# scientific reports



# **Risk-based bridge life cycle OPEN cost and environmental impact assessment considering climate change effects**

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**To enhance sustainability and resilience against climate change in infrastructure, a quantitative evaluation of both environmental impact and cost is important within a life cycle framework. Climate change effects can lead performance deterioration in bridge components during their operational phase, highlighting the necessity for a risk-based evaluation process aligned with maintenance strategies. This study employs a two-phase life cycle assessments (LCA) framework. First, risk assessments are conducted to evaluate the impact of climate change on steel plate girder bridges and prestressed concrete (PSC) girder bridges under identical structural conditions. The reduction in flexural strength of steel plate girders and PSC girders due to changes in environmental variables such as temperature and relative humidity, induced by various climate change scenarios, was evaluated analytically. Subsequently, life cycle environmental impact and cost assessments were performed, including maintenance outcomes derived from risk assessments. The findings revealed that the environmental impact and cost could increase by approximately 12.4% when climate change is considered, compared to scenarios where it is not taken into account. Sensitivity analyses were performed to identify the key factors influencing environmental impact and cost. The analysis determined that the frequency of preventive maintenance, the recycling rate, and environmental cost coefficient weight in the life cycle assessment significantly affected the results.**

**Keywords** Life cycle assessment, Environmental impact, Life cycle cost, Climate change, Risk assessment, Bridge deterioration

In the context of global climate change, the relationship between the infrastructure industry and its environmental impact has been a focal point of discussions, addressing key concerns such as energy consumption and greenhouse gas emissions<sup>1</sup>. The crux of mitigating the environmental impact within the infrastructure industry lies in the engineer's commitment to sustainable decisions in conceptual design and maintenance strategies. Life cycle assessment (LCA) serves as a robust method to evaluate the environmental impact of infrastructure over its entire lifespan<sup>[2](#page-15-1)</sup>. Through LCA, infrastructure designers and decision-makers can assess the environmental and economic feasibility of their projects in the concept design phase<sup>[3](#page-15-2)</sup>, fostering more sustainable approaches to maintenance and material selection.

Infrastructures inevitably experience structural performance deterioration over time due to their long design lives, and this deterioration is significantly influenced by both the environmental conditions encountered and ongoing climate changes<sup>[4](#page-15-3)</sup>. Climate change further complicates the life cycle management of infrastructure, as structures are subject to accelerated deterioration under evolving environmental conditions<sup>[5](#page-15-4)</sup>. High temperatures, increased relative humidity, and intensified chloride exposure can significantly degrade construction materials, posing long-term risks to infrastructure performance<sup>[6](#page-15-5)[,7](#page-15-6)</sup>. Given the longevity of infrastructure, these environmental variables heighten the need for a robust maintenance plan integrating climate-adaptive management strategies<sup>[8](#page-15-7)</sup>. Without maintenance plans considering climate change effects, infrastructure could experience reduced structural integrity, leading to increased repairs that, paradoxically, generate further emissions and exacerbate climate impacts.

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This study explores the influence of climate change and material selection on the environmental sustainability of bridge infrastructure. The life cycle environmental impact and cost of two bridge design alternatives are quantified utilizing different primary construction materials, namely, a prestressed concrete (PSC) bridge and a steel plate bridge, under identical site conditions. The PSC girder bridge represents a highway bridge currently under construction in South Korea, while the steel plate girder bridge serves as an alternative design that meets the same code requirements. Corrosion was a mainly focus of the analysis, given its strong correlation with climate change<sup>9,10</sup> and its common role as a major degradation factor for both bridge types in this study. A two-phase LCA framework was applied: First, a time-variant risk assessment was conducted, accounting for the deterioration of structural components caused by corrosion considering climate change effects. The influence of corrosive environments on life cycle management was explored under four distinct climate change scenarios. Second, integrating maintenance action data from the risk assessment into the LCA allowed for the quantification of environmental impacts and economic feasibility. The LCA results provide insights into the significance of integrated maintenance strategies, material recycling, and the potential future environmental impacts. Additionally, the proposed bridge LCA framework can support infrastructure designers to extend analysis to other limit states, such as shear and fatigue, and additional performance degradation models.

## **Literature review**

Extensive studies have compared the LCA of various building and infrastructure designs. Gervásio and da Silva[11](#page-16-0) were early researchers who applied the LCA approach to compare the environmental impact and life cycle cost (LCC) of two alternative solutions: steel composite structures and concrete structures. The comparative analysis revealed that while the concrete solution could be cheaper, the environmental life-cycle analysis favored the steel solution, which achieved a slightly better global score. Du et al.<sup>12</sup> provided a guide on using LCA to support bridge procurement decisions by comparing the environmental impact, cumulative energy demand, and atmospheric emission for five different bridge designs. Their sensitivity analysis on monetary weighting systems and input uncertainties showed significant volatility in the outcomes. The results emphasize that clear specifications of key variables are necessary before selecting the most LCA-feasible proposal. Nouri et al.<sup>13</sup> found that the design of structures that are more resistant to environmental load can increase initial costs compared to those that are not, but contribute to reducing structural damage, reducing long-term damage costs and reducing total life cycle costs. Desai and Bheemrao<sup>14</sup> raised questions about the criteria for material selection in the construction industry and highlighted the importance of LCA-based sustainable construction material selection by comparing the environmental impact of two-story concrete residential buildings using steel or timber reinforcement. Katebi et al[.15](#page-16-4) performed LCA for roof systems using eight different types of materials to identify a design that minimizes the environmental impact and pinpoint the main sources of pollution. They revealed that the suitability of environmental impact and economic feasibility does not always accompany them. Huang et al.<sup>[16](#page-16-5)</sup> used LCA to analyze the comprehensive environmental performance of expressway construction and developed an environmental price model. Their study revealed that material selection is a crucial factor in infrastructure construction.

Moreover, numerous studies have reported that changes in environmental factors (e.g., temperature, relative humidity, chloride, etc.) due to the climate crisis can accelerate the corrosive deterioration of materials in infrastructure. Lin and Wang<sup>17</sup> confirmed through accelerated corrosion experiments on carbon steel that a temperature change of 10 ℃ can lead to a weight loss of about 32% due to corrosion. Konovalova[18](#page-16-7) found that a temperature increase of 10 ℃, based on accelerated corrosion experiments on iron-carbon alloys, increases the corrosion rate by 2.3 times. Stewart et al.<sup>[6](#page-15-5)</sup> analyzed that, based on  $CO_2$  emission scenarios, the loss of reinforcing steel rebar in concrete structures due to chloride-induced corrosion could be as high as 9.5%. Xie et al.<sup>19</sup> found that, under global warming scenarios, the amount of chloride ions on the surface of reinforcing steel rebar could increase by 6–15%. Consequently, attention must be given to maintenance to ensure the preservation of infrastructure performance throughout its lifetime during the climate crisis. Increased maintenance actions following this attention could lead to higher emissions, potentially exacerbating climate change. Qiao et al.<sup>20</sup> emphasized the need for maintenance strategies to be transformed with climate change for optimization of LCC. Cadenazzi et al.[21](#page-16-10) and Han et al.[22](#page-16-11) reported that notable differences can emerge in the environmental impact and cost throughout the life cycle if technical actions such as member replacement or reinforcement are integrated into the maintenance strategies. However, thorough investigations into maintenance strategies that respond to the effects of climate change are still lacking.

This study addresses a crucial gap in sustainable infrastructure management by examining the combined impact of climate change and material selection on the life cycle environmental impact of bridge designs. While previous studies have assessed bridge life cycles, they often overlook how climate-adaptive maintenance strategies and material choices intersect to influence long-term environmental outcomes. Employing a proposed two-phase LCA framework, this research quantifies the effects of climate-driven corrosion on material degradation and integrates these into a cost-benefit analysis of maintenance actions. This approach offers a more comprehensive view of how adaptive management and optimized material selection can improve environmental impacts and economic feasibility under varying climate scenarios.

# **Risk assessment considering climate change effects Limit state function**

In this study, the risk-based LCA framework was presented with a focus on the flexural capacity for the limit state Strength I according to the AASHTO LRFD bridge design specification<sup>[23](#page-16-12)</sup>. The bending failure of the girder, caused by the combined effects of dead load and traffic load, is the most representative limit state considered in bridge design and risk assessment. Numerous previous studies on bridge reliability and risk assessment <span id="page-2-0"></span>have primarily focused on evaluating the bending strength limit state of the girder $24,25$  $24,25$ . Risk assessments were conducted on the flexural strength of the maximum positive moment section under the AASHTO design truck load, extending up to the target service life of 100 years. The limit state function for the flexural capacity of the bridge superstructure, varying with time *t*, is defined as follows:

$$
g = \varphi \cdot M_n(t) - \left(\gamma_p \cdot M_{DL} + 1.75 \cdot M_{LL}\right) \tag{1}
$$

where,  $\varphi$  is the resistance factor.  $M_n(t)$  is the time-variant nominal flexural resistance.  $\gamma_p$  is the load factor for a permanent load. *MDL* represents the nominal bending moments due to permanent load. *MLL* is the nominal bending moments due to live load. The failure probability *Pf* defined as the probability that *g* will be less than zero and was calculated using Monte Carlo simulations (MCS). The number of MCS samples was set to 500,000, as determined by a convergence check.

#### **Deterioration modeling**

As deterioration progresses over time, *M<sup>n</sup>* (*t*) in Eq. ([1\)](#page-2-0) gradually decreases, and the failure probability of the bending member increases. This section explains the mechanisms that contribute to the declining strength of PSC girder bridges and steel plate girder bridges.

#### *PSC girder bridge*

The primary sources of deterioration in PSC girder bridges are (1) the cross-section loss of reinforcement steel (including steel rebar and prestressing steel) due to corrosion, and (2) the long-term stress loss of prestressing steel. In practice, most composite PSC girders have their neutral axes within the slab when subjected to flexural limits. Consequently, the *Mn,psc* (*t*), nominal flexural resistance of PSC girder for time *t*, considering area loss of reinforcement steel due to corrosion and long-term stress loss of prestressing steel can be defined as Eq. ([2](#page-2-1)), which is a modification of the design equation from  $AASHTO<sup>23</sup>$  $AASHTO<sup>23</sup>$  $AASHTO<sup>23</sup>$ .

<span id="page-2-1"></span>
$$
M_{n, psc}(t) = \sum_{i} A_{ps,i}(t) \cdot f_{ps,i}(t) \cdot \left(d_{ps,i} - \frac{a}{2}\right) + \sum_{j} A_{s,j}(t) \cdot f_s \cdot \left(d_{s,j} - \frac{a}{2}\right)
$$
(2)

where,  $A_{ps,i}$  (*t*),  $f_{ps,i}$  (*t*) and  $d_{ps,i}$  are area, tensile stress and distance from extreme compression fiber to the centroid of *i*-th prestressing steel, respectively.  $A_{s,j}$  (*t*) and  $d_{s,j}$  are area and distance from extreme compression fiber to the centroid of *j*-th steel rebar, respectively. *fs* is tensile stress of steel rebar and *a* is the depth of the equivalent stress block. *fps,i* (*t*) with long-term stress loss of prestressing steel can be calculated according to AASHTO LRFD bridge design specification<sup>23</sup>.

The most severe corrosion category of reinforcement steel is primarily caused by chloride penetration $26$ . Therefore, reinforcement steel under chloride-induced corrosion is accounted for herein. Corrosion of prestressing steel and steel rebars under the concrete cover occurs when the chloride concentration at the steel surface exceeds a threshold of chloride concentration,  $C_{cr}$ . The chloride concentration in depth *x* at *t*,  $C(x, t)$ , is given empirically by Fick's second law of diffusion as follows $27$ :

$$
C(x,t) = C_0 \cdot \left[1 - \text{erf}\left(\frac{x}{2 \cdot \sqrt{D_c \cdot t}}\right)\right]
$$
 (3)

<span id="page-2-2"></span>where  $C_0$  is the chloride ion concentration on the surface of concrete. erf  $(\cdot)$  is the Gauss error function, and *Dc* is the diffusion coefficient of chloride ion for inland or marine environment. The chloride transfer in concrete is closely related to environmental conditions and ambient temperature, and these factors can be accelerated by climate change<sup>19</sup>. The corrected chloride concentration, accounting for temperature and relative humidity, is proposed in de Medeiros-Junior et al.<sup>28</sup> as shown in Eq. [\(4\)](#page-2-2). This was achieved by applying the modified diffusion coefficient proposed in Saetta et al.<sup>29</sup>.

$$
C(x,t) = C_0 \cdot \left[1 - \text{erf}\left(\frac{x}{2 \cdot \sqrt{f_T(T) \cdot f_t(t) \cdot f_h(RH) \cdot D_c \cdot t}}\right)\right]
$$
(4)

<span id="page-2-5"></span><span id="page-2-4"></span><span id="page-2-3"></span>where,  $f_T(T)$  is the correction factor representing the influence of temperature (*T*) according to Arrhenius' law, as shown in Eq. [\(5\)](#page-2-3). *f<sup>t</sup>* (*t*) is the correction factor representing the influence of the equivalent maturation time, as shown in Eq. [\(6\)](#page-2-4). *f<sup>h</sup>* (*RH*) is the correction factor representing the influence of relative humidity (*RH*), as shown in Eq. [\(7](#page-2-5)).

$$
f_T(T) = \exp\left[\frac{E_{a,dif}}{R_g} \cdot \left(\frac{1}{T_0} - \frac{1}{T}\right)\right]
$$
\n(5)

$$
f_t(t) = \zeta + (1 - \zeta) \cdot \left(\frac{28}{t_e}\right)^{1/2} \tag{6}
$$

$$
f_h(RH) = \left[\lambda + (1 - \lambda) \cdot \frac{(1 - RH)^4}{(1 - RH_c)^4}\right]^{-1}
$$
 (7)

where,  $T_0$  is the reference temperature (=296 K).  $E_{a,dif}$  is the activation energy of diffusion (=44.6 kJ/mol)<sup>[30,](#page-16-19)[31](#page-16-20)</sup>.  $R_q$  is the universal gas constant gas constant (=8.314 J/mol·K).  $\zeta$  is the coefficient defined as the ratio between the diffusion coefficient for  $t \to \infty$ , which varies from 0 to 1. According to Oh and Jang<sup>32</sup>, the effect of the time dependence of chloride diffusion for well-cured concrete is not significant. Therefore, constant value 1 is used for *ζ* herein. *te* is the actual time of exposure to chloride (days). *RHc* is the humidity at which *Dc* drops halfway between its maximum and minimum value  $(=0.75)^{33}$ .  $f_T(T)$  increases exponentially as *T* increases.  $f_h(RH)$ increases with *RH* and converges to 1 when *RH* becomes 1. The mean, coefficient of variation (COV), and distribution type of parameters related to chloride-induced corrosion are shown in Table [1.](#page-3-0)

<span id="page-3-1"></span>It is assumed that corrosion of steel reinforcement starts when the  $C(x, t)$  at the level of each steel reinforcement reaches a *Ccr*. *A<sup>s</sup>* (*t*) was determined using the non-uniform loss model of steel reinforcement affected by pitting corrosion<sup>34</sup>. This model is also applicable to the calculation of  $A_{ps}$  ( $t$ )<sup>24</sup>. The corrosion pit depth  $p(t)$  and remaining area of the corroded steel reinforcement can be estimated as follows:

$$
p(t) = 0.0116 \cdot (t - t_i) \cdot i_{corr} \cdot R \tag{8}
$$

$$
i_{corr} = \frac{0.378 \cdot (1 - w/c)}{x}
$$
 (9)

$$
A(t) = \begin{cases} \frac{\pi \cdot D_0^2}{4} - A_1 - A_2, & p(t) \le \frac{\sqrt{2}}{2} \cdot D_0 \\ A_1 - A_2, & \frac{\sqrt{2}}{2} \cdot D_0 < p(t) \le D_0 \\ 0, & p(t) > D_0 \end{cases}
$$
(10)

$$
A_1 = 0.5 \cdot \left[ 2 \cdot \arcsin\left(\frac{a}{D_0}\right) \cdot \left(\frac{D_0}{2}\right)^2 - a \cdot \left| \frac{D_0}{2} - \frac{p(t)^2}{D_0} \right| \right] \tag{11}
$$

$$
A_2 = 0.5 \cdot \left[ 2 \cdot \arcsin\left(\frac{a}{2 \cdot p(t)}\right) \cdot p(t)^2 - a \cdot \frac{p(t)^2}{D_0} \right] \tag{12}
$$

$$
a = 2 \cdot p(t) \sqrt{1 - \left\{ \frac{p(t)}{D_0} \right\}^2}
$$
\n(13)

where,  $t_i$  is time of corrosion initiation when  $C(x, t)$  exceeds  $C_{cr}$ ,  $i_{corr}$  is corrosion current density estimated as proposed by Vu and Stewart<sup>35</sup>, as shown in Eq. ([9](#page-3-1)). *R* is coefficient representing ratio between maximum and average corrosion penetration in the range of 4 to 8.  $w/c$  is a water/cement ratio (=0.5).  $D_0$  is initial diameter of reinforcement steel.

#### *Steel plate girder bridge*

The time-variant nominal flexural resistance of steel plate girder bridge *Mn,sp* (*t*) regarding strength degradation is defined based on  $AASHTO<sup>23</sup>$  $AASHTO<sup>23</sup>$  $AASHTO<sup>23</sup>$  as follows:

$$
M_{n,sp}(t) = \begin{cases} M_p(t) & D_p(t) \leq 0.1 \cdot D_t \\ M_p(t) \cdot \left[ 1.07 - 0.7 \cdot \frac{D_p(t)}{D_t} \right] & Otherwise \end{cases}
$$
(14)

where,  $M_p(t)$  is a time-variant plastic moment,  $D_p(t)$  is a distance from the top of the concrete deck to the neutral axis of the composite section at the plastic moment calculated every time step, and *Dt* is total depth of the composite section. The detail equations of *M<sup>p</sup>* (*t*) according to the location of the neutral axis of the composite section are included in AASHTO[23](#page-16-12). In this paper, it is assumed that *M<sup>p</sup>* (*t*) gradually decreases with time due to cross-sectional loss due to corrosion. The thickness of the web and bottom plate, which are the main corrosion locations of the steel plate girders<sup>36</sup>, decreases with corrosion progress. In addition, to account for painting loss during the corrosion process, the coating degradation model for steel plates proposed by Kere and Huang $37$  was applied. By using this model in conjunction with the corrosion depth over time, the equivalent corrosion depth resulting from painting loss can be calculated. Time-variant corrosion depth of steel member,  $d_p(t_{corr})$ , can be modeled by power law relationship as follows<sup>38</sup>:

<span id="page-3-0"></span>

Parameter	Mean	<b>COV</b>	<b>Distribution type</b>	Reference
$C_{cr}$	3.35 $\text{kg/m}^3$	0.375	Truncated normal (at $0.35 \text{ kg/m}^3$ )	Val and Stewart <sup>63</sup>
$C_0$ (marine)	$2.95 \text{ kg/m}^3$	0.7	Lognormal	Val and Stewart <sup>63</sup>
$C_0$ (inland)	$1.15 \text{ kg/m}^3$	0.5	Lognormal	Val and Stewart <sup>63</sup>
$D_c$ (marine)	$212.9$ mm <sup>2</sup> /year	1.62	Lognormal	Soliman et al. <sup>64</sup>
$D_c$ (inland)	76.9 mm <sup>2</sup> /year	1.22	Lognormal	Soliman et al. <sup>64</sup>

**Table 1**. Parameters associated with the chloride-induced corrosion.

$$
d_p(t_{corr}) = p \cdot (t_{corr})^q \tag{15}
$$

where, *tcorr* is time after corrosion initiation, *p* and *q* are corrosion parameters determined from the regression analysis of experimental data, which have three corrosive environmental conditions, namely, rural, urban, and marine environments<sup>38</sup>.

The actual environmental corrosive conditions for field-exposed materials are dynamic and influenced by multiple environmental factors. According to Cai et al.[9](#page-15-8) , *RH* and *T* were identified as the most influential factors for steel corrosion. Therefore, this study focuses on correcting the effects of *T* and *RH* on steel corrosion. A revised corrosion model that accounts for the integrated effects of the dynamic corrosive environment is presented as follows<sup>[9](#page-15-8)</sup>:

$$
d_p(t_{corr}) = R(T, RH) \cdot R(S) \cdot R(Cl) \cdot p \cdot (t_{corr})^q \tag{16}
$$

$$
R(T, RH) = \int_{T_{min}}^{T_{max}} \int_{RH_{min}}^{RH_{max}} r(T, RH) \cdot h(RH|T) \cdot dRH \cdot f(T) \cdot dT \tag{17}
$$

$$
r(T, RH) = \exp\left[\alpha \cdot (RH - RH_{ref})\right] \cdot \exp\left[\frac{E_{a,cor}}{R_g} \cdot \left(\frac{1}{T_{ref}} - \frac{1}{T}\right)\right]
$$
(18)

where,  $r(T, RH)$  is accelerating factor for *T* and *RH*, which increases exponentially as *T* and *RH* increase<sup>[39](#page-16-28)</sup>. *f* (*T*) is the probability density functions (PDF) of *T*. *h* ( $RH|T|$ ) is the conditional PDF of  $RH$  given *T*.  $\alpha$  is constant (=0.1141) estimated from experimental data of Huang et al.<sup>40</sup>.  $RH_{ref}$  is a reference relative humidity, which is also the average relative humidity. *Ea,cor* is the activation energy of corrosion (=50.4 kJ/mol) according to Konovalova[18.](#page-16-7) *Tref* is a reference temperature and also the average temperature.

#### **Climate change scenarios**

Four representative scenarios were selected based on the AR6 Synthesis Report<sup>41</sup>: scenario 0 (Without climate change); scenario 1 (Low/SSP1-2.6/RCP2.6); scenario 2 (Intermediate/SSP2-4.5/RCP4.5); scenario 3 (Very high/SSP5-8.5/RCP8.5). scenarios 1 to 3 predict climate change based on different trajectories of greenhouse gas emissions: achieving net zero emissions by 2070, maintaining the current level by 2050, and doubling the current level by 2050, respectively. Based on the predicted trends in average temperature changes for each climate change scenario in the report, future temperature data for the next 100 years were generated by adjusting the annual average temperature with constant COV. In addition, relative humidity has been found to exhibit a clear cointegration with temperature<sup>42</sup>. Therefore, in this study, the probability distribution and correlation between environmental variables were derived using atmospheric data collected by the Korea Meteorological Administration[43](#page-16-32) over 10 years from 2014 to 2023. The probability distribution of temperature in South Korea was investigated to follow a normal distribution<sup>44</sup>. Moreover, the probability distribution of relative humidity, when averaged over time and space, also approaches a normal distribution, with the bimodality diminishing<sup>45</sup>. Table [2](#page-4-0) shows the mean, standard deviation, and correlation coefficient of the collected environmental variable data. As previously mentioned, due to the strong correlation between temperature and relative humidity, future relative humidity data for the next 100 years were sampled using multivariate Gaussian distributions to account for the interdependence with generated future temperature data. Based on the generated environmental variable data, the correction factors for chloride penetration and steel corrosion were derived. Figure [1](#page-5-0) shows the changes in the mean values of generated future temperature and relative humidity data over 100 years.

#### **Maintenance strategy**

Throughout the use stage, the maintenance strategies for the bridge were broadly categorized into preventive maintenance (PM), such as patching and repainting, and essential maintenance (EM), such as girder replacement and reconstruction<sup>46</sup>. The demands for maintenance action according to environmental conditions and climate change scenarios were estimated based on the risk assessment results. Since the PM interval ( *tm*) may vary depending on the rules or budgets of the bridge operators, LCA was performed by changing *tm* to 5, 10, and 15 for the parameter study. EM of girder replacement was set to occur when the failure probability  $(P_f)$  reached the target failure probability ( $P_{f,target}$ ) of 0.6% according to AASHTO MBE<sup>47</sup>. The corrosion rate can be initialized by PM, and the corrosion damage is initialized by EM. By performing risk assessment while changing the climate change scenario and *tm*, the number of PM and EM occurrences in the use stage was calculated probabilistically. This increase in the number of maintenance cycles directly led to an increase in environmental impact and cost.

<span id="page-4-0"></span>

**Table 2**. Environment variables in South Korea (2014–2024).

<span id="page-5-0"></span>

**Fig. 1**. Change in the future environmental variable. (**a**) temperature. (**b**) relative humidity.

## **Life cycle assessment**

The bridge LCA was carried out following ISO 14040  $48$ , 14044  $49$ , and 21930  $50$ . The system boundary was defined as cradle-to-cradle, including the potential benefits and loads of a circular economy. The production and construction stage (stage A) were calculated based on the bridge design bill of quantities (BOQ). Stage A includes the production of construction materials, but excludes material transport to the construction site and the operation of construction equipment. The use stage (stage B) was determined according to the number of maintenance actions (i.e., PM and EM) obtained from the risk assessment results. Stage B covers the production of maintenance materials anticipated over the life cycle but does not include transport to the site or maintenance equipment operation. The end-of-life stage (stage C) is calculated from the amount of waste generated, excluding the portion that is recycled, from the combined total of the amounts from stage A and stage B. Stage C includes equipment operation for bridge demolition and waste landfilling. For the assessment of the circular economy (stage D), adherence to ISO 20915<sup>51</sup> is crucial for accurately evaluating industry-specific data and the losses and benefits associated with recycling processes.

### **Cost assessment**

The bridge LCC includes material purchase cost, construction cost, maintenance action cost, deconstruction cost, disposal cost, recycling cost. For the realistic assessment of the bridge LCC in South Korea, the material price database<sup>52</sup>, annual maintenance cost model, and deconstruction, disposal, and recycling costs<sup>53</sup> were applied. The annual maintenance cost model covers all costs associated with actions taken throughout the entire lifetime of the bridge, addressing issues such as crack, efflorescence, peeling, corrosion, painting, and other maintenance actions occurring in the superstructure (slab and girder) and the substructure (abutment and pier). Due to the high uncertainty of bridge maintenance costs, which are influenced by factors like location, damage condition, and maintenance period, maintenance costs in LCC at the design phase was simply assessed using generalized statistical data. The superstructure replacement cost is estimated to be 133% of the initial superstructure construction cost<sup>[22](#page-16-11)[,54](#page-16-43)</sup>. Labor costs for construction workers were calculated at 33% of the total project budget<sup>55</sup>. Table [3](#page-6-0) shows the surveyed unit price of the bridge LCC.

#### **Environmental impact assessment**

The OpenLCA<sup>56</sup>, utilizing the ecoinvent database (DB)<sup>57</sup>, was employed for the environmental impact assessment. In cases where the product information was absent in the ecoinvent DB, the industrial environmental product

<span id="page-6-0"></span>

**Table 3**. Unit price of bridge LCC.

<span id="page-6-1"></span>

**Table 4.** MVCs per reference unit for impact categories $62$ .

declaration (EPD) DB registered in PASS<sup>[58](#page-17-5)</sup>, the LCA software developed by Korea Environmental Industry and Technology Institute (KEITI), was utilized. Environmental data related to all included materials throughout the bridges' lifetimes were gathered and characterized using the ReCiPe 2016 Midpoint (E) method<sup>59</sup>, yielding results for the six environmental impact categories:

- Global warming potential (GWP): It primarily relates to the emission of greenhouse gases such as carbon dioxide, methane, and nitrous oxide, contributing to the rise in Earth's temperatures.
- Abiotic depletion potential (ADP): It assesses the extent to which the production of a substance uses up Earth's non-living resources (such as metals and fossil fuels) and the risk of depleting these resources.
- Ozone depletion potential (ODP): It measures the ability of a chemical substance to degrade the Earth's ozone layer. Substances like chlorofluorocarbons (CFCs) belong to this category.
- Acidification potential (AP): It measures the ability of emissions to form acid rain when released into the atmosphere. Sulfur oxides (SOx) and nitrogen oxides (NOx) are primarily responsible.
- Photochemical ozone creation potential (POCP): It measures the capacity of certain chemicals to form ground-level ozone through photochemical reactions in the atmosphere. Ground-level ozone is a major component of smog, harmful to human respiratory health, and can damage crops and materials.
- Eutrophication potential (EP): This indicator quantifies the degree to which a substance can increase nutrient levels (particularly nitrogen and phosphorus) in aquatic environments, excessively promoting algae growth and eventually leading to oxygen depletion in water bodies.

Environmental impacts, considering loss and benefits from the recycling of steel, were determined based on data provided by Korean industries and environmental reports<sup>[60](#page-17-7)</sup>. The recycling rate of steel was commonly applied as 40.2%, reflecting the proportion of steel scrap used in crude steel production in South Korea in 2021<sup>61</sup>.

Additionally, a monetary valuation was conducted to provide an intuitive understanding of the environmental impact. Amadei et al.<sup>62</sup> collected monetary valuation coefficients (MVCs) for environmental impacts. Table [4](#page-6-1) shows the reference units, maximum, minimum, and average values of MVCs according to the six impact categories.

Figure [2](#page-7-0) illustrates the proposed framework of the risk-based bridge LCA method considering the climate change effect. The initial quantities based on the BOQ are calculated from the design that meets the target performance of the bridge. From the initial quantities, the cost and environmental impact during the stage A can be determined. Next, based on environmental conditions and climate change scenarios, a risk assessment is performed to account for the deterioration of member strength over time. Using the results of the risk assessment, the life cycle PM and EM action demands are estimated, leading to costs and environmental impacts during the stage B. In the stage C, the environmental impact and cost of treating residual waste, excluding recycling, are calculated. Finally, in the stage D, the potential benefits in terms of environmental impact and cost, based on the material recycling rate, are determined. All results from stages A to D are combined to calculate the total environmental impact and LCC over the bridge's lifespan.

# **Case study**

# **Designed bridges**

Case studies were conducted to verify the feasibility of the proposed framework in this study. The target bridges are the PSC girder bridge currently under construction in South Korea and the steel plate girder bridge, which was considered as a design alternative. Figure [3](#page-7-1) shows the side view and cross-sections of girders to be evaluated.

<span id="page-7-0"></span>

**Fig. 2**. Conceptual procedure of the proposed risk-based LCA considering climate change.

<span id="page-7-1"></span>



The target bridge for design consists of four girders with a total width of 12.3 m and a total length of 175 m (5 spans at 35 m each). The BOQ of the two bridge design alternatives is attached in Supplementary Material.

Properties for calculating *Mn,psc* (*t*) are as follows: *fps,i* (0) = 893 MPa, 871 MPa, 852 MPa, 830 MPa, 807 MPa for the *i*-th prestressing steel, respectively. Each PSC tendon consists of 12 strands with a diameter of 12.9 mm.  $D_{s,0} = 15.9$  mm.  $f_s = 300$  MPa. The numbers for the *i*-th prestressing steel are indicated in the red box in Fig. [3](#page-7-1)b. HSB380, bridge structural steel, was applied to the main member of the steel plate girder bridge.

### **Risk assessment**

The maximum load effect including permanent load and live load acting on the flexural member calculated according to the bridge property and AASHTO design truck is 10,203 kN·m and 7,273 kN·m on the PSC girder bridge and steel plate girder bridge, respectively. The environmental condition in the risk assessment is assumed as marine where the most vulnerable condition for deterioration of chloride penetration and steel corrosion.

Figure [4](#page-8-0) shows the correction factors for concrete chloride penetration for each climate change scenario over time. The correction factors proportionally accelerate the corrosive deterioration according to the severity of climate change. Compared to scenario 0, climate change scenario 3 may increase the comprehensive correction factor  $f_T(T) \cdot f_t(t) \cdot f_h(RH)$  by 29%, 15%, and 48%, respectively.

Figure [5](#page-9-0) shows the results of time-varying risk assessments and the corresponding variation of chloride concentration in the PSC girder bridge with a *tm* of 10 years. As shown in Fig. [5a](#page-9-0), compared to scenario 0, chloride concentration increased by up to 8%, 25%, and 61% for each respective climate change scenario. The failure probability also increased by up to 50% (from 0.2 to 0.3%) over 100 years, depending on the severity of climate change, as shown in Fig. [5b](#page-9-0). However, the impact of climate change on the increase in the failure probability was not significant enough to necessitate additional EM demand. Therefore, the final life cycle EM demand was evaluated as zero.

Figure [6](#page-9-1) shows the correction factor *R* (*T, RH*) by climate change scenarios. *R* (*T, RH*) changed sensitively to the severity of climate change compared to other correction factors due to the influence of high correlation

<span id="page-8-0"></span>

**Fig. 4**. Correction factors for concrete chloride penetration. (**a**)  $f_T(T)$ . (**b**)  $f_h(RH)$ . (**c**)  $f_T(T)f_t(t)f_h(RH)$ .

<span id="page-9-0"></span>

**Fig. 5**. Times-varied risk assessment results of the PSC girder bridge with  $t_m$  of 10 years. (**a**)  $C(x, t)$ . (**b**) failure probability.

<span id="page-9-1"></span>

**Fig. 6**. Correction factors for steel corrosion *R(T*, *RH)*.

between *T* and RH. *R* (*T, RH*) increased in climate change scenario 3 by 32% compared to scenario 0, which directly means a 32% increase in corrosion depth.

Figure [7](#page-10-0) shows the results of the time-varied risk assessment for the steel girder bridge with a  $t_m$  of 10 years. As shown in Fig. [7a](#page-10-0), the corrosion growth rate could accelerate according to climate change scenarios due to the correction factor. Figure [7](#page-10-0)b illustrates the failure probability of each climate change scenario. The failure probability increases according to the severity of the climate change scenarios. Since failure probability exceeded the *Pf,target* during the target life in scenarios 1 to 3, the number of EM requirements was added once as shown in Fig. [7](#page-10-0)b.

<span id="page-10-0"></span>

**Fig.** 7. Times-varied risk assessment results of steel girder bridge with  $t_m$  of 10 years. (a)  $d_p(t)$ . (b) failure probability.

<span id="page-10-1"></span>

**Fig. 8**. Environmental impact of PSC girder bridge (scenario 0–3).

# **Environmental impact**

Figure [8](#page-10-1) showcases the environmental impact during the life cycle of the PSC bridge with a *tm* of 10 years, considering the PM and EM determined by risk assessment. Since the number of maintenance actions is identical across all scenarios, with zero EM and nine PMs, the evaluation results for environmental impact remain consistent. The environmental impact during the stage B arises from these nine PM actions and constitutes under 3% of the total depending on the category. This is a very low value compared to stage A, which accounts for 85 to 99%, or stage D, which accounts for up to 32%.

Figure [9](#page-11-0) shows the life cycle environmental impact of the steel plate girder bridge assessed based on the number of PM and EM demands with a *tm* of 10 years. When only PM such as repainting was considered, the stage B accounted for a very small proportion of less than 1% of the total environmental impact, as shown in Fig. [9a](#page-11-0). However, the demand for additional EM due to the effects of climate change significantly increased the environmental impact of stage B, and accounted for a proportion of 8–11% of the total depending on the category, as shown in Fig. [9b](#page-11-0).

Figure [10](#page-12-0) compares the life cycle environmental impacts of scenarios 0 and 3 of the two bridge designs. When comparing the environmental impact of each material during the stage A, the steel components showed a dominant contribution to ADP and EP, with steel accounting for around 90% of these impacts. As a result, the steel plate girder bridge exhibited higher ADP and EP values compared to the PSC girder bridge across all scenarios. On the other hand, the PSC girder bridge had higher environmental impact scores in the remaining four categories, mainly due to the significant proportion of the environmental impact attributed to concrete, which ranged between 20 and 40%. In Scenario 3, however, due to the requirement for EM on the steel girder bridge throughout its life cycle, the steel girder bridge was evaluated to have higher environmental impacts in all categories. This increase was driven by the additional steel demands necessitated by girder replacement, significantly raising the overall environmental impact of the steel bridge by approximately 8–10%. Therefore, considering the impact of climate change in stage B is crucial, as it is directly correlated with the environmental conditions that accelerate structural deterioration and increase the EM demand.

The sensitivity analysis concerning the recycling rate was also performed for scenario 3 of the two bridge designs. We investigated the influence of potential benefits from recycling on the environmental impact of bridges by changing the recycling rate of steel from 0 to 100%, as depicted in Fig. [11.](#page-12-1) The analysis shows that the EP varies up to 4.8 times, and GWP 1.6 times depending on the recycling rate. Specifically, in GWP, the comparative results could reverse based on the recycling rate. These fluctuations highlight the substantial environmental benefits of recycling. In addition, the higher quantity of steel in the steel plate girder bridge leads to a larger fluctuation in environmental impacts relative to the PSC girder bridge, dependent on the recycling rate. This variability confirmed that the recycling rate could significantly influence the LCA, underscoring the importance of applying an appropriate recycling rate.

# **Economic feasibility**

Table [5](#page-13-0) displays the LCC calculation results for scenarios 0 and 3 of the two bridge designs. The change in the number of EM requirements leads to a reversal in the total LCC for the two bridges. Similar to environmental

<span id="page-11-0"></span>

**Fig. 9**. Environmental impact of steel girder bridge. (**a**) scenario 0. (**b**) scenario 1–3.

<span id="page-12-0"></span>

**Fig. 10**. Comparison of Environmental Impacts. (**a**) scenario 0. (**b**) scenario 3.

<span id="page-12-1"></span>



impacts, considering climate change effects at the stage B is a significant variable in design. Table [6](#page-13-1) illustrates the conversion of the environmental impacts in Fig. [10](#page-12-0) into environmental costs using the MVCs from Table [4.](#page-6-1) As with LCC, the change in the number of EM, due to consideration of the impact of climate change, results in a reversal of the environmental costs of the two bridges. Figure [12](#page-13-2) compares LCC and the maximum environmental cost. Compared to LCC, the environmental cost can be evaluated up to 71%, and its proportion is expected to increase significantly in the future if the importance of environmental impact increases from the present.

# **Effect of PM interval**

Changes in the frequency of PM according to maintenance strategies can profoundly influence the failure probability of the bridges. Additional risk assessments were performed with *tm* of 5 years and 15 years (i.e., 10 years *±* 50%). In the LCC, PM costs were calculated in proportion to the change in the number of PMs

<span id="page-13-0"></span>

**Table 5**. Summary of LCC.

<span id="page-13-1"></span>

	PSC bridge scenario 0-3	Steel bridge scenario 0	<b>Steel bridge</b> scenario 3
Max(10 <sup>8</sup> KRW)	35.6	35.3	39.6
Min $(10^8$ KRW)	1.25	1.16	1.29
Average $(10^8$ KRW $)$ 11.4		10.8	12.1

**Table 6**. Summary of environmental costs.

<span id="page-13-2"></span>



according to *tm* for each case. Figure [13](#page-14-0) illustrates the failure probability evaluated based on the change in *t<sup>m</sup>* for each bridge design. For the PSC girder bridge, the failure probability did not exceed the *Pf,target*, even when the *tm* was reduced to 5 years. However, when the *tm* was increased to 15 years, the failure probability exceeded *Pf,target* once in scenario 3, resulting in the occurrence of EM demand.

For the steel plate girder bridge, with an initially assumed *tm* of 10 years, the failure probability exceeded *Pf,target* twice in climate change scenarios 1 to 3. However, when the *tm* was reduced to 5 years, it did not exceed  $P_{f,target}$  in any climate change scenario. Conversely, increasing  $t_m$  to 15 years resulted in the failure probability exceeding *Pf,target* in scenarios 0 and 1 to 3 once or twice, respectively, necessitating additional EM demand.

Figure [14](#page-14-1) shows the environmental cost including life cycle PM and EM demands, as well as the total cost combined with LCC based on change in  $t_m$ . In scenario 0, the shorter the  $t_m$  the more economical the steel plate girder bridge becomes compared to the PSC girder bridge. This is because, in the absence of EM demands, the steel plate girder bridge is evaluated to have lower LCC and environmental costs than the PSC girder bridge. From the results of the steel plate girder bridge in scenario 0 and the PSC girder bridge in scenario 3, the total cost was found to be lowest when *tm* is 10 years, characterized by a low number of PMs and no EM being induced. Furthermore, in both scenarios 0 and 3, the total cost of the two bridge types is reversed depending on the change in *tm*. This highlights that identifying the optimal point to reduce the number of PMs without inducing EM can significantly improve the environmental impact and LCC of the bridge.

### **Conclusion**

This study systematically evaluated the life cycle environmental and economic impacts of two bridge designs-PSC girder and steel plated girder bridges-under varying climate change scenarios. Correction factors for concrete chloride penetration and steel corrosion, derived from weather statistics in South Korea and IPCC climate change scenarios, were integrated into the risk assessment to assess the climate change effect. In both

<span id="page-14-0"></span>

**Fig. 13**. Risk assessment results according to PM interval. (a) PSC girder bridge ( $t_m = 5$  years). (b) PSC girder bridge ( $t_m$  = 15 years). (c) steel plate girder bridge ( $t_m$  = 5 years). (d) steel plate girder bridge ( $t_m$  = 15 years).

<span id="page-14-1"></span>

**Fig. 14**. Total Cost according to PM interval. (**a**) PSC girder bridge. (**b**) steel plate girder bridge.

bridge design cases, the failure probability increased with the severity of the climate change scenario. The risk assessment results suggest that climate change may impact the failure probabilities and maintenance demands. Moreover, the analysis underscored the critical role of the recycling rate in determining the environmental impacts. Based on the proposed framework, an optimized maintenance strategy can be established during the climate crisis, taking into account both cost and environmental impact. The BOQ, material, and construction unit prices used in this study were based on data from South Korea. As the LCA results may change with future improvements in the database or modifications to the manufacturing process, it is advisable to focus more on trends rather than specific figures. The key findings include:

- In the case of steel plate girder bridges with a PM interval of 10 years, the failure probability, exacerbated by steel corrosion correction for climate change effects, exceeded the target failure probability of 0.6% over the lifetime. This increased failure probability necessitated a demand for life cycle EM to address these risks. On the other hand, for PSC bridges under the same conditions, the expected change in failure probability due to climate change was relatively insignificant, and the need for EM remained unchanged.
- A sensitivity analysis of the life cycle environmental impact relative to the recycling rate was performed. In the environmental impacts of EP and GWP, fluctuations of up to 4.8 times and 1.6 times, respectively, were observed. In particular, in GWP, the environmental impact of the two bridge designs could be reversed depending on the recycling rate. This result quantitatively suggests that the outcomes of LCA can be significantly improved by increasing the recycling rate.
- The environmental impact of PM during the use stage (stage B) is less than 3% for the PSC girder bridge and less than 1% for the steel girder bridge, representing a very small portion of the overall environmental impact categories. However, the environmental impacts and costs associated with EM can significantly influence the overall LCA results. Therefore, it is crucial to predict and incorporate the potential demand for EM, in addition to PM, by employing a risk-based LCA that considers the climate change effect during the design phase.
- When the maximum MVCs were applied, environmental costs were evaluated at up to 71% of LCC. If the weight for environmental impacts increases in the future, environmental costs may become a more important evaluation index than LCC as the relative proportion of LCC increases.
- Changes in the PM interval is directly related to the failure probability of the bridge. Decreasing the PM interval increases PM demand but decreases EM demand while increasing the PM interval has the opposite effect. Regardless of the climate change scenario, the total cost of the PSC girder bridge and the steel girder bridge is observed to reverse depending on changes in the PM interval. However, a shorter PM interval does not always guarantee better LCC and environmental outcomes. From the results of the steel plate girder bridge in Scenario 0 and the PSC girder bridge in Scenario 3, the total cost was found to be lowest when the PM interval was set to 10 years. This interval is characterized by a low number of PM activities and the absence of EM demands. Therefore, it is crucial to optimize the bridge life cycle maintenance demand based on risk evaluation. This optimization can significantly improve the life cycle environmental impact and LCC assessment results.

This study is limited to examining the effects of corrosion and flexural performance deterioration due to climate change. Given that the proposed bridge LCA framework has proven applicable, future research should expand to include limit states such as shear and fatigue, as well as other deterioration phenomena, such as shrinkage, frost damage, and etc. Also, additional life cycle environmental impact and cost assessments should be conducted incorporating material transport, construction equipment operation, and maintenance measures based on realworld industrial data.

### **Data availability**

Additional data can be made available at request to the corresponding author, with exception of confidential data from LCA model applied in this study.

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# **Author contributions**

All authors contributed to the conception and design of the study. Material preparation, data collection, and analysis were performed by Sang Hyeon Lee and Lee-Sak An. Sang Hyeon Lee wrote the first draft of the manuscript, and Lee-Sak An and Ho-Kyung Kim revised it. Ho-Kyung Kim supervised the project. All authors read and approved the final manuscript.

# **Declarations**

# **Competing interests**

The authors declare no competing interests.

# **Additional information**

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