Bacteriophages Active against Bacteroides fragilis in Sewage-Polluted Waters

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Twelve strains of different Bacteroides species were tested for their efficiency of detection of bacteriophages from sewage. The host range of several isolated phages was investigated. The results indicated that there was a high degree of strain specificity. Then, by using Bacteroides fragilis HSP 40 as the host, which proved to be the most efficient for the detection of phages, feces from humans and several animal species and raw sewage, river water, water from lagoons, seawater, groundwater, and sediments were tested for the presence of bacteriophages that were active against B. fragilis HSP 40. Phages were detected in feces of 10% of the human fecal samples tested and was never detected in feces of the other animal species studied. Moreover, bacteriophages were always recovered from sewage and sewage-polluted samples of waters and sediments, but not from nonpolluted samples. The titers recovered were dependent on the degree of pollution in analyzed waters and sediments.

Over 100 different types of virus occur in domestic sewage (21). The survival and transmission of viruses in sewagecontaminated water have been recognized as potential public health problems (4, 21, 29, 30). Currently available technology for virus isolation from water has the following shortcomings. Practical methods for the detection of viruses of primary concern such as hepatitis A virus, rotavirus, and Norwalk virus are not available; and techniques for both the recovery and the identification of human enteric viruses have limited sensitivity, are time-consuming and expensive, and require highly skilled labor and sophisticated laboratory facilities (30).

Although in some situations various bacterial indicator systems have proved to be useful for assessing the virological quality or safety of a wide variety of waters (11, 24, 30), recent epidemiological and microbiological findings have raised serious concerns about the adequacy of fecal indicator bacteria for predicting the virological quality of water (6, 7, 13, 20, 22, 23, 37, 39). Therefore, indicators more closely related to enteric viruses would be preferable for the evaluation of virological water quality (5, 41). Bacteriophages may play a valuable role in this regard because their structure, composition, morphology, and size closely resemble those of enteric viruses. Several investigators have proposed the use of coliphages as virus indicators (3, 25, 35, 49-52). Replication of coliphages in estuary water during the summer months (54) and the existence of many coliphages whose natural habitats appear to be the aquatic environment (48), however, represent the serious shortcomings of coliphages as indicators of enteric viruses.

In this study the potential of bacteriophages of Bacteroides spp. as indicators of human enteric viruses was examined. Bacteroides ssp. are strict anaerobes and are a major component of human feces (19, 45). It is unlikely that they occur naturally in the environment, and consequently, the phages infecting them have the potential for being good virus indicators. The purpose of this study, therefore, was to investigate the presence of bacteriophages of Bacteroides spp. in human feces, sewage, and sewage-contaminated waters.

MATERIALS AND METHODS

Bacterial strains. Twenty-five strains of Bacteroides were used in this study. Strains HSP 39, HSP 40, and HSP 44 were obtained from the Microbiology Service at the Hospital de Sant Pau (Barcelona, Spain). They were isolated from peritoneal exudates and were classified as Bacteroides fragilis according to substrate utilization, products of fermentation determined by gas-liquid chromatography, growth in 20% bile, and indole production (12). B. fragilis C-17898, C-13970, A-22055, C-19075, C-13700-2, C-6814, C-13182, C-1416, C-8797, C-14021, and C-13525 were isolated from clinical specimens; and strains BBE-18 and B-1 were isolated from feces and were kindly provided by Garcia Rodriguez (Department of Microbiology, Faculty of Medicine, University of Salamanca, Salamanca, Spain). The remaining strains B. hypermegas DSM-1672, B. thetaiotaomicron DSM 2079, B. vulgatus DSM 1447, B. fragilis DSM 2151, and B. ovatus DSM ¹⁸⁹⁶ were obtained from the German Collection of Microorganisms; B. distasonis ATCC 8503, B. eggerthii ATCC 27754, B. uniformis ATCC 8492, and B. microfusus ATCC ²⁹⁷² were obtained from the American Type Culture Collection (Rockville, Md.).

Media and growth conditions. Modified blood agar base (MBAB) contained the following per liter of distilled water: 40 g of blood agar base no. 2 (Oxoid Ltd., London, England), 0.05 g of L-cysteinium chloride monohydrate (E. Merck AG, Darmstadt, Federal Republic of Germany), 0.12 g of $MgSO_4 \cdot 7H_2O$ (Merck), and 0.05 g of CaCl₂ (Merck). Modified brucella broth (MBB) contained the following per liter: 28 g of brucella broth (ADSA, Barcelona, Spain), 0.05 g of L-cysteinium chloride monohydrate, 0.12 g of MgSO₄. $7H₂O$, 0.05 g of CaCl₂. Soft agar for the overlay technique was prepared by adding ⁵ g of agar per liter to MBB. MBAB-S was MBAB containing 0.05% potassium sorbate and was adjusted to pH 5.7 to avoid spore germination.

When solid media were used, cultures were incubated inside anaerobic jars (GasPak; BBL Microbiology Systems, Cockeysville, Md.). For incubation of liquid cultures, screwcap tubes or bottles filled with MBB were used; this allowed us to avoid the use of anaerobic jars.

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Sample	TABLE 1. Efficiency of different Bacteroides strains for the isolation of Bacteroides bacteriophages from sewage B. fragilis HSP 40	B. fragilis HSP 44	39 B. fragilis HSP	B. fragilis DSM 2151	ovatus DSM 1896 æ,	B. vulgatus DSM 1447	MPN of the following strains/100 ml ^a : hypermegas DSM 1672 \mathbf{B}	B. microfusus ATCC 29728	eggerthii ATCC 27254 \mathbf{g}	B. distasonis ATCC 8503	B. thetaiotaomicron DSM 2079	B. uniformis ATCC 8492

TABLE 1. Efficiency of different Bacteroides strains for the isolation of Bacteroides bacteriophages from sewage

 $a \rightarrow$, Not tested; 0, values of <3.

Collection and processing of samples. Surface water samples were collected in sterile glass bottles submerged about 20 cm below the water surface. Groundwater samples were taken from 20-m-deep wells. All samples were placed at 4°C and examined within 6 h after collection.

Sediment samples were collected with a Van Veen dredge. The sediment samples were carefully deposited on a clean surface. The top layer (2 to 3 cm) of sediment was scooped into sterile plastic bags, sealed, and stored at 4°C until they were processed. Samples were tested within 12 h of collection. Samples showing layer disruption were discarded. Then, 100 g of sediment was suspended in 100 ml of buffer $(Na_2HPO_4, 7 g; KH₂PO₄, 3 g; NaCl, 5 g; 0.1 M MgSO₄, 10$ ml; 0.01 M CaCl₂, 10 ml, distilled water to 1,000 ml) (14) for phages and vigorously shaken for 2 h at 4°C. The gross particles were allowed to sediment. Phages in the supernatant were counted by the most probable number (MPN) method, as described below for the water samples.

For feces analysis, fresh feces were collected in sterile bottles. Then, buffer for phages was added to make 0.1 g of fecal matter per ml of suspension. Then, phages were counted as described below for the water samples.

Sampling sites. Raw sewage was collected from a city sewage system, the average MPNs of fecal coliforms and fecal streptococci of which were 3.1×10^7 and 8.2×10^6 , respectively. Besos and Llobregat are two highly polluted, small rivers that flow through densely populated areas north and south of Barcelona. The mean MPNs of fecal coliforms and fecal streptococci per 100 ml were 3.08×10^7 and $4.22 \times$ 10⁶, respectively, for the Besos River and 8.62 \times 10⁶ and 1.32×10^6 , respectively, for the Llobregat River. Polluted seawater was collected from an area which receives a large amount of untreated sewage from Barcelona; the mean MPNs of fecal coliforms and fecal streptococci per ¹⁰⁰ ml were 1.3×10^4 and 7.2×10^3 , respectively. Numbers are mean values obtained twice a month over the previous 4 years. Contaminated sediments were collected either from the Besos River or from the coastal area from which seawater samples were taken.

Groundwater samples were taken from a 20-m-deep aquifer from which enteroviruses, fecal coliforms, and fecal clostridia have occasionally been isolated.

Water and sediment samples free of human pollution were obtained far $(>10 \text{ km})$ from any human settlement in the natural park Delta de l'Ebre. Water and sediment collected in eutrophic lagoons from the park were nearly anaerobic. Other samples were taken from nonpolluted mountain streams near Manlleu, which is 70 km northeast of Barcelona.

Phage assays. Enumerations of phages by plaque counts (PFU) were conducted by the double-agar layer technique (1). Tubes containing 2.5 ml of soft MBB agar kept at 45°C were inoculated with 10⁷ CFU of the indicator bacteria and a fraction not exceeding 1.0 ml of the sample to be analyzed. After mixing, the contents were poured onto a plate of MBAB and incubated inside GasPak (BBL) jars at 37°C. Plaques were counted after 18 h of incubation. When water samples were analyzed by this method, samples were filtered through membrane filters (pore size, $0.22 \mu m$; Millipore Corp., Bedford, Mass.) before they were mixed with the host.

Phages were also enumerated by a modification of the MPN method for coliphage described by Kott (34). Each of three screw-cap tubes (30 ml) containing 10 ml of 2.5-stretch MBB was inoculated with ¹⁰ ml of the sample and ¹⁰ ml of the host. A second set of three tubes containing ¹⁹ ml of single-strength MBB was inoculated with ¹⁰ ml of the host and ¹ ml of the sample, and a third set of three tubes containing ²⁰ ml of single-strength MBB was inoculated with 10 ml of the host and 0.1 ml of the sample. The tubes were incubated at 37°C for 36 to 48 h. Following incubation, 5 ml from each tube was centrifuged at 2,000 \times g for 15 min, and the supernatant was either mixed with 2 ml of chloroform or filtered through a 0.22 - μ m-pore-size membrane filter. Fi-

 $a +$, Host supportive of phage growth; $-$, host not supportive of phage growth.

^b Strains HSP 39, C-19075, C-13700-2, C-13182, C-1416, C-8797, C-14021, C-13525, BBE-18, and B-1.

^e Strains B. ovatus DSM 1896, B. vulgatus DSM 1447, B. hypermegas DSM 1672, B. microfusus ATCC 29728, B. eggerthii ATCC 27754, B. distasonis ATCC 85023, B. thetaiotaomicron DSM 2079, and B. uniformis ATCC 8492.

nally, a loopful from each tube was spotted onto overlayed host lawns plated on either MBAB-S when the sample was treated with chloroform or MBAB when the sample was filtered. Plates were incubated anaerobically at 37°C for 16 h. The results of the tests were recorded and computed by the use of the MPN table for coliform bacteria (26). Sewage and highly polluted samples were diluted, and series of three tubes were inoculated with 1 ml of each dilution into single-strength MBB tubes.

Phage strains and preparation of phage stocks. Phages were purified by three consecutive single-plaque isolations for the preparation of phage stocks. High-titer phage stocks were prepared by the soft-agar plate method described by Adams (1). Approximately 10^7 PFU were combined with 10^8 sensitive bacteria in tubes containing 2.5 ml of soft MBB agar and then poured onto MBAB plates. Following ¹⁶ ^h of anaerobic incubation at 37°C, the phages were harvested by adding 3 ml of the buffer for phages (described above) to each plate. After ¹ h at 4°C, the buffer was collected and centrifuged at $2,000 \times g$ for 15 min. The supernatant was then filtered through a 0.22 - μ m-pore-size membrane and stored at 4^oC.

TABLE 3. Comparison of phage counts obtained by the double-agar layer (PFU) and the MPN techniques

Sample	MPN upper limit	MPN/100 ml	MPN lower limit	PFU/100 ml
	3.3×10^{4}	3.5×10^{3}	6.4×10^{2}	1.0×10^3
\mathfrak{D}	3.8×10^{4}	1.1×10^{4}	1.5×10^{3}	1.0×10^{2}
3	2.4×10^{4}	4.6×10^{3}	7.1×10^{2}	8.0×10^{2}
4	4.4×10^{4}	1.5×10^{4}	3.0×10^{3}	3.0×10^{2}
5	1.3×10^{3}	2.4×10^{2}	36	
6	2.1×10^{5}	2.0×10^{4}	7.0×10^{3}	1.0×10^{2}
7	2.4×10^{4}	4.6×10^{3}	7.1×10^{2}	6.0×10^{2}
8	4.4×10^{5}	1.1×10^{5}	3.0×10^{4}	1.4×10^{3}
9	3.8×10^{3}	9.3×10^{2}	1.5×10^{2}	1.5×10^{2}
10			< 0.5	

Determination of host range. The Bacteroides strains were grown in MBB and overlayed on MBAB plates. Then, drops of phage stocks containing 10^8 to 10^9 PFU/ml were spotted onto the respective plates. The plates were incubated anaerobically at 37°C for 16 h, after which areas of phage lysis within the bacteria lawns were clearly visible.

RESULTS

Strain choice. A total of ¹³ sewage samples were tested for the presence of bacteriophages by using 12 different strains as indicator hosts (Table 1). In every sample, B. fragilis HSP 40 was the host that allowed for the greatest recovery of bacteriophages.

Host range. During this study 25 bacteriophages infecting B. fragilis HSP 40 were isolated, and their host range was assayed among 25 strains of different Bacteroides species. None of the bacteriophages were active against Bacteroides spp. other than B. fragilis (Table 2). Moreover, all of the bacteriophages had a rather restricted host range among the B. fragilis strains tested (Table 2).

To ascertain the identity of strain HSP 40, six bacteriophages were isolated on type strain B. fragilis DSM ²¹⁵¹ (ATCC 25285). Two of them were active against strain HSP 40. Both were added to the host range study. Moreover, one of the phages isolated on strain HSP 40 was active against strain DSM 2151.

Method optimization. Ten samples were analyzed for bacteriophages active against B . fragilis HSP 40 by using the double-agar layer technique to determine PFU and by the MPN method adapted to bacteriophages. The MPN method was more efficient than the double-agar layer technique

TABLE 4. Counts obtained following removal of residual bacteria by two methods after phage enrichment

		$MPN/100$ ml by:	
Sample	Method $1a$	Method 2 ^b	
	1.1×10^{4}	1.1×10^{4}	
2	2.4×10^{3}	2.1×10^{3}	
3	2.4×10^{2}	9.3×10^{2}	
4	6.4×10^{2}	1.2×10^{3}	
5	2.4×10^{2}	43	
6	43	23	
	23	23	
8	23	15	
9	93	15	
10	1.1×10^{3}	9.3×10^{2}	

^a Method 1, Treatment with chloroform and plating on MBAB-S.

^b Method 2, Filtering through a membrane (pore size, 0.22 μ m) and plating on MBAB.

TABLE 5. Bacteriophages active against B. fragilis HSP ⁴⁰ in feces of various animals

Animal species	No. tested	No. $(\%)$ of positive tests	
Human	40	4(10)	
Cow	40	0(0)	
Pig	50	0(0)	
Rabbit	21	0(0)	
Mouse	28	0(0)	
Hen	20	0(0)	
Ouail	10	0(0)	

(Table 3) with respect to both the counts of phages and the percentage of samples from which phages were recovered.

Because the purified bacteriophages described in Table 2 were not inactivated by chloroform and because the filtering of ^a fraction from each MPN tube was expensive and time-consuming, chloroform treatment and centrifugation were used instead of filters to obtain phage suspensions that were free of bacteria. This method, however, described by Adams (1) did not work because under anaerobic conditions bacterial growth occurred in the spots, even after treatment with chloroform. Bacteria surviving chloroform treatment were found to be anaerobic spore formers. To eliminate this problem, potassium sorbate was added to MBAB medium and the pH was lowered. An optimal concentration of potassium sorbate of 0.05% at pH 5.7 (MBAB-S medium) was determined. Growth of B. fragilis HSP 40 was uninhibited, while spore germination was prevented. Results in Table 4 indicate that after bacteriophage enrichment, treatment with chloroform and the use of MBAB-S medium gave results similar to those obtained by filtering through a 0.22 - μ m-pore-size membrane.

Bacteriophages in feces. Feces from humans and several homeothermic animals, which are more likely to contribute to water pollution, were analyzed for the presence of bacteriophages active against B. fragilis HSP 40 (Table 5). Bacteriophages were isolated from 4 of 40 humans studied. No phages were recovered from the feces of any of the animals tested. Successive surveys of two of the humans from whom phages were found in the feces revealed a variable pattern of phage excretion. Incidence and levels of phages in the feces ranged from 24 to 2.4 \times 10⁸/g in one and from 0 to more than 2.4×10^8 /g in the other.

Bacteriophages in environmental samples. A total of ¹⁶³ environmental samples were tested for the presence of bacteriophages active against B. fragilis HSP 40. Bacteriophages were consistently found in sewage-polluted waters and sediments (Table 6). They were detected in all the samples from raw sewage and highly polluted rivers, with

values that approached $10⁵$ MPN of bacteriophages per 100 ml and with average values of about $10⁴$ and a median count of 4.0×10^3 . The percentage of positive isolations decreased to 77.2% in seawater near Barcelona, where the greatest counts reached ¹⁰³ MPN per ¹⁰⁰ ml. Mean and median values were $10²$ and 4, respectively. Phages were isolated in 100% of river and 91% of marine sediments from the rivers and coastal area studied. Also, phages were detected in 4 of 19 groundwater samples, corresponding to an aquifer from which enteroviruses have occasionally been isolated; bacteriophages were isolated from an enrichment of 100 ml of groundwater.

Bacteriophages infecting B. fragilis HSP 40 were never recorded from samples of waters and sediments that had not received sewage pollution (Table 6).

DISCUSSION

One of the problems in detecting bacteriophages in the environment is the lack of an universal plating indicator for particular bacterial species. Moreover, some strains consistently give higher counts than others (15, 27). Our results are in good agreement with those of previous reports because B. fragilis HSP 40 consistently gave higher counts than the other Bacteriodes strains tested. Therefore, strain HSP 40 was used for the enumeration of bacteriophages in feces and environmental samples. The percentage of human feces samples from which phages were isolated and the phage numbers in positive samples were higher than those found by other investigators (8, 32, 40, 45). Therefore, strain HSP 40 seems to be an efficient indicator strain for the detection and enumeration of bacteriophages active against B. fragilis in environmental samples.

The host range of phages isolated by using B. fragilis HSP 40 as the host was very narrow. This high specificity of bacteriophages infecting Bacteroides spp. has already been observed by other investigators (8, 15, 33, 40). No polyvalent phages active against B . fragilis, which are so frequently observed with coliphages (17, 18, 43, 53), seemed to occur in Bacteroides spp. This high strain specificity may be the reason for the failure to isolate phages from the feces of the nonhuman animals tested. In addition, this high specificity may indicate that the similarity among strains of Bacteroides infecting different animal species is very low. It can therefore be concluded that bacteriophages active against B. fragilis are good indicators of human fecal pollution. Moreover, according to the narrow host range shown by the majority of bacteriophages infecting Bacteroides spp., results of this study confirm the classification of strain HSP 40 as B. fragilis.

Although the recovery of Bacteroides phages from raw sewage has already been reported (8, 15, 40), no data have

TABLE 6. Levels of bacteriophages active against B. fragilis HSP ⁴⁰ in water and sediments

Samples	No. of samples	% Samples positive for phages	Maximum value/100 ml	Minimum value/100 ml	Mean value/100 ml	Median value/100 ml	
Sewage	33	100	1.1×10^{5}		6.2×10^{3}	1.1×10^3	
River water ^{a}	22	100	1.1×10^{5}	93	1.6×10^{4}	4.6×10^{3}	
River sediment ^a		100	4.6×10^{5}	90	1.08×10^{5}	4.3×10^{3}	
Seawater ^a	22	77.2	1.1×10^3	\leq 3	1.2×10^{2}		
Marine sediment ^a	12	91.0	43	$<$ 3	13.4		
Groundwater ^a	19	21.0	NK ^c				
Nonpolluted ^b	50	0					

^a Samples from areas with sewage pollution.

 b Water and sediments from areas without known sewage pollution.</sup>

 c NK, Not known, because we only tested whether they were present in 100 ml.

been available on levels of recovery. By using the MPN method, a method which consistently gave values significantly higher than those by PFU counts, bacteriophages active against B. fragilis HSP ⁴⁰ were isolated from all sewage samples tested. The values, which ranged from ⁷ to 1.1×10^5 MPN of phages per 100 ml, were significantly higher than those reported for human enteric viruses in sewage around the world, which have ranged from 0 to 2.0 \times $10³$ per 100 ml (2, 9, 13, 28, 31, 36, 38, 42, 44, 46, 47, 55), and lower (about 2 log units) than those found for coliphages in sewage (16, 25, 35). Although the percentage of humans excreting bacteriophages that are active against B. fragilis HSP 40 is low, which is in agreement with data from other investigators (31), the amounts excreted are so high that they can account for the values recovered from sewage. Moreover, they were isolated from many different types of sewage-polluted waters and sediments, although either the percentage of positive samples or the phage counts decreased as we moved further away from the pollution source. On the contrary, phages were never detected in waters or sediments without known sewage pollution.

These data indicate that bacteriophages active against B. fragilis can be considered human sewage indicators, a concept that has been proposed by other investigators for coliphages (10). Further studies about the behavior of B. fragilis bacteriophages in the environment need to be done to ascertain whether they deserve to be considered as potential surrogate indicators of human viruses in the environment.

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LITERATURE CITED

- 1. Adams, M. H. 1959. Bacteriophages. John Wiley & Sons, Inc., New York.
- 2. Agbalika, F. P., P. Hartemann, and J. M. Foliguet. 1984. Trypsin treated MalO4: a sensitive cell line for isolating enteric viruses from environmental samples. Appl. Environ. Microbiol. 47:378-380.
- 3. Ayres, P. A. 1977. Coliphages in sewage and the marine environment, p. 275-298. In F. A. Skinner and J. M. Shewan (ed.), Aquatic microbiology. Academic Press, Inc., New York.
- 4. Berg, G. (ed.). 1967. Transmission of viruses by the water route. John Wiley & Sons, Inc., New York.
- 5. Berg, G. (ed.). 1978. The indicator system, p. 1-13. In Indicators of-yiruses in water and food. Ann Arbor Science Publishers, Inc., Ann Arbor, Mich.
- 6. Berg, G., D. R. Dahling, G. A. Brown, and D. Berman. 1978. Validity of fecal coliforms, total coliforms, and fecal streptococci as indicator of viruses in chlorinated primary sewage effluents. Appl. Environ. Microbiol. 36:880-884.
- 7. Berg, G., and T. G. Metcalf. 1978. Indicators of viruses in waters, p. 267-296. In G. Berg (ed.), Indicators of viruses in water and food. Ann Arbor Science Publishers, Inc., Ann Arbor, Mich.
- 8. Booth, S. J., R. L. Van Tassell, J. L. Johnson, and T. D. Wilkins. 1979. Bacteriophages of Bacteroides. Rev. Infect. Dis. 1:325- 334.
- 9. Buras, N. 1976. Concentration of enteric viruses in wastewater and effluent. A two year survey. Water Res. 10:295-298.
- 10. Cabelli, V. J. 1978. New standards for enteric bacteria, p. 223-271. In R. Mitchell (ed), Water pollution microbiology.

John Wiley & Sons, Inc., New York.

- 11. Cabelli, V. J., A. P. Dufour, L. J. MacCabe, and M. A. Levin. 1982. Swimming-associated gastroenteritis and water quality. Am. J. Epidemiol. 115:606-616.
- 12. Cato, E. P., and J. L. Johnson. 1976. Reinstatement of sp. rank for B. fragilis, B. ovatus, B. distasonis, B. thetaiotaomicron, and B. vulgatus: designation of neotype strains for B. fragilis. Int. J. Syst. Bacteriol. 26:230-237.
- 13. Cliver, D. 0. 1975. Viruses associated with wastewater solids. Environ. Lett. 10:215-223.
- 14. Clowes, R. C., and W. Hayes (ed.). 1968. Experiments in microbial genetics, p. 188. Blackwell Scientific Publications, Ltd., Oxford.
- 15. Cooper, S. W., E. G. Szymczak, N. V. Jacobus, and F. P. Tally. 1984. Differentiation of Bacteroides ovatus and Bacteroides thetaiotaomicron by means of bacteriophage. J. Clin. Microbiol. 20:1122-1125.
- 16. Dhillon, T. S., Y. S. Chan, S. M. Sun, and W. S. Chau. 1970. Distribution of coliphages in Hong Kong sewage. Appl. Microbiol. 20:187-191.
- 17. Dhillon, T. S., and E. K. S. Dhillon. 1972. Studies on bacteriophage distribution. II. Isolation and host range based classification of phages active on three species of Enterobacteriaceae. Jpn. J. Microbiol. 16:297-306.
- 18. Dhillon, T. S., E. K. S. Dhillon, H. C. Chau, W. K. Li, and A. H. C. Tsang. 1976. Studies on bacteriophage distribution. Virulent and temperate bacteriophage content of mammalian feces. Appl. Environ. Microbiol. 32:68-74.
- 19. Geldreich, E. E. 1978. Bacterial populations and indicator concepts in feces, sewage, stormwater and solid wastes, p. 51-97. In G. Berg (ed.), Indicators of viruses in water and food. Ann Arbor Science Publishers, Inc., Ann Arbor, Mich.
- 20. Gerba, C. P., S. M. Goyal, N. L. Labelle, I. Cech, and G. F. Bogdan. 1979. Failure of indicator bacteria to reflect the occurrence of enterovirus in marine waters. Am. J. Public Health 69:1116-1119.
- 21. Gerba, C. P., C. Wallis, and J. L. Melnick. 1975. Viruses in water: the problem, some solutions. Environ. Sci. Technol. 9:1122-1126.
- 22. Goyal, S. M. 1983. Indicators of viruses, p. 211-230. In G. Berg (ed.), Viral pollution of the environment, vol. 1. CRC Press, Inc., Boca Raton, Fla.
- 23. Goyal, S. M., W. N. Adams, M. L. O'Malley, and D. W. Lear. 1984. Human pathogenic viruses at sewage sludge disposal sites in the middle Atlantic region. Appl. Environ. Microbiol. 48:758-763.
- 24. Grabow, W. 0. K., J. S. Burger, and E. M. Nupen. 1980. Evaluation of acid-fast bacteria, Candida albicans, enteric viruses and conventional indicators for monitoring wastewater reclamation systems. Prog. Water Technol. 12:803-817.
- 25. Grabow, W. 0. K., P. Coubrough, E. M. Nupen, and B. W. Bateman. 1984. Evaluation of coliphages as indicators of the virological qualdity of sewage-polluted water. Water S.A. 10:7-14.
- 26. Greenberg, A., et al. (ed.). 1975. Standard methods for the examination of water and wastewater. American Public Health Association, Washington, D.C.
- 27. Havelaar, A. H., and W. H. Hogeboom. 1983. Factors affecting the enumeration of coliphages in sewage and sewage polluted waters. Antonie van Leeuwenhoek 49:387-397.
- 28. Hejkal, T. W., E. M. Smith, and C. P. Gerba. 1984. Seasonal occurrence of rotavirus in sewage. Appl. Environ. Microbiol. 47:588-590.
- 29. Hoadley, A. W., and S. M. Goyal. 1976. Public health implications of the application of wastewater to land, p. 101-132. In R. L. Sanks and T. Asans (ed.), Land treatment and disposal of municipal and industrial wastewater. Ann Arbor Science Publishers, Inc., Ann Arbor, Mich.
- 30. IAWPRC Study Group of Water Virology. 1983. The health significance of viruses in water. Water Res. 17:121-132.
- 31. Irving, L. G., and F. A. Smith. 1981. One year survey of enteroviruses, adenoviruses and reoviruses isolated from ef-

fluent of an activated sludge purification plant. Appl. Environ. Microbiol. 41:51-59.

- 32. Kai, S., S. Watanabe, K. Furune, and A. Ozawa. 1985. Bacteroides bacteriophages isolated from human faeces. Microbiol. Inmunol. 29:895-899.
- 33. Keller, R., and N. Traub. 1974. The characterization of a Bacteroides fragilis bacteriophage recovered from animal sera: observations on the nature of Bacteroides bacteriophage carrier culture. J. Gen. Virol. 24:179-189.
- 34. Kott, Y. 1966. Estimation of low number of E. coli bacteriophage by use of the most probable number method. Appl. Microbiol. 14:141-144.
- 35. Kott, Y., N. Roze, S. Sperber, and N. Betzer. 1974. Bacteriophages as viral pollution indicators. Water Res. 8:165-177.
- 36. Lucena, F., A. Bosch, R. Girones, and J. Jofre. 1983. Virus enterics a aigues urbano-industrials. Treb. Soc. Cat. Biol. 35: 71-79.
- 37. Lucena, F., M. Finance, J. Sancho, J. Jofre, and L. Schwarzbrod. 1982. Viral pollution determination of superficial waters (river water and sea water) from the urban area of Barcelona (Spain). Water Res. 16:173-177.
- 38. Martins, M. T., L. A. Soares, E. Marques, and A. G. Moline. 1983. Human enteric virus isolated from influents of sewage treatment plants in S. Paulo. Water Sci. Technol. i5:69-73.
- 39. Marzouk, Y., S. M. Goyal, and C. P. Gerba. 1980. Relationship of viruses and indicator bacteria in water and wastewater of Israel. Water Res. 14:1585-1590.
- 40. Nacescu, N., H. Brandis, and H. Werner. 1972. Isolation of two Bacteroides fragilis phages from sewage and detection of lysogenic B. fragilis strains. Zentralbl. Bakteriol. Parasitenkd. Infektionskr. Hyg. Abt. ¹ Orig. Reihe A 219:522-529.
- 41. 0. M. S. 1979. Les virus humains dans ^l'eau, les eaux usees et le sol. Rapport d'un group scientifique no. 639. O.M.S. Geneva, Switzerland.
- 42. Payment, P., R. Ayache, and M. Trudel. 1983. A survey of enteric viruses in domestic sewage. Can. J. Microbiol. 29: 111-119.
- 43. Primrose, S. B., N. D. Seeley, K. B. Logan, and J. W. Nicolson. 1982. Methods for studying aquatic bacteriophage ecology. Appl. Environ. Microbiol. 43:694-701.
- 44. Ruiter, G. G., and R. Fujioka. 1978. Human enteric viruses in sewage and their discharge into the ocean. Water, Air, Soil Pollut. 10:95-103.
- 45. Salyers, A. A. 1984. Bacteroides of the human lower intestinal tract. Annu. Rev. Microbiol. 38:293-313.
- 46. Sattar, S. A., and J. C. N. Westwood. 1976. Comparison of four eluents in the recovery of indigenous viruses from raw sewage. Can. J. Microbiol. 22:1586-1589.
- 47. Schwartzbrod, L., F. Lucena, and C. Finance. 1979. Etude quantitative de la population virable dans ^l'afluent d'une station d'epuration d'eaux residuaires. J. Fr. Hydrol. 10:7-20.
- 48. Seeley, N. D., and S. B. Primrose. 1980. The effect of temperature on the ecology of aquatic bacteriophages. J. Gen. Virol. 46: 87-95.
- 49. Simková, A., and J. Cěrvenira. 1981. Coliphages as ecological indicators of enteroviruses in various water systems. Bull. W.H.O. 59:611-618.
- 50. Søgaard, H. 1983. Relationship between the occurrence of coliphages and E.coli in Danish marine bathing areas. Zbl. Bakt. Hyg. I. Abt. Orig. B. 177:394-401.
- 51. Stetler, R. E. 1984. Coliphages as indicators of enteroviruses. Appl. Environ. Microbiol. 47:319-324.
- 52. Suner, I., and I. Pinol. 1966. Coliform bacteriophages and marine water contamination. Adv. Water Pollut. Res. 3:105- 111.
- 53. Tomas, J.+ M. Regue, R. Pares, J. Jofre, and W. W. Kay. 1984. P1 bacteriophage and tellurite sensitivity in Klebsiella pneumoniae and Escherichia coli. Can. J. Microbiol. 30:830-836.
- 54. Vaughn, J. M., and T. G. Metcalf. 1974. Coliphages as indicators of enteric viruses in shellfish arnd shellfish raising estuarine waters. Water Res. 8:613-616.
- 55. Wellings, F. M., A. L. Lewis, C. W. Mountain, and L. V. Pierce. 1975. Demonstration of virus groundwater after effluent discharge onto soil. Appl. Environ. Microbiol. 29:751-757.