

Benzothiazole Degradation by *Rhodococcus pyridinovorans* Strain PA: Evidence of a Catechol 1,2-Dioxygenase Activity

Nicolas Haroune,¹ Bruno Combourieu,¹ Pascale Besse,¹ Martine Sancelme,¹ Thorsten Reemtsma,² Achim Kloepper,² Amer Diab,³ Jeremy S. Knapp,³ Simon Baumberg,⁴ and Anne-Marie Delort^{1*}

Laboratoire de Synthèse et Etude de Systèmes à Intérêt Biologique, UMR 6504 CNRS-Université Blaise Pascal, 63177 Aubière Cedex, France¹; Department of Water Quality Control, Technical University of Berlin, Sekr KF 4, 10623 Berlin, Germany²; and Division of Microbiology, School of Biochemistry and Molecular Biology,³ and School of Biology,⁴ University of Leeds, Leeds, LS2 9JT, United Kingdom

Received 24 June 2002/Accepted 9 September 2002

The pathway for biodegradation of benzothiazole (BT) and 2-hydroxybenzothiazole (OBT) by *Rhodococcus pyridinovorans* strain PA was studied in detail. The kinetics of biodegradation were monitored by in situ ¹H nuclear magnetic resonance (NMR) in parallel with reversed-phase high-performance liquid chromatography (HPLC). Successive oxidations from BT to OBT and then from OBT to dihydroxybenzothiazole were observed. Further insight was obtained by using a mutant strain with impaired ability to grow on BT and OBT. The precise structure of another intermediate was determined by in situ two-dimensional ¹H-¹³C NMR and HPLC-electrospray ionization mass spectrometry; this intermediate was found to be a ring-opening product (a diacid structure). Detection of this metabolite, together with the results obtained by ¹H and ¹⁹F NMR when cells were incubated with 3-fluorocatechol, demonstrated that a catechol 1,2-dioxygenase is involved in a pathway for biodegradation of BTs in this *Rhodococcus* strain. Our results show that catechol 1,2-dioxygenase and catechol 2,3-dioxygenase activities may both be involved in the biodegradation of BTs depending on the culture conditions.

Benzothiazoles (BTs), a group of xenobiotic compounds containing a benzene ring fused with a thiazole ring, are manufactured worldwide for a variety of applications. Depending on the substituent of the thiazole ring, BTs are used as fungicides in lumber and leather production (30), as antifungal agents (6), as slimicides in the paper and pulp industry (27), and as chemotherapeutic agents (7). Methabenzthiazuron is used as a herbicide (25, 36). These applications indicate that BTs have a wide spectrum of biological activity. Aminobenzothiazole is an intermediate in dye production. Finally, mercaptobenzothiazole and its derivatives are used to promote vulcanization in rubber manufacture. This industrial process produces other BTs, such as benzothiazole-2-sulfonate and 2-hydroxybenzothiazole (OBT), as by-products. Released from rubber products or from BT production plants, these compounds have been detected in industrial wastewaters and in various environmental compartments (20, 30) and are of concern for the aquatic environment due to their limited biodegradability and potential toxicity (17, 23).

Only a few bacterial isolates have been shown to degrade BTs in pure culture (15, 16). Gaja and Knapp (21) described *Rhodococcus* strain PA, which grows on BT as a sole source of carbon, nitrogen, and energy, and strain TA, which grows on 2-aminobenzothiazole. Two strains, *Rhodococcus erythropolis* BTS1 and *Rhodococcus rhodochrous* OBT18, were shown to degrade BT and OBT (2, 17, 18). Benzothiazole-2-sulfonate is

also degraded by *R. erythropolis* (18), and 2-aminobenzothiazole is degraded by *R. rhodochrous* (24).

There have been few reports on the metabolic pathways involved in biodegradation of BTs. Recently, we used ¹H nuclear magnetic resonance (NMR) spectroscopy to elucidate the structures of intermediates in the biodegradation of BT and OBT by *R. rhodochrous* and *R. erythropolis* (2) and the structures of intermediates in the biodegradation of 2-aminobenzothiazole by *R. rhodochrous* (24). We showed that a common biodegradative pathway led to hydroxylation of the benzene ring at position 6 (2, 24). Various NMR techniques were used. (i) In situ ¹H NMR was performed directly with culture media at natural abundance and allowed biodegradation kinetics to be monitored; this approach was used previously to study microbial biodegradation of xenobiotic compounds (3, 5, 8–10, 13, 14, 29). (ii) ¹H-¹⁵N gradient heteronuclear multiple-bond correlation (HMBC) at natural abundance helped us to identify precisely the hydroxyl group position on the benzene ring by taking advantage of the information contained in ¹H-¹⁵N scalar couplings (2, 24, 26).

In this paper we describe a detailed study of the pathway of biodegradation of BT and OBT by *Rhodococcus pyridinovorans* strain PA in which a mutant strain with impaired ability to grow on BT and OBT was used. Our data demonstrate that a catechol 1,2-dioxygenase is involved in the biodegradation of BTs by this *Rhodococcus* strain.

MATERIALS AND METHODS

Chemicals. BT and OBT were purchased from Aldrich. Tetradeuterated sodium trimethylsilylpropionate (TSPd₄) was purchased from Eurisotop (Saint Aubin, France), and 3-fluorocatechol (3FC) was purchased from Acros Organics (Geel, Belgium).

* Corresponding author. Mailing address: Laboratoire de Synthèse et Etude de Systèmes à Intérêt Biologique, UMR 6504 CNRS-Université Blaise Pascal, 63177 Aubière Cedex, France. Phone: 33 4 73 40 77 14. Fax: 33 4 73 40 77 17. E-mail: amdelort@chimtp.univ-bpclermont.fr.

Dihydroxybenzothiazole (diOBT) was obtained as previously described (2).

Identification of *R. pyridinovorans* strain PA. *Rhodococcus* strain PA, which was isolated and described by Gaja and Knapp (21), was identified here on the basis of 16S rRNA sequences kindly provided by Leonid Kulakov and Mike Larkin, Queens University, Belfast, Northern Ireland. For amplification of 16S rRNA genes, the following primers were used: forward primer Eubac27F (5'-A GAGTTTGATCCTGGCTCAG-3') and reverse primer Eubac1492R (5'-AAG GAGGTGATCCAGCCGA-3'). The PCR regimen was as follows: denaturation at 95°C for 5 min; denaturation at 94°C for 45 s, annealing at 57°C for 30 s, and extension at 72°C for 1 min for 32 amplification cycles; and a final step consisting of 8 min at 72°C. The products were allowed to stand at 8°C.

Mutagenesis. Mutants of *R. pyridinovorans* strain PA were obtained by chemical mutagenesis performed with *N*-methyl-*N'*-nitro-*N*-nitrosoguanidine (MNNG) by the method of Delić et al. (12), except that the final concentration of MNNG was 40 µg/ml. Mutants with altered ability to utilize OBT were isolated on the basis of slow or no growth on OBT mineral salts agar.

Growth conditions. Unless otherwise indicated, *R. pyridinovorans* strain PA and mutant strain N-120-8 were grown in 100-ml portions of Trypticase soy broth (TSB) (bioMérieux, Marcy l'Etoile, France) in 500-ml Erlenmeyer flasks incubated at 30°C and 200 rpm. The cells were harvested after 20 h of culture. Cells for enzyme assays and some NMR studies were grown for ca. 48 h on liquid mineral salts medium containing (per liter) 1 g of (NH₄)₂SO₄, 1 g of K₂HPO₄, 1 g of KH₂PO₄, 4 mg of FeCl₃ · 6H₂O, and 40 mg of MgSO₄ · 7H₂O; the pH was adjusted to 6.7, and sodium acetate (10 mM), sodium benzoate (5 mM), or OBT (2 mM) was added as the sole source of carbon and energy.

Incubation with xenobiotic compounds. Cells were centrifuged at 8,000 × *g* for 15 min at 5°C. The pellet was washed twice with buffer (1 g of K₂HPO₄ per liter, 1 g of KH₂PO₄ per liter, 4 mg of FeCl₃ · 6H₂O per liter, 40 mg of MgSO₄ · 7H₂O per liter; pH 6.7) and finally resuspended in this buffer or in water for mass analysis (5 g [wet weight] of cells in 100 ml of buffer). The resting cells were incubated with 3 mM OBT or 1 mM BT in 500-ml Erlenmeyer flasks at 30°C with agitation (200 rpm). Negative controls lacking substrate or cells were incubated under the same conditions. Samples (1 ml) were taken every 30 min, centrifuged at 12,000 × *g* for 5 min, and prepared for high-performance liquid chromatography (HPLC) or ¹H NMR analysis.

HPLC analyses. HPLC analyses were performed by using a Waters 600E chromatograph with a Waters 486 UV detector set at 295 nm and a reversed-phase column (Interchrom Nucléosil C₁₈; 5 µm; 250 by 4.6 mm; Interchim) at room temperature. The mobile phase was acetonitrile-water (20:80, vol/vol), and the flow rate was 1 ml/min.

NMR spectroscopy. All ¹H and ¹³C NMR spectra were recorded with a Bruker Avance 300 spectrometer at 300.13 and 75.47 MHz, respectively, at 298 K by using a 5-mm triple-tuned ¹H-¹³C-¹⁵N probe equipped with a z-gradient coil. The ¹H and ¹³C 90° pulse lengths were 7.5 and 8.6 µs, respectively.

¹H NMR spectroscopy. Water resonance was suppressed by using the classical double pulsed-field gradient echo sequence WATERGATE. A total of 128 scans were collected (relaxation delay, 5 s; acquisition time, 3.64 s; spectral window, 3,420 Hz; 32,000 data points). A 1-Hz line-broadening procedure was applied before Fourier transformation, and baseline correction was performed on spectra before integration with Bruker software. TSPd₄ was used as an internal reference for chemical shift (0 ppm) and for quantification of metabolites, as previously described (9).

¹H-¹³C HSQC experiments. Gradient-enhanced heteronuclear single quantum coherence (HSQC) was achieved by using 128 (*t*₁) × 1,024 (*t*₂) data points, and the number of free induction decays added for each *t*₁ increment was 300. The acquisition time was 143 ms for spectral widths of 3,600 and 18,100 Hz in F2 and F1, respectively. Zero-filling to 512 points and π/2 shifted sine window function in *t*₁ and π/3 shifted squared sine function in *t*₂ were applied prior to two-dimensional Fourier transformation.

¹H-¹³C gradient HMBC experiments. A low-pass J filter (145 Hz) was used. The delay to allow ³J_{CH} correlations was set at 50 ms. Typically, 1,024 data points with 32 scans for each of 128 *t*₁ increments were acquired with spectral widths of 3,600 Hz in F2 and 19,600 Hz in F1. The required acquisition time was 142 ms. A recycle delay of 2 s was used. Zero-filling to 512 points and an exponential filtering (2 Hz) window in *t*₁ and an exponential filtering (1 Hz) window in *t*₂ were applied prior to two-dimensional Fourier transformation. The ¹³C chemical shifts were referenced to TSPd₄.

For both HSQC and HMBC experiments, water resonance was suppressed by using a 2-s presaturation period during the recycle delay.

¹⁹F NMR spectroscopy. ¹⁹F NMR measurements were obtained with a Bruker Avance 400 spectrometer at 376.49 MHz and 298 K by using a 5-mm QNP ¹³C/³¹P/¹⁹F-¹H probe equipped with a z-gradient coil. The ¹⁹F 90° pulse length was 13.5 µs.

The spectral width used was 75,188 Hz. The number of data points used for acquisition was 131,072. About 19,500 scans were recorded. A 1.5-s recycle delay was used. No ¹H decoupling was achieved. Chemical shifts were reported relative to CFC₃ used as an external reference.

Enzyme assays. Catechol 1,2-dioxygenase and catechol 2,3-dioxygenase activities were assayed by the method of Gibson (22) by using cell extracts of *R. pyridinovorans* PA grown on TSB (bioMérieux), nutrient broth (Oxoid Ltd., Basingstoke, Hants, United Kingdom), acetate mineral salts, benzoate mineral salts, or OBT mineral salts. Cell suspensions were harvested by centrifugation, washed twice in distilled water, and resuspended in phosphate buffer (2 g of KH₂PO₄ per liter; pH 7) to which acetone (20%, vol/vol) was added. Cells were broken by ultrasonication (Sanyo Soniprobe150; amplitude, 10 to 15 µm) for 2.5 min; each 0.5-min burst was followed by a 2-min cooling period. Unbroken cells and cell fragments were removed by centrifugation (15 min at 12,000 × *g* and 4°C); the supernatant fluid was kept on ice until it was assayed. The protein concentrations of cell extracts were measured by the bicinchoninic acid method (Pierce, Rockford, Ill.).

MS. HPLC-mass spectrometry (MS) analyses were performed with a system consisting of an HP 1100 LC and Quattro LC tandem mass spectrometer (Micromass, Manchester, United Kingdom) as described elsewhere (31). Analytes were separated by ion pair chromatography on a phenylhexyl column (length, 150 mm; inside diameter, 2 mm; 3 µm; Phenomenex) at 40°C with eluents A (methanol-H₂O, 20:80) and B (methanol-H₂O, 70:30), both containing 5 mM tributylamine (32). Gradient elution was as follows: 0 min with 20% eluent B, 13 min with 85% eluent B, 14 min with 85% eluent B, 15 min with 20% eluent B, and 24 min with 20% eluent B. The triple-stage quadrupole mass spectrometer was equipped with an electrospray probe and was operated in the negative ion mode with a cone voltage of 26 kV. For MS-MS experiments, a collision energy of 20 eV was used.

RESULTS AND DISCUSSION

Identification of *R. pyridinovorans* strain. The *Rhodococcus* strain PA 16S rRNA sequence (1,384 bases) was obtained (GenBank database accession number AJ457068) and compared to 16S rRNA sequences in the GenBank databases (<http://www.ncbi.nlm.nih.gov/GenBank/index.html>) by using the program BLASTn 2.0 (1). The sequence of *Rhodococcus* strain PA was most similar to the sequence of *R. pyridinovorans* strain RO4 (accession number AF459741.1; 100% identical over the 1,384 bp sequenced for strain PA). On the basis of this similarity, we believe that strain PA belongs to the species *R. pyridinovorans*. Strain RO4 of this species (isolated on the basis of its ability to degrade pyridine) was described by Yoon et al. (37). *Rhodococcus* strain PA also showed high degrees of similarity to strains of *Rhodococcus roseus* (accession number X80624.1; 99.6% similar over the 1,384 bp sequenced, differing in only 5 bp) and *Rhodococcus rhodochrous* (accession number X79288.1; 97.2% similar over the 1,384 bp sequenced, differing in only 39 bp).

BT degradation by *R. pyridinovorans* strain PA. HPLC chromatograms and in situ ¹H NMR spectra collected during the biodegradation of BT (3 mM) by TSB-grown resting cells of *R. pyridinovorans* are shown in Fig. 1. TSB was used as it allowed easy production of cells, resulted in very large amounts of biomass, and consequently permitted rapid transformation of BT. As Fig. 1A shows, the concentration of BT (retention time, 45 min) decreased with time, and BT completely disappeared after 1 h. A new metabolite appeared after 30 min of incubation; the retention time of this compound (25 min) was similar to that of OBT (2), and its identity was confirmed by coinjection with authentic OBT. OBT was not present after 22 h. After 3 h of incubation, a new product appeared. The retention time of this compound (8 min) was shorter than that of OBT, indicating greater polarity, and was identical to that of diOBT

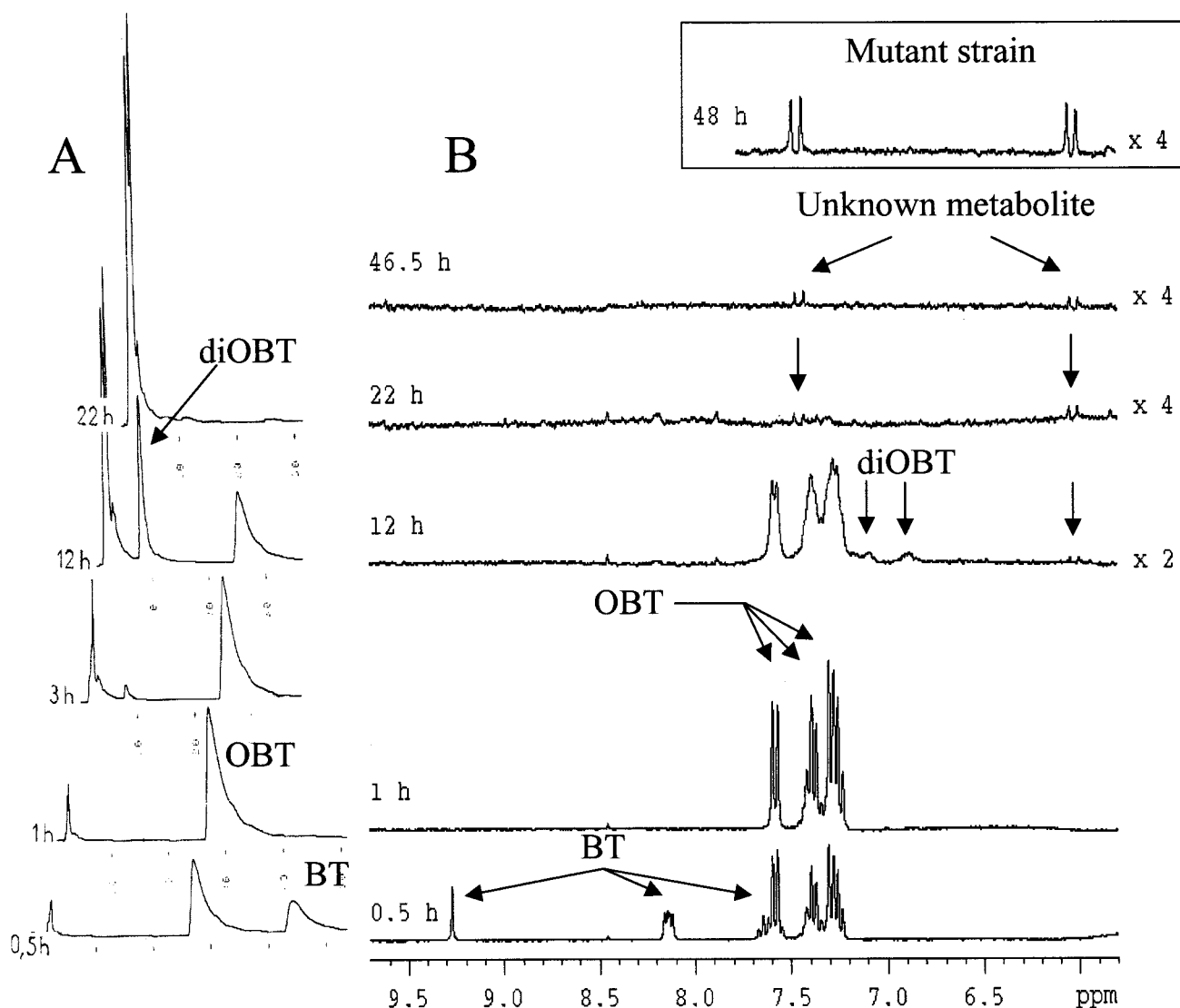


FIG. 1. Degradation of 3 mM BT by resting cells of *R. pyridinovorans* strain PA. (A) HPLC chromatograms. (B) In situ ^1H NMR spectra. The inset shows a ^1H NMR spectrum of *R. pyridinovorans* mutant strain N-120-8 incubated with 3 mM OBT for 48 h on the same scale.

(2). This assignment was confirmed by coinjection of diOBT onto the HPLC column. The concentration of diOBT increased until 12 h and then decreased with time.

^1H NMR signals corresponding to those of BT [δ ppm: 7.57 (t), 7.65 (t), 8.14 (2xd), 9.28 (s)] decreased with time (Fig. 1B), and new signals in the aromatic region appeared after 30 min; these new signals included two triplets at 7.24 and 7.38 ppm and two doublets at 7.29 and 7.58 ppm. They were similar to the signals observed previously during the degradation of BT by *R. erythropolis* and *R. rhodochrous* (2) and were assigned to OBT. The OBT intensity decreased, and OBT finally disappeared by 22 h. At 12 h, fleeting signals corresponding to those of diOBT (doublet at 7.08 ppm, doublet of doublet at 6.86 ppm, and doublet at 7.15 ppm) were detected in the NMR spectra; this ^1H NMR fingerprint was fully analyzed by Besse et al. (2). The detection of OBT and diOBT by ^1H NMR is consistent with the results obtained by HPLC.

From 12 to 46.5 h, new, small ^1H NMR signals that reso-

nated at 6.03 and 7.45 ppm were observed, and they corresponded to the signals of an unknown metabolite. The coupling constant ($J = 13.5$ Hz) and the chemical shifts of these two protons are consistent with the presence of an ethylenic bond in the structure. Therefore, the compound could correspond to a ring-opening compound, such as a dicarboxylic acid compound resulting from an intradiol (*ortho*) cleavage of a catechol derived from BT (Fig. 2). However, the concentration of this metabolite was far too low to allow confirmation of this hypothetical structure. Attempts to isolate the compound were unsuccessful.

Elucidation of the structure of the unknown metabolite. In order to increase the concentration of the unknown metabolite, we constructed mutants of *R. pyridinovorans* strain PA by chemical mutagenesis with MNNG. Mutant isolate N-120-8 was chosen for further study; it was unable to degrade BT and grew more slowly than the wild-type strain when OBT was used as the sole carbon and nitrogen source. This indicates that the

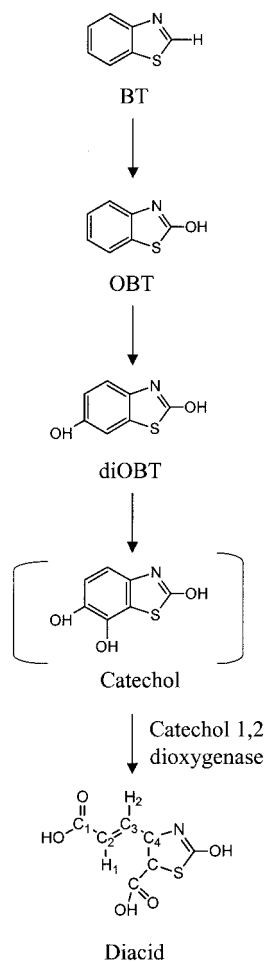


FIG. 2. Pathway of biodegradation of BT by *R. pyridinovorans* strain PA.

initial oxidation step (BT to OBT) is catalyzed by a different enzyme than the subsequent hydroxylation steps.

The kinetics of degradation of OBT (3 mM) by *R. pyridinovorans* strain PA and mutant strain N-120-8 were compared by performing ^1H NMR and HPLC. For mutant strain N-120-8, HPLC showed that the OBT concentration decreased with time, while the concentration of diOBT increased (data not shown). The ^1H NMR signals corresponding to those of OBT (two triplets at 7.24 and 7.38 ppm and two doublets at 7.29 and 7.58 ppm) decreased with time and were not present after 21 h of incubation; this is consistent with HPLC data. NMR could not detect diOBT due to its lower sensitivity than HPLC. Finally, the ^1H NMR signals ($\delta = 6.03$ and 7.45 ppm) of the unknown metabolite observed in the case of BT degradation were clearly detected (Fig. 1B, inset). It appears that the new metabolites observed during incubation of the wild-type and mutant strains with BT and OBT, respectively, were the same and that the unknown metabolite accumulated to a greater extent (2.5-fold) with the mutant strain, allowing complementary NMR and MS experiments to be performed.

Samples collected after 21 to 100 h of incubation of mutant strain N-120-8 with OBT were pooled and concentrated by

freeze-drying (after 21 h OBT and diOBT were absent from the samples, simplifying further NMR and MS analyses).

First, in situ ^1H - ^{13}C NMR experiments were performed with the concentrated sample without previous purification. On the ^1H - ^{13}C HSQC spectrum two crosspeaks were detected, indicating that there were $^1J_{\text{H-}^{13}\text{C}}$ couplings between the protons resonating at $\delta = 6.03$ and 7.45 ppm and the corresponding carbons resonating at 129.7 and 125.0 ppm, respectively. These data confirm that the protons are bound to ethylenic carbons. In the next experiment, ^1H - ^{13}C HMBC was designed to select ^1H - ^{13}C correlations between ethylenic protons and quaternary carbons ($^2J_{\text{H-}^{13}\text{C}}$, $d_6 = 50$ ms). The long-range heteronuclear shift correlations seen on the two-dimensional NMR spectrum demonstrated that the proton resonating at 7.45 ppm is close to a carboxylic group (C-1, $\delta^{13}\text{C} = 177.2$ ppm) and thus is assigned to H-1, while the proton resonating at $\delta = 6.03$ ppm is two bonds distant from the quaternary carbon (C-4) resonating at 137.1 ppm and can be assigned to H-2 (Fig. 2).

The combined results of these two NMR experiments are consistent with the proposed diacid structure (Fig. 2). However, NMR data gave no indication about the structure of the thiazole ring; therefore, liquid chromatography-electrospray ionization tandem MS was performed with the same sample to elucidate this aspect of the molecule. The daughter ion spectrum of the metabolite ($R_t = 7.1$ min) showed five diagnostically important signals; besides a weak signal for the molecular anion, which is consistent with the molecular formula of the suggested product (m/z 214), we recorded fragment ions generated by single and twofold decarboxylation (m/z 170 and m/z 126, respectively), which clearly indicate the presence of two carboxylate subunits in this metabolite. Loss of an ethylenic subunit likely results in an m/z of 98. Finally, the anion at m/z 42 is due to NCO^- (31). These two final fragments confirm that the thiazole ring with the hydroxy group in position 2 was not altered during oxygenation.

OBT and diOBT were previously observed as intermediates in two other *Rhodococcus* strains able to degrade BT (and OBT), namely, *R. erythropolis* BTS1 and *R. rhodochrous* OBT18 (2, 18), suggesting that there is a common pathway in these various rhodococci. Successive hydroxylation steps are observed from BT to OBT and then from OBT to diOBT, both of which could result from action of monooxygenases, which are probably distinct, as the mutant strain cannot catalyze the first step.

Very importantly, the presence of a dicarboxylic acid intermediate derived from OBT was detected, demonstrating that there is opening of the benzene ring of the OBT (while it is still attached to the thiazole ring) by *ortho* (intradiol) catechol cleavage.

Incubation in the presence of 3FC. Although the putative BT-catechol intermediate could not be detected by either HPLC or NMR because insufficient compound accumulated, the presence of the dicarboxylic intermediate is consistent with intradiol opening of the benzenic ring by a catechol 1,2-dioxygenase.

To confirm the activity of this enzyme, a specific inhibitor, 3FC (1 mM), was added to an incubation mixture containing *R. pyridinovorans* strain PA and BT (1 mM). The time courses of the BT, OBT, and diOBT concentrations (as determined by HPLC and in situ ^1H NMR) in the absence and in the presence

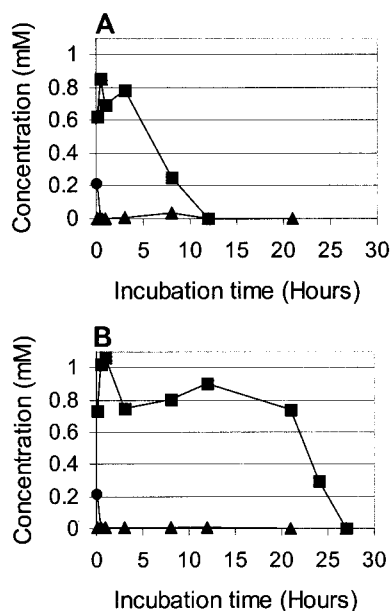


FIG. 3. Degradation of BT by *R. pyridinivorans* strain PA in the absence of 3FC (A) and in the presence of 3FC (B). Time courses of the concentrations of BT (●), OBT (■), and diOBT (▲) were determined from ^1H NMR spectra.

of 3FC are shown in Fig. 3A and B, respectively. The concentration of the diacid intermediate was too low to measure with the NMR spectra.

3FC had a clear effect on the transformation of OBT into diOBT (Fig. 3B); OBT was exhausted within 27 h in the presence of the inhibitor and within only 12 h in its absence. Also, diOBT did not accumulate in the incubation medium in the presence of the inhibitor (Fig. 3B). The transformation of BT into OBT was so rapid (less than 30 min) that it was not possible to see the influence of 3FC (Fig. 3). In a complementary experiment performed with more BT (3 mM) and fewer cells (0.5 g/liter), a clear reduction in the BT transformation rate was observed in the presence of 3FC. In the absence of 3FC, BT was exhausted after 30 min, while BT disappeared only after 70 min in the presence of 3FC (data not shown).

When incubation of 3FC alone with *R. pyridinivorans* strain PA was monitored by *in situ* ^1H NMR, the signals of 3FC ($\delta = 6.75$ ppm, multiplet) decreased with time, while new signals (5.95 ppm, doublet; 6.95 to 7.30 ppm, unresolved signals) appeared; the chemical shifts of these signals and the coupling constant of the doublet ($J_{1\text{H}-1\text{H}} = 11.2$ Hz) are consistent with the presence of 2-fluoromuconate. The structure of this compound was confirmed by ^{19}F NMR. Two main signals were detected at -123.7 and -111.8 ppm in the ^{19}F NMR spectrum recorded after 49 h of incubation; these signals were assigned to the fluoride anion (F^-) (the presence of which proves that defluorination occurred) and 2-fluoromuconate, respectively. The signal corresponding to 3FC (-140.4 ppm) was no longer present. These assignments were based on the chemical shift data and on the coupling constant ($^3J_{19\text{F}-1\text{H}} = 20$ Hz; 2-fluoromuconate) reported by Boersma et al. (4). These data confirmed that there was an intradiol cleavage catechol dioxygenase

TABLE 1. Specific activities of catechol 1,2-dioxygenase and catechol 2,3-dioxygenase on pyrocatechol in cell extracts of wild-type *R. pyridinivorans* PA^a

Growth substrate	Protein concn in cell extract (mg/ml)	Sp act (nmol/min/mg of protein) of ^a :	
		Catechol 1,2-dioxygenase	Catechol 2,3-dioxygenase
Acetate	0.9	31	ND ^b
Benzoic acid	4.7	870	ND
Nutrient broth	0.61	29	ND
TSB	2.7	7.9	ND
OBT	0.3	ND	0.48

^a The values are the means of at least three determinations, all of which gave very similar activities.

^b ND, no detectable activity.

activity in TSB-grown cells. 3FC inhibited BT degradation, but accumulation of the putative BT-catechol was not observed.

Enzyme assays. As explained above, we worked with TSB-grown cells to get enough biomass and thus enough intermediates for analysis by NMR spectroscopy. Our data clearly revealed that there was catechol 1,2-dioxygenase activity; this finding appeared to conflict with previous reports (21) of a catechol 2,3-dioxygenase activity. Therefore, we tested the effects of various media (compared with TSB and OBT) on the catechol 1,2-dioxygenase and catechol 2,3-dioxygenase activities. As TSB contains aromatic compounds, like amino acids, that could induce dioxygenase activities, we also tested nutrient broth, a benzoate-containing medium, and an acetate-containing growth medium (without aromatic compounds).

Conventional enzyme assays were conducted with cell extracts of *R. pyridinivorans* strain PA grown on various media. The specific activities of catechol 1,2-dioxygenase and catechol 2,3-dioxygenase acting on pyrocatechol as a substrate (Table 1) confirmed the presence of catechol 1,2-dioxygenase activity in cells grown on TSB, the medium used for growth of cells in all experiments on biotransformation of BTs reported so far. However, the activity in TSB-grown cells was low (similar to the activity in cells grown in nutrient broth or acetate mineral salts medium). This activity can be regarded as basal activity compared to the very high, induced levels of catechol 1,2-dioxygenase activity found in benzoate-grown cells. Interestingly, cells grown on OBT mineral salts medium did not contain detectable catechol 1,2-dioxygenase activity but did have catechol 2,3-dioxygenase activity, which was not observed in cells grown on any other substrate. The assays performed with OBT-grown cells confirmed previous findings (19, 21) obtained with OBT-grown rhodococci.

Complementary NMR experiments were then performed with resting cells grown on TSB, acetate, and OBT mineral salts media and then incubated with OBT (3 mM). OBT was completely degraded after 24, 21, and 4 h, respectively. In resting cells grown on TSB, OBT, or acetate (data not shown), one-dimensional NMR spectra showed the presence of the signals assigned to the diacid compound resulting from catechol 1,2-dioxygenase activity. These experiments proved that this enzyme was active under these various conditions.

Enzyme assays could not detect catechol 1,2-dioxygenase activity in OBT-grown cells, but this could have been due to a

problem of sensitivity or specificity, as pyrocatechol, not the putative BT-catechol, was the substrate for enzyme assays.

Both catechol 1,2-dioxygenase and catechol 2,3-dioxygenase have been reported in many *Rhodococcus* strains (33). Sometimes both activities are present in the same strain and can act simultaneously. For example, Warhurst et al. (34) showed that *R. rhodochrous* NCIMB 13259 could convert styrene to 3-vinylcatechol, which was metabolized by a catechol 1,2-dioxygenase to 2-vinyl-*cis,cis*-muconate, a dead-end product. At the same time a catechol 2,3-dioxygenase activity was also present, and it was proposed that this activity was part of a productive pathway by which 3-vinylcatechol was fully degraded. Similar observations were made by the same authors (35) with other aromatic compounds. More recently, Dean-Ross et al. (11) studied degradation of anthracene by an unidentified *Rhodococcus* species and proposed that degradation occurs via 1,2-dihydroxyanthracene, further metabolism of which proceeds via both *ortho* cleavage and *meta* cleavage. Moody et al. (28) obtained similar results for a related organism, *Mycobacterium* sp. strain PYR-1. Anthracene appears to be metabolized via three routes, but the product of the *ortho* cleavage [3-(2-carboxyvinyl)naphthalene-2-carboxylic acid] appears to accumulate, while the *meta* cleavage product (6,7-benzocoumarin) disappears from the culture medium and is assumed to be a transient intermediate and substrate for ring fission enzymes.

Conclusions. The pathways for degradation of BTs are still largely unknown; we describe here a detailed study of the degradation of BT (and OBT) by *R. pyridinovorans* strain PA. Identification of various metabolites was made possible by the use of complementary analytical tools, such as HPLC, in situ ^1H NMR, ^1H - ^{13}C HSQC and HMBC, in situ ^{19}F NMR, and HPLC-electrospray ionization MS. Finally, the use of a mutant strain allowed us to increase the concentration of one of the metabolites.

Considering the metabolites identified, a biodegradative pathway can be proposed for the degradation of BT (and OBT) by *R. pyridinovorans* strain PA (Fig. 2). A dicarboxylic acid produced by *ortho* cleavage of a putative BT-catechol has been identified, and the involvement of a catechol 1,2-dioxygenase in cleavage of the benzene ring of BT is reported here for the first time. However, in *R. pyridinovorans* PA it is not yet certain whether this dicarboxylic acid produced by *ortho* cleavage of OBT is part of a productive pathway for BT degradation or, in common with the examples described above, is a by-product of a basal activity of catechol 1,2-dioxygenase. The role in BT degradation of the catechol 2,3-dioxygenase activity, which is induced by growth on OBT, is similarly unclear. Both these questions are currently being investigated.

ACKNOWLEDGMENTS

The sequence of the 16S rRNA of strain PA was kindly determined by L. Kulakov and M. J. Larkin of Queens University, Belfast, Northern Ireland, whose assistance is gratefully acknowledged. We also thank Lynsey Farrell for her participation in this work.

N. Haroune is a recipient of a grant from the Ministère de l'Éducation Nationale et de la Recherche. A. Diab is grateful to the Egyptian Ministry of Education for financial support.

REFERENCES

- Altschul, S. F., W. M. Gish, W. Miller, E. W. Myers, and D. J. Lipman. 1990. Basic local alignment search tool. *J. Mol. Biol.* **215**:403–410.
- Besse, P., B. Combourieu, G. Boyse, M. Sancelme, H. De Wever, and A. M. Delort. 2001. Long-range ^1H - ^{15}N heteronuclear shift correlation at natural abundance: a tool to study benzothiazole biodegradation by two *Rhodococcus* strains. *Appl. Environ. Microbiol.* **67**:1412–1417.
- Besse, P., B. Combourieu, P. Poupin, M. Sancelme, N. Truffaut, H. Veschambre, and A. M. Delort. 1998. Degradation of morpholine and thiomorpholine by an environmental *Mycobacterium* involves a cytochrome P450. Direct evidence of the intermediates by *in situ* ^1H -NMR. *J. Mol. Biocatal. B Enzymes* **5**:403–409.
- Boersma, M. G., T. Y. Dinarevia, W. J. Middelhoven, W. J. H. Van Berkel, J. Doran, J. Vervoort, and I. M. C. M. Rietjens. 1998. ^{19}F nuclear magnetic resonance as a tool to investigate microbial degradation of fluorophenols to fluorocatechols and fluoromuconates. *Appl. Environ. Microbiol.* **64**:1256–1263.
- Brecker, L., and D. W. Ribbons. 2000. Biotransformations monitored *in situ* by proton nuclear magnetic resonance spectroscopy. *Trends Biochem.* **18**:197–202.
- Bujdakova, H., K. Kralova, and E. Sidoova. 1994. Antifungal activity of 3-(2-alkylthio-6-benzothiazolylaminomethyl)-2-benzothiazolinethiones *in vitro*. *Pharmazie* **49**:375–376.
- Bujdakova, H., T. Kuchta, E. Sidoova, and A. Gvozdjakova. 1993. Anti-*Candida* activity of four antifungal benzothiazoles. *FEMS Microbiol. Lett.* **112**:329–334.
- Combourieu, B., P. Besse, M. Sancelme, J. P. Godin, A. Monteil, H. Veschambre, and A. M. Delort. 2000. Common degradative pathways of morpholine, thiomorpholine, and piperidine by *Mycobacterium aurum* MO1: evidence from ^1H -nuclear magnetic resonance and ion spray mass spectrometry performed directly on the incubation medium. *Appl. Environ. Microbiol.* **66**:3187–3193.
- Combourieu, B., P. Besse, M. Sancelme, H. Veschambre, A. M. Delort, P. Poupin, and N. Truffaut. 1998. Morpholine degradation pathway of *Mycobacterium aurum* MO1: direct evidence of intermediates by *in situ* ^1H nuclear magnetic resonance. *Appl. Environ. Microbiol.* **64**:153–158.
- Combourieu, B., P. Poupin, P. Besse, M. Sancelme, H. Veschambre, N. Truffaut, and A. M. Delort. 1998. Thiomorpholine and morpholine oxidation by a cytochrome P450 in *Mycobacterium aurum* MO1. Evidence of the intermediates by *in situ* ^1H NMR. *Biodegradation* **9**:433–442.
- Dean-Ross, D., J. D. Moody, J. P. Freeman, D. G. Doerge, and C. E. Cerniglia. 2001. Metabolism of anthracene by a *Rhodococcus* species. *FEMS Microbiol. Lett.* **204**:205–211.
- Delić, V., D. A. Hopwood, and E. J. Friend. 1971. Mutagenesis by *N*-methyl-*N'*-nitro-*N*-nitrosoguanidine (NTG) in *Streptomyces coelicolor*. *Mutat. Res.* **9**:167–182.
- Delort, A. M., and B. Combourieu. 2000. Microbial degradation of xenobiotics, p. 411–430. *In* J. N. Barbotin and J. C. Portais (ed.), *NMR in microbiology: theory and applications*. Horizon Scientific, Wymondham, Norfolk, United Kingdom.
- Delort, A. M., and B. Combourieu. 2001. *In situ* ^1H -NMR study of the biodegradation of xenobiotics: application to heterocyclic compounds. *J. Ind. Microbiol. Biotechnol.* **26**:2–8.
- De Wever, H., P. Besse, and H. Verachtert. 2001. Microbial transformations of 2-substituted benzothiazoles. *Appl. Microbiol. Biotechnol.* **57**:620–625.
- De Wever, H., S. de Cort, I. Noots, and H. Verachtert. 1997. Isolation and characterization of *Rhodococcus rhodochrous* for the degradation of the wastewater component 2-hydroxybenzothiazole. *Appl. Microbiol. Biotechnol.* **47**:458–461.
- De Wever, H., and H. Verachtert. 1997. Biodegradation and toxicity of benzothiazoles. *Water Res.* **31**:2673–2684.
- De Wever, H., K. Vereecken, A. Stolz, and H. Verachtert. 1998. Initial transformations in the biodegradation of benzothiazoles by *Rhodococcus* isolates. *Appl. Environ. Microbiol.* **64**:3270–3274.
- De Wever, H. 1995. Biodegradability of benzothiazoles. Ph.D. thesis. Katholieke Universiteit Leuven, Leuven, Belgium.
- Fiehn, O., T. Reemtsma, and M. Jekel. 1994. Extraction and analysis of various benzothiazoles from industrial wastewater. *Anal. Chim. Acta* **295**:297–305.
- Gaja, M. A., and J. S. Knapp. 1997. The microbial degradation of benzothiazoles. *J. Appl. Microbiol.* **83**:327–334.
- Gibson, D. T. 1970. Assays of enzymes of aromatic metabolism. *Methods Microbiol.* **6A**:463–478.
- Gold, L. S., T. H. Slone, B. R. Stern, and L. Bernstein. 1993. Comparison of target organs of carcinogenicity for mutagenic and non-mutagenic chemicals. *Mutat. Res.* **286**:75–100.
- Haroune, N., B. Combourieu, P. Besse, M. Sancelme, and A. M. Delort. 2001. ^1H NMR: a tool to study the fate of pollutants in the environment. *C. R. Acad. Sci. Paris Chim.* **4**:759–763.
- Hartley, D., and H. Kidd. 1987. *The agrochemical handbook*. The Royal Society of Chemistry, Nottingham, United Kingdom.
- Martin, G. E., and C. E. Hadden. 2000. Long-range ^1H - ^{15}N heteronuclear shift correlation at natural abundance. *J. Nat. Prod.* **63**:543–585.
- Meding, B., K. Toren, A. T. Karlberg, S. Hagberg, and K. Wass. 1993.

- Evaluation of skin symptoms among workers at a Swedish paper mill. *Am. J. Ind. Med.* **23**:721–728.
28. **Moody, J. D., J. P. Freeman, D. G. Doerge, and C. E. Cerniglia.** 2001. Degradation of phenanthrene and anthracene by cell suspensions of *Mycobacterium* sp. strain PYR-1. *Appl. Environ. Microbiol.* **67**:1476–1483.
 29. **Poupin, P., N. Truffaut, B. Combourieu, P. Besse, M. Sancelme, H. Veschambre, and A. M. Delort.** 1998. Degradation of morpholine by an environmental strain of *Mycobacterium* involves a cytochrome P450. *Appl. Environ. Microbiol.* **64**:159–165.
 30. **Reemtsma, T., O. Fiehn, G. Kalnowski, and M. Jekel.** 1995. Microbial transformations and biological effects of fungicide-derived benzothiazoles determined in industrial wastewater. *Environ. Sci. Technol.* **29**:478–485.
 31. **Reemtsma, T.** 2000. Determination of 2-substituted benzothiazoles of industrial use from water by liquid chromatography/electrospray ionization tandem mass spectrometry. *Rapid Commun. Mass Spectrom.* **14**:1612–1618.
 32. **Storm, T., T. Reemtsma, and M. Jekel.** 1999. Use of volatile amines as ion pairing agents for the high-performance liquid chromatographic-tandem mass spectrometric determination of aromatic sulfonates in industrial wastewater. *J. Chromatogr. A* **854**:175–185.
 33. **Warhurst, A. M., and C. A. Fewson.** 1994. Biotransformations catalysed by the genus *Rhodococcus*. *Crit. Rev. Biotechnol.* **14**:29–73.
 34. **Warhurst, A. M., K. F. Clarke, R. A. Hill, R. A. Holt, and C. A. Fewson.** 1994. Metabolism of styrene by *Rhodococcus rhodochrous* NCIMB 13259. *Appl. Environ. Microbiol.* **60**:1137–1145.
 35. **Warhurst, A. M., K. F. Clarke, R. A. Hill, R. A. Holt, and C. A. Fewson.** 1994. Production of catechols and muconic acids from various aromatics by the styrene-degrader *Rhodococcus rhodochrous* NCIMB 13259. *Biotechnol. Lett.* **16**:513–516.
 36. **Wegler, R., and L. Eue.** 1977. *Chemie der Pflanzenschutz- und Schädlingsbekämpfungsmittel*, vol. 5. Herbizide. Springer-Verlag, Berlin, Germany.
 37. **Yoon, J.-H., S.-S. Kang, Y.-G. Cho, S. T. Lee, Y. H. Kho, C.-J. Kim, and Y.-H. Park.** 2000. *Rhodococcus pyridinivorans* sp. nov., a pyridine-degrading bacterium. *Int. J. Syst. Evol. Microbiol.* **50**:2173–2180.