee K, Smith JL, Last JA. Absence of respiratory inflammatory reaction of elemental sulfur using the California pesticide illness database and a mouse model. *J Agromedicine.* 2005;10:41–47.
79. Teyssie R, Manangama G, Baldi I, et al. Assessment of residential exposures to agricultural pesticides: a scoping review. *PLoS One.* 2020;15:e0232258.
80. *Perils of Pesticides Address to Pacific Lutheran University* 1989. Chavez Foundation; 1989. Available at: <https://chavezfoundation.org/speeches-writings/#1549063588679-ed96425e-7969>. Accessed 26 October 2023.
81. Schwartz NA, von Glasco CA, Torres V, Ramos L, Soria-Delgado C. “Where they (live, work and) spray”: pesticide exposure, childhood asthma and environmental justice among Mexican-American farmworkers. *Health place.* 2015;32:83–92.
82. Alavanja MC, Sandler DP, McMaster SB, et al. The agricultural health study. *Environ Health Perspect.* 1996;104:362–369.
83. Liebman AK, Wiggins MF, Fraser C, Levin J, Sidebottom J, Arcury TA. Occupational health policy and immigrant workers in the agriculture, forestry, and fishing sector. *Am J Ind Med.* 2013;56:975–984.
84. Gunier RB, Bradman A, Harley KG, Eskenazi B. Will buffer zones around schools in agricultural areas be adequate to protect children from the potential adverse effects of pesticide exposure? *PLoS Biol.* 2017;15:e2004741.
85. *CalEnviroScreen 4.0*. California Office of Environmental Health Hazard Assessment; 2021. Available at: <https://oehha.ca.gov/calenviroscreen/report/calenviroscreen-40>. Accessed 26 October 2023.
86. Li L, Johnson B, Segawa R. Empirical relationship between use, area, and ambient air concentration of methyl bromide. *J Environ Qual.* 2005;34:420–428.