A Network Toxicology Approach for Mechanistic Modelling of Nanomaterial Hazard and Adverse Outcomes

Giusy del Giudice, Angela Serra, Alisa Pavel, Marcella Torres Maia, Laura Aliisa Saarimäki, Michele Fratello, Antonio Federico, Harri Alenius, Bengt Fadeel, and Dario Greco*

Hazard assessment is the first step in evaluating the potential adverse effects of chemicals. Traditionally, toxicological assessment has focused on the exposure, overlooking the impact of the exposed system on the observed toxicity. However, systems toxicology emphasizes how system properties significantly contribute to the observed response. Hence, systems theory states that interactions store more information than individual elements, leading to the adoption of network based models to represent complex systems in many fields of life sciences. Here, they develop a network-based approach to characterize toxicological responses in the context of a biological system, inferring biological system specific networks. They directly link molecular alterations to the adverse outcome pathway (AOP) framework, establishing direct connections between omics data and toxicologically relevant phenotypic events. They apply this framework to a dataset including 31 engineered nanomaterials with different physicochemical properties in two different in vitro and one in vivo models and demonstrate how the biological system is the driving force of the observed response. This work highlights the potential of network-based methods to significantly improve their understanding of toxicological mechanisms from a systems biology perspective and provides relevant considerations and future data-driven approaches for the hazard assessment of nanomaterials and other advanced materials.

1. Introduction

Chemical safety assessment, comprising the evaluation of exposure hazard and risk, is essential for safeguarding human and environmental health.^[1] Hazard assessment relies on model systems to perform an initial evaluation of potential adverse effects of chemicals.

Traditionally, hazard evaluation focused on the compound characteristics, overlooking the context of the exposure that occurs in real-life situations. However, a growing perspective suggests that the molecular configuration of the biological system exposed (here defined as all the active genes and present epigenetic modifications at the physiological stage) can influence the observed phenotype.^[2] This is especially relevant for the emerging class of advanced materials, such as engineered nanomaterials (ENMs), where extrinsic descriptors (namely those properties which are affected by the test system exposed), take different values depending on external conditions.^[3] Indeed, in contemporary toxicology, the

G. del Giudice, A. Serra, A. Pavel, M. Torres Maia, L. A. Saarimäki, M. Fratello, A. Federico, D. Greco Finnish Hub for Development and Validation of Integrated Approaches

(FHAIVE), Faculty of Medicine and Health Technology Tampere University Tampere 33520, Finland E-mail: dario.greco@tuni.fi G. del Giudice, A. Serra, L. A. Saarimäki, A. Federico, D. Greco

Division of Pharmaceutical Biosciences, Faculty of Pharmacy University of Helsinki Helsinki 00790, Finland

The ORCID identification number(s) for the author(s) of this article can be found under https://doi.org/10.1002/advs.202400389

© 2024 The Author(s). Advanced Science published by Wiley-VCH GmbH. This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

DOI: 10.1002/advs.202400389

A. Serra, A. Federico, D. Greco Tampere Institute for Advanced Study **Tampere University** Tampere 33100, Finland H. Alenius Human Microbiome Research Program (HUMI) University of Helsinki Helsinki 00014, Finland H. Alenius, B. Fadeel Institute of Environmental Medicine Karolinska Institutet Stockholm 171 77, Sweden D Greco Institute of Biotechnology University of Helsinki Helsinki 00790, Finland

observed toxicities are often assumed to be primarily influenced by the dose and the intrinsic properties of the ENMs exposure (i.e., its physicochemical characteristics and all the properties that are not dependent on the test system exposed), while the features of the exposed biological system (test system) are referred to as "biological descriptors" of the exposure.^[4] However, understanding the system effects is important for selecting relevant models and to correctly characterize the hazard.^[5] Thus far, a plethora of omics (e.g., transcriptomics, proteomics) data have been generated to investigate the molecular responses to compounds in various test systems.^[6] However, the impact of the biological system on the observed toxicity is frequently overlooked.

The view of living organisms as integrated systems of dynamic and interrelated components has shaped the field of systems biology.^[7] This holistic paradigm also includes interactions between organisms with xenobiotics, which is the conceptual foundation of systems toxicology.^[8] The systems theory paradigm states that the interactions store more information than the individual elements of the system (the whole is bigger than the sum¹). In this light, molecular profiling of systems alterations are interpreted using analytical approaches with increasing complexity, often based on network theory.^[9,10] We previously described a systems toxicology approach for ENM grouping and prioritization, demonstrating that including molecular alterations of biological systems after the exposure provides better molecular proxies of toxicity.^[11] However, traditional approaches in mechanistic toxicogenomics do not explicitly require modeling the interactions between genes, and are based on the assessment of a linear representation of the response.^[12] Established methods to characterize chemicals mechanism of action (MOA) using omics data, preprocess and filter data in order to select the most relevant alterations and characterize them functionally. A common strategy uses gene ontology and pathway databases to interpret these alterations.^[13] Similarly, other studies interpreted differentially expressed genes and dose responsive genes using the adverse outcome pathway (AOP) framework.^[4,14] Such representations, however, are often difficult to associate with system-level effects, and require interpretation and manual reconstruction of the events. We recently demonstrated on a larger set of transcriptomic alterations associated with ENM exposures that, despite the heterogeneity of individual profiles, conserved mechanisms of gene regulation underlie the response to nanoparticulate exposure across multiple species.^[6] The patterns of co-regulation can be represented as connections between genes and highlighted by network models. We previously used co-expression network inference to characterize the MOA of nanomaterials.^[15] However, also in this case, molecular alterations were interpreted via functional annotation, requiring the MOA to be manually reconstructed. Therefore, translating gene-level data into interpretable biological responses is needed in order to explain the observed toxicity in a mechanistic fashion.

Systems effects in toxicology are usually expressed as a causal chain of molecular, cellular and systemic events linking exposures with adverse outcomes (i.e., functional and apical endpoints). These representations are organized in the AOP framework, a multiscale concept connecting early molecular initiating events (MIEs) to adverse outcomes (AOs) through a causally linked chain of events, referred to as key events (KEs).^[16] In hazard assessment, AOPs are usually selected a priori based

on the end-point of interest. However, the structure of the AOP and the presence of shared events in multiple pathways allows the representation of the entire AOP as a comprehensive network of events which can be applied to a variety of stressors.^[17-22] The use of a network of events allows the simultaneous investigation of multiple toxicity mechanisms and the prediction of multiple adverse outcomes. However, limitations still exist in connecting this conceptual framework with molecular alterations (as deduced from omics data). The recently curated molecular annotation of KEs represents an essential step in connecting AOPs to measurable and mechanistic information as derived from toxicogenomics data.^[14,23] Indeed, the annotation provides sets of genes associated to each event, which has been manually curated on a selection of taxonomically relevant AOPs. The annotation was based on information derived from pathways, phenotypes and gene ontologies, and provides a robust and functionally relevant molecular dimension to the AOP framework.

Here, we describe a novel network toxicology framework that directly interprets the molecular response to chemicals as a chain of toxicologically relevant events. Our framework exploits network properties to contextualize the toxicological responses with respect to the test system. We showcase examples based on a comprehensive set of 31 industrially relevant ENMs of varying chemistries, which are representative of a range of different physicochemical properties and hazard potentials. Our dataset was previously tested on both in vitro and in vivo test systems, with global expression levels of mRNA, miRNA, and proteins previously assessed from two human cell lines (monocyte-like THP-1 cells and bronchial epithelial BEAS-2B cells), mRNA expression from mouse lung tissues, protein corona profiles, and comprehensive characterization of all the ENMs.^[11] While other studies considering both in vitro and in vivo assays focused on developing in vitro-to-in vivo extrapolation models for doseresponse evaluation, this study focuses on exploiting network and systems theory to contextualize the molecular response to ENMs with respect to the test system used and interpreting it as a coherent chain of causative events leading to toxicologically relevant effects.

2. Results and Discussion

2.1. A Novel Network-Based Framework Links Toxicogenomics Data to AOPs

Representing the response to an exposure as the result of the biological system-chemical interactions requires modeling approaches that scale from molecular alterations to phenotypic effects in a system-dependent manner. Different biological systems respond differently to the same stimulus, due to their distinct molecular buildup (genotype and epigenotype).^[24] This is reflected by the differences in the expression profiles, both in terms of amplitude of the response, as well as coordinated expression patterns (co-expression) across genes.^[25] The response mechanism to an exposure is investigated as all the statistically significant changes at a cellular and molecular level induced by the substance. In transcriptomic studies, this is usually represented as lists of differentially expressed genes (DEGs) which, in turn, are interpreted via functional annotation and/or enrichment (**Figure 1**).^[26]





3 2 1 n Representation of the Interpretation of the MoA reconstruction molecular alteration toxicological effect Arbitrary interpretation **Functional annotation** of retrieved functions (lists of pathways) Differential analysis (lists of genes) AOP and KE annotation/ Arbitrary interpretation of enrichment retrieved events (lists of AOPs/KEs) Transcriptomic data generation and preprocessing Arbitrary interpretation of Community detection functional enrichment retrieved functions for (lists of pathways) each module Network inference Reconstruction of an AOP-based network of alterations

Figure 1. Approaches for the analysis and interpretation of the MOA of ENMs starting from toxicogenomics data. In step 0, toxicogenomics data are generated and preprocessed. In step 1, traditional methods represent molecular alterations in terms of lists of differentially expressed genes (DEGs), while other studies (including this study) model them as co-expression networks. In step 2, previous approaches reconstruct the mechanism of action (MOA) via functional annotation or enrichment of adverse outcome pathway (AOP) associated genes, leading to an arbitrary interpretation of the retrieved functions (step 3). The proposed study allows a reconstruction of an AOP based network of alterations, requiring minimal interpretation with respect to previous approaches.

The obtained results (lists of annotated pathways) require manual interpretation in order to reconstruct the MOA and connect functions with toxicologically relevant endpoints. While previous studies have linked the transcriptomic profile of ENMs exposure to AOPs, reconstructing a unique mechanism of response from lists of events still requires manual interpretation.

Here, we developed a framework that exploits network properties and the AOP annotation to convert molecular alterations into chains of events across organismal complexity levels (from molecular to tissue/organism). The aim is to interpret input toxicogenomics data exploiting the connection between events (representative of the mechanism) and genes (representative of the test system) in order to reconstruct a model of AOP-based mechanism of action (**Figure 2**). This alternative approach is completely data driven, and requires minimum interpretation with respect to previous methods, converting omics data in a set of consecutive events that are more readily assessed.

We tested our framework by characterizing the effects of 31 ENMs in three biological systems (the human monocytic cell line THP-1, the human bronchial epithelial cell line BEAS-2B, and lung tissue from C57BL/6 mice exposed by oropharyngeal aspiration). While it is not feasible to characterize the response to all available ENMs in a single study, we selected a dataset as it is representative of different physicochemical characteristics (e.g.

shape, size, core chemistry, and surface functionalisation) with a range of hazard potentials.^[27,28] This allows to investigate materials with various characteristics, as well as to compare exposures while minimizing confounding factors deriving from different experimental designs and nanoparticles manufacturing differences. From the original dataset we identified 93 experimental settings (test system, time, dose, and materials exposed) which we modeled inferring co-expression networks. In this study, an *exposure* will be considered as a combination of these experimental variables. In order to infer a network for each exposure, the top and bottom 100 deregulated genes with respect to the control were selected, and a statistical test (i.e., the Hotelling test) was used to assess shared patterns in the expression of each gene pair.

We systematically evaluated the differences arising from the traditional MOA investigation and the use of network-based approaches. All the results for each exposure are reported in the supplementary materials and the comparison results are described in the following paragraph (Supplementary Files, available at https://doi.org/10.5281/zenodo.10390383 in the "enrichment_results" and "network_comparison_results" folders, respectively). In order to identify significantly represented KEs in the exposure co-expression networks, we developed a topological enrichment method. The method consists in comparing the co-expression network in input (based on experimental data) with





Figure 2. Workflow converting molecular profiles in networks based on the AOP framework. Network models of the exposure (co-expression networks or preselected edgelists) are used as input of the framework. In step 1, a topological enrichment of the adverse outcome pathway (AOP) network is performed based on the input data. In step 2, possible molecular initiating events (MIEs, blue squares) and adverse outcomes (AOs, red squares) connecting to the enriched events (KEs, green squares) are prioritized based on topological properties of the AOP network and molecular alterations. In step 3, the list of MIEs, KEs and AOs is mapped on the AOP network to form an AOP-based interpretation of the mechanism of action (MOA). The results can be visualized as a network or in tabular form, and interpreted independently or compared with other exposure mechanisms.

the AOP network built by combining all the MIEs, KEs, and AOs as they are linked in the AOP-wiki. In detail, our strategy identifies KEs as enriched in the co-expression network if their genes are more densely connected in the AOP network with respect to random sets of genes of the same size (cf. Methods) (Figure 2). The enrichment results were then filtered to exclude KEs for which the expression of the annotated genes did not reach a significant alteration, ensuring that the enriched KE had a relevant effect in the biological system under study. The filtered results were then mapped on the AOP network and the MOA was reconstructed. To do so, the most probable MIEs and AOs within the AOP network were selected based on topological and molecular information. The workflow is depicted in Figure 2 (refer to Methods for details).

2.2. Network-Based Approach can be used to Mechanistically Interpret Toxicogenomics Data

In order to analyze the results produced by our framework, we selected three case studies on ENMs exposures which were previously described in,^[27,28] demonstrating that the proposed approach can recapitulate the main findings while providing additional knowledge with respect to previous studies relying on DEG analysis. Previous in vivo studies on the same dataset revealed that the multi-walled carbon nanotubes (MWCNTs) triggered the most prominent changes in the transcriptome amongst all studied materials, as well as an extensive eosinophil infiltration in mouse lungs.^[28] Here, we evaluated the reconstructed mechanism of pristine MWCNTs in the mouse lung using as an input the co-expression network derived from exposure data. Our novel network-based framework reported a response mechanism mostly centered on inflammation and reactive oxygen species (ROS) production, which has been largely discussed in the field (Supplementary Figure 1).^[29-34] The enrichment of "frustrated phagocytosis" is a relevant event that plays a pivotal role in responses to high aspect ratio materials. For instance, frustrated phagocytosis has been implicated in the prolonged production of proinflammatory cytokines and ROS,^[35] as well as poor clearance of inhaled particles.^[36] All of these factors contribute to the pathogenic potential of MWCNT exposures and favor the development of pulmonary fibrosis, which is a known long term consequence of MWCNT exposure (Supplementary Figure 1).^[37] Our results also highlighted the enrichment of "decreased fibrinolysis", which tightly controls the enzymatic process ensuring the breakdown of fibrin, hence controlling blood clot formation. While the disruption of fibrinolysis is associated with inflammatory processes, direct interactions with ENMs and the coagulation system have also been reported.^[38,39] These mechanisms on fibrinolysis may also play a part in the profibrotic effects of MWCNTs.^[40] Interestingly, the retrieved mechanism also suggested a substantial epigenetic regulation (ncRNA expression alteration, peroxisome activator receptor promoter demethylation) which has been explored in independent studies on the same material.^[41,42] The discussed fibrosis and inflammatory events were well captured by the in vivo model, but were less evident in THP-1 cells. Interestingly, only the BEAS-2B cells displayed enrichment for the modulation of the extracellular matrix composition and for alterations of the TGF- β dependent fibrosis pathway.

The same dataset revealed the most pronounced inflammation in the CuO-exposed group, associated with high levels of neutrophil infiltration in mouse lungs.^[28] Specifically, greater toxicity was observed for the pristine and amino-functionalized forms of CuO.^[27] This is also evident in the present study, where the reconstructed mechanisms of response reported apoptosis and decreased cell proliferation in both cell lines (Supplementary Table 1). Hence, while the pristine CuO enriched both events associated with inflammation in all the test systems and cell viability, the CuO-NH₂ showed a massive effect on cell proliferation in the two cell lines, and impact on the DNA repair, chromosome stability and altered microtubule dynamic in the in vivo model. Interestingly, the predicted MIEs included alteration of the oxidative status in all the test systems, and alteration of ion pumps and receptors in both lung and BEAS-2B. This is in line with the other findings that report a toxicity mechanism





Figure 3. Reconstructed mechanism of response to titanium dioxide (TiO₂) exposures with varying shapes, i.e., spherical (A) versus rod-shaped (B). Enriched key events (KEs), molecular initiating events (MIEs) and adverse outcomes (AOs) are reported in green, blue, and red, respectively. Dashed lines connect events which are not formally linked in the AOP-Wiki database, but are related to common biological processes (not provided as output of the framework).

mainly based on oxidative insults.^[43] Similar results were also observed in the A549 cell line, revealing an acute genotoxicity of nano-sized CuO, which is well captured by our framework.^[43-45]

We also investigated the response mechanism to titanium dioxide (TiO₂) in the mouse lung, and compared the effects of variations in the material geometry (Figure 3). Previous studies demonstrated that the spherical TiO₂ or rod-shaped induce different responses, albeit without describing the specific mechanism.^[28] Our framework highlights the effect of the rod shape on the cell membrane, with initiating events comprising narcosis and decompartmentalization (Figure 2B). Alterations of the cell membrane also affect lipid composition and induce the activation of AKT2 signaling, which regulate many processes including metabolism, proliferation, cell survival, growth and angiogenesis.^[46] The spherical TiO₂, instead, showed its main effect on inflammation, ROS production, ion balance alteration, and effects on the calcitonin gene-related peptide, inducing respiratory alterations (Figure 2A and Supplementary Files, available at https://doi.org/10.5281/zenodo.10390383 in the "enrichment_results" folder). Moreover, the framework captured the genotoxicity associated with TiO₂ nanoparticles, one of the concerns behind the recent and much-debated EU ban on their use as food additives.^[47] The effects of rod-shaped TiO₂ have been scarcely described to date; however, other studies on rodshaped nanocarriers and rod-shaped silica and silver nanoparticles reported a significant effect of shape, leading to increased cell uptake, especially by epithelial cells.^[48-50] Notably, BEAS-2B response mechanisms to TiO_2 showed substantially more alterations with respect to the other test systems.

Finally, we evaluated the results obtained by our framework when giving in input lists of dose-dependent genes, which is a relevant scenario for regulatory purposes. We selected dosedependent data obtained from the analysis of epithelial primary cell lines derived from an asthmatic patient, subsequently exposed to CuO in an air liquid interface protocol (GSE127773).^[51] In this case, the enriched events captured the cilia impairment which is a typical characteristic of epithelial cells in the lungs of asthmatic patients, as well as an evident cytotoxic effect of epithelial cells (Supplementary table 2). Importantly, our findings, based solely on the expression data, are confirmed by the in vitro cytotoxicity assays results reported in the original study.

In sum, we have showcased a selection of the results, using MWCNTs, CuO, and TiO_2 as exemplars, to demonstrate the potential of our framework by validating the retrieved ENMs mechanism of action against previous studies, literature knowledge and in vitro assays. By achieving comparable conclusions in a data-driven manner, we show how our approach can alleviate the

ADVANCED SCIENCE NEWS ______ www.advancedsciencenews.com

problem of toxicogenomics data interpretability when profiling the molecular response to chemical nanomaterial exposure.

2.3. Representing the MOA as a Network Increases the Information Content

Systems biology relies on the principle that connections store more information than isolated pathways or gene sets.^[12] We previously demonstrated that different transcriptomic profiles in response to nanoparticulate share similar regulatory mechanisms.^[6] Gene-to-gene connections can be modeled through the construction of co-expression networks, where genes are linked based on the presence of a statistically significant co-expression. These relationships may originate from shared functional attributes, involvement in common pathways, participation in protein complexes, or engagement in transcriptional regulation.^[52] Therefore, we tested whether the response mechanism extracted from co-expression networks via our framework would provide better qualitative and quantitative results than the one relying solely on gene sets. Since network approaches can be used on entire sets of transcripts while gene expression analysis is based on the selection of differentially expressed genes, we purposefully inferred the network filtering genes based on their expression changes, so that no bias derived from an increased input could affect the results.

In order to evaluate how effectively the network based strategy performs with respect to the traditional representation of ENM responses (Figure 2), we compared the detected mechanisms when levels of neutrophil infiltration induced by ENM exposure are compared (^[11], (Supplementary Figure 2,3)). To ensure an unbiased comparison, we used as input our framework lists of DEGs and performed a classic enrichment of the KE, and compared the retrieved mechanisms. In both cases, our framework detects inflammatory events associated with exposures with low neutrophil infiltration (Supplementary Figure 2,3). Early pathological alterations in the lung are usually linked to injury of pulmonary endothelial cells and vessel destabilization.^[53] The network-based approach did capture this element, together with TGF- β activation, when medium neutrophil infiltration is induced (Supplementary Figure 2). The gene set based approach, instead, captures the TGF- β activation for high infiltration levels only (Supplementary Figure 3). Interestingly, exposures resulting in high neutrophil infiltration induce "oxidation of membrane lipids" and "NRF2 depression" in network reconstructed mechanisms, but not in the ones derived by gene sets (Supplementary Figure 2). NF-E2-related factor-2 is a regulator of cellular antioxidant responses, largely expressed by neutrophils and only recently identified as a key player in the regulation of inflammasome activation.^[54] The coexistence of oxidative stress markers, lipid peroxidation, inhibition of NFkB, and a high neutrophil infiltration (as evidenced by the network approach) is well known as being prognostic of chronic and deregulated inflammation and has recently been associated with lung function deterioration in COVID-19.[55]

In order to produce a quantitative evaluation of the use of network models to represent exposure responses, we defined measures of information content and compared the results obtained from gene sets and co-expression networks based strategies (Figure 4). Detailed results of the framework for each set of DEGs is reported in the supplementary files (available at https://doi.org/10.5281/zenodo.10390383 in the "network_comparison_results" folder). When profiling the MOA of an exposure, information gain can be defined as the ability to reconstruct a complete mechanism connecting the initiating event(s) to observable outcome(s). In this light, we defined the "AOP completeness score" as a measure of the ability to observe a complete chain of events connecting one or more MIE and one or more AO (Figure 4A, and refer to Methods). A desirable gain can also be expressed in terms of less uncertainty of the observed mechanism, indicating a more predictable or homogeneous exposure response. Therefore, we defined the "Key event connectivity score" as the ratio between the portion of connected KEs against the unconnected ones characterizing the MOA of an exposure. For the same exposure, the DEGs and the co-expression networks were used as input to obtain two independent models of the MOA. In both cases, we demonstrated that both the "AOP completeness score" and the "Key event connectivity score" showed higher values when a co-expression network was used as input (Figure 4, Supplementary Files available at https://doi.org/ 10.5281/zenodo.10390383 in the "network_comparison_results" folder). It is important to note that the AOP network was solely built based on the case-effect relationships as retrieved from the AOP-wiki database. Therefore, the inference of the co-expression network is completely unrelated to it, suggesting that the higher value of the "Key event connectivity score" only depends on the evaluation of gene-gene relationships. Taken together, these results demonstrate that a network-based representation of toxicogenomics data systematically outperforms a traditional analysis based on the evaluation of individual molecules. For this reason we focus on the results obtained from co-expression networks only.

2.4. Network-Based Models of Exposure Highlight Mechanisms Related to Different Hazard Levels

A current need in toxicological assessment is to prioritize highly hazardous materials for comprehensive testing. However, previous efforts showed that toxic responses cannot be clearly distinguished when analyzing whole transcriptomic profiles or by relying only on specific physicochemical properties.^[56] Given the increase in information content of the network approaches, we investigated whether it was possible to identify differences in nanomaterials exposures with varying toxicity levels. First, we grouped the 93 exposures based on the toxicity endpoints we had collected in ref. [11]. Briefly, for each of the 31 materials included in the original dataset, in vitro assays were performed in order to assess cytotoxicity, genotoxicity and immunotoxicity. In Fortino et al.,^[11] the results were homogenized and clustered, which resulted in a general toxicity score for each exposure and, finally, in three toxicity groups (no-to-low, medium, and high hazard). Second, for each group, we extracted overrepresented edges in the co-expression networks of the same group (namely edges which are significantly more present in a group of networks with respect to others). This allows the extraction of characteristics which are representative of that group, but not of others. Finally, the edges were interpreted in the AOP context using our framework. Our

SCIENCE NEWS _____ www.advancedsciencenews.com





Figure 4. Graphical representation of the defined information score based on the possibility of reconstructing the mechanism of action (MOA) of the compound. A) The "AOP completeness score" was defined as a measure of the ability to observe a complete chain of events connecting one or more molecular initiating events (MIEs) and one or more adverse outcomes (AOs). B) The "Key event connectivity score" was defined as the ratio between the portion of connected key events (KEs) against the unconnected ones characterizing the MOA of an exposure. For each score, pie charts representing the results in the information metric values when mechanisms of response are reconstructed based on lists of differentially expressed genes (DEGs) and co-expression networks.

results showed that low hazard exposures (as defined in ref. [11]) were associated with oxidative stress, recruitment of inflammatory cells, and decreased lung function (Supplementary Figure 4). In intermediate hazard exposures, cell proliferation alterations emerged, as well as cell activation, and cytoskeleton modifications (Supplementary Figure 4). Among the most relevant MIEs represented in medium hazard nanomaterial exposures, we found frustrated phagocytosis and tubulin binding (Supplementary Figure 4). Previous studies on high aspect ratio materials proved that mesothelial cells initiate pro-inflammatory responses when retaining materials, even after weeks, and can play an important role in stimulation of tumor growth.[57,58] Highly hazardous materials mainly enriched adenomas and carcinomas as AOs (Supplementary Figure 4). Interestingly, the relative representation of AOs was the highest in the events enriched by highly toxic compounds, with the amount of molecular initiating events following the opposite trend (Figure 5). This suggests that a possible difference between low, medium and highly toxic materials is related to the induction of alterations which are indicative of an adverse outcome. It is known that different stimuli can alter a cell state either in an irreversible or reversible manner.^[24] In the first case, the molecular response facilitates the transition of the cell from the steady state to a new stable state, whereas in the second case a temporary unstable state is reached that may be reverted to the initial one, once the stimulus is removed.^[24] In the context of an AOP, reaching an adverse outcome represents a multiscale

event implying that alterations have happened across biological levels of organization (molecular, cellular, tissue, organism). Tissue responses, for example, underlie cellular and molecular complex events that need to happen in order for the macroscopic effect to emerge. Our hypothesis echoes the "hierarchical oxidative stress model" of Nel et al., where the nanoparticle capability of inducing different levels of ROS, triggers molecular and cellular events with increasing toxicity potential.^[59] Our results suggest that highly hazardous materials may induce an alteration that quickly spreads across biological levels and moves towards a new phenotypic state which is, therefore, harder to revert.

2.5. Network Properties Underline the Impact of the Exposed Biological System

The results discussed above pointed towards differences between the test systems. Therefore, we decided to investigate the overall impact of the test system on all the 93 exposures we previously defined (i.e., dose, time, materials, and test system) by clustering the inferred co-expression networks based on their topological properties. Our analysis showed that biological systems have a significant impact on the response to ENM exposure, with exposures to different materials in the same test system clustering together (**Figure 6A**). Furthermore, exposures in the same biological system share portions of the response identified as the set of







Figure 5. Distribution of molecular initiating events (MIEs), key events (KEs) and adverse outcomes (AOs) across engineered nanomaterials (ENMs) toxicity classes (low, medium, high). Data are reported as a percentage of the total enriched events.

statistically overrepresented edges in the networks inferred from the transcriptomics data profiled in the same biological system. These overrepresented edges underlie functions which are representative of the system under evaluation (Figure 6C–E), such as inflammogenic functions for THP-1 cells (Figure 6C), phagocytosis and endocytosis for BEAS-2B cells (Figure 6E). However, for mouse lungs, this behavior is less evident, with less overrepresented edges and functions shared across the exposures. This is possibly due to the cytological complexity of the tissue, as well as to the different growth conditions of the test system which may affect the biocorona formation on ENMs^[60] (Figure 6D).

In order to evaluate the impact of the biological system on the molecular response, we clustered the reconstructed AOP-based response mechanisms for each exposure. We observed that the exposures cluster based on the induced biological response, rather than the exposed material (Supplementary Figure 5). Our results indicated that even when exposed to the same ENM, the observed response mechanism is a function of the interaction between the compound and the test system. Indeed, the 93 exposures clustered in two groups, almost completely separating the THP-1 and BEAS-2B exposures (Supplementary Figure 5). When a major inflammatory response was generated, THP-1 responses were closer to the responses in mouse lungs. On the contrary, when cell stress responses, mainly based on DNA damage, were activated, exposures in BEAS-2B overlapped more with the lung mechanism. Lung exposures split in the two groups according to the nature of the tissue response (Supplementary Figure 5). These results do not imply the validity of the in vitro or in vivo models, but highlight the effect of different test models provide on the MOA of nanomaterials, which needs to be considered when assessing the hazard.

Several previous efforts have focused on the possibility of identifying physicochemical similarities between the ENMs to predict common responses.^[61] Our results suggest that in vitro test systems, due to their molecular buildup, usually show only portions of the complete response, as already described.^[14] On the other hand, in complex systems, such as the lung tissue, a core mechanism drives the response, but it is difficult to dissect the contribution of specific cytological components. This supports the idea that, in order to correctly estimate hazards, information on the test system must be considered, as the observed mechanism of action depends on the test system used for the assessment even in comparable experimental settings (same material, same or equivalent doses and time of exposure). To further investigate this phenomenon, for each ENM in the original dataset, we compared the similarity of the in vitro systems and the in vivo counterpart. This is an important step in light of efforts to develop alternative methods to animal testing, where the in vivo system is often considered as the benchmark for new approach methodologies (NAMs). Other investigators have compared animal experiments with non-animal-based alternatives in order to assess in vitro-to-in vivo extrapolation (IVIVE).^[62] However, defining IVIVE relationships requires a focus on doses and physiologically based kinetic models, which is out of the scope of this study. Here, we focused on the impact of the in vitro or in vivo test system with respect to the observed response. In order to exploit the network properties to assess the similarities or differences between in vitro and in vivo systems, we defined a "biological system similarity score" as the comparison between the response mechanism in pairs of systems, with respect to the complete AOP network (refer to Methods for further details). More than half of the ENMs showed a higher similarity between mouse lungs and THP-1 cells, while the remaining exposures equally split in the other two categories



ADVANCED SCIENCE

www.advancedscience.com



Figure 6. Biological systems influence the response to ENM exposures. A) Clustering of the 93 exposures co-expression networks based on topological properties (edge betweenness, Jaccard, Hamming distance, SMC distance between the existing edges of the networks as well as a percentage value of shared edges between each pair of input networks). Each measure is converted into a distance when possible, and a consensus is taken as input for the clustering. Clustering results are reported as a heatmap, where the individual 93 exposures and the 5 clusters are plotted on the y axis and on the x axis, respectively. B) Similarity between biological systems in exposures of the same material. In teal, exposures where the highest similarity is between lung and THP-1, in violet when the highest value is between BEAS-2B and lung, in gray when the in vitro systems are not representative of the in vivo counterpart. Engineered nanomaterial (ENM) chemistry has been annotated. C–E) Top fifteen enriched pathways of the overrepresented edges for each biological system, with their respective enrichment score (ES). The enrichment was performed using a weighted Kolmogorov-Smirnov test against the information in the KEGG pathways database. The statistical significance of the enrichment analysis was estimated by permutation analysis over 100 random shuffles of the edge sets. The p-values were corrected for multiple comparisons using the false discovery rate (FDR) method and setting 0.05 as the significance threshold.

representing mouse lungs and BEAS-2B cells, and THP-1 cells and BEAS-2B cells, respectively (Figure 6B). This is not unexpected, as many ENMs induce an inflammatory response that is well captured by cell lines such as THP-1,^[63] suggesting that specific properties of the cell lines allow the mechanistic translatability across test systems. However, in 25% of the cases, the THP-1 and BEAS-2B in vitro systems shared more similarities than the in vivo system, suggesting that the in vitro assessed hazard was less overlapping with tissue-level observations in one quarter of the exposures. Thus, selecting the appropriate cell line(s), or utilizing a panel of cell lines of varying origin, may be important. $^{[2,64]}$

2.6. The Molecular Buildup Influences the Applicability Domain of Hazard Tests

Given the remarkable impact of the test systems on the observed response mechanism, we investigated whether the molecular buildup of the biological system would affect the assessable







Figure 7. Explorable AOP network is determined by the biological system. Subgraph of the AOP network containing the events enriched by the 31 engineered nanomaterials (ENMs) in the three biological systems: mouse lungs (orange), BEAS-2B cells (blue), and THP-1 cells (teal). Event type has been reported in the legend.

hazard with respect to the complete set of KEs in the AOP framework. An AOP represents the toxicologically relevant events to be tested in order to mechanistically assess the hazard of chemicals. To this aim, we combined the AOP-based response mechanisms of all 93 exposures in the different test systems, and represented them as a portion of the AOP network (**Figure 7**).

Mapping patterns of molecular alterations to the AOP has been shown to highlight similarities across biological systems.^[14] This is also observed in the present study, where a core portion of the AOP network is enriched by all the test systems, covering events related to oxidative stress, NADH-ubiquinone and ROS production, TGF- β and cytokine release. The OECD has recently concluded that nanotoxicity related to human health mainly arises from inflammation, oxidative stress, and cytotoxicity (https://one.oecd.org/document/ENV/CBC/MONO(2022)3/ en/pdf). These patterns were overlapping across test systems, suggesting that some early KEs involved in the initial assessment of hazard for ENMs, can indeed be tested in vitro while being representative of the in vivo counterpart. However, our results also show that test systems enrich different portions of the AOP network, de facto limiting the range of the assessable response mechanisms. THP-1 exposures specifically enriched pathways

related to immunity and adipogenesis. Furthermore, the airway hyperresponsiveness is a specific mechanistic chain visible only in this cell line, as well as infertility and placental insufficiency. BEAS-2B enriched KEs associated with neuronal degeneration, lysosomal dysfunction, as well as events specific to epithelial cells. Importantly, BEAS-2B enriched lung fibrosis and collagen deposition, which is a relevant long-term consequence of ENM exposure, but is not visible in all the test systems. Overall, molecular events such as nuclear receptor activations appear to be better observable in vitro. Surprisingly, the chain of NFkB related KEs, a pivotal element of ENM-induced responses, are only enriched by in vitro systems. Exposures in the lung, on the contrary, enriched calcium homeostasis related pathways, and the complete mechanistic chain connecting nuclear receptors with PPAR gamma and adipogenesis. Internalization of the compounds and DNA damage associated KEs are extensively observable in the lung, and partially shared with BEAS-2B. Given the experimental design of this dataset,^[27,28] this is a strong indicator that the differences in the observed events emerged because of the selected test system. The test system impact on the emerging phenotype upon compound exposure highlights the importance of considering both the exposed system and the compound in toxicological settings. A similar concept has been addressed by regulatory agencies as "the appropriateness of test species".^[65] However, this often only refers to the species-specific effects of ENMs. In contrast, we show here how the molecular buildup of different test systems impacts both the observed toxic effect and the detectable MOA in terms of toxicity testing.

Importantly, our results show that while some core events of ENMs exposure response are observable both in vitro and in vivo, the choice of the test system impacts the overall assessable response. While the applicability domain of a test system typically refers to the types of chemicals that can be tested, the present results suggest that exposures are not defined exclusively by the intrinsic features of the test substance. Our study indicates that in vitro systems can better assess molecular events, but they may fail to capture entire mechanistic chains, which are observable in the in vivo setting. These considerations suggest that our framework could be used to contextualize existing molecular toxicological evidence with respect to the biological system. The impact of the molecular buildup of the test system informs on the consistency and biological concordance criteria which are usually evaluated to weight the relevance of omics data for regulatory hazard assessment.

3. Conclusions

The ultimate aim of the 21st century chemical assessment is to build a mechanistic understanding of the chemical-biological interactions. This can be achieved by using the AOP framework, including knowledge concerning MIEs and KEs with their underlying data, to produce models able to associate unstudied substances with known AOPs mechanisms.[66] In this study, we presented a novel network-based framework for linking toxicogenomics data to AOPs and for interpreting the response mechanisms of biological systems to nanomaterial exposures. Our framework acknowledges the inherent variability in how different biological systems respond to the same stimuli (exemplified here by ENMs) due to their unique molecular buildup. It enables the conversion of molecular alterations into chains of events across different levels of organismal complexity, ranging from molecular to tissue/organism events. This eases the challenges of interpreting omic responses and provides a robust and generalizable strategy to link molecular evidence with toxicologically relevant phenotypes.

To validate our framework, we conducted an extensive analysis of 93 exposures to 31 ENMs in three distinct biological systems. This dataset was selected for its comprehensive coverage of industrially relevant ENMs with varying physicochemical characteristics and well-defined hazard levels.^[11] Importantly, while we presented and validated our framework on ENMs, we strongly believe that it can seemingly be used on a variety of chemical exposures. Indeed, AOPs are stressoragnostic, and can easily be applied to chemical and not chemical stressors, as already reported.^[67] Similarly, omics data can be obtained from a plethora of exposure conditions, which expands the applicability of our methodology beyond nanomaterial exposure.

We systematically compared the outcomes of our networkbased approach with traditional MOA investigations. Our results revealed that the network-based approach provided a deeper insight into the observed response. Representing MOAs as networks offers a framework to interpret multiscale complex events, extrapolating substantial more information as compared to other analytical frameworks for interpretation of toxicogenomics data. While network-based approaches are already well established in other fields of life sciences, systems toxicology applications are still in their infancy, with network properties hardly being exploited to solve toxicologically relevant issues.^[9,10] This study underscores for the first time the potential of network theory in improving toxicogenomics data interpretability and its contribution to hazard assessment.

We also addressed the crucial issue of hazardous materials identification, a challenging task that traditional toxicogenomics methods struggle with. By applying our framework to ENMs exposures inducing various levels of toxicity, we highlighted common features and distinct molecular response mechanisms in each group. This approach allowed us to connect highly hazardous materials with molecular alterations associated with more apical events in the AOP framework. Importantly, this suggests that toxicity could be re-defined as the ability of inducing a multiscale alteration reaching a new stable patho/physiological state (e.g., an AO) in a particular experimental setting. In this light, no exposure would be "inert"; instead, it simply induces an alteration that can be easily reverted to homeostasis. While dosing and temporal information cannot be considered by the current framework, we believe that future studies can integrate this information. This key insight, and the link to the molecular layers, are important inputs to the future design of quantitative AOPs.

Importantly, we showed that the choice of the test system profoundly influences the observed response mechanism, emphasizing the need to consider the interaction between the compound and the biological system when assessing hazards. While this notion is well established in toxicology, this study highlights the importance of the molecular buildup (defined by the genotype and epigenotype) of different test systems in toxicological settings and its impact on the observable toxic effects and mechanisms of action. From a practical study design point-of-view, this means that toxicologists need not only to describe the features of the test materials and model systems^[68] but also need to consider whether biological responses are assessable or not given the choice of the model system.^[2]

In conclusion, our network-based framework represents a significant step forward in understanding the complex relationship between chemical exposures and biological responses. It has the potential to contribute to toxicological assessments by providing a more comprehensive and informative perspective on toxicogenomics data interpretation. One of the limitations of this work is represented by the scarce number of endorsed AOPs, which are for the majority of cases (394 out of 443) still under development or review. However, as the structure of the AOP framework and its molecular annotation will be unchanged and enriched over time, the approach described in this study will still provide a way to systematically map and interpret omics data for regulatory purposes. As the biomedical knowledge encoded in the AOP framework will increase, this approach will be also potentially useful to further characterize pathophysiological mechanisms of disease. Our observations could be seamlessly integrated into future safeand-sustainable-by-design frameworks, offering a powerful tool for researchers and regulators alike.



In conclusion, our network-based approach overcomes some limitations of current omics based approaches, increasing interpretability and extrapolation of the effect of the specific exposure components, while also representing the first example of the potential contribution of systems and network theory to the hazard assessment of ENMs and, potentially, of other advanced materials.

4. Experimental Section

Data and statistical analysis: Datasets selection and preprocessing: The case studies are based on the publicly available datasets GSE148705 and GSE157266. These datasets we previously generated cover a broad panel of ENMs including metal, metal oxide, and carbon-based ENMs with different core (Ag, Au, TiO₂, CuO, nanodiamond and multiwalled carbon nanotubes) and surface chemistries (COOH, NH₂, or PEG). Each material was tested at equipotent doses (EC₁₀) in the two different human cell lines (THP-1 and BEAS-2B), and at a dose of 10 µg/day for four consecutive days in mice. For further details, refer to Gallud et al.^[27] and Kinaret et al.^[28]

For the dose dependent analysis, the dataset GSE127773 was used.^[51] Transcriptomic data were preprocessed with the EUTOPIA software as previously described.^[6,69]

For the GSE157266 dataset, mouse genes have been converted to their orthologs, excluding the genes with no 1:1 relationship between the mouse and human genome. In these datasets, for each ENM exposure, three replicates were present.

The final number of genes shared among all the datasets were 11972. In the preprocessing phase, the MDS highlighted outlier samples both in the BEAS-2B and Mouse datasets. Therefore the samples 'RNA_2_NTND_24_0_1', 'RNA_2_NTQD_24_0_1', 'RNA_2_NTCuO_24_0_1', 'RNA_2_NTND_24_0_3' and 'NS6_1_4' have been discarded from the datasets, respectively.

AOP network and annotation: In order to link the transcriptomic alteration induced by ENMs exposure, we exploited the AOP-gene annotation curated in ref. [14]. The gene sets associated with each KE in the AOP were previously described in ref. [23]. Briefly, a molecular annotation of the events as retrieved from the AOP-wiki was performed using a combination of natural language processing and manual curation. As a result, all the events (MIEs, KEs, and AOs) are associated with lists of genes based on their association with known pathways, phenotypes and gene ontologies. Only taxonomically relevant AOPs were selected, and the presence of redundant events was manually curated. The robustness and functional relevance of the molecular annotation was evaluated with a variety of experiments in ref. [14]. The AOP network used in this framework was obtained by exploiting the connections between KEs present in the AOP-Wiki database (https://aopwiki.org/, downloaded on 26.10.2022). Furthermore, the exact annotation and AOP network used in this study, has been made available (cf. Data and code availability statement).

Computational framework: Framework input: Network inference, differential and dose-dependent analysis

In the current study, input data for the framework were either in the form of a list of genes to be enriched or of a network (Figure 2).

Differential expression analysis was performed using the "limma" R package to compute the gene expression difference between exposed and controls, and correcting the P-value using the Benjamini–Hochberg. The normalized and corrected expression matrix was filtered according to standard thresholds (p-value < 0.05 and $|\log FC| > 0.58$).

In order to infer co-expression networks for each exposure, we first defined gene nodes that would be shared in all the transcriptomic profiles. We selected genes present in all the studied dataset, where for human and mouse a conversion between orthologs has been performed, and mapped them to Ensembl ID. For each individual exposure, we ranked the genes based on both fold change and p-values, and selected the 100 top and bottom deregulated genes. The union resulted in 3061 genes. Due to the limited sample size (3 replicates for each condition), for each pair of genes we performed a Hotelling test (using the R package Hotelling, version 1.0), to identify differences in their multivariate means, which is routinely applied for multivariate analysis of gene expression.^[70,71] To mitigate the risk of false positives resulting from multiple comparisons, we applied a correction to the p-values using the False Discovery Rate (FDR). Edges were ranked according to their p-value from the lowest to the highest. Only edges with a p-value lower than the 5% percentile of the distribution were included in the network.

This was repeated for each triple resulting in 93 binary networks. Information on the p-value was stored as an edge attribute in the exposure specific network (log), while information on the fold change was stored as node attribute and used in further steps of the analysis for filtering. The code to infer network starting from expression data is made available (cf. Data and code availability statement).

In order to provide dose-dependent genes as an input, we followed the procedure outlined in the BMDx tool (version commit February 2022).^[72] In summary, multiple models were fitted and evaluated, with the optimal model chosen based on the Akaike information criterion. Estimates for effective doses (BMD, BMDL, and BMDU) were computed assuming constant variance. The benchmark response was determined using the standard deviation method with a benchmark response factor (BMRF) of 1.349, indicating a minimum 10% deviation from control levels. Genes were considered relevant only if they exhibited lack-of-fit P > 0.01 and had estimated BMD, BMDL, and BMDU values. Any genes with BMD or BMDU values exceeding the highest exposure dose were excluded. Additionally, genes were excluded if the ratio between predicted doses exceeded suggested thresholds (BMD/BMDL > 20, BMDU/BMDL > 20, and BMDU/BMDL > 40).

AOP gene sets enrichment and topological enrichment: In order to obtain enriched KEs for input lists of genes, we performed an enrichment of the differentially expressed genes. Differential expression analysis was performed as described under the *"Framework input: Network inference, differential and dose-dependent analysis"* section. Then, enrichment was performed against KEs associated gene sets using Fisher's exact test implemented in the *enrich* function of the R package "bc3net".^[73] The set of genes associated with each key event were originally derived from Saarimäki et al.^[14] Given the redundancy of some events in the AOP-wiki database, the jaccard index between sets of genes annotated to each KE was used as a similarity measure in order to identify redundant events. In such cases, the events and respective sets of genes were merged and used as clustered KEs. Gene sets of each test system and genes from clustered KEs were merged, and unique genes were used to prepare the background of the enrichment.

In case a network is used as input for the framework (both in the cases in which we analyzed a co-expression network or overrepresented set of edges) a topological enrichment was developed and used to find overrepresented KEs in the AOP network. First, all shortest paths between all nodes on the input network were estimated with the sknetwork.path.get_shortest_path function.^[74] For each gene set (comprising the genes associated with a given KE) the average shortest path between each gene pair on the co-expression network was calculated. 1000 *t*-tests were performed by comparing the average shortest path of the known gene set versus randomly drawn gene sets of the same size. After estimating the average shortest path length of the known gene set divided by the average shortest path length of the random gene set was lower than 1.

Reconstruction of the AOP based mechanism of response: In order to represent the MOA of the exposure as a network of KEs, we created a framework which filters the most relevant KEs, prioritizing "plausibile" MIEs and AOs, and finally maps it back to the AOP network. In this context, plausibility was defined both as a measure of proximity to an enriched KE on the AOP network, as well as the presence of statistically altered genes in the annotated gene set of the event.

First, the enrichment p-values were corrected using the "fdr" method with the p.adjust function of the R package "stats". KEs with more than 1000 genes annotated and adjusted p-value higher than 0.05 were discarded. On the filtered results, only events with a p-value lower than the



SCIENCE NEWS

5-th percentile of the distribution were selected. Finally, KE where genes annotated had a fold change lower than 0.58 were discarded.

Second, probable MIEs and AOs were prioritized based on the expression of the annotated genes, the closeness on the AOP network, and the size of the annotated gene set. Closeness was computed with the shortest.paths function in the "igraph" R package.^[75] Events with no annotated gene differentially expressed were discarded. MIEs and AOs were ranked based on the p-value of the expression of the annotated genes and the distance from the enriched KE. In case more than 20 events were prioritized, 20 was considered as an arbitrary cut-off. All the prioritized events were mapped on the AOP network and plotted using the "igraph" R package.^[75] In order to ensure complete usability and reproducibility of the framework, all the code and custom functions have been made available (cf. Data and code availability statement).

Network comparison: Definition of information measures: In order to quantitatively compare our framework results obtained using either coexpression networks or lists of differentially gene sets as an input, we defined two measures of information. First, we defined information as the capability of reconstructing an AOP based mechanism from a MIE to an AO, which we defined as the "AOP completeness score". In this light, we computed, for each reconstructed network of ENM response, the transitivity, betweenness, shortest paths and diameter using the homonym functions of the igraph R package.^[75] When multiple values were returned (e.g., betweenness and shortest paths), the mean of the values distribution was considered. These measures indicate a more connected network, thus a better possibility of reconstructing a unified mechanism of response between causally linked events. For this reason, we considered a greater (mean) value as an indication that a longer mechanism could be completed. When comparing the results of two mechanisms, we evaluated as more informative the response network having at least three out of four higher values of connectedness.

Second, we considered the possibility of having a more predictable or homogeneous mechanism, which can be a desirable gain as the information content is less uncertain. We defined it as The "Key event connectivity score", namely the ratio between the portion of connected KEs against the unconnected ones (isolated nodes) characterizing the MOA of an exposure.

Computation of a biological system similarity score between network responses: We investigated whether the obtained mechanism of response could be used to derive the similarities between exposures of the same material across biological systems. We defined this measure as the consensus of four difference parameters. First, we evaluated how similarly the two networks cover a portion of the complete AOP network. To determine the coverage of the AOP by each network, we calculated the delta of the ratio between the number of Key Events (KEs) enriched by each network and the total number of KEs in the complete AOP. In essence, this quantifies how much each network contributes to the enrichment of events within the AOP relative to the others.

Second, we considered the overlap between the two response networks, computing a jaccard index of the KEs nodes and edges.

Finally, we computed the ratio of the diameters of the biggest component (computed with the diameter function in the R package "igraph"), as it represents the largest retrievable mechanism in each response network.^[75] The mean of all the normalized scores (between 0 and 1) was used as a similarity measure between the networks.

Network analysis: Clustering strategies for co-expression networks and AOP events networks: Co-expression networks were clustered using the functions developed in VOLTA, a python package for co-expression network analysis.^[76] Briefly, we first computed edge based similarities between each network pair using the get_edge_similarity function. The get_edge_similarity.estimate_similarities_edges from VOLTA is a wrapper function that returns the consensus between the jaccard similarity, jaccard distance, kendall rank coefficient for the edges ranked based on betweenness, hamming distance and SMC similarity. Where applicable the similarities are converted into a distance by taking 1-x. These distances were used as input values for the clustering, which was also performed based on the VOLTA functionalities. In detail, the consensus_clustering function was used based on the three individual clusterings (hierarchical, affinity

In order to cluster the AOP based mechanisms of response, we first computed a similarity score as the mean of the jaccard index between networks nodes and edges. This similarity reflects the number of shared nodes and edges between the response mechanisms. For each pair of networks, the similarity was converted in distance. Finally, the clustering was performed using the "hclust" function from the R package stats, with the ward.D2 method, and results plotted as a dendogram.

Identification of overrepresented structures: In order to highlight edges overrepresented in groups of networks, we exploited the function get_statistical_overrepresented_edges from VOLTA in the pattern matching module. Briefly, the function finds edges that are statistically overrepresented in a cluster (group of networks) based on a hypergeometric function and Benjamin-Hochberg correction. Edges associated with specific biological systems, nanoparticle surface modifications, and toxicity endpoints were retrieved in the same way.

Edge based enrichment analysis: Classical enrichment analysis is used to identify classes of genes that are over-represented in large sets. However, when performing enrichment analysis over the nodes co-expression network, the classical algorithm does not consider if the genes connected by edges in the network are part of the same pathways. In order to consider gene-gene association we performed enrichment of network edges against gene pairs derived from human pathways.

To create the background, we collected data for pathway gene sets in the KEGG databases.^[77] Background edges were created based on the annotation of genes to the same gene set. Input edges were ranked from the most to the least co-deregulated, based on their p-value. In case of edges present in multiple exposures (like overrepresented edges), the individual p-values were combined using the sumlog method. The enrichment analysis was performed by means of a weighted Kolmogorov-Smirnov test, to identify edge sets that are over-represented on top of the rank. The statistical significance of the enrichment analysis was estimated by permutation analysis over 100 random shuffles of the edge sets. The p-values were corrected for multiple comparisons using the false discovery rate (FDR) method and setting 0.05 as the significance threshold. The analysis was performed by using the code originally developed in ref. [78].

Supporting Information

Supporting Information is available from the Wiley Online Library or from the author.

Acknowledgements

This study was supported by the Academy of Finland project UNICAST NANO (322761), by the European Research Council (ERC) programme, Consolidator project ARCHIMEDES (101043848), and the EU FP7 project "NANOSOLUTIONS" under Grant Agreement No. 309329. This study was supported by the EU project "INSIGHT" (101137742) and the "Comp-SafeNano" project (101008099). A.S. and A.F. were supported by the Tampere Institute for Advanced Study.

Conflict of Interest

The authors declare no conflict of interest.

Author Contributions

G.-d.G.: Data curation, Formal analysis, Investigation, Validation, Methodology, Software, Visualization, Writing – original draft, Writing – review & editing. A.S.: Formal analysis, Validation, Methodology, Software, Supervision, Writing – original draft, Writing – review & editing. A.P.: Methodol-

SCIENCE NEWS



www.advancedscience.com

ogy, Software, Validation, Writing – review & editing. M.T.M.: Formal analysis, Writing – original draft, Writing – review & editing. L.A.S.: Investigation, Writing – original draft, Writing – review & editing. M.F.: Methodology, Software, Writing – review & editing. A.F.: Supervision, Writing – review & editing. H.A.: Investigation, Writing – review & editing. B.F.: Investigation, Writing – review & editing. D.G.: Conceptualization, Supervision, Funding acquisition, Project administration, Resources, Writing – original draft, Writing – review & editing.

Data Availability Statement

The data that support the findings of this study are openly available in Zenodo at [https://doi.org/10.5281/zenodo.10359584], reference number [80].

Keywords

adverse outcome pathways, complex systems, engineered nanomaterials, network theory, systems toxicology, toxicogenomics

Received: January 10, 2024 Revised: May 10, 2024 Published online: June 25, 2024

- [1] J. V. Tarazona, M. M. Vega, *Toxicology* **2002**, *181*, 187.
- [2] G. Del Giudice, G. Migliaccio, N. D'Alessandro, L. A. Saarimäki, M. Torres Maia, M. E. Annala, J. Leppänen, L. Möbus, A. Pavel, M. Vaani, A. Vallius, L. Ylä-Outinen, D. Greco, A. Serra, *Front. toxicol.* **2023**, *5*, 1294780.
- [3] E. Wyrzykowska, A. Mikolajczyk, I. Lynch, N. Jeliazkova, N. Kochev, H. Sarimveis, P. Doganis, P. Karatzas, A. Afantitis, G. Melagraki, A. Serra, D. Greco, J. Subbotina, V. Lobaskin, M. A. Bañares, E. Valsami-Jones, K. Jagiello, T. Puzyn, *Nat. Nanotechnol.* **2022**, *17*, 924.
- K. Jagiello, S. Halappanavar, A. Rybińska-Fryca, A. Williams, U. Vogel, T. Puzyn, Small (Germany) 2021, 17, e2003465.
- [5] I. Suciu, D. Pamies, R. Peruzzo, P. H. Wirtz, L. Smirnova, G. Pallocca, C. Hauck, M. T. D. Cronin, J. G. Hengstler, T. Brunner, T. Hartung, I. Amelio, M. Leist, Arch. Toxicol. 2023, 97, 2035.
- [6] G. Del Giudice, A. Serra, L. A. Saarimäki, K. Kotsis, I. Rouse, S. A. Colibaba, K. Jagiello, A. Mikolajczyk, M. Fratello, A. G. Papadiamantis, N. Sanabria, M. E. Annala, J. Morikka, P. A. S. Kinaret, E. Voyiatzis, G. Melagraki, A. Afantitis, K. Tämm, T. Puzyn, D. Greco, *Nat. Nanotechnol.* 2023, *18*, 957.
- [7] S. Kesić, Saudi J. Biol. Sci. 2016, 23, 584.
- [8] S. J. Sturla, A. R. Boobis, R. E. FitzGerald, J. Hoeng, R. J. Kavlock, K. Schirmer, M. Whelan, M. F. Wilks, M. C. Peitsch, *Chem. Res. Toxicol.* 2014, *27*, 314.
- [9] A.-L. Barabási, N. Gulbahce, J. Loscalzo, Nat. Rev. Genet. 2011, 12, 56.
- [10] A. L. Hopkins, Nat. Chem. Biol. 2008, 4, 682.
- [11] V. Fortino, P. A. S. Kinaret, M. Fratello, A. Serra, L. A. Saarimäki, A. Gallud, G. Gupta, G. Vales, M. Correia, O. Rasool, J. Ytterberg, M. Monopoli, T. Skoog, P. Ritchie, S. Moya, S. Vázquez-Campos, R. Handy, R. Grafström, L. Tran, D. Greco, *Nat. Commun.* **2022**, *13*, 3798.
- [12] A. Sewer, M. Talikka, F. Martin, J. Hoeng, M. C. Peitsch, in (Ed: I. Y. Abdurakhmonov), *Bioinformatics in the era of post genomics and big data. InTech*, **2018**.
- [13] G. Scala, A. Serra, V. S. Marwah, L. A. Saarimäki, D. Greco, BMC Bioinformatics 2019, 20, 79.

- [14] L. A. Saarimäki, J. Morikka, A. Pavel, S. Korpilähde, G. Del Giudice, A. Federico, M. Fratello, A. Serra, D. Greco, Adv. Sci. 2023, 10, e2203984.
- [15] P. Kinaret, V. Marwah, V. Fortino, M. Ilves, H. Wolff, L. Ruokolainen, P. Auvinen, K. Savolainen, H. Alenius, D. Greco, ACS Nano 2017, 11, 3786.
- [16] G. T. Ankley, R. S. Bennett, R. J. Erickson, D. J. Hoff, M. W. Hornung, R. D. Johnson, D. R. Mount, J. W. Nichols, C. L. Russom, P. K. Schmieder, J. A. Serrrano, J. E. Tietge, D. L. Villeneuve, *Environ. Toxi*col. Chem. **2010**, 29, 730.
- [17] L. Margiotta-Casaluci, S. F. Owen, B. Huerta, S. Rodríguez-Mozaz, S. Kugathas, D. Barceló, M. Rand-Weaver, J. P. Sumpter, *Sci. Rep.* 2016, 6, 21978.
- [18] C. A. LaLone, D. L. Villeneuve, J. Wu-Smart, R. Y. Milsk, K. Sappington, K. V. Garber, J. Housenger, G. T. Ankley, *Sci. Total Environ.* 2017, 584, 751.
- [19] M. M. Angrish, C. A. McQueen, E. Cohen-Hubal, M. Bruno, Y. Ge, B. N. Chorley, *Toxicol. Sci.* 2017, 159, 159.
- [20] D. Knapen, M. M. Angrish, M. C. Fortin, I. Katsiadaki, M. Leonard, L. Margiotta-Casaluci, S. Munn, J. M. O'Brien, N. Pollesch, L. C. Smith, X. Zhang, D. L. Villeneuve, *Environ. Toxicol. Chem.* **2018**, *37*, 1723.
- [21] M. M. Angrish, J. P. Kaiser, C. A. McQueen, B. N. Chorley, *Toxicol. Sci.* 2016, 150, 261.
- [22] D. Knapen, L. Vergauwen, D. L. Villeneuve, G. T. Ankley, *Reprod. Toxicol.* 2015, 56, 52.
- [23] L. A. Saarimäki, M. Fratello, A. Pavel, S. Korpilähde, J. Leppänen, A. Serra, D. Greco, *Sci. Data* 2023, 10, 409.
- [24] J. S. Fleck, J. G. Camp, B. Treutlein, Science 2023, 381, 733.
- [25] R. Q. Figueiredo, S. D. Del Ser, T. Raschka, M. Hofmann-Apitius, A. T. Kodamullil, S. Mubeen, D. Domingo-Fernández, *BMC Bioinformatics* 2022, 23, 231.
- [26] B. Alexander-Dann, L. L. Pruteanu, E. Oerton, N. Sharma, I. Berindan-Neagoe, D. Módos, A. Bender, *Mol. Omics* 2018, 14, 218.
- [27] A. Gallud, M. Delaval, P. Kinaret, V. S. Marwah, V. Fortino, J. Ytterberg, R. Zubarev, T. Skoog, J. Kere, M. Correia, K. Loeschner, Z. Al-Ahmady, K. Kostarelos, J. Ruiz, D. Astruc, M. Monopoli, R. Handy, S. Moya, K. Savolainen, B. Fadeel, *Adv. Sci.* **2020**, *7*, 2002221.
- [28] P. A. S. Kinaret, J. Ndika, M. Ilves, H. Wolff, G. Vales, H. Norppa, K. Savolainen, T. Skoog, J. Kere, S. Moya, R. D. Handy, P. Karisola, B. Fadeel, D. Greco, H. Alenius, *Adv. Sci* **2021**, *8*, 2004588.
- [29] A. A. Sallam, M. M. Ahmed, M. A. El-Magd, A. Magdy, H. I. Ghamry, M. Y. Alshahrani, M. F. Abou El-Fotoh, *Molecules* 2022, 27, 2117.
- [30] X. He, S.-H. Young, D. Schwegler-Berry, W. P. Chisholm, J. E. Fernback, Q. Ma, Chem. Res. Toxicol. 2011, 24, 2237.
- [31] B. Hindman, Q. Ma, Arch. Toxicol. 2019, 93, 887.
- [32] P. Møller, D. V. Christophersen, D. M. Jensen, A. Kermanizadeh, M. Roursgaard, N. R. Jacobsen, J. G. Hemmingsen, P. H. Danielsen, Y. Cao, K. Jantzen, H. Klingberg, L.-G. Hersoug, S. Loft, *Arch. Toxicol.* 2014, *88*, 1939.
- [33] L. A. Saarimäki, P. A. S. Kinaret, G. Scala, G. del Giudice, A. Federico, A. Serra, D. Greco, *NanoImpact* 2020, 20, 100274.
- [34] P. A. S. Kinaret, G. Del Giudice, D. Greco, Nano Today 2020, 35, 100945.
- [35] M. S. P. Boyles, L. Young, D. M. Brown, L. MacCalman, H. Cowie, A. Moisala, F. Smail, P. J. W. Smith, L. Proudfoot, A. H. Windle, V. Stone, *Toxicol. In Vitro* 2015, *29*, 1513.
- [36] F. A. Murphy, A. Schinwald, C. A. Poland, K. Donaldson, Part. Fibre Toxicol. 2012, 9, 8.
- [37] M. Alswady-Hoff, J. S. Erdem, M. Aleksandersen, K. H. Anmarkrud, Ø. Skare, F.-C. Lin, Simensen, Y. J. Arnoldussen, V. Skaug, E. Ropstad, S. Zienolddiny-Narui, *Int. J. Mol. Sci.* 2022, 23, 6005.
- [38] R. L. Medcalf, J. thromb. Haemost. 2007, 5, 132.
- [39] A. N. Ilinskaya, M. A. Dobrovolskaia, Nanomedicine 2013, 8, 969.

ADVANCED SCIENCE NEWS

www.advancedsciencenews.com



www.advancedscience.com

- [40] M. Schuliga, C. Grainge, G. Westall, D. Knight, Int. J. Biochem. Cell Biol. 2018, 97, 108.
- [41] A. Malur, A. Mohan, R. A. Barrington, N. Leffler, A. Malur, B. Muller-Borer, G. Murray, K. Kew, C. Zhou, J. Russell, J. L. Jones, C. J. Wingard, B. P. Barna, M. J. Thomassen, Am. J. Respir. Cell Mol. Biol. 2019, 61, 198.
- [42] A. A. Shvedova, N. Yanamala, E. R. Kisin, T. O. Khailullin, M. E. Birch, L. M. Fatkhutdinova, *PLoS One* **2016**, *11*, e0150628.
- [43] E. Moschini, G. Colombo, G. Chirico, G. Capitani, I. Dalle-Donne, P. Mantecca, Sci. Rep. 2023, 13, 2326.
- [44] H. L. Karlsson, P. Cronholm, J. Gustafsson, L. Möller, Chem. Res. Toxicol. 2008, 21, 1726.
- [45] H. L. Karlsson, J. Gustafsson, P. Cronholm, L. Möller, *Toxicol. Lett.* 2009, 188, 112.
- [46] P. Abeyrathna, Y. Su, Vasc. Pharmacol. 2015, 74, 38.
- [47] M. Younes, G. Aquilina, L. Castle, K.-H. Engel, P. Fowler, M. J. Frutos Fernandez, P. Fürst, U. Gundert-Remy, R. Gürtler, T. Husøy, M. Manco, W. Mennes, P. Moldeus, S. Passamonti, R. Shah, I. Waalkens-Berendsen, D. Wölfle, E. Corsini, M. Wright, *EFSA J.* **2021**, *19*, e06585.
- [48] D. Acharya, K. M. Singha, P. Pandey, B. Mohanta, J. Rajkumari, L. P. Singha, *Sci. Rep.* **2018**, *8*, 201.
- [49] Y. Zhao, Y. Wang, F. Ran, Y. Cui, C. Liu, Q. Zhao, Y. Gao, D. Wang, S. Wang, Sci. Rep. 2017, 7, 4131.
- [50] M. J. Mitchell, M. M. Billingsley, R. M. Haley, M. E. Wechsler, N. A. Peppas, R. Langer, *Nat. Rev. Drug Discov.* **2021**, *20*, 101.
- [51] I. Kooter, M. Ilves, M. Gröllers-Mulderij, E. Duistermaat, P. C. Tromp, F. Kuper, P. Kinaret, K. Savolainen, D. Greco, P. Karisola, J. Ndika, H. Alenius, ACS Nano 2019, 13, 6932.
- [52] S. van Dam, U. Võsa, A. van der Graaf, L. Franke, J. P. de Magalhães, Brief. Bioinform. 2018, 19, 575.
- [53] L. Ding, J. Yang, C. Zhang, X. Zhang, P. Gao, Front. Med. 2021, 8, 616200.
- [54] D. G. Helou, B. Noël, F. Gaudin, H. Groux, Z. El Ali, M. Pallardy, S. Chollet-Martin, S. Kerdine-Römer, J. Immunol. 2019, 202, 2189.
- [55] M. Laforge, C. Elbim, C. Frère, M. Hémadi, C. Massaad, P. Nuss, J.-J. Benoliel, C. Becker, Nat. Rev. Immunol. 2020, 20, 515.
- [56] S. Halappanavar, L. Rahman, J. Nikota, S. S. Poulsen, Y. Ding, P. Jackson, H. Wallin, O. Schmid, U. Vogel, A. Williams, *NanoImpact* 2019, 14, 100158.
- [57] S. E. Mutsaers, Respirology 2002, 7, 171.
- [58] A. Schinwald, K. Donaldson, Part. Fibre Toxicol. 2012, 9, 34.
- [59] A. Nel, T. Xia, L. Mädler, N. Li, *Science* **2006**, *311*, 622.
- [60] M. C. Lo Giudice, L. M. Herda, E. Polo, K. A. Dawson, Nat. Commun. 2016, 7, 13475.
- [61] V. Stone, S. Gottardo, E. A. J. Bleeker, H. Braakhuis, S. Dekkers, T. Fernandes, A. Haase, N. Hunt, D. Hristozov, P. Jantunen, N.

Jeliazkova, H. Johnston, L. Lamon, F. Murphy, K. Rasmussen, H. Rauscher, A. S. Jiménez, C. Svendsen, D. Spurgeon, G. A. Oomen, *Nano Today* **2020**, *35*, 100941.

- [62] M. Yoon, A. Efremenko, B. J. Blaauboer, H. J. Clewell, Toxicol. In Vitro 2014, 28, 164.
- [63] K. Bhattacharya, G. Kiliç, P. M. Costa, B. Fadeel, Nanotoxicology 2017, 11, 809.
- [64] I. Hansjosten, J. Rapp, L. Reiner, R. Vatter, S. Fritsch-Decker, R. Peravali, T. Palosaari, E. Joossens, K. Gerloff, P. Macko, M. Whelan, D. Gilliland, I. Ojea-Jimenez, M. P. Monopoli, L. Rocks, D. Garry, K. Dawson, P. J. F. Röttgermann, A. Murschhauser, C. Weiss, *Arch. Toxicol.* **2018**, *92*, 633.
- [65] OECD. Important Issues on Risk Assessment of Manufactured NanomaterialsSeries on the Safety of Manufactured Nanomaterials No. 103, OECD Environment, Health and Safety Publications, 2022.
- [66] S. Brescia, C. Alexander-White, H. Li, A. Cayley, *Toxicol. Res.* 2023, 12, 1.
- [67] L.-A. Clerbaux, J. Filipovska, P. Nymark, V. Chauhan, K. Sewald, M. Alb, M. Sachana, A. Beronius, M.-J. Amorim, C. Wittwehr, ALTEX -Alternatives to animal experimentation 2024, 41, 233.
- [68] M. Faria, M. Björnmalm, K. J. Thurecht, S. J. Kent, R. G. Parton, M. Kavallaris, A. P. R. Johnston, J. J. Gooding, S. R. Corrie, B. J. Boyd, P. Thordarson, A. K. Whittaker, M. M. Stevens, C. A. Prestidge, C. J. H. Porter, W. J. Parak, T. P. Davis, E. J. Crampin, F. Caruso, *Nat. Nanotechnol.* **2018**, *13*, 777.
- [69] V. S. Marwah, G. Scala, P. A. S. Kinaret, A. Serra, H. Alenius, V. Fortino, D. Greco, Source Code Biol. Med. 2019, 14, 1.
- [70] J. Curran, T. Hersh 2021. Hotelling: Hotelling's T² Test and Variants (1.0) [Computer software], CRAN, https://CRAN.R-project.org/ package=Hotelling.
- [71] Y. Lu, P.-Y. Liu, P. Xiao, H.-W. Deng, Bioinformatics 2005, 21, 3105.
- [72] A. Serra, L. A. Saarimäki, M. Fratello, V. S. Marwah, D. Greco, *Bioinformatics* 2020, 36, 2932.
- [73] R. de Matos Simoes, F. Emmert-Streib, PLoS One 2012, 7, e33624.
- [74] T. Bonald, N. de Lara, Q. Lutz, B. Charpentier, Journal of Machine Learning Research: JMLR 2020, 21, 1.
- [75] G. Csardi, T. Nepusz, InterJournal, Complex Systems 2006, 1695, 1.
- [76] A. Pavel, A. Federico, G. Del Giudice, A. Serra, D. Greco, Bioinformatics 2021, 37, 4587.
- [77] M. Kanehisa, M. Furumichi, Y. Sato, M. Ishiguro-Watanabe, M. Tanabe, Nucleic Acids Res. 2021, 49, D545.
- [78] J. Han, X. Shi, Y. Zhang, Y. Xu, Y. Jiang, C. Zhang, L. Feng, H. Yang, D. Shang, Z. Sun, F. Su, C. Li, X. Li, *Sci. Rep.* **2015**, *5*, 13044.
- [79] A. Pavel, G. del Giudice, A. Federico, A. Di Lieto, P. A. S. Kinaret, A. Serra, D. Greco, *Brief. Bioinform.* 2021, 22, 1430.
- [80] G. del Giudice, D. Greco, In Advanced Science. Zenodo (2023), https: //doi.org/10.5281/zenodo.10359584.