Relative Abundances of Proteobacterial Membrane-Bound and Periplasmic Nitrate Reductases in Selected Environments⁷†

D. Bru, A. Sarr, and L. Philippot*

INRA, University of Burgundy, Soil and Environmental Microbiology, CMSE, 17 rue Sully, B.P. 86510, 21065 Dijon Cedex, France

Received 21 March 2007/Accepted 3 July 2007

Dissimilatory nitrate reduction is catalyzed by a membrane-bound and a periplasmic nitrate reductase. We set up a real-time PCR assay to quantify these two enzymes, using the *narG* **and** *napA* **genes, encoding the catalytic subunits of the two types of nitrate reductases, as molecular markers. The** *narG* **and** *napA* **gene copy numbers in DNA extracted from 18 different environments showed high variations, with most numbers ranging from 2** \times 10² to 6.8 \times 10⁴ copies per ng of DNA. This study provides evidence that, in soil samples, the number **of proteobacteria carrying the** *napA* **gene is often as high as that of proteobacteria carrying the** *narG* **gene. The high correlation observed between** *narG* **and** *napA* **gene copy numbers in soils suggests that the ecological roles of the corresponding enzymes might be linked.**

Nitrate in the environment can be either assimilated by plants and microorganisms or reduced to nitrite by one of two microbial dissimilatory processes: denitrification or dissimilatory reduction of nitrate to ammonium. Nitrate reduction by denitrification is of great importance since the produced nitrite is then reduced to N_2O or N_2 gases, which can lead to considerable nitrogen losses in agriculture and emissions of greenhouse gases (6, 13, 28). The reduction of nitrate present in contaminated water caused by the nitrate-reducing bacteria living in the human digestive tract is a potential health problem. As nitrite enters the bloodstream, it reacts with hemoglobin to form methemoglobin, blocking oxygen transport and causing a disease commonly called "blue baby syndrome" (26). Two types of dissimilatory nitrate reductase, differing in their locations, were characterized: a membrane-bound (Nar) and a periplasmic (Nap) nitrate reductase (2, 16, 29). The membrane-bound nitrate reductase is present in proteobacteria, firmicutes, actinobacteria, and even archaea, whereas the periplasmic nitrate reductase is present only in proteobacteria (21, 24). Nitrate-reducing proteobacteria can harbor Nar or Nap or both (18, 29). In contrast to that of Nar, the physiological role of Nap is still unclear and seems to differ between bacteria (11, 24). Thus, one proposed role for Nap is to support anaerobic metabolism as an alternative to Nar (1, 8). It has also been proposed that Nap facilitates the switch from aerobic respiration to denitrification (27) or scavenges nitrate in some pathogenic bacteria (23). The importance and diversity of the bacteria containing Nar have been extensively studied using both cultivation-based and direct molecular approaches (3, 5, 7, 17, 19, 22). However, only a few studies have focused on bacteria containing Nap (4, 9, 25). In this study, we investigated

Corresponding author. Mailing address: INRA, University of Burgundy, Soil and Environmental Microbiology, CMSE, 17 rue Sully, B.P. 86510, 21065 Dijon Cedex, France. Phone: 33 3 80 69 33 46. Fax: 33 3 80 69 32 24. E-mail: philippo@dijon.inra.fr.

the relative abundances of the two types of nitrate reductases in various environments, using a real-time PCR-based assay.

Primer design, standard curves, and real-time PCR procedures. In order to quantify the two types of nitrate reductases, a new real-time PCR assay was set up, using the *narG* and *napA* genes, encoding the catalytic subunits of the membranebound and periplasmic nitrate reductases, respectively, as molecular markers. All available sequences were aligned, and *narG* and *napA* primer sets specific to the proteobacteria were designed. The degenerated narG-f (5'-TCGCCSATYCCGGC SATGTC-3'), narG-r (5'-GAGTTGTACCAGTCRGCSGAY TCSG-3'), V17m (5'-TGGACVATGGGYTTYAAYC-3'; modified after reference 9), and napA4r (5--ACYTCRCGHGCVG TRCCRCA-3') primers were used to amplify fragments of 173 (for *narG*) and 152 (for *napA*) bp. Serial dilutions of linearized plasmids containing the *narG* and *napA* genes from *Pseudomonas aeruginosa* PAO1 were used to generate standard curves. The real-time PCR assays were carried out with a 20 - μ l reaction volume containing SYBR green PCR master mix (ABsoluteTM QPCR SYBR ROX Mix; Abgene, France), 2 M of each primer, 100 ng of T4 gp32 (QBiogene, France), and 1.25 μ l of template DNA (2 to 12.5 ng). Thermocycling conditions for *narG* were as follows: 15 min at 95°C; 6 cycles consisting of 30 s at 95°C and 30 s at 63°C, with a touchdown of -1 °C by cycle; and 40 cycles consisting of 30 s at 95°C, 30 s at 58°C, 30 s at 72°C, and 30 s at 80°C. Conditions for *napA* were similar except that the annealing temperature was set at 61°C. All real-time PCRs were performed with an ABI Prism 7900 sequence detection system (Applied Biosystems). Quantification of the 16S rRNA gene was performed as described previously (15). DNAs extracted from triplicate samples from 18 different environments, such as soils (agricultural, industrial, or glacier), river sediments, waters, or biofilms, were used as templates.

Evaluation of assay specificity and sensitivity. A functional gene pipeline interactive tool (http://flyingcloud.cme.msu.edu /fungene/) was used for in silico evaluation of primer specificity. Searches for the *narG* and *napA* primer sequences among 36 and 40 *narG* and *napA* sequences from complete genomes of proteobacteria showed that 78, 70, 86, and 86% did not

[†] Supplemental material for this article may be found at http://aem .asm.org/.
^{\sqrt{v}} Published ahead of print on 13 July 2007.

Strain or isolate	Phylum	Nitrate reductase activity	Nitrate reductase enzymes	O-PCR result for:	
				narG	napA
Bacillus senegalensis M1518	<i>Firmicutes</i>		ND, ND		
Bacillus cereus M944 ND	<i>Firmicutes</i>		Nar, ND		
Bacillus thuringiensis	<i>Firmicutes</i>		Nar, ND		
Bacillus megaterium	Firmicutes		Nar, ND		
Streptomyces bluensis	Actinobacteria		ND, ND		
Agrobacterium tumefaciens	Alphaproteobacteria		$-$, Nap		
Bradyrhizobium japonicum 562	Alphaproteobacteria		ND, ND		
Bradyrhizobium japonicum USDA 110	Alphaproteobacteria		$-$, Nap		$^{+}$
Hyphomicrobium denitrificans DSM 1869	Alphaproteobacteria		ND, ND		
Sinorhizobium morelense SN611	Alphaproteobacteria		ND, ND		
Sinorhizobium meliloti 1021	Alphaproteobacteria		$-$, Nap		
Rhizobium meliloti 50	Alphaproteobacteria		ND, ND		
Paracoccus denitrificans Pd1222	Alphaproteobacteria		Nar, Nap		
Alcaligenes faecalis ATCC 8750	Betaproteobacteria		ND, ND		
Alcaligenes eutrophus H16	Betaproteobacteria		Nar, Nap		$^{+}$
Achromobacter cycloclastes ATCC 21921	Betaproteobacteria		ND, ND		
Escherichia coli JM109	Gammaproteobacteria		Nar, Nap		
Pseudomonas denitrificans CCUG 2519	Gammaproteobacteria		ND, ND		
Pseudomonas fluorescens C7R12	Gammaproteobacteria	$^+$	Nar, $-$	$^+$	

TABLE 1. Bacterial strains used in this study to test the specificities of the *narG* and *napA* primers*^a*

a Q-PCR, quantitative PCR; ND, not determined; -, not present.

exhibit any mismatch with the narG-f, narG-r, V17m, and napA4r primers, respectively. Primer specificity was further confirmed experimentally using a collection of 19 strains (Table 1). Four firmicutes and one actinomycete nitrate-reducing strain were selected as negative controls, and 14 nitrate-reducing strains belonging to the alpha-, beta-, and gammaproteobacteria were selected as positive controls. None of the gram-positive nitrate reducers, which were used as negative controls, gave an amplicon. For *Proteobacteria*, an absence of PCR products with both *narG* and *napA* primers was recorded only with *Alcaligenes faecalis* ATCC 8750. Sequence analysis of 84 and 79 *narG* and *napA* real-time PCR products from agricultural soils (Côte Saint André and Yvetot), glacier soil (Rotmoosfermer), cave biofilm (Padirac), and river phototrophic biofilm (Garonne) revealed that all sequences were related to the *narG* or *napA* genes. A high level of diversity among the sequences of the real-time PCR products was observed, with identities as low as 67% for *narG* and 69% for *napA* to the sequences used for the designs of the primers (see Fig. S3 and S4 in the supplemental material). This indicates that our newly developed real-time PCR systems are suitable for general detection of proteobacterial nitrate reductase genes. However, 3 out of 84 *narG* sequences fell into a cluster containing only *narG* from actinobacteria, indicating that the designed *narG* primers were not entirely specific to proteobacteria.

The detection limit of our assay was around 10 copies per ng of template DNA, and no signal was detected in the no-template control. The PCR efficiencies of the *narG* and *napA* real-time PCR assays were 86 and 83%, respectively. Genomic DNA from *Pseudomonas aeruginosa* PAO1, for which the theoretical 16S rRNA, *narG*, and *napA* gene copy numbers per ng of DNA were calculated, was used as an external control. The presence of PCR inhibitors coextracted with DNA was tested as described previously (12), and this test did not reveal any significant inhibition. Therefore, our assays based on novel primer sets narG-f–narG-r and V17m–napA4r provide an efficient and sensitive method for quantifying either Nar or Nap in environmental samples.

Quantification of *narG* **and** *napA* **genes.** To compare with accuracy the numbers of genes in the different environments, results were expressed as gene copy numbers per ng of extracted DNA. Quantification of the *narG* and *napA* genes revealed high variations between environments, with most numbers ranging from 2×10^2 to 6.8×10^4 copies per ng of DNA (Fig. 3). The highest copy numbers for both *narG* and *napA* genes were observed in the river sediment samples. The *napA* gene copy numbers were lower than those of *narG* in most of the freshwater samples, whereas similar *narG* and *napA* gene copy numbers were observed in all soil samples except the Yvetot soil and the Rotmoosfermer glacier soil (Fig. 1). The numbers of 16S rRNA genes were between 1 and 3 logs higher than those of *narG* or *napA* genes. However, up to 12 copies of 16S rRNA may be found in the same bacterial genome (10), while only 1 *napA* copy and a maximum of 3 *narG* copies have been identified in the same strain (18). Hence, our study provides evidence that in soils, the numbers of proteobacteria containing periplasmic nitrate reductase are similar to the numbers of proteobacteria with the membranebound nitrate reductase. This is consistent with the results from Roussel-Delif et al. (25) and Carter et al. (4), which showed that a large proportion of nitrate-reducing, gram-negative isolates contain Nap. Unfortunately, it is not possible to conclude from our study whether the similar *narG* and *napA* copy numbers in many samples resulted from a majority of nitrate reducers possessing both types of nitrate reductase or from similar numbers of bacteria possessing either Nar or Nap. The *narG* genes from proteobacteria were mainly targeted in our assay, whereas these are also present in gram-positive bacteria and archaea, in contrast to *napA*. Therefore, the fact that the *narG* gene copy numbers were either similar to or higher than the *napA* gene copy numbers indicates that Nar is probably predominant in the environment.

FIG. 1. Abundance of 16S rRNA, *narG*, and *napA* genes expressed as gene copy numbers per ng of extracted DNA, analyzed from three independent replicates per site. Error bars indicate standard deviations. S indicates significant differences (*P* 0.05) between the *narG* and *napA* gene copy numbers (Student's *t* test) for each site.

Interestingly, a high correlation coefficient of 0.9 between *narG* and *napA* gene copy numbers was calculated for the soil and sediment samples, whereas this coefficient was only 0.4 for samples from river biofilm and water samples (Fig. 2). In contrast, the correlation between *narG* or *napA* and 16S rRNA gene copy numbers did not exceed 0.5. The physiological role of Nap is still unclear and probably differs between strains (11, 21, 24). The high correlation observed between the numbers of *narG* and *napA* genes but not with 16S rRNA suggests that the ecological role of Nap in the majority of soil proteobacteria might be similar or complementary to that of Nar. Our results also showed that the abundances of the two types of nitrate reductases differed between environments, which could be due to the selection of nitrate reducers in some habitats. Since detection of functional genes is only a weak hint of the presence of the corresponding activity (20), investigation of the relative contributions of the two types of nitrate reductases to the total nitrate reduction activity in the different environments is of interest. Unfortunately, the activities of the two types of nitrate reductase are simultaneously monitored by the nitrate reduction assay developed by Kandeler (14), and the activities of Nar and Nap can be distinguished only on bacterial

FIG. 2. Correlation between *narG* and *napA* gene copy numbers in the soil and sediment samples (white triangles) and in water and biofilm samples (black circles). Each replicate of the 18 selected environments is plotted.

isolates (4). In the future, integrated studies are needed to compile further information on the physiology, diversity, and distribution of nitrate reducers for a more comprehensive understanding of nitrate reduction in the environment.

Nucleotide sequence accession numbers. The GenBank accession numbers for the sequences reported here are EF217059 to EF217221.

We are grateful to A. Brauman, G. Braker, F. Garabetian, S. Hallin, E. Kandeler, S. Nazareth, and B. Stres for providing reference strains or environmental samples. We also thank S. Henry and S. Hallet for their assistance.

This work was supported by the French Ministry of Research (ACI PNBC MUTEN).

This work is dedicated to the memory of Frédéric Marmont and Ivan Mahne.

REFERENCES

- 1. **Bedzyk, L., T. Wang, and R. W. Ye.** 1999. The periplasmic nitrate reductase in *Pseudomonas* sp. strain G-179 catalyzes the first step of denitrification. J. Bacteriol. **181:**2802–2806.
- 2. **Berks, B. C., S. J. Ferguson, J. W. B. Moir, and D. J. D. Richardson.** 1995. Enzymes and associated electron transports systems that catalyse the respiratory reduction of nitrogen oxides and oxyanions. Biochim. Biophys. Acta **1232:**97–173.
- 3. **Brunel, B., J. D. Janse, H. J. Laanbroek, and J. D. Woldendorp.** 1992. Effect of transient oxic conditions on the composition of the nitratereducing community from the rhizosphere of *Typha angustipholia*. Microb. Ecol. **24:**51–61.
- 4. **Carter, J. P., Y. H. Hsaio, S. Spiro, and D. J. Richardson.** 1995. Soil and sediment bacteria capable of aerobic nitrate respiration. Appl. Environ. Microbiol. **61:**2852–2858.
- 5. **Che`neby, D., S. Hallet, M. Mondon, F. Martin-Laurent, J. C. Germon, and L. Philippot.** 2003. Genetic characterization of the nitrate reducing community based on *narG* nucleotide sequence analysis. Microb. Ecol. **46:**113–121.
- 6. **Conrad, R.** 1996. Soil microorganisms as controllers of atmospheric trace gases (H2, CO, CH4, OCS, N2O, and NO). Microbiol. Rev. **60:**609–640.
- 7. **Deiglmayr, K., L. Philippot, U. A. Hartwig, and E. Kandeler.** 2004. Structure and activity of the nitrate-reducing community in the rhizosphere of *Lolium perenne* and *Trifolium repens* under long-term elevated atmospheric pCO₂. FEMS Microbiol. Ecol. **49:**445–454.
- 8. **Delgado, M. J., N. Bonnard, A. Tressierra-Ayala, E. J. Bedmar, and P.** Müller. 2003. The *Bradyrhizobium japonicum napEDABC* genes encoding the periplasmic nitrate reductase are essential for nitrate respiration. Microbiology **149:**3395–3403.
- 9. **Flanagan, D. A., L. G. Gregory, J. P. Carter, A. Karakas-Sen, D. J. Richardson, and S. Spiro.** 1999. Detection of genes for periplasmic nitrate reductase in nitrate respiring bacteria and in community DNA. FEMS Microbiol. Lett. **177:**263–270.
- 10. **Fogel, G. B., C. R. Collins, J. Li, and C. F. Brunk.** 1999. Prokaryotic genome size and SSU rDNA copy number: estimation of microbial relative abundance from a mixed population. Microb. Ecol. **38:**93–113.
- 11. **Gonzalez, P. J., C. C. I. Moura, C. D. Brondino, and J. J. G. Moura.** 2006.

Bacterial nitrate reductases: molecular and biological aspects of nitrate reduction. J. Inorg. Biochem. **100:**1015–1023.

- 12. **Henry, S., D. Bru, B. Stres, S. Hallet, and L. Philippot.** 2006. Quantitative dectection of the *nosZ* gene, encoding nitrous oxide reductase, and comparison of the abundance of 16S rRNA, *narG*, *nirK*, and *nosZ* genes in soils. Appl. Environ. Microbiol. **72:**5181–5189.
- 13. **Hofstra, N., and A. F. Bouwman.** 2005. Denitrification in agricultural soils: summarizing published data and estimating global annual rates. Nutr. Cycl. Agroecosyst. **72:**267–278.
- 14. **Kandeler, E.** 1995. Nitrate reductase activity, p. 176–179. *In* F. Schinner, R. Öhlinger, E. Kandeler, and R. Margesin (ed.), Methods in soil biology. Springer, Berlin, Germany.
- 15. **Lopez-Gutierrez, J. C., S. Henry, S. Hallet, F. Martin-Laurent, G. Catroux, and L. Philippot.** 2004. Quantification of a novel group of nitrate-reducing bacteria in the environment by real-time PCR. J. Microbiol. Methods **57:** 399–407.
- 16. **Moreno-Vivian, C., P. Cabello, M. Martinez-Luque, R. Blasco, and F. Castillo.** 1999. Prokaryotic nitrate reduction: molecular properties and functional distinction among bacterial nitrate reductases. J. Bacteriol. **181:**6573– 6584.
- 17. **Nijburg, J., M. J. L. Coolen, S. Gerards, P. J. A. K. Gunnewiek, and H. J. Lannbroek.** 1997. Effects of nitrate availability and the presence of *Glyceria maxima* on the composition and activity of the dissimilatory nitrate-reducing bacteria community. Appl. Environ. Microbiol. **63:**931–937.
- 18. **Philippot, L.** 2002. Denitrifying genes in bacterial and Archaeal genomes. Biochim. Biophys. Acta **1577:**355–376.
- 19. **Philippot, L.** 2005. Tracking nitrate reducers and denitrifiers in the environment. Biochem. Soc. Trans. **33:**204–207.
- 20. **Philippot, L., and S. Hallin.** 2005. Finding the missing link between diversity and activity using denitrifying bacteria as a model functional community. Curr. Opin. Microbiol. **8:**234–239.
- 21. **Philippot, L., and O. Hojberg.** 1999. Dissimilatory nitrate reductases in bacteria. Biochim. Biophys. Acta **1446:**1–23.
- 22. **Philippot, L., S. Piutti, F. Martin-Laurent, S. Hallet, and J. C. Germon.** 2002. Molecular analysis of the nitrate-reducing community from unplanted and maize-planted soil. Appl. Environ. Microbiol. **68:**6121–6128.
- 23. **Potter, L., H. Angove, D. Richardson, and J. Cole.** 2001. Nitrate reduction in the periplasm of Gram-negative bacteria. Adv. Microb. Physiol. **45:**51–112.
- 24. **Richardson, D. J., B. C. Berks, D. A. Russel, S. Spiro, and C. J. Taylor.** 2001. Functional, biochemical and genetic diversity of prokaryotic nitrate reductase. Cell. Mol. Life Sci. **58:**165–178.
- 25. **Roussel-Delif, L., S. Tarnawski, J. Hamelin, L. Philippot, M. Aragno, and N. Fromin.** 2005. Frequency and diversity of nitrate reductase genes among nitrate-dissimilating Pseudomonas in the rhizosphere of perennial grasses grown in field conditions. Microb. Ecol. **49:**63–72.
- 26. **Shearer, L. A., J. R. Goldsmith, C. Young, O. A. Kearns, and B. R. Tamplin.** 1972. Methemoglobin levels in infants, in an area with high nitrate water supply. Am. J. Public Health **62:**1174–1180.
- 27. **Siddiqui, R. A., U. Warnecke-Eberz, A. Hengsberger, B. Schneider, and B. Friedrich.** 1993. Structure and function of a periplasmic nitrate reductase in *Alcaligenes eutrophus* H16. J. Bacteriol. **175:**5867–5876.
- 28. **Tiedje, J. M.** 1988. Ecology of denitrification and dissimilatory nitrate reduction to ammonium, p. 179–244. *In* A. J. B. Zendher. (ed.), Biology of anaerobic microorganisms. John Wiley & Sons, Inc., New York, NY.
- 29. **Zumft, W. G.** 1997. Cell biology and molecular basis of denitrification. Microbiol. Mol. Biol. Rev. **61:**533–536.