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A Critical Comparison of Murine Pathology and Epidemiological Data of TCDD, PCB126, and PeCDF

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Abstract

2,3,7,8-Tetrachlorodibenzo-*p*-dioxin (TCDD, or dioxin) and dioxin-like compounds (DLCs) induce numerous toxicities, including developmental, endocrine, immunological, and multi-organ carcinogenic, in animals and/or humans. Multiple studies completed by the National Toxicology Program (NTP) focused on the effects caused in Harlan Sprague-Dawley rats by specific DLCs, among them the prototypical dioxin, TCDD. Because humans are exposed daily to a combination of DLCs, primarily via ingestion of food, the Toxic Equivalency Factor (TEF) was developed in order to evaluate health hazards caused by these mixtures. Herein we review the pathological effects reported in humans exposed to TCDD; 3,3′,4,4′,5-pentachlorobiphenyl (PCB 126); and 2,3,4,7,8, pentachlorodibenzofuran (PeCDF) and compare them to similar changes seen in NTP murine studies performed with the same compounds. While there were differences in specific pathologies observed, clear consistency in the target organs affected (liver, oral cavity, cardiovascular system, immune system, thyroid, pancreas, and lung) could be seen in both human studies and rodent toxicity and carcinogenicity investigations.

Keywords

Carcinogenesis; dioxin; human; PCB126; PeCDF; rodent; TCDD

Introduction

2,3,7,8-Tetrachlorodiobenzo-*p*-dioxin (TCDD, or dioxin) was classified in 1997 as a "Group 1 carcinogen (carcinogenic to humans, non-genotoxic carcinogen)" by the International Agency for Research on Cancer (IARC) ([http://www.iarc.fr/ENG/Databases/index.php\)](http://www.iarc.fr/ENG/Databases/index.php). Hazards associated with TCDD do not constitute a novel finding, because its toxicologic properties have been known since the 1950's (Watanabe et al., 1999). The top three sources

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Although the existence of low levels of TCDD and dioxins is widespread, only a limited number of human studies have been conducted to measure effects on various systems of the body due to the limited number of populations that have been highly exposed. TCDD became well known as a contaminant of the herbicide, Agent Orange, used in the Vietnam War and was found in Times Beach, Missouri; Love Canal, New York; and Seveso, Italy following an industrial explosion in 1976 (Bertazzi et al.*,* 2001; Schecter et al., 2006). Most recent studies indicate that exposure is associated with an increase in all cancers combined and several specific cancers including rectal cancer, lung cancer, Hodgkin's disease, non-Hodgkin's lymphoma, and myeloid leukemia (ATSDR, 1998; Bertazzi et al., 2001).

3,3′,4,4′,5-Pentachlorobiphenyl (PCB 126) was produced as a component of dielectric insulating fluid for transformers and capacitors prior to 1977 (NTP, 2004a). The creation and utilization of this chemical ceased because of residue found in the environment, although it is still discharged into the air and waterways by the disposal of objects that contain it (NTP, 2004a). Bioaccumulation of PCB 126 is prevalent in human and animal tissues with biological responses comparable to those evoked by TCDD (NTP, 2004a). Information on health effects of PCBs is available from studies of exposure in the workplace, consumption of contaminated rice oil in the Japanese "Yusho Accident" and the Taiwanese "Yu-Cheng Accident," consumption of contaminated fish, general environmental exposures, and food products of animal origin. Evidence suggests that PCBs can produce several toxicities, such as skin lesions, dental abnormalities, immune deficiency, and/or reproductive abnormalities, and increase the risk of developing cardiovascular and/or liver disease and diabetes (Carpenter, 2006; Schecter et al., 2006). PCB126 has been listed as a Group 2A carcinogen (probably carcinogenic to humans) by the IARC (<http://www.iarc.fr/ENG/Databases/index.php>).

2,3,4,7,8,-Pentachlorodibenzofuran (PeCDF) has been made intentionally, only for the use of scientific research, but is found in the environment due to discharge from sources of combustion and incineration (NTP, 2004c). As one of the high-potency dioxin-like compounds (DLCs), PeCDF shows such high relative toxicity that it is similar to the most potent dioxin, TCDD. Toxic effects observed following exposure to PeCDF, as well as the structurally similar polychlorinated dibenzodioxins (PCDD), include developmental and reproductive alterations, immunotoxicity, teratogenicity, carcinogenicity, and lethality (ATSDR, 1998; Birnbaum, 1994; Poland and Knutson, 1982). PeCDF is listed by the IARC as a Group 3 carcinogen that is not classifiable as to carcinogenicity in humans [\(http://www.iarc.fr/ENG/Databases/index.php\)](http://www.iarc.fr/ENG/Databases/index.php).

Studies suggest that the similar toxic and carcinogenic effects of TCDD and DLCs including PCB126 and PeCDF are mediated through their binding to the aryl hydrocarbon receptor (AhR) (ATSDR, 1998; ATSDR, 2000), which performs three chief functions: generating cellular signal transduction, binding DNA, and activating transcription (IARC, 1997). Upon binding of TCDD, PCB126, or PeCDF to AhR in cells, AhR-dioxin complex forms, cause a cascade of events involving gene and mRNA regulation, protein synthesis, biochemical modifications, and cell-growth stimulus. The eventual result is the formation of lesions, including cancer (Mandal, 2005).

The National Toxicology Program (NTP) conducted four murine carcinogenesis bioassays of TCDD in Swiss-Webster mice (NTP, 1982a), Osborne-Mendel rats, B6C3F₁mice (NTP, 1982b), and female Harlan Sprague-Dawley rats (NTP, 2004b). Two carcinogenesis studies of PCB 126 (NTP, 2004a) or PeCDF (NTP, 2004c) in female Harlan Sprague-Dawley rats

were also reported by the NTP. These studies of dioxin and DLCs are listed in Table 1 (web address of NTP long-term study reports/abstracts: [http://ntp.niehs.nih.gov/go/reports\)](http://ntp.niehs.nih.gov/go/reports). In these investigations, increases occurred in incidences of neoplastic effects, such as cholangiocarcinoma and/or hepatocellular adenoma, squamous-cell carcinoma of the oral cavity and uterus, and cystic keratinizing epithelioma of the lung (Brix et al., 2004; Hailey et al., 2005; Jokinen et al., 2003; Nyska et al., 2005; Tani et al., 2004; Walker et al., 2005, 2006; Yoshizawa et al., 2005a, b).

This review, one of a series of works highlighting specific findings from these NTP dioxin Toxic-Equivalency-Factor (TEF) evaluative studies, focuses on comparison of effects observed in rodents to those reported in human epidemiologic studies induced by exposure to TCDD, PCB 126, and PeCDF. Our objective is to provide understanding of possible concordance of pathologic and neoplastic effects reported in rodents and humans.

Results and Discussion

In this literature review, we compare lesions observed in rats, mice, and humans exposed to TCDD, PCB 126, and PeCDF. Pathological changes induced by these compounds, seen in various NTP studies, are compared to effects reported in epidemiological investigations (Table 2). We systematically analyze information gleaned for target organ(s) and summarize toxicity and carcinogenicity reported within the various bodily systems.

Alimentary System

Liver—Several case reports of hepatomegaly, alterations in hepatic enzymes, hepatitis, and/ or chronic liver disease have been reported among human populations exposed accidentally and occupationally to TCDD and PCBs, including those in the BASF, Seveso, Yusho, and Yucheng accidents (Calvert et al., 1992; McGregor et al., 1998; Yu et al., 1997; Zober et al., 1994, 1997/1998); these are summarized in Table 2. Changes in liver function and structure and increased liver size have consistently been reported in murine investigations (ATSDR, 1998; ATSDR, 2000). In the NTP studies of TCDD, PCB126, and PeCDF, the incidences of several kinds of hepatocellular alterations were increased, such as toxic hepatopathy, hepatocellular hypertrophy, multinucleated hepatocytes, fatty change, necrosis, altered foci, nodular hyperplasia, inflammation, pigmentation, portal fibrosis, bile-duct changes (bile-duct hyperplasia, cysts, cholangiofibrosis) and/or oval-cell hyperplasia (Hailey et al., 2005; NTP, 2004a, b, c). The incidences of hepatocellular adenoma, carcinoma, cholangioma, cholangicarcinoma, and/or hepatocholangiocarcinoma were increased in these NTP rodent studies (Hailey et al.*,* 2005; NTP, 1982b, 2004a, b, c; Walker et al.*,* 2006). In epidemiologic studies of human cohorts exposed to PCBs, liver/biliary/gallbladder cancer was observed only in one study (Brown et al.*,* 1987) in which the data failed to demonstrate a dose-response relationship. Twelve other investigations of workers exposed to PCBs at equivalent levels failed to detect an elevated risk of liver/biliary/gallbladder cancer at any exposure level (Golden et al.*,* 2003); however, increases in hepatobiliary cancers have been observed in residents exposed to TCDD during the Seveso Accident (ATSDR, 1998; Pesatori et al., 2003).

Oral cavity and tooth—Earlier studies revealed that teeth in both mouse and rat are sensitive to TCDD and PCBs during development; broken teeth, periodontitis, gum pigmentation, and other dental aberrations were reported (Alaluusua et al., 1993; Kattainen et al., 2001; Kiukkonen et al., 2002; Lukinmaa et al., 2001; Miettinen et al., 2002; Partanen et al., 1998). Human studies revealed that childhood exposure causes dental aberrations and lesions and gingival squamous hyperplasia; they correlated with animal studies showing evidence of dental sensitivity to TCDD and PCBs (Table 2; Alaluusua et al., 2004; Guo et al., 1999, 2004; Hashiguchi et al., 1997). By comparison, following chronic exposure of adult SD rats to TCDD

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and PCB126 in the NTP studies, the incidences of gingival squamous hyperplasia and/or squamous-cell carcinoma in the oral cavity increased significantly (NTP, 2004a). These results indicate that dioxin and DLCs target the gingiva of the oral cavity, in particular the junctional epithelium of molars (Yoshizawa et al., 2005a). One hypothesis asserts that, during chronic exposure, the action of dioxin and DLCs involves a disruption of retinoid action leading to altered growth and differentiation of the oral gingival epithelium that results in development of gingival squamous hyperplasia and, ultimately, neoplasia (Yoshizawa et al., 2005a). Dental lesions like those seen in developmental-exposure studies were not noted in the NTP chronic studies of TCDD and PCB126.

Exocrine pancreas—Although many comparative studies of toxicities induced by dioxins in animals and humans have been published, exocrine pancreatic toxicity has not been entirely clarified in humans (Table 2). Pancreatic cancer was increased in workers exposed to PCBs at a transformer manufacturing plant in Canada (ATSDR, 2000;Yassi et al.*,* 1994). No statistically significant increased mortality resulted from pancreatic cancer in the Seveso (Bertazzi et al., 2001), Yusho (Kuratsune et al., 1987), and Yucheng (Guo et al., 1999) accidents. In the NTP studies of TCDD, PCB126, and PeCDF, acinar-cell vacuolation, atrophy, inflammation, and/ or arteritis developed at a high incidence (NTP, 2004a;Nyska et al., 2005). Rare occurrences of pancreatic acinar-cell adenomas and carcinomas were noted. These acinar-cell lesions might be related to a direct effect of TCDD on the pancreas. The increase in cytochrome-P450 1A1 and decrease in cholecystokinin-receptor expressions in vacuolated acinar cells may be involved in the pathogenesis of pancreatic lesions and tumors by initiating proliferation (Yoshizawa et al., 2005b).

Stomach and intestine—A retrospective study of BASF employees exposed to TCDD after a 1953 incident involving a chemical reactor revealed a high incidence of appendicitis (Zober et al., 1994). Clinical observations suggestive of gastrointestinal damage have been reported in workers exposed to airborne PCBs and those individuals in the Yusho accident exposed to PCBs (ATSDR, 2000). Follow-up studies of the Seveso accident that caused TCDD exposure showed an increased incidence of rectal cancer (Bertazzi et al., 2001; Pesatori et al., 2003). An elevation in rectal cancer was found in capacitor-manufacturing plants in the United States, but follow-up evaluation of 2588 total workers after 7 years of observation detected no additional deaths from this cancer (Brown, 1987). In the NTP study of TCDD (Table 2), the incidence of squamous-cell hyperplasia in the forestomach was increased, probably due to the disruption of retinoid metabolism (Van Birgelen et al., 1995). No intestinal tumors have been induced by TCDD and PCBs in the rodent models.

Cardiovascular System—Suggestive but inconclusive evidence has implicated adverse cardiovascular effects in humans exposed to relative high concentrations of dioxins (ATSDR, 1998). Increased deaths from chronic heart disease were observed among the Seveso cohort and in the medical study conducted by the National Institute for Occupational Safety and Health (NIOSH) (Bertazzi et al., 2001; Pesatori et al., 2003). Several occupational-exposure studies have investigated the possible relationship(s) between PCB exposure and increased risk of cardiovascular disease or altered blood pressure (ATSDR, 2000).

Dioxins have been found to produce effects on the cardiovascular system in the rodent models (Table 2). One study reported that rats treated with TCDD showed hemorrhages in the brain as well as the spinal cord, and an increase occurred in the incidence of mesenteric and thoracic periarteritis accompanied by thrombosis, above background levels. In the same study, with females only, the incidence of myocardial degeneration also increased above background (Kociba et al., 1978). In C57B1/6J mice, treatment with dioxins created a high mean tail-cuff blood pressure and an elevated urinary excretion of vasoactive eicosanoids (Dalton et al.*,* 2001). When *ApoE* (−/−) mice were exposed to dioxins, a quicker progression occurred towards

the creation of atherosclerotic lesions (Dalton et al.*,* 2001). NTP investigations in the Harlan SD rats showed increased cardiomyopathy and chronic active arteritis, seen mostly in the mesentery and pancreas of rats dosed with TCDD, PCB 126, and PeCDF (Jokinen et al., 2003;NTP, 2004a,b,c). The cardiomyopathy and arteritis observed were similar to some lesions studied in humans (Jokinen et al., 2003). For example, upon observation of humans with idiopathic or toxic cardiomyopathy, multiple foci of myocardial fibrosis could be distinguished (Schoen et al., 1994). Rat arteritis documented in the NTP studies (NTP, 2004a;Jokinen et al., 2003) was best compared to spontaneous polyarteritis nodosa seen in humans (Schoen et al., 1994).

Endocrine System

Adrenal cortex—TCDD and PCBs have been shown to accumulate in the adrenal glands when incorporated into the body (Li and Wang, 2005). Because adrenal steroids exert profound influences on glucose tolerance, insulin sensitivity, lipid metabolism, obesity, vascular function, and cardiac remodeling, the potential association exists of adrenocortical alterations with increased diabetic and cardiovascular risk among highly exposed people (Li and Wang, 2005). In the NTP studies of TCDD, PCB126, and PeCDF (Table 2), increased incidences of adrenal cortical atrophy, cytoplasmic vacuolation, cystic degeneration, hyperplasia and/or tumors occurred (NTP, 2004a). Cortical atrophy and cytoplasmic vacuolation constituted prominent effects and may reflect continued stress in these rats, leading to depletion of corticosteroid hormones or activation of some other unknown mechanism(s) (Frith et al., 2000; Sapolsky et al., 1987). Hormonal manipulation can be used to induce cortical adenoma (Cohen et al., 1957). Cystic degeneration is considered to be a continuum of cytoplasmic vacuolation; in particular, severe cell loss appears with the resultant formation of cystic spaces. This lesion occurs predominantly in aging female, particularly Sprague-Dawley (SD), rats (Frith et al., 2000). Administration of estrogen in several forms was found to cause both benign and malignant neoplasms (Dunning et al., 1953). Clarification is needed to determine whether effects of estrogenic activity and/or antiestrogenic activity of TCDD and DLCs are related to the induction of cystic degeneration and adrenal carcinogenesis.

Thyroid gland—In our literature review, we examined effects of PCBs and dioxin on the thyroid gland, such as increases in TSH, T_3 , and/or T_4 in various populations, including men, women, children, infants, and, particularly, veterans of the Vietnam War (Table 2; ATSDR, 2000;Pavuk et al., 2003;Zober et al., 1994,1997/1998). An increased prevalence of combined thyroid disorders or diseases (goiter, thyrotoxicosis, hypothyroidism, and thyroid adenoma) was reported in German and American workers (ATSDR, 2000;Zober et al., 1994, 1997/1998). Overall, results of studies that examined dioxin and thyroid function have provided little and inconsistent evidence of long-term dioxin effects on the thyroid in humans. An increase in thyroid tumors was observed in a large cohort of workers exposed to chlorophenoxy herbicides, including those contaminated with TCDD (Saracci et al., 1991). Moreover, a possible association between exposure to dioxin and thyroid tumors has been implicated by epidemiological investigations related to the Seveso accident (Pesetori et al., 2003).

TCDD and PCBs are thought to disrupt thyroid hormonal homeostasis in the endocrine system by the induction of phase II enzymes, the UDP-glucuronosyltransferase(UGT)s (Kohn, 2000; Kohn et al., 1996; Schuur et al., 1997; Sewell et al., 1995; van Birgelen et al., 1995; van der Plas et al., 2001). The manufacture and discharge of thyroxine (T_4) are under the direction of the thyroid-stimulating hormone (TSH) (Tani et al., 2004), which receives both negative and positive control from the hypothalamus, pituitary, and thyroid by the thyrothropin-releasing hormone, TSH itself, T_4 , and triiodothyronine (Tani et al., 2004). TCDD induces the production of UGT-1 mRNA through an aryl hydrocarbon receptor-dependent transcriptional mechanism (Bock et al., 1998; Yuch et al., 2003). A decrease in the serum T_4 levels brought about by

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TCDD-induced UGT could possibly lead to a decline in the negative feedback inhibition on the pituitary gland, which would cause an increase in TSH resulting in hyperstimulation of the thyroid follicular cells (Sewall et al.*,* 1995). Follicular hypertrophy induced by TCDD and PCBs in SD rats is characterized by smaller follicles and increased height of lining epithelial cells with no evidence of hyperplasia (Tani et al., 2004). Several studies have revealed associations among $TCDD$, T_4 , and thyroxine-binding globulin (TBG) and documented thyroid disease in humans following exposure (Zober et al., 1994).

Endocrine pancreas—Evidence exists that environmental exposure to TCDD and PCBs contributes to the incidence of diabetes (ATSDR 1998; McGregor et al., 1998; Steenland et al., 1999). A highly significant relationship between dioxin exposure and the severity of diabetes was noted in the United States Air Force personnel who dropped Agent Orange contaminated with dioxin during the Vietnam War (Henriksen et al., 1997). Epidemiologic analysis of the Seveso accident also showed a significantly higher incidence of diabetes in exposed individuals (Bertazzi et al., 2001; Pesatori et al., 2003). In the present NTP studies of TCDD, PCB126, and PeCDF (Table 2), no related lesions in endocrine pancreas were noted (NTP, 2004a). Discrepancy exists among reports of exocrine and endocrine pathological effects in rats and humans, including the diabetic condition; however, potential long-term functional disturbances in the glucose metabolism in rats exposed to these compounds should be further investigated.

Female Genital System

In the previous NTP investigations using SD rats, the range of treatment-related changes in the uterine system comprised active inflammation, cystic endometrial hyperplasia, squamous metaplasia, and uterine carcinoma in the PeCDF study and squamous-cell carcinoma of the uterus in the TCDD study (Table 2). The span of alterations seen with the different compounds may suggest more than one mechanism, such as estrogenic activity and/or antiestrogenic activity of PeCDF and TCDD, as well as local disruption of retinoid homeostasis via vitamin A deficiency. TCDD and DLCs were documented as inducers of squamous metaplasia and squamous-cell carcinoma in other organs, including the oral cavity and lung (Walker et al, 2006;Yoshizawa et al., 2005b). Evidence shows that TCDD and DLCs can increase the incidence and severity of endometriosis in monkeys (Rier and Foster, 2002;Rier et al., 1993, 2001) and promote the growth or survival of endometrial tissue implanted in a surgicallyinduced murine model of endometriosis (Cummings et al., 1996,1999;Johnson et al., 1997). The inappropriate and uncontrolled growth of endometrial cells, normally confined to the lining of the uterus, outside the womb (Birnbaum and Cummings, 2002), leads to endometriosis. In the present NTP rat studies of TCDD and DLCs, no incidences of endometriosis were noted. Although studies with sufficient statistical powers have been lacking, and non-differential misclassifications of disease status have hindered clarification, a potential connection exists between exposure to TCDD and DLCs and the increasing prevalence of endometriosis in humans, as cohort and clinical studies conducted in Israel, Italy, and the United States have demonstrated (Arisawa et al., 2005;Eskenazi et al., 2000,2002;Louis et al., 2005). In the Seveso Women's Health Study, the first investigation to explore the relation between TCDD exposure and endometriosis in a large population of women with a wide range of exposures, women with serum TCDD levels of 100 pg/ml (ppt) or higher showed a doubling, but stastistically insignificant, risk for endometriosis without evidence of any dose-response relationship (Eskenazi et al., 2002). Moreover, the results of an epidemiologic study in Belgium revealed no statistically significant association between exposure to DLCs and the occurrence of endometriosis in infertile women (Pauwels et al., 2001). Clear evidence of a role for dioxins in the increased incidences of endometriosis will require careful cohort studies of women with and without endometriosis, documenting exact exposures.

Male Genital System

TCDD has been responsible for altering testosterone, follicle-stimulating hormone, and luteinizing hormone in the male genital system; existing and half-life dioxin levels present within the serum have been positively associated with luteinizing hormone and inversely correlated to testosterone (Egeland et al., 1994). Essential for the continuance of libido, testosterone is responsible for the commencement and safeguarding of spermatogenesis; elevated levels of luteinizing hormone and follicle-stimulating hormone can cause low sperm counts (Morrow et al., 1986). In comparison to research observed in animals, human investigations have shown that dioxin manipulates the hypothalamic-pituitary axis as well as testosterone production (Egeland et al., 1994). In a study in rats, dioxin reduced testosterone levels (Mebus et al., 1987; Moore and Peterson, 1988). No evidence exists that decline in testosterone is a consequence of an augmentation in testosterone catabolism or secretion (Kleeman et al., 1990). On the other hand, proof has shown that dioxin reduces synthesis of testosterone (Kleeman et al., 1990). Research in rats indicated that dioxin has the ability to slow the receptiveness of the pituitary gland to testosterone and Ley-dig cells to stimulation of luteinizing hormone (Bookstaff et al., 1990; Moore et al., 1992). Data gathered from a human study suggested a similar pathway of dioxin activity due to weakened pituitary and Leydigcell receptiveness (Egeland et al., 1994). No toxicologic data from the male reproductive system were available from the NTP studies (NTP, 1982a)

Immune and Hematopoietic Systems

The effects of TCDD and PCBs on immunotoxicity and hematopoietic toxicity in rodents, seen in the NTP studies, and humans are summarized in Table 2.

In the NTP studies of TCDD, PCB126, and/or PeCDF, thymic and/or splenic lymphoid atrophy was noted as an immunosuppressant effect (Table 2). Lymphocytes are known to be one of the primary targets of TCDD and PCB immunotoxicity via AhR (Halperin et al., 1998;Kerkvliet, 2002;Watanabe et al., 1999). Both humoral and cell-mediated immunity were suppressed following acute or chronic exposure of rodents to low levels of TCDD. Data from several hostresistance models indicated that TCDD increased the susceptibility of animals to a variety of infectious diseases, prevented the rejection of transplanted tumors, and increased tumor growth and metastasis, presumably through alteration in immune function. The greater sensitivity of lymphocytes in humans than that in mice and rats seen in an *in vitro* study of CYP1A1 induction suggested higher susceptibility of the human immune system to TCDD (Nohara et al., 2006). As exposure to PCBs can suppress both antibodies (immunoglobulins) and the cellular immune response, frequent infections can be a direct result (Carpenter, 2006). Studies have clearly shown that persons exposed to TCDD and PCBs exhibit a greater incidence of numerous kinds of infections, including those of the respiratory tract, skin, and ear (ATSDR, 1998, 2000;Carpenter, 2006;Guo et al., 2004;Weisglas-Kuperus et al., 2000,2004;Zober et al., 1994).

An excess of lymphohemopoietic neoplasms was found in the Seveso population and several cohort studies, as shown in Table 2–4; the higher risk of Hodgkin's disease, non-Hodgkin's lymphoma, and myeloid leukemia was noted (ATSDR, 1998, 2000; Bertazzi et al., 2001; Pesatori et al., 2003). A case-control study nested in the IARC international cohort provided weak evidence of a dose-response relationship with estimated TCDD exposures (IARC, 1997). In the previous NTP rodent studies, although positive dose-related trends were observed in the incidences of either lymphoma or leukemia in female B6C3F1 mice, TCDD has not been carcinogenic for the lymphohematopoietic system (NTP, 1982b).

Integumentary System

Chloracne and skin hyperpigmentation caused by TCDD and DLCs constitute well known side effects in several dioxin-exposure accidents (Table 2), such as those at Seveso, Yusho, and Yu-Cheng (Guo et al., 1999;2004;Schecter et al., 2006;Urabe and Asahi, 1985). Recently, a notorious event occurred when Ukraine President Viktor Yushchenko was stricken with facial chloracne resulting from deliberate poisoning with TCDD during his presidential campaign (Holt, 2005;Schecter et al.*,*2006;Sterling and Hanke, 2005).

In order to elucidate the effects of TCDD on the integumentary system, researchers used rhesus monkeys to study morphological changes that appeared during dosing with 500 ppt of TCDD for nine months. Upon necropsy, squamous metaplasia and keratinization were seen within the sebaceous glands and hair follicles. Keratinization was also observed in the Meibomian glands of the eyelids. Thickening of the fingernails and toenails was reported (Allen et al.*,* 1977).

In the rodent NTP studies, TCDD was carcinogenic for female Swiss-Webster mice causing fibrosarcomas in the integumentary system (Table 2; NTP, 1982a). Few reports have documented increased incidences of fibrosarcoma in dioxin and PCB human epidemiologic studies (ATSDR, 1998,2000). Significant associations have been reported between the occurrence of malignant melanoma and PCB exposure in humans (Golden et al., 2003). The incidence of melanoma was reportedly increased, relative to national rates, among white ranchhand veterans who sprayed TCDD-contaminated herbicides in Vietnam. Researchers know that genetic susceptibility and environmental factors contribute to the initiation and progression of melanoma. Excessive ultraviolet exposure is considered the main etiological factor in its initiation; however, epidemiological and experimental evidence has suggested that exposure to environmental carcinogens contributes to the development of melanoma. Exposure to environmental chemicals, especially TCDD, activates the aryl-hydrocarbon-receptor pathway and contributes to melanoma progression, specifically through stimulation of the expression and activity of the matrix metalloproteinases (Villano et al., 2006).

While many small clinic-based breast cancer studies have been published (Table 2), some of which suggest a dose-dependent relation to total PCB exposure (ATSDR, 2000;Golden et al., 2003), most large retrospective epidemiologic investigations have not demonstrated a relationship between elevated total PCB levels and breast cancer. While the human data of the relationship between TCDD exposure and the occurrence of breast cancer have not been clarified, an earlier epidemiologic study of the Seveso population suggested a possible insignificant decrease in the incidence of breast cancer (Huff et al., 1994). A subsequent epidemiologic investigation demonstrated a significant increase in breast cancer in association with elevated serum levels of TCDD (Birnbaum and Fenton, 2003). Due to its antiestrogenic properties, TCDD inhibited spontaneous and carcinogen-induced mammary tumor formation and growth in rodent models (Safe et al., 1999).

Nervous System

In utero exposure and exposure of infants and young children to PCBs, including those reported in the Yucheng accident, have been linked to adverse effects on intellectual functioning, (ATSDR, 2000; Carpenter, 2006; Guo et al., 2004); delay in neurodevelopment, neurobehavioral changes (hypotony, kyperactivity, lower mean intelligence quotients, altered latencies, and amplitudes of auditory event-related potentials), and mental disorders were documented. Similar changes in cognitive and motor behavior have been found in mice and rats following single-dose exposure and monkey offspring born to mothers chronically exposed during pregnancy (Brouwer et al., 1998). Symptoms of neurological disorders, such as peripheral nervous signs, lower limb weakness, muscle pains, sleepiness, headaches, dizziness, and nausea were also reported in some adult epidemiologic studies of TCDD (ATSDR,

1998; Watanabe et al., 1999; Zober et al., 1994, 1997/1998). Relatively little information is available on neurological effects of PCBs in adult humans (ATSDR, 2000). In the rodent NTP studies of TCDD, PCB126, and PeCDF, no clinical neurobehavioral changes and no neurological histopathological changes were seen (Hill et al., 2003; NTP, 1982a,b, 2004a,b,c)

Respiratory System

The respiratory system has exhibited effects induced by TCDD and PCBs, which are summarized in Table 2. The risks of pulmonary cancer were found to be increased during 20 years that passed after 1976, when numerous people were exposed accidentally to TCDD at a chemical plant in Seveso, Italy (Bertazzi et al., 2001;Pesatori et al., 2003). Additional documentation indicated pulmonary diseases and cancers following exposure to TCDD during the accident that occurred in 1953 at BASF, also a chemical company (Zober et al., 1994, 1997/1998). Not only has TCDD caused pulmonary diseases, but PCBs have also increased the risks of pulmonary diseases and cancers (ATSDR, 2000). For example, the Yusho accident in Japan and the Yucheng accident in Taiwan caused pulmonary diseases and cancers due to the contamination of rice oil with PCBs (ATSDR, 2000;Kikuchi, 1984;Nakanishi and Shigematsu, 1991); in both accidents more frequent or severe respiratory infections and chronic bronchitis accompanied by persistent cough and production of sputum were noted (ATSDR, 2000). The NTP animal study showed that treatment with TCDD, PCB 126, PeCDF, and the mixture of all three induced several forms of pulmonary diseases, such as cystic keratinizing epithelioma, bronchiolar metaplasia, and squamous metaplasia of the alveolar epithelium. Potential mechanisms leading to altered differentiation and/or proliferation of bronchiolar and alveolar epithelia might have operated via induction of CYP1A1 or disruption of retinoid metabolism; a significant toxicity to humans of TCDD, PCB 126, PeCDF, and the mixture of all three was likely indicated.

Urinary System

In humans, although abnormal urinary porphyrin patterns have been reported, few publications have implicated dioxin-induced renal effects. The porphyrogenic effect of TCDD was reported in female rats treated chronically for 45 weeks with 0.01, 0.10, and 1.00 *μ*g/kg TCDD; the dose of 0.01 *μ*g/kg/week resulted in significantly enhanced coproporphyrinuria after 3 months and at 10 months with a significant rise in uroporphyrin (Cantoni et al., 1981). Additionally, the previous two-year rat study showed increased urinary excretion of porphyrins in female rats exposed to 0.10 *μ*g/kg TCDD (Kociba et al., 1978). In the present NTP studies of TCDD, PCB126, and PeCDF, the incidences and/or grade levels of nephropathy were increased (Table 2; NTP, 2004a,b,c). In humans, mechanisms and extrapolations have not been clarified; the kidney historically has not been a target of TCDD-induced neoplasia, although direct responsiveness to TCDD is known to induce renal biochemical alterations (Pegg et al., 1976; Walker et al., 1995).

Skeletal System

In the 14-year follow-up study of the Yucheng Accident (Guo et al., 1999), the incidences of arthritis and herniated disks were found to be higher in the exposed population than the control population (Table 2). In persons exposed in the Yusho accident, an elevated incidence of joint inflammation was documented (Kuratsune, 1980). Rheumatoid arthritis is characterized by proliferation of synoviocytes that produce the critical proinflammatory cytokine, IL-1beta; TCDD was shown to upregulate, via AhR, the mRNA level of IL-1beta in the human fibroblastlike line of synoviocytes (Tamaki et al., 2006). In Lewis rats, TCDD inhibited bone growth, bone modeling, and mechanical strength of bone in a dose-dependent manner (Jamsa et al., 2001) without a direct effect on increased osteoclastic bone resorption (Ilvesaro et al., 2005).

No reports have been published, however, of TCDD-and DLC-induced skeletal lesions in the present NTP studies of TCDD, PCB126, and PeCDF (Table 2; NTP, 1982a,b, 2004a,b,c).

Other Systems

TCDD also has been linked to an increase in cancer mortality. In one human study, the subjects worked in twelve plants within the United States that manufactured chemicals contaminated with TCDD. Serum TCDD levels were two to three times higher than those observed in the general public. A comparison of the human study to animal research revealed that these levels were similar to those that produced cancer in animals (Steenland et al., 1999). A comparable increase in soft-tissue sarcoma was present in the Seveso population and the TCDD-exposed manufacturing workers in the United States (Fingerhut et al., 1991; Frumkin, 2003); however, no such increase was found in the other cohort studies (IARC, 1997; McGregor et al., 1998). Investigations of Vietnam veterans have not demonstrated any increase in soft-tissue sarcoma (Frumkin, 2003). Additionally, several reevaluations of occupational and accidental exposures to PCBs have not shown any excess in overall cancer mortality or mortality emanating from specific cancer sites (Bosetti et al., 2003). An excessive risk for soft-tissue sarcoma in humans, based on a small number of deaths, has been reported (Table 2–6) (ATSDR, 1998). Data for incidences of soft-tissue sarcoma were generally not available (IARC, 1997); however, no reports have been published of TCDD- and DLC-induced soft-tissue sarcoma in the present NTP studies of TCDD, PCB126, and PeCDF (Table 2–6; NTP, 1982a,b, 2004a,b,c).

Conclusions

By altering the functioning of many different organ systems, documented in rodent studies and human populations, exposures to TCDD and PCBs constitute risk factors for a large number of human diseases. In studies of human populations exposed to PCBs, exposure is to a mixtures of PCBs that includes both dioxin-like PCBs (e.g. PCB126 and PCB118) and non-dioxin like compounds (e.g. PCB153). We have summarized findings from the NTP studies and investigations reported by numerous other researchers and shown comparisons between effects manifested in rodents and humans. In the NTP murine studies, TCDD, PCB126, and/or PeCDF induced neoplastic lesions (hepatocellular and/or cholangiolar, pancreatic, thyroidal, and pulmonary) and nonneoplastic effects (hepatocellular and/or cholangiolar lesions in the liver, squamous hyperplasia in the oral cavity, cardiovascular damage, thyroidal hypertrophy, and immunosuppression) similar to those reported in humans. While differences in specific pathologies were observed, clear consistency could be seen in the target organs affected (liver, oral cavity, cardiovascular system, immune system, thyroid, pancreas, and lung) in both human and murine studies.

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Abbreviations

AhR

aryl hydrocarbon receptor

ATSDR

Agency for Toxic Substances and Disease Registry

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CYP

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Table 1

Associated doses used in the NTP two-year studies of dioxin and dioxin-like compounds.

a
Toxic equivalency factor (Van den Berg et al., 1998, 2006).

b To humans, TCDD is classified as carcinogenic chemical, PCB126 as probable carcinogenic chemical; PeCDF is not classifiable as to carcinogenicity by International Agency for Research on Cancer (IARC) (<http://monographs.iarc.fr/ENG/Classification/index.php>).

c National Toxicology Program (NTP) study report number.

Hashiguchi et al., 1995, 1997;

Table 2
Comparison of toxic effects and carcinogenesis of TCDD and PCBs in rodent NTP studies and human epidemiology. NIH-PA Author ManuscriptNIH-PA Author Manuscript

Comparison of toxic effects and carcinogenesis of TCDD and PCBs in rodent NTP studies and human epidemiology.

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2003; Rix et al.,

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 b National Toxicology Program (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

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 $\ensuremath{^c}\xspace$ Osborne Mendel rats. *c*Osborne Mendel rats.

 $d_{\rm BCGF1}$ mice,

 $e_{\rm Swiss\mbox{-}Webster}$ mice. e Swiss-Webster mice.

 $f_{\rm No}$ available information published. *f* No available information published.

 ${}^a\!{\rm Harlan}$ Sprague Dawley rats. *a*Harlan Sprague Dawley rats.

 b Vational Toxicology Program (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

 $\emph{``No available information published}.$ *c*No available information published.

 $a_{\rm Harlan}$ Sprague-Dawley rats. *a*Harlan Sprague-Dawley rats.

 b National Toxicology Program (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

 $\prescript{c}{\operatorname{Osborne}}$ Mendel rats. *c*Osborne Mendel rats.

 $d_{\rm B6C3F1}$ mice.

 ${}^e\!N$ o available information published. *e*No available information published.

 ${}^d\!H\!arl$ an Sprague-Dawley rats. *a*Harlan Sprague-Dawley rats.

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 $b_{\rm National\ Toricology\ Program}$ (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

 $^{\prime}$ B6C3F1 mice.

 d_No available information published. *d*_{No available information published.}

 $a_{\rm Harlan}$ Sprague-Dawley rats. *a*Harlan Sprague-Dawley rats.

 $b_{\mbox{\scriptsize\sc National}}$ Toxicology Program (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

 $c_{\rm Swiss\mbox{-}Webster\,mice.}$ *c*Swiss-Webster mice.

 $d_{\rm No}$ available information published. *d*_{No available information published.}

 $a_{\rm Harlan}$ Sprague Dawley rats. *a*Harlan Sprague Dawley rats.

 $b_{\rm National~Toxiology~Program}$ (NTP) study report number. *b*National Toxicology Program (NTP) study report number.

 ${}^c\!$ No available information published. *c*No available information published.