## Biodegradation of Polystyrene, Poly(methyl methacrylate), and Phenol Formaldehyde

DAVID L. KAPLAN,<sup>†</sup>\* ROY HARTENSTEIN, AND JIM SUTTER

State University of New York, College of Environmental Science and Forestry, Syracuse, New York 13210

Received for publication 15 June 1979

The biodegradation of three synthetic <sup>14</sup>C-labeled polymers, poly(methyl methacrylate), phenol formaldehyde, and polystyrene, was studied with 17 species of fungi in axenic cultures, five groups of soil invertebrates, and a variety of mixed microbial communities including sludges, soils, manures, garbages, and decaying plastics. Extremely low decomposition rates were found. The addition of cellulose and minerals failed to increase decomposition rates significantly.

Until recently, biodegradation of synthetic polymers had been assessed either by plate tests, with visual evaluation, or by observations on changes in physical characteristics of plastics placed into soil or aquatic systems (9, 13). The physical characteristics, such as hardness or tensile strength, may relate to shifts in conformation or bond breakage at vulnerable regions of macromolecules but do not necessarily provide evidence for biodegradation which would lead to assimilation and respiration. Similarly, the plating assays do not provide direct evidence for metabolic involvement. Neither type of assay is sufficiently sensitive to detect chemical decomposition processes which may be occurring at a very slow rate.

More recently, <sup>14</sup>C-labeled plastics have been synthesized to study their degradation (1, 6) and provide sensitivity better than 0.001%. Guillet et al. (6) studied the biodegradation of '4C-labeled polystyrene in soil before and after photodegradation. Tsuchii et al. (16) followed the microbial decomposition of styrene oligomers. In this study we examined a wide range of microbiological systems to determine whether a potential existed for degrading '4C-labeled polystyrene, poly(methyl methacrylate), and phenol formaldehyde.

Fungi were grown in 2.5% malt extract on orbital shakers at 28 to  $30^{\circ}$ C. They were harvested aseptically with cheesecloth and homogenized in 0.85% potassium chloride with a Waring blender. Suspensions, 1.0 ml, were added to 125-ml Erlenmeyer flasks containing  $0.045 \mu\text{Ci}$ of one of the polymers and either 2.0% malt extract or <sup>100</sup> mg of cellulose with <sup>10</sup> ml of fungal medium. The fungal medium consisted of 2.0 g of  $NH_4H_2PO_4$ , 0.6 g of  $KH_2PO_4$ , 0.4 g of  $K_2HPO_4$ , 0.3 g of Na<sub>2</sub>HPO<sub>4</sub>, 0.2 g of MgSO<sub>4</sub>, 0.7  $g$  of CaCl<sub>2</sub>, 12.0 mg of FeSO<sub>4</sub>  $\cdot$  7H<sub>2</sub>O, 6.6 mg ZnSO<sub>4</sub> $\cdot$  $7H<sub>2</sub>O$ , 5.0 mg of  $MnSO<sub>4</sub>·H<sub>2</sub>O$ , 1.0 mg of  $CoCl<sub>2</sub>·$  $6H<sub>2</sub>O$ , 1.0 mg of CuSO<sub>4</sub>  $\cdot$  5H<sub>2</sub>O, 0.1 mg of thiamine  $\cdot$ HCI, and 1.0 g of yeast extract in 1.0 liter of distilled water adjusted to pH 5.0 (modified from 2). Flasks were incubated at 28 to  $30^{\circ}$ C for 5 weeks without shaking. Vials (diameter, 5.5 by 1.5 cm) were suspended with wire from the lip of the flasks, which were sealed tightly with a rubber stopper. A strip of filter paper (3 by <sup>3</sup> cm) with 0.05 ml of <sup>20</sup> N NaOH was placed in each vial to trap  ${}^{14}CO_2$ . The strips were changed twice weekly during the 5-week period.

The fungi used were as follows. Coriolus hirsutus (Wolf. ex Fr.) Quel, Gloeophyllum trabeum (Pers. ex Fr.) Murr., Coriolus versicolor (L. ex Fr.), Poria placenta (Fr.) Cke., Bjerkandera adusta Willd. ex Fr., Daedalea quercina L. ex Fr., and Phellinus pini (Thore ex Fr.) A. Ames were obtained from Frances Lombard, Forest Products Laboratory, Madison, Wis.; Au-<br>reobasidium pullulans (DeBary) Arnaud. reobasidium pullulans (DeBary) Fomes annosus (Fr.) Cke., Peniophoragigantea (Fr.) Massee, Fomes everhartii (Ell. and Gell.) var. Schr. and Spauld., and Poria xantha (Fr.) Uce were obtained from R. Zabel of this College and are named as in Hepting (8); Aspergillus fumigatus Fres., Paecilomyces varioti Bainer, Trichoderma koningii Dud., Penicillium variable Sopp, and Aspergillus niger van Tiegh were obtained from R. Ziobro and C. J. Wang of this college.

Five groups of soil invertebrates were tested to determine whether they harbored procaryotic organisms capable of degrading one or more of the three plastics: an isopod (Isopoda), Oniscus asellus L.; a millipede (Diplopoda), Diploiulus sp.; a snail (Gastropoda), Oxychilus draparnaldi Beck; the slugs (Gastropoda) Limax maximus Linne and Deroceras reticulatum Muller;

t Present address: Environmental Protection Group, U.S. Army Natick Research and Development Command, Natick, MA 01760.

and the earthworms (Oligochaeta) Eisenia foetida (Savingny), Eudrilus eugeniae Kingberg, and a Pheretima spp. The animals were collected near the college and used on the day of their capture. Experiments were run for 14 days in 125-ml Erlenmeyer flasks and traps were changed on days 1, 2, 4, 7, 10, and 14. The invertebrates were fed  $0.045 \mu$ Ci of each polymer. The polymers were offered to the earthworms on 1.5 g of horse manue, to the isopod and millipede on a section (2 by 2 cm) of decayed box elder leaf (Acer negundo) and to the slugs and snail on a piece (2 by 2 cm) of lettuce (Lactuca sativa). These quantities of food and polymer were consumed within week <sup>1</sup> of testing.

Mixed microbial systems in the form of soils, manures, decaying plastics, and garbage were obtained locally and used fresh. Activated and anaerobic sludges were obtained from the Limestone-Meadowbrook and Ley Creek Wastewater Treatment Plants, Onondaga County, N.Y., and from the pulp sludge from a paper mill, respectively. Samples weighing 10 g were tested in 125 ml flasks containing  $0.045 \mu$ Ci of one of the polymers, with and without 100 g of cellulose and <sup>10</sup> ml of salts solution adjusted to pH 7.0. The salts solution consisted of 1.0 g of  $K_2HPO_4$ , 1.0 g of  $KH_2PO_4$ , 0.1 g of MgSO<sub>4</sub>, 0.1 g of NaCl, and 3.0 <sup>g</sup> of NH4H2PO in <sup>1</sup> liter of tap water. A new supply of cellulose and salts was added every 4 weeks during the 11-week test period. Flasks were flooded with 5 N  $H<sub>2</sub>SO<sub>4</sub>$  to release residual  ${}^{14}CO_2$  into the traps at the end of the experiment.

Phenol formaldehyde  $(0.0027 \mu\text{Ci/mg})$  was synthesized from <sup>14</sup>C-ring-labeled phenol (purchased from Amersham, Arlington Heights, Ill.) with formaldehyde in a stepwise polymerization<br>reaction (15). Poly(methyl methacrylate) methacrylate) (0.0095  $\mu$ Ci/mg) was synthesized by bulk radical polymerization with azobisisobutyronitrile from <sup>[14</sup>C]methyl methacrylate monomer (purchased from Amersham, Arlington Heights, Ill.) (15). Polystyrene (0.0017  $\mu$ Ci/mg) was synthesized by emulsion polymerization of [8-<sup>14</sup>C]styrene (purchased from Research Products Intern. Corp., Elk Grove Village, Ill.) with persulfate (15). All polymers were in granular form and less than 0.5 mm in diameter.

Radioactivity was measured in a Beckman LS 100C-liquid scintillation counter in a solution containing 5.0 <sup>g</sup> of PPO (2,5-diphenyloxazole), 0.4 <sup>g</sup> of POPOP [1-4-bis-(5-phenyloxazolyl)-benzene], 6.5 ml of monoethanolamine, 500 ml of toluene, and methanol to <sup>1</sup> liter. Each vial was corrected for quench with an external standard and for base line rates of  ${}^{14}CO_2$  release from sterile controls consisting of the polymers in 10 ml of distilled water. All fungal and mixed microbial systems were run with two or three repetitions, whereas invertebrates were run with four or five repetitions.

Fungi. Fungi in axenic cultures demonstrated very limited ability to degrade the polymers during 35 days. As a group, the 17 different fungi degraded from 0 to 0.29%, 0 to 0.17%, and 0 to 0.24% of poly(methyl methacrylate), phenol formaldehyde and polystyrene, respectively. A. pullulans, commonly associated with decaying acrylic paints (14, 19), converted only about 0.1% of the poly(methyl methacrylate) polymer to  ${}^{14}CO_2$  in 35 days. Despite the ability of many of these fungi to degrade the complex aromatic plant polymer lignin (3; D. L. Kaplan and R. Hartenstein, Siol Biol. Biochem., in press), they were unable to decompose the synthetic plastic polymers.

Invertebrates. Neither the eight soil invertebrates nor the microbes egested by them in their fecal pellets were able to degrade any of the polymers. Soil invertebrates are also unable to degrade lignin (12).

Mixed microbial communities. During <sup>11</sup> weeks in silt loam, cow manure, activated sludge, or decaying plastics (a mixture of plastics and adhering debris in various stages of decomposition collected from fields) or 5 weeks in anaerobic sludge, pulp mill sludge, horse manure, garbage, garden soil, or farm soil, all three polymers were highly recalcitrant to biological decay. Figure <sup>1</sup> presents data on the decomposition of polystyrene in the silt loam and in activated sludge from the aeration tanks. In the different microbial systems, total decomposition of the polymers during 5 or 11 weeks ranged from 0.04 to 0.57%, 0 to 0.15%, and 0 to 0.15% for polystyrene, phenol formaldehyde, and poly(methyl methacrylate), respectively.

It is clear that all three polymers are highly recalcitrant to biological decay. Polystyrene appeared to be destroyed most rapidly, but even here (Fig. 1) the trivial percentage decomposed in any system, together with the decomposition kinetics of a rapid early output of  ${}^{14}CO_2$  followed by virtual cessation of activity despite periodic renewal of nutrients, suggests that only oligomeric residues or trace impurities were destroyed. Guillet et al. (6) also found virtually no biodegradation of polystyrene and less than  $0.01\%$  <sup>14</sup>CO<sub>2</sub> during 8 weeks in garden soil. Rates and total quantities of  $^{14}CO_2$  production from poly(methyl methacrylate) and phenol formaldehyde in this study were even lower than for polystyrene. Albertsson and Ranby (1) found 0.005 to 0.10% degradation per month for polyethylenes. In comparison to these synthetic polymers, rates of decompositions of lignins in soils range from about 2 to  $15\%$  (4, 7, 11) during 1



FIG. 1. Decomposition of polystyrene in soil and activated sludge (aeration tank). Symbols: 0, soil;  $\bullet$ , soil with cellulose and salts;  $\triangle$ , sludge;  $\blacktriangle$ , sludge with cellulose and salts.

month and in axenic cultures of white-rot basidiomycetes up to 7% (Kaplan and Hartenstein, in press). The rates at which cellulose and chitin decompose vary with environmental circumstances (5), but cellulose in horse manure may decompose as rapidly as 2% per day (17), and the enzymatic activity of cellulose (18), like that of chitinase (10), is measurable by chemical rather than more sensitive radiochemical procedures in hours instead of weeks or months.

The results of this study, in which numerous heterogeneous microbial communities failed to effect biodegradation of the plastics tested, suggest that the evolution of the required biochemical catalysts within these microbial communities has not yet developed.

This study was supported by the National Science Foundation Research Applied to National Needs program.

## LITERATURE CITED

- 1. Albertsson, A. -C., and B. Ranby. 1975. Biodegradation of synthetic polymers: the C-14 method applied to polyethylenes, p. 743-751. Proceedings of the Third International Biodegradation Symposium. Applied Science Publishers, London.
- 2. Ander, P., and K. E. Ericksson. 1976. The importance of phenol oxidase activity in lignin degradation by the white-rot fungus Sporotrichum pulverulentum. Arch. Microbiol. 109:1-8.
- 3. Christman, R. F., and R. T. Oglesby. 1971. Microbio-

logical degradation and the formation of humus, p. 769- 798. In D. V. Sarkanen and C. H. Ludwig (ed.), Lignins occurrence, formation, structure, and reactions. Wiley-Interscience, New York.

- 4. Crawford, D. L, R. L. Crawford, and A. L. Pometto, HI. 1977. Preparation of specifically labeled '4C-(lignin) and '4C-(cellulose)-lignocelluloses and their decomposition by microflora of soil. Appl. Environ. Microbiol. 33:1247-1251.
- 5. Dickinson, C. H., and G. J. F. Pugh (ed.). 1974. Biology of Plant Litter Decomposition, vol. <sup>1</sup> and 2. Academic Press Inc., New York.
- 6. Guillet, J. E., T. W. Regulski, and T. B. McAneney. 1974. Biodegradability of photodegraded polymers. II. Tracer studies of biooxidation of Ecolyte PS polystyrene. Environ. Sci. Technol. 8:923-925.
- 7. Haider, K., J. P. Martin, and E. Rietz. 1977. Decomposition in soil of "C-labeled coumaryl alcohols; free and linked into dehydropolymer and plant lignins and model humic acids. Soil Sci. Soc. Am. J. 41:556-562.
- 8. Hepting, G. H. 1971. Diseases of forest and shade trees of the United States. Agriculture handbook 386, U.S. Department of Agriculture, Forest Service, Washington,  $D.C.$
- 9. Howard, P. H., J. Saxena, P. R. Durkin, and L-T. Ou. 1975. Review and evaluation of available techniques for determining persistence and routes of degradation of chemical substances in the environment. EPA-560/5- 75-006. U.S. Environmental Protection Agency, Washington, D.C.
- 10. Jeuniaux, C. 1966. Chitinases. Methods Enzymol. 8:644- 650.
- 11. Kirk, T. K., W. J. Connors, R. D. Bleam, W. F. Hackett, and J. G. Zeikus. 1975. Preparation and microbial decomposition of synthetic (<sup>14</sup>C)lignins. Proc. Natl. Acad. Sci. 72:2515-2519.
- 12. Neuhauser, E. F., R. Hartenstein, and W. J. Connors. 1978. The role of soil macroinvertebrates in the degradation of vanillin, cinnamic acid and lignins. Soil Biol. Biochem. 10:431-435.
- 13. Potts, J. E., R. A. Clendinning, W. B. Ackart, and W. D. Neighisch. 1973. The biodegradability of synthetic polymers. Polymer Sci. Technol. 3:61-79.
- 14. Schmitt, J. A., D. E. Padgett, and J. B. Achmoody. 1976. Laboratory and successional studies with Aureobasidium pullulans. J. Paint Technol. 48:35-42.
- 15. Sorenson, W. R., and T. W. Campbell. 1968. Preparative methods in polymer chemistry, 2nd ed. Interscience, New York. 504 pp.
- 16. Tsuchii, A., T. Suzuki, and Y. Takahara. 1977. Microbial degradation of styrene oligomer. Agric. Biol. Chem. 41:2417-2421.
- 17. Waksman, S. A., T. C. Cordon, and N. Hulpoi. 1939. Influence of temperature upon the microbiological processes in composts of stable manure. Soil Sci. 47:83-113.
- 18. Whitaker, D. R. 1971. Cellulases, p. 273-290. In P. D. Boyer (ed.), The enzymes, 3rd ed., vol. V. Academic Press Inc., New York.
- 19. Winters, H., I. R. Isguith, and M. Goll. 1975. A study of the ecological succession in biodeterioration of a vinyl acrylic paint film. Dev. Ind. Microbiol. 17:167-171.