# Development and Application of Real-Time PCR Assays for Quantification of *erm* Genes Conferring Resistance to Macrolides-Lincosamides-Streptogramin B in Livestock Manure and Manure Management Systems $\nabla$

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**Erythromycin and tylosin are commonly used in animal production, and such use is perceived to contribute to the overall antimicrobial resistance (AR) reservoirs. Quantitative measurements of this type of AR reservoir in microbial communities are required to understand AR ecology (e.g., emergence, persistence, and dissemination). We report here the development, validation, and use of six real-time PCR assays for quantifying six classes of** *erm* **genes (classes A through C, F, T, and X) that encode the major mechanism of resistance to macrolides-lincosamidesstreptogramin B (MLSB). These real-time PCR assays were validated and used in quantifying the six** *erm* **classes in five types of samples, including those from bovine manure, swine manure, compost of swine manure, swine waste lagoons, and an Ekokan upflow biofilter system treating hog house effluents. The bovine manure samples were found to contain much smaller reservoirs of each of the six** *erm* **classes than the swine manure samples. Compared to the swine manure samples, the composted swine manure samples had substantially reduced** *erm* **gene abundances (by up to 7.3 logs), whereas the lagoon or the biofilter samples had similar** *erm* **gene abundances. These preliminary results suggest that the methods of manure storage and treatment probably have a substantial impact on the persistence and decline of MLS<sub>B</sub> resistance originating from food animals, thus likely affecting the dissemination of such resistance genes into the environment. The abundances of these** *erm* **genes appeared to be positively correlated with those of the** *tet* **genes determined previously among these samples. These real-time PCR assays provide a rapid, quantitative, and cultivation-independent measurement of six major classes of** *erm* **genes, which should be useful for ecological studies of AR.**

There is a growing interest in ecological studies of antimicrobial resistance (AR) owing to the increasing concern over the potential risk associated with AR originating from animals intended for food (2, 4, 12, 20, 47). Although mostly commensals, the microbes in the intestines and manures of food animals (estimated at  $>10^{10}/g$  of manure) can serve as larger reservoirs of AR genes than pathogens (16, 32). These large AR gene reservoirs likely increase both dissemination of AR genes to the environment and resistance gene transfer, not only among commensals (22) but also to pathogens (45). To assess the potential risk associated with AR originating from agricultural use of antibiotics, these resistance gene reservoirs need to be measured. Whole-community analysis was proposed as providing the greatest ability to assess the resistance gene reservoir in a microbial community (16).

Both tylosin and erythromycin belong to the structurally distinct, yet functionally related, macrolide-lincosamide-streptogramin B  $(MLS_B)$  superfamily of antibiotics. Erythromycin is used on both human and food animals, whereas tylosin is exclusively used on food animals (23). In fact, tylosin is one of the most commonly used antimicrobials in poultry, swine, and

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beef cattle (42). The use of tylosin on animals significantly increased the resistance by gut commensal bacteria to  $MLS_B$ (9, 17). Resistance to tylosin in a food animal production environment was found to be encoded by *erm* genes (1, 18, 19, 43). The *erm* genes encode 23S rRNA adenine-specific  $N_6$ methyltransferases, which methylate the 23S rRNA of bacteria (28). Such methylation results in decreased binding of all  $MLS_B$  drugs to their target (bacterial ribosomes) and thus resistance to all MLS<sub>B</sub> antibiotics. The *erm* genes are among the most common AR genes of MLS<sub>B</sub>, and 32 classes of *erm* genes ( $\geq 80\%$  amino acid sequence identity within each class) have been identified and sequenced to date among many different genera of bacteria (http://faculty.washington.edu/marilynr /ermwebA.pdf) (29). Additionally, *erm* genes are among the most common acquired resistance genes in bacteria and the only genes conferring resistance to  $MLS_B$  currently found in anaerobes (28, 31).

Given the difficulties in cultivating most of the bacteria in intestines and manures of mammalians (44), DNA-based techniques, especially PCR, are often used to examine resistance genes in these microbial communities. Both PCR and real-time PCR have been used in detecting and quantifying, respectively, *tet* genes in various environments (3, 8, 34, 47). These studies yielded interesting new knowledge on the distribution and reservoirs of many *tet* gene classes in several types of microbial communities. Because of both the widespread use of erythro4408 CHEN ET AL. **APPL. ENVIRON.** MICROBIOL.





*a* For touchdown PCR. See the text for details.<br>*b* Numbered according to the  $erm(A)$  gene in *Staphylococcus aureus* transposon Tn554 (GenBank accession no. X03216).

<sup>c</sup> Numbered according to the *erm*(B) gene in *Streptococcus pneumoniae* transposon Tn*IS45* (GenBank accession no. X52632).<br><sup>d</sup> Numbered according to the *erm*(C) gene on plasmid pT48 (from *S. aureus* strain T48) (GenB

<sup>e</sup> Numbered according to the erm(F) gene on plasmid pBF4 (from *Bacteroides fragilis*) (GenBank accession no. M14730).<br><sup>f</sup> Numbered according to the erm(T) gene on plasmid p121BS (from *Lactobacillus* spp.) (GenBank acce

mycin and tylosin and cross selection among different  $MLS_B$ , *erm* genes are among the most widely distributed AR genes (28, 29). However, their distribution and abundance in entire microbial communities, including animal manure (the major reservoir of *erm* genes derived from food animals), remain to be determined. Although a few publications reported the detection of *erm* genes in pathogenic isolates (5, 6, 11, 25, 35, 36), no PCR-based assay has been reported to quantify the *erm* gene reservoirs in entire microbial communities.

To complement the emerging efforts to understand AR ecology and dynamics, we are undertaking an effort to develop capabilities for quantitative measurements of AR gene reservoirs in entire microbiomes. In a previous study (47), we developed three real-time PCR assays that permit quantification of 10 major *tet* gene classes present in entire microbiomes. In this report, we described the development of six real-time PCR assays specific for *erm*(A), *erm*(B), *erm*(C), *erm*(F), *erm*(T), and *erm*(X) and their utility in quantifying the reservoirs of these *erm* genes present in bovine manures, swine manures, composted swine manures, swine waste lagoons, and an Ekokan upflow biofilter (EUB) system.

## **MATERIALS AND METHODS**

**Bacterial strains and growth conditions.** *Staphylococcus aureus*::Tn*554* carrying *erm*(A) and *Bacillus subtilis* carrying *erm*(C) on plasmid pE194 (15) were grown in trypticase soy broth containing 15 µg/ml erythromycin. Three *Escherichia coli* DH5 $\alpha$ strains carrying *erm*(B), *erm*(F), and *erm*(X) on plasmids pJIR229 (kindly provided by M. C. Roberts, University of Washington), pFD292, and pFK12 (kindly provided by A. Tauch, Universität Bielefeld, Germany), respectively, were grown in Luria-Bertani (LB) broth containing 30 µg/ml erythromycin. Overnight cultures were centrifuged, and the biomass was resuspended in Tris-EDTA buffer. These cell suspensions and an aliquot of plasmid p121BS (43) carrying *erm*(T) were used as positive controls in optimizing respective PCR assays.

**Microbial community samples and DNA extraction.** In addition to the same sets of community DNA samples (being stored at  $-80^{\circ}$ C in separate aliquots) previously used in the development of real-time PCR assays specific for *tet* genes (47), another eight fresh bovine manure samples collected from a beef herd in Ohio were added to the bovine manure set. As described for the previous sets of DNA samples (47), the community DNA from these eight bovine manure samples was extracted using the RBB+C method, which was shown to substantially increase DNA yields (48). The quality and quantity of these DNA samples were also determined by agarose gel electrophoresis and fluorospectrometry (47). In total, 55 samples belonging to five types were analyzed: samples from bovine manure  $(n = 16)$ , swine manure  $(n = 10)$ , compost of swine manure  $(n = 13)$ , lagoons with swine manure  $(n = 6)$ , and throughout an EUB system treating swine manure  $(n = 10)$ .

**Phylogenetic analysis of** *erm* **gene sequences, primer design, and specificity tests.** Thirty-two classes of *erm* genes have been identified so far. All the *erm* gene sequences belonging to these 32 classes currently available in GenBank were retrieved and then aligned using ClustalX (40). We attempted to design a single primer pair that permits detection of all known *erm* genes by PCR, but such a universal primer pair is not possible, because of the high degrees of sequence divergence among *erm* classes (data not shown). Thus, we chose to design specific primers for individual *erm* gene classes. The classes chosen were A, B, C, F, T, and X, because they, based on previous studies of resistant bacterial isolates, are common and/or have been detected in bacteria of animal origin. The sequences of these six classes were dereplicated after alignment using ClustalX (40). Using the *erm*(Y) gene from *Staphylococcus aureus* as an outgroup, a neighbor-joining tree was inferred using the program TreeCon as described previously (46). Each class of sequences was separated, and one classspecific primer pair was designed using the approach described previously (47). The *erm*(C)-specific primer pair reported by Chung et al. (10) matches all the known *erm*(C) sequences and allows for suitable amplicon length. Thus, it was used in real-time PCR to quantify *erm*(C). All the primers used in this study are described in Table 1.

"Regular" PCR was done using a PTC-100 thermocycler (MJ Research, Waltham, MA) with 50- $\mu$ l volumes containing 1 $\times$  PCR buffer (20 mM Tris-HCl [pH 8.4] and 50 mM KCl), 200  $\mu$ M deoxynucleoside triphosphates, 500 nM of each primer, 1.75 mM  $MgCl<sub>2</sub>$ , 670 ng/ $\mu$ l bovine serum albumin, 1.0  $\mu$ l community DNA (about 50 ng) or cell suspensions (the positive controls), and 1.25 U Platinum *Taq* DNA polymerase (Invitrogen Corporation, Carlsbad, CA), which allows hot-start PCR. After an initial denaturation at 94°C for 4 min, five cycles of touchdown PCR (denaturation at 94°C for 30 s, annealing for 30 s with a 1°C-per-cycle decrement from 5°C above to the final annealing temperature indicated in Table 1, and extension at 72°C for 1 min) was performed, followed by 30 regular cycles of PCR (94°C for 30 s, 30 s at the respective annealing temperature, and 72°C for 45 s) and a final extension for 7 min at 72°C. The

TABLE 2. Affiliations of the sequenced *erm* genes as determined by comparison to GenBank sequences

erm gene	erm clone	GenBank accession no. DO887617	Prevalence 5/5	Most similar match (GenBank accession no.)	Identity $(\% )$ 95.5
erm(A)	$ermA-SM1$			<i>Oceanobacillus iheyensis</i> HTE831 $erm(A)$ gene (BA000028)	
erm(B)	$ermB-BM3$	DO887618	6/7	Streptococcus pyogenes erm $(B)$ gene (AJ972606) and 20 other $erm(B)$ sequences in GenBank	100
	$ermB-BM5$	DO887619	1/7	<i>Enterococcus faecium erm</i> $(B)$ gene $(AY827541)$ and three other $erm(B)$ sequences in GenBank	99.2
erm(C)	$ermC-SM1$	DO887620	3/3	<i>Staphylococcus hyicus</i> plasmid pSES21 erm(C) (Y09003)	99.5
erm(F)	$ermF-SM2$	DO887621	5/5	Bacteroides fragilis R plasmid pBF4 erm(F) (M14730) and three other $erm(F)$ sequences in GenBank	98.7
erm(T)	$ermT-BM6$	DO887622	7/7	Lactobacillus fermentum plasmid pLME300 erm(T) (AJ188494) and two other $erm(T)$ sequences in GenBank	99.7
erm(X)	$ermX-BM1$	DO887623	6/6	Corynebacterium striatum strain M82B R plasmid pTP10 $erm(X)$ (AF024666)	99.4

above-mentioned optimal annealing temperatures were predetermined by gradient PCR using a RoboCycler (Stratagene, La Jolla, CA) and the plasmid DNA or the DNA from the pure cultures. No-template controls were included in parallel.

To confirm primer specificity, the PCR products from one community DNA sample were cloned into the TOPO-TA cloning vector (Invitrogen). Randomly selected clones were sequenced (one strand) by the Plant and Microbe Genome Facility at The Ohio State University. Base-calling examination and comparisons with GenBank sequences were performed as described previously (47). The BLASTn search output alignments were also examined for the presence of breakage, which can indicate chimeric sequences.

**Preparation of sample-derived real-time PCR standards.** One sample-derived standard was prepared for each of the six *erm* classes from each of the two sets of community DNA: (i) the DNA extracted from the bovine manure samples and (ii) the DNA derived from the swine manure, swine waste lagoon, swine manure compost, and EUB system samples, as done previously (47) with minor modification. Instead of amplifying the target *erm* genes from individual community DNA samples and then pooling the PCR products together, we amplified the *erm* genes by using respective specific primers and a DNA mixture containing approximately 100 ng of the individual DNA samples within each sample set. Then, the PCR product was purified using a QIAquick PCR purification kit (QIAGEN, Inc., Valencia, CA) and quantified using a Quant-iT kit (Invitrogen) as done previously (47). One sample-derived real-time PCR standard was also prepared from each set of the samples by using the pooled DNA and the universal bacterial primer pair 27f/1525r (21) for quantification of total bacteria by realtime PCR. The conditions of this PCR are the same as those described elsewhere (49). For each sample-derived standard, copy number concentration was calculated based on the length of the PCR product and the mass concentration. Tenfold serial dilutions were made in Tris-EDTA prior to real-time PCR (47). In total, 14 real-time PCR standards were prepared from the two sets of community DNA samples for the seven (six for *erm* gene classes and one for total bacteria) real-time PCR assays. Each of these standards was used in respective real-time PCR assays.

**Real-time PCR.** The conditions of the real-time PCR assays of *erm* genes were the same as those of the regular PCR described above, with the following exceptions: decreased primer concentrations (250 nM each) and inclusion of  $0.133\times$  of SYBR green I (Molecular Probes, Eugene, OR) and 30 nM of the reference dye ROX (Stratagene). As done previously (47), the thermal profiles consisted of four segments: (i) initial denaturation at 95°C for 4 min; (ii) five touch-down cycles of 94°C for 30 s, 5°C above the respective annealing temperature (Table 1) for 30 s with a 1°C decrement per cycle, and 72°C for 40 s; (iii) 45 cycles of 94°C for 30 s, the respective annealing temperature for 30 s, 72°C for 30 s, and 18 s at the temperature for fluorescence acquisition  $(T_{FA})$  (Table 1); and (iv) 95°C for 2 min, 55°C for 30 s, and 95°C for 30 s. Fluorescence data were collected at the 72 $\degree$ C and  $T_{FA}$  steps (end point) of the third segment and during the ramping from 55°C to 95°C (all point) of the last segment. Quantification of total bacteria was performed as described by Nadkarni et al. (24), except for the use of the sample-derived standards instead of genomic DNA from a single bacterial strain. All the real-time PCR assays were performed using an Mx3000p real-time PCR system (Stratagene). Baseline and threshold calculations were performed with Mx3000p software using the fluorescence signals acquired at *T*<sub>FA</sub>, at which primer dimmers completely denatured and did not adversely affect the quantification accuracy. Following real-time PCR, the products were confirmed by agarose gel electrophoresis and exclusion at T<sub>FA</sub> of fluorescence resulting from possible primer dimmers was verified by melting curve analysis (except for the real-time PCR assays for total bacteria, which employed the universal TaqMan probe) (24). All the real-time PCRs were done in triplicate for both the standards and the microbial community DNA samples.

Exactly as described previously (47), each of the real-time PCR assays for *erm* genes were validated by quantifying a series of known copies of *erm* gene standards spiked into a swine manure community DNA sample against respective sample-derived real-time PCR standards. The detection limit of each realtime PCR assay was also determined from the serial dilutions of the samplederived standard templates (47). Following these validation experiments, the abundance of each *erm* gene class present in each community DNA sample was quantified against that of its respective sample-derived standard by using the real-time PCR conditions described above. The identity of each sample was concealed during real-time PCR for a blind test and was revealed only after quantification was completed to avoid influencing results. For ease of description, *erm* gene abundance expressed as the number of copies  $g^{-1}$  (or copies ml<sup>-1</sup>) in the case of liquid samples) is referred to as absolute abundance, whereas abundance expressed as the number of *erm* copies per million copies (cpmc) of total bacterial *rrs* genes is referred to as relative abundance. The absolute abundance was calculated by multiplying the number of copies per real-time PCR and the number of reactions that can be done with the DNA derived from 1 gram or ml of each sample (47), while the relative abundance was calculated by dividing the absolute abundance of each *erm* gene class by the corresponding total bacterial abundance (the number of *rrs* gene copies per g or ml of sample) in each sample and then multiplying by 1 million.

The real-time PCR assays were also used to determine the prevalence of each *erm* gene class among different types of samples. A sample was considered positive for an *erm* gene when at least two of the three replicate real-time PCRs yielded a threshold cycle value and a PCR product of the expected size (based on the agarose electrophoresis) in the respective real-time PCR assay of that sample. The prevalence of the *erm* gene among each type of sample was calculated as the percentage of samples that yielded the expected PCR product.

**Statistical analysis.** The data were  $log_{10}$  transformed and analyzed using the Mixed Procedure of SAS 9.1 (SAS Institute, Cary, NC). Least-squares means (LSM) were calculated for all the data sets. Mean separation was conducted by using Fisher's protected least-significant-difference test, with significance declared at *P* values of  $\leq 0.05$ . The absolute abundance and relative abundance of *erm* genes were graphed as boxes and whiskers by using GraphPad Prism 4 (GraphPad Software, San Diego, CA).

**Nucleotide sequence accession numbers.** The *erm* gene sequences determined in this study have been deposited in GenBank under the accession numbers listed in Table 2.

## **RESULTS**

**Primer specificity and** *erm* **gene diversity.** Our in silico analysis suggested that no universal *erm* primer is possible, due to sequence divergency (Fig. 1), but at least one primer pair can be designed for each *erm* gene class. In this study, one specific primer pair was designed for each of the six classes of *erm*



FIG. 1. A neighbor-joining tree of six classes of *erm* genes. The tree was inferred from DNA sequences, and it was arbitrarily rooted with the *erm*(Y) gene of *Staphylococcus aureus*. Bootstrap values were calculated from 100 replicates, and the number at each node indicates the number of times that the node was supported in the bootstrap analysis. The bar represents a 0.1 estimated change per nucleotide. Each primer pair listed in Table 1 targets a corresponding cluster.

genes:  $erm(A)$ ,  $erm(B)$ ,  $erm(C)$ ,  $erm(F)$ ,  $erm(T)$ , and  $erm(X)$ . Because of the high degree of sequence similarity within these classes, degenerate bases were needed only for the *erm*(A) and *erm*(X)-specific primers (Table 1). Amplification of the intended *erm* gene by PCR using each of the primer pairs from the pure culture carrying the target *erm* gene and from one of the community DNA samples all produced a band of the expected size (data not shown). The sequencing of the randomly selected clones from each clone library of the community DNA produced sequences that match known *erm* sequences of the intended class in GenBank with high sequence identity (Table 2). The sequence identities with known *erm* gene sequences are within the homology ranges proposed for each class (31). Additionally, none of our sequences was broken into two segments in the BLASTn search alignments, suggesting a very low probability of chimeric sequences among our *erm* sequences. Collectively, these sequencing results confirmed the specificity of the *erm* class-specific primers and their utility with complex microbiome DNA samples. As indicated by clone ermA-SM1,

these primer pairs may also prime amplification of heretofore unidentified members of the respective *erm* gene classes. These preliminary results suggest a greater diversity of *erm* genes than what has been identified in bacterial isolates.

**Validation of real-time PCR assays and quantification of** *erm* **genes.** The accuracy of each real-time PCR assay was validated by quantifying known numbers of respective *erm* gene templates mixed into microbiome DNA samples, essentially as described previously (47). When the copy numbers of the *erm* genes spiked into the samples were plotted against the corresponding copy numbers of the *erm* genes quantified in the validation experiments, after correction for the background numbers of the *erm* genes present in the microbiome DNA itself, high *r* <sup>2</sup> values over at least 5 orders of magnitude were obtained for all six assays (Fig. 2). However, the slopes of all six regression plots are less than 1, suggesting suppression of PCR amplification of the targets in the presence of community DNA, especially in the case of *erm*(X). The value of each slope, referred to as the "suppression coefficient" hereafter, was fac-



FIG. 2. Validation curves plotting the copy numbers of the spiked *erm* gene standard (*x* axis) against the corresponding quantification values for that *erm* gene, after correction for background copies present in the community DNA sample, which contained (in numbers of copies/reaction) the following:  $erm(A)$ ,  $4.18 \times 10^3$ ;  $erm(B)$ ,  $6.32 \times 10^4$ ;  $erm(C)$ ,  $88.6$ ;  $erm(F)$ ,  $11.4$ ,  $erm(T)$ ,  $208$ ; and  $erm(X)$ , 75.4. Error bars indicate standard deviations  $(n = 3)$ . gDNA, genomic community DNA.

tored into the calculation of abundance of the respective *erm* class in all the samples by dividing the quantification values by the respective suppression coefficient. These real-time PCR assays detected fewer than 10 *erm* gene copies per real-time PCR (data not shown).

Different classes of the *erm* genes differed in prevalence among the five types of samples analyzed (Fig. 3). The *erm*(B) genes are the most prevalent and were detected in all 55 samples, while *erm*(C) genes were the least prevalent. The prevalence of *erm*(F) was only slightly lower than that of *erm*(B). The five different types of samples also differed in prevalence of different *erm* gene classes (Fig. 3). The swine manure samples have much higher prevalences of the *erm*

genes than the bovine manure samples. Compared to the swine manure samples, the composted manure samples, but not the lagoon samples and the samples collected from the Ekokan biofilter system, had lower prevalences of *erm*(C), *erm*(F), and *erm*(X).

The absolute abundances of the *erm* genes varied considerably among the five types of samples as well as among different *erm* gene classes within individual sample types (Fig. 4 and Table 3). LSM for the bovine manure samples were substantially  $($  >4.0 logs,  $P < 0.001$ ) lower for *erm* genes of all six classes than LSM for the swine manure samples. Compared to the swine manure samples, the composted swine manure samples had substantially reduced  $(P < 0.001)$  *erm* gene abun-



FIG. 3. Prevalence (percent positive samples) of *erm* genes in the five types of samples analyzed. Prevalence is indicated as a percentage of positive samples among all the samples within each sample type. BM, fresh bovine manure samples ( $n = 16$ ); SM, fresh swine manure samples ( $n = 10$ ); Cp, composted swine manure samples  $(n = 13)$ ; Lgn, samples from lagoons receiving hog house effluent  $(n = 6)$ ; EUB, samples from an EUB system treating hog house effluent and a lagoon receiving its effluent  $(n = 10)$ .

dances: *erm*(A), by 2.3 logs; *erm*(B), by 5.7 logs; *erm*(C), by 5.0 logs; *erm*(F), by 4.3 logs; *erm*(T), by 7.1 logs; and *erm*(X), by 7.3 logs. In contrast to the composted swine manure samples, the lagoon samples and the EUB samples did not exhibit significantly reduced *erm* gene abundances, except for the lagoon samples, which had 1.5 logs fewer  $erm(A)$  ( $P = 0.03$ ) and 2.3 logs fewer *erm*(T) ( $P < 0.001$ ), and the Ekokan biofilter samples, which had 2.7 logs fewer *erm*(T) ( $P < 0.0001$ ) than the swine manure samples.

The total *erm* gene reservoir consisting of the six *erm* gene classes is summarized in Table 3. Collectively, the swine manure samples had almost 6 logs more of the *erm* genes than the bovine manure samples. In the swine manure samples, *erm*(B) was the most abundant, accounting for approximately 72% of the *erm* gene reservoir, whereas *erm*(B) and *erm*(F) accounted for approximately 56% and 43%, respectively, of the *erm* gene reservoir in the bovine manure samples. Among the three types of treated swine manure samples, composted swine ma-



FIG. 4. Box-and-whisker plots of absolute abundance of *erm* genes. All *erm* data are expressed as log<sub>10</sub> numbers of copies per gram (wet weight) or ml of samples. See the legend to Fig. 3 for the acronyms of the sample types. Error bars indicate maximum and minimum values, horizontal lines indicate median values, and boxes indicate values between the 25th and 75th percentiles. The value above each box-and-whisker plot is the LSM for each type of sample. n.d., not detected.

Sample $type^a$	Total erm genes (no. of copies/g)	Proportion $(\%)^b$ of indicated gene class						
		erm(A)	erm(B)	erm(C)	erm(F)	erm(T)	erm(X)	
<b>BM</b> <b>SM</b> Cp Lgn <b>EUB</b>	$9.99 \times 10^3$ $3.63 \times 10^{9c}$ $1.60 \times 10^5$ $2.67 \times 10^{10c}$ $1.07 \times 10^{9c}$	$0.11 \pm 23.75$ $0.08 \pm 0.16$ $8.80 \pm 5.48$ $0.03 \pm 0.08$ $0.52 \pm 1.92$	$56.48 \pm 37.21$ $71.53 \pm 14.81$ $37.71 + 22.16$ $99.33 \pm 1.55$ $93.74 \pm 15.86$	$0.01 \pm \langle 0.01$ $0.00 \pm 0.00$ $0.06 \pm 10.29$ $0.00 \pm 0.00$ $0.02 \pm 0.13$	$42.65 \pm 37.85$ $2.45 \pm 9.56$ $10.83 \pm 6.14$ $0.60 \pm 1.32$ $5.23 \pm 14.35$	$0.74 \pm 3.89$ $25.36 \pm 15.40$ $30.42 \pm 15.67$ $0.02 \pm 0.06$ $0.19 \pm 0.51$	$0.01 \pm 0.01$ $0.58 \pm 0.56$ $12.17 \pm 22.35$ $0.02 \pm 0.16$ $0.30 \pm 0.57$	

TABLE 3. Proportion of each *erm* gene class (percentage of total *erm* genes) in the *erm* gene reservoirs consisting of the six *erm* gene classes among the five types of samples (based on LSM for absolute abundance)

*<sup>a</sup>* BM, bovine manure; SM, swine manure; Cp, compost; Lgn, lagoon.

 $<sup>b</sup>$  Data shown are means  $\pm$  standard deviations.</sup>

 $c$  Significantly ( $P < 0.05$ ) higher than values for BM and Cp.

nure samples had at least 4 logs fewer ( $P < 0.05$ ) *erm* genes than the swine manure samples collected from the lagoons and the Ekokan biofilter system. The proportions of each of the six *erm* gene classes also differed among the four types of swine manure samples. Relative to the swine manure samples, the composted swine manure samples had increased proportions of *erm*(A), *erm*(F), *erm*(T), and *erm*(X) but decreased proportions of *erm*(B). Interestingly, in both the lagoon and the Ekokan biofilter samples, *erm*(B) accounts for most (approximately 99% and 94%, respectively) of the *erm* gene reservoir.

In order to determine the relative abundances of individual *erm* gene classes, the total bacteria in each sample were quantified and expressed as the number of *rrs* copies per gram or ml of the sample. The total bacterial abundances in the bovine manure, swine manure, compost, lagoon, and Ekokan biofilter system samples were  $2.53 \times 10^9$ ,  $4.86 \times 10^{10}$ ,  $2.74 \times 10^8$ ,  $6.20 \times$  $10^{10}$ , and  $1.58 \times 10^{10}$  rs copies g<sup>-1</sup> or ml<sup>-1</sup>, respectively. Different *erm* gene classes exhibited different relative abundances (Fig. 5). Apparently, *erm*(B) had the greatest ( $P < 0.01$ ) relative abundance among the six classes measured, reaching as high at  $4.27 \times 10^5$  cpmc (equivalent to  $42.7\%$  of the total bacterial *rrs* copies) in the lagoon samples (Fig. 5B). The other five *erm* gene classes were much less abundant, amounting to fewer than  $2.0 \times 10^4$  cpmc (or 2.0% of the total bacterial *rrs* copies). The relative abundances of all six *erm* gene classes also varied considerably among the five different types of manure samples. The bovine manure samples had the lowest relative abundances of all six *erm* gene classes, several orders of magnitude lower than those found in the four types of swine manure samples. Except for the *erm*(C) genes, the swine manure samples had considerably greater ( $P \leq 0.0001$ ) relative abundances for all the *erm* genes analyzed than the bovine manure samples (Fig. 5).

The three types of treated swine manure samples had variable relative abundances of *erm* genes (Fig. 5). Except for the *erm*(A) and *erm*(C) genes, the composted swine manure samples had lower relative abundances of  $erm(B)$  (by 2.38 logs,  $P <$ 0.001),  $erm(F)$  (by 2.26 logs,  $P < 0.0001$ ),  $erm(T)$  (by 2.03 logs,  $P < 0.005$ ), and *erm*(X) (by 1.1 logs,  $P < 0.01$ ) than the swine manure samples. On the contrary, the lagoon and the Ekokan biofilter samples were found to have relative abundances for all the quantified *erm* genes similar  $(P > 0.05)$  to those for the swine manure samples, except those for *erm*(B), which were more than 2 logs lower ( $P < 0.0001$ ) in these two types of treated manure samples than in the swine manure samples.

### **DISCUSSION**

Using a similar approach reported previously (47), we developed six real-time PCR assays that permit quantification of six major classes of *erm* genes in animal manures, which constitute the major reservoirs of AR originating from food animal production. As done previously (47), we used respective sample-derived standards in validating each of the real-time PCR assays and in quantifying the *erm* gene classes. The use of such standards is important because differences in target sequence diversity (and thus differences in amplification efficiencies) between the samples and the real-time PCR standards can lead to inaccuracies (24). By using sample-derived standards, the real-time PCR assays were validated, and the target genes quantified, against potentially all target *rrs* or *erm* genes present in the samples analyzed rather than a few selected  $MLS_B$ -resistant laboratory strains, which may or may not be present in the samples. Additionally, the validation against sample-derived standards eliminates the effects of cell lysis and DNA recovery, which probably vary with different DNA extraction methods and stains used. All the assays were shown to be robust with five types of samples and specific for the intended *erm* gene classes. As validated against known copies of respective *erm* genes spiked into microbiome DNA, these realtime PCR assays were found to be precise and accurate (Fig. 2). Given the physiochemical and microbial complexity of the samples tested in this study, these real-time PCR assays may be applicable to other types of samples, such as soil, aquatic, and sewage samples.

The validation experiments identified different suppression coefficients, as indicated by the different slopes (Fig. 2), for different real-time PCR assays. All the suppression coefficients are smaller than 1, suggesting that the targets in the standards are amplified more efficiently than those present in the samples. This is probably attributable to a higher complexity of the DNA templates in the community DNA than in the real-time PCR standards. The differences in suppression coefficients among the real-time PCR assays may be explained, at least partially, by differences in efficiencies in primer annealing and in secondary structures of the amplicons and primers. Since suppression coefficients vary for different real-time time PCR assays, even with the same set of samples, new real-time PCR assays should be validated and the suppression coefficients factored into quantification by real-time PCR for improved accuracy.

Different *erm* gene classes had considerably variable abun-



FIG. 5. Box-and-whisker plots of relative abundance of *erm* genes. All *erm* data are expressed as  $log_{10}$  numbers of transformed *erm* cpmc of total 16S *rrs* genes. See the legend to Fig. 3 for the acronyms of the sample types. See the legend to Fig. 4 for a detailed explanation of the plots.

dances within each of the five types of samples. However, several general trends appeared to be evident. Except for *erm*(T) being slightly more abundant than *erm*(F) in the compost samples, *erm*(B) and *erm*(F) exhibited the greatest abundances in nearly all the sample types (Fig. 4 and 5). This is in concordance with cultivation-based studies that revealed the distribution of *erm*(B) and *erm*(F) genes in the highest numbers of bacterial genera (21 and 20 genera, respectively) (30). Moreover, *erm*(B) was also found in 95% of the erythromycin-resistant enterococci isolated from three swine farms (17). Hence, the high abundance of these two *erm* genes is associated, or at least concurrent, with their wide occurrence in a large number of bacterial populations. The finding that these two *erm* genes often reside on mobile genetic elements (31) may explain, at least partially, their wide distribution and high abundance in microbiomes. Of interest is the high prevalence (Fig. 3) and abundance (Fig. 4E) of *erm*(T) found in the swine manure samples. This gene was first identified in a *Lactobacillus* strain from swine manure in 2001 (43) and has been identified only in *Streptococcus pasteurianus* (41) and *Streptococcus bovis* (38) as well as on plasmids p121BS in a *Lactobacillus* sp. (43), pLME300 in *Lactobacillus fermentum* (14), and pGT633 in *Lactobacillus reuteri*. Since none of these species is typically predominant in manure, other species may be the host for the *erm*(T) gene. Further studies are needed to confirm whether the *erm*(T) gene is universally abundant in swine manures. It also remains to be determined why only swine manures had large *erm*(T) gene reservoirs. On the other end of the spectrum, *erm*(C) exhibited the lowest abundance in all of the five types of samples. Its wide occurrence (so far found in 16 bacterial genera) (http://faculty .washington.edu/marilynr/ermweb4.pdf) and frequent residence on mobile genetic elements (31) seem to contradict its low abundance in these samples of animal origin.

All of the swine manure samples had significantly greater abundances of all six classes of *erm* genes than the bovine manure samples, either in absolute terms or in relative terms (Fig. 4 and 5). This is consistent with the previous finding with respect to *tet* genes present in these samples. The use of antibiotics (including tetracycline and erythromycin) in these swine farms was suggested to be the main contributing factor (47), but further studies are required, perhaps through the examination of conventional and organic swine farms, to determine if these differences can be attributed solely to differences in the use of erythromycin (or tylosin) or if differences in community composition in fecal microflora also play a role. The preliminary results also suggest that treatment of hog house effluents by either the EUB system or the lagoons tends not to appreciably reduce the *erm* gene reservoirs. These findings are consistent with the previous study of *tet* gene abundance (47) and corroborate a recent report (13) describing the low efficacy of eliminating erythromycin-resistant enterococci in an urban wastewater treatment plant. More in-depth studies are required to assess reduction and dissemination of AR in various types of wastewater treatment facilities and to identify potential factors that can affect the reduction and dissemination of AR during the treatment processes.

AR derived from food animals disseminates into the environment primarily via application and discharge of animal manures. Therefore, AR dissemination to the environment can be prevented or minimized through adequate and proper treatment and management of animal manures. In our previous study (47), we found that composted swine manure samples had substantially reduced *tet* gene reservoirs, whereas the samples from the lagoons and the EUB system did not. Interestingly, the *erm* genes quantified in this study also showed such a trend (Fig. 4). Thus, the results of these two independent studies tend to support our previous hypothesis that composting can effectively reduce AR to a variety of antimicrobials (47). However, it remains to be determined why composting, but not using lagoons or the Ekokan system, can effectively reduce AR reservoirs in animal manures. Inactivation of anaerobic *erm*-carrying bacteria during the composting process, which is largely aerobic and often has a thermophilic phase, and the low efficiency of lateral gene transfer in a solid-compost matrix to bacteria that can survive the composting process may collectively contribute to effective AR reduction during composting.

The prevalence of AR to a particular antimicrobial has been exclusively reported as a percentage of resistant isolates. The prevalences of individual resistant bacterial species obtained from swine farms are rather high and vary considerably. For instance, approximately 55% of the *Campylobacter coli* isolates (26, 39) and up to 98% of the airborne enterococcus isolates (7) from conventional swine farms were found to be resistant to erythromycin. Jackson et al. (17) also found that 95% of the erythromycin-resistant enterococci isolated from three swine farms carry *erm*(B). The relative abundances of all six classes of *erm* genes as determined in this study were much lower (less than 43%). These results further indicate that resistance prevalence likely varies among different bacterial species within a microbiome and imply that prevalence data obtained from specific cultivated species probably are not reflective of the overall prevalence in entire microbiomes.

Among the five different types of samples analyzed, patterns and/or magnitudes of differences differed between the relative and the absolute abundance measurements (Fig. 4 and 5). There are at least two possible explanations for such a disparity. First, because total bacterial abundance in the samples also affects the relative *erm* gene abundance data, any difference in total bacterial abundance will change the relative *erm* gene abundance. Second, different types of samples may harbor different bacterial populations carrying different *erm* genes, conceivably possessing different ecologies, and such differences could contribute to the incongruity observed between absolute

and relative *erm* gene abundance. Further studies are needed to test the latter hypothesis.

Interestingly, we noticed rather similar patterns of abundance between the *tet* genes determined previously (47) and the *erm* genes determined in this study among these sets of samples. Multidrug resistance is common among resistant isolates due to the occurrence of multiple resistance genes on the same mobile elements (27, 33, 37). It has been shown that the *erm*(F) gene is often associated with conjugative transposons and linked to  $tet(M)$ ,  $tet(Q)$ , and  $tet(X)$  (29), whereas  $erm(B)$  is often linked with *tet*(M) on Tn*917*-like conjugative transposons (30, 33) and with *tet*(Q) (29). Some staphylococci strains were also found to have both *tet* and *erm* genes (5). Although staphylococci are not likely predominant microbes in the samples analyzed, this type of linkage between *tet* and *erm* genes may also exist in other microbes. We postulate that the similar patterns of abundance observed between *tet* and *erm* genes in these five types of samples are probably attributable, at least partially, to the physical linkage between these two types of resistant genes and/or the carriage of these genes by the same bacteria. The results of this study provided some preliminary community-level clues that *erm* and *tet* genes, and maybe other AR genes, may be linked together and/or carried by the same bacteria, so they can exhibit similar dynamics in microbial communities. Consequently, reduction of one type of AR by a specific manure treatment may indicate the reduction of other types of AR present in the manure. Further studies are needed to test this hypothesis.

In conclusion, we developed and validated real-time PCR assays that can be used to accurately quantify the reservoirs of six classes of *erm* genes present in manure, compost, lagoon, and bioreactor samples. Our preliminary data suggest that different *erm* genes may have different reservoir sizes in microbial communities and different methods of manure treatments may have different efficiencies in reducing *erm* gene abundance. These should be evaluated in greater detail so that effective mitigation strategies can be developed to reduce dissemination of AR originating from food animals. Additionally, AR prevalence determined from bacterial isolates probably does not reflect the overall prevalence or abundance in the microbial community where the isolates are isolated.

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