

SYNTHESIS

Evolutionary consequences of fishing and their implications for salmonJeffrey J. Hard,¹ Mart R. Gross,² Mikko Heino,^{3,4,5} Ray Hilborn,⁶ Robert G. Kope,¹ Richard Law⁷ and John D. Reynolds⁸

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Keywords

adaptation, fitness, heritability, life history, reaction norm, selection, size-selective mortality, sustainable fisheries.

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Received: 3 November 2007

Accepted: 29 January 2008

doi:10.1111/j.1752-4571.2008.00020.x

Abstract

We review the evidence for fisheries-induced evolution in anadromous salmonids. Salmon are exposed to a variety of fishing gears and intensities as immature or maturing individuals. We evaluate the evidence that fishing is causing evolutionary changes to traits including body size, migration timing and age of maturation, and we discuss the implications for fisheries and conservation. Few studies have fully evaluated the ingredients of fisheries-induced evolution: selection intensity, genetic variability, correlation among traits under selection, and response to selection. Most studies are limited in their ability to separate genetic responses from phenotypic plasticity, and environmental change complicates interpretation. However, strong evidence for selection intensity and for genetic variability in salmon fitness traits indicates that fishing can cause detectable evolution within ten or fewer generations. Evolutionary issues are therefore meaningful considerations in salmon fishery management. Evolutionary biologists have rarely been involved in the development of salmon fishing policy, yet evolutionary biology is relevant to the long-term success of fisheries. Future management might consider fishing policy to (i) allow experimental testing of evolutionary responses to exploitation and (ii) improve the long-term sustainability of the fishery by mitigating unfavorable evolutionary responses to fishing. We provide suggestions for how this might be done.

Introduction

Anadromous salmonids (Table 1) migrate through freshwater and marine habitats, where they grow to maturity before homing to natal rivers for reproduction (Quinn 2005). Their high nutritional quality and relative ease of capture have subjected them to substantial human exploitation, through commercial, recreational and aboriginal fisheries. Demographic and stock-recruitment relationships for salmon are often used by fisheries managers to set exploitation levels with the objective of a maximum

sustainable yield (Ricker 1958, 1969; Walters and Martell 2004). But rarely are the evolutionary responses of salmon considered in the setting of exploitation levels or in the methods and timing of capture. Even though no single study has yet conclusively demonstrated fisheries-induced evolutionary changes in exploited fish in the wild, theoretical and empirical evidence for fisheries-induced selection pressures is strong (e.g. Ricker 1981; Heino 1998; Law 2000; Carlson et al. 2007), and there is a growing body of evidence suggesting that evolutionary changes in fish life histories may already be widespread (e.g. Ricker

Table 1. Prominent life history traits of the primary salmonids considered in this paper for evidence of fisheries-induced evolution. Most anadromous forms that spend more than a single season at sea are vulnerable to extensive fishing.

Species (common names)	Scientific name	Migration	Reproduction	Age structure
Atlantic salmon	<i>Salmo salar</i>	Anadromous	Iteroparous	Variable (MSW)
Sea trout/brown trout	<i>Salmo trutta</i>	Anadromous/FW resident	Iteroparous	Variable (MSW)
Chinook salmon	<i>Oncorhynchus tshawytscha</i>	Anadromous	Semelparous	Variable (MSW)
Chum salmon	<i>Oncorhynchus keta</i>	Anadromous	Semelparous	Variable (MSW)
Coho salmon	<i>Oncorhynchus kisutch</i>	Anadromous	Semelparous	Simple (~16 months at sea)
Pink salmon	<i>Oncorhynchus gorbuscha</i>	Anadromous	Semelparous	Fixed (2 years)
Sockeye salmon	<i>Oncorhynchus nerka</i>	Anadromous/FW resident*	Semelparous	Variable (MSW)
Cutthroat trout	<i>Oncorhynchus clarki</i>	Anadromous/FW resident†	Iteroparous	Variable
Steelhead/rainbow trout	<i>Oncorhynchus mykiss</i>	Anadromous/FW resident	Iteroparous	Variable (MSW)
Brook charr	<i>Salvelinus fontinalis</i>	Anadromous/FW resident	Iteroparous	Variable
Lake whitefish	<i>Coregonus clupeaformis</i>	FW resident	Iteroparous	Variable
European grayling	<i>Thymallus thymallus</i>	FW resident	Iteroparous	Variable

FW, freshwater; MSW, multi-sea winter.

*Freshwater resident form = kokanee.

†All but the coastal subspecies exhibit the freshwater resident form only.

1981; Law 2000; Kuparinen and Merilä 2007; Edeline et al. 2007; International Council for the Exploration of the Sea (ICES) 2007; Swain et al. 2007). Moreover, evolutionary changes in fish life histories could affect viability and future yield in the fisheries, which is the opposite of that desired in management (Heino 1998; Law 2000; Conover and Munch 2002; de Roos et al., 2006).

Concerns about the potential evolutionary effects of salmon fishing are now a century old, but relatively few studies of these effects are available, and none of these investigations provides direct evidence for fisheries-induced evolution (Table 2). Stone (1880, 1882) and Rutter (1904) appear to have been the first to speculate in the literature that salmon fisheries might enhance the representation of smaller, younger male breeders and that removal of larger adults could lead to reductions in adult size as well as yield. Smith (1920) was concerned that removal of immature salmon in ocean fisheries would reduce future yields, presumably through earlier maturation, but Miller (1957) argued that the high plasticity of salmonid growth and maturation would render inert any selection imposed by fishing.

In the intervening century, such general concerns have persisted (Birkeland and Dayton 2005; Law 2007; Fenberg and Roy 2008; Hutchings and Fraser 2008), but salmon fishery management seldom incorporates evolutionary considerations in practice. In this review, we discuss what is known about the evolutionary consequences of fishing for salmon and address three central questions: First, what are the likely genetic consequences for salmon exposed to fishing, and what is the evidence? Second, do these consequences matter, when considered with other factors influencing viability? Finally, what is the lesson for

management – how hazardous is it to ignore evolutionary considerations in salmon fishery management?

Fishing as an agent of change for salmonid life histories

Fishing practice

Salmon are extensively exploited by fisheries. For some populations, commercial and recreational fishing for anadromous salmon kills over 80–90% of individuals (Hankin and Healey 1986; Walters 1986; Heard 1991; Hilborn and Walters 1992; Pacific Salmon Commission (PSC) 2007). Historically, anadromous salmon were intercepted in high-seas fisheries as well as in coastal and riverine fisheries both in the Pacific and in the Atlantic. High-seas salmon fisheries in the Pacific have been prohibited since the 1990s and have been strongly restricted in the Atlantic; salmon are also by-catch in other fisheries. In high-seas fisheries, both immature and maturing individuals were killed, whereas terminal fisheries in estuaries and freshwater killed maturing individuals during their spawning migrations.

In recent decades, catches of Atlantic salmon have continued to decline, reaching their lowest levels in history. Productivity in nearly all populations is limited by high rates of marine mortality (International Council for the Exploration of the Sea (ICES) 2006). For Pacific salmon, catches have generally increased since the 1980s around the northern Pacific rim, with the exception of stocks in western Alaska (declining since the 1990s) and in southern British Columbia (declining since the 1980s) and farther south (declining since the 1930s). Increases in catch have been influenced by increasing hatchery production

Table 2. Summary of studies that have evaluated trends in size and life history of exploited salmonid populations potentially affected by fishing. Putative factors are the primary ones identified by the authors. Nearly all studies evaluated phenotypic trends or estimated norms of reaction, and therefore the primary causal factors for these patterns could not be ascertained. The table does not include modeling investigations of fishing-induced evolution specific to salmonids, such as Hard (2004); Hard (in press) for Chinook salmon or Thériault et al. (in press) for brook charr.

Species	Traits examined	Location (period)	Putative factors	Evidence for potential evolutionary response	References
Atlantic salmon	Body weight, run timing	Ireland (1926–1999)	F, E	↓ In weight, delay in run timing	Quinn et al. (2006)
	Body weight	Wales, UK (1907–1977)	F, E	↓ In weight, ↓ in MSW adults, ↑ incidence of grilse	Gee and Milner (1980)
	Body weight, age	Quebec, Canada (1859–1983)	F	↓ In weight, ↑ in age at maturation, ↓ in iteroparity	Bielak and Power (1986)
	Body weight, age	Maritime provinces, Canada (1954–1973)	F	Pop. variation in age & weight neg. correlated with fishing rate	Schaffer and Elson (1975)
	Body weight	North Sea – Norway and Scotland (1965–1993)	E	↓ In weight	Friedland et al. (2000)
	Age at maturation	Maritime provinces, Canada (1965–1972)	F	Changes in age at maturation, ↓ in MSW adults, ↑ incidence of grilse	Ritter and Newbould (1977); Paloheimo and Elson (1974); Ritter et al. (1986)
	Age at maturation	Scotland, UK (1872–1993)	E	Variable trends in incidence of grilse	Summers (1995)
	Body weight, age	Norway and NW Russia (1980–1994)	F	Variable trends in weight & size at age, ↓ in spawner age	Jensen et al. (1999)
	Allele frequency	Spain (1988–2000)	F	Generally stable frequency of common <i>MEP-2*</i> allele	Consuegra et al. (2005)
	Body weight	Spain (1988–2000)	F	Trend toward ↓ spawner body weight	Consuegra et al. (2005)
	Age at maturation	Spain (1988–2000)	F	Trend toward ↓ sea age of spawners	Consuegra et al. (2005)
	Run timing	Spain (1945–2000)	F	Delays in median timing of capture	García de Leániz et al. (1992, 2001); Consuegra et al. (2005)
	Body length and weight	Spain (1945–2000)	F	↓ In weight & length of harvested fish	García de Leániz et al. (1992, 2001)
	Degree of iteroparity	Spain (1945–2000)	F	↓ Longevity, ↓ frequency of iteroparity	García de Leániz et al. (1992, 2001); Consuegra et al. (2005)
	Age at maturation	Spain (1945–2000)	F	↑ In smolt age, ↓ in sea age, ↑ frequency of grilse	García de Leániz et al. (1992, 2001)
	Age at maturation	Quebec, Canada (1967–1984)	F	↑ Frequency of mature male residents	Caswell et al. (1984); Montgomery et al. (1986)

Table 2. Continued

Species	Traits examined	Location (period)	Putative factors	Evidence for potential evolutionary response	References
Chinook salmon	Body weight	British Columbia, Canada (1951–1975)	F	↓ In mean weight (24 of 24 groups)	Ricker (1981)
	Body weight	British Columbia, Canada (1951–1991)	F, E	Variable trends in mean weight, with some ↓ showing reversals	Ricker (1995)
	Body weight	West coast N. America (1975–1993)	E	Variable trends in mean weight, with ↓ predominant	Bigler et al. (1996)
	Body length	West coast N. America (1979–1993)	E	↓ In mean weight	Bigler et al. (1996)
	Age at maturation	West coast N. America (1975–1993)	E	Variable trends in mean age, with ↓ predominant	Bigler et al. (1996)
	Spawn timing	Puget Sound, WA, USA (1960–2000)	H, E	Significant advances in spawn timing	Quinn et al. (2002)
	Body length and weight	British Columbia, Canada (1951–1981)	E	Predominantly negative trends in size, depending on period	Healey (1986)
	Body length	Yukon River, AK (1970–2004)	F or E	↓ Trends in relative abundance of large spawners (4 of 7 groups)	Hyer and Schleusner (2005)
	Body length	British Columbia, Canada (1951–1975)	F	↓ In mean weight (40 of 48 groups)	Ricker (1981)
	Body weight	British Columbia, Canada (1951–1991)	F, E	Weak, variable trends in mean weight (most groups)	Ricker (1995)
Chum salmon	Body weight	West coast N. America (1975–1993)	E	↓ In mean weight	Bigler et al. (1996)
	Body length	West coast N. America (1979–1993)	E	↓ In mean length	Bigler et al. (1996)
	Body length and weight	British Columbia, Canada (1951–1981)	E	Variable trends in size (mostly negative), depending on period	Healey (1986)
	Age at maturation	West coast N. America (1975–1993)	E	↑ In mean age	Bigler et al. (1996)
	Body length	Hokkaido, Japan (1992–1997) Kurile Islands, Russia	E, H	↓ In size at maturation & ↑ in age at maturation	Ishida et al. (1993, 1995); Kaeriyama and Katsuyama (2001); Eggers et al. (2005); Kaev (2000); Kaev and Romasenko (2003)
	Age at maturation, body length	Hokkaido, Japan (1962–1997)	E, H	↓ In size at maturation & ↑ in age at maturation	Morita et al. (2005); Morita and Fukuwaka (2006, 2007)
	Body weight	British Columbia, Canada (1951–1975)	F	↓ In mean weight in most areas	Ricker and Wickett (1980); Ricker (1981)
	Body weight	British Columbia, Canada (1951–1991)	E	↓ In mean weight (56 of 60 groups)	Ricker (1981)
	Body weight	British Columbia, Canada (1951–1991)	F, E	↓ In mean weight for most areas (except in north)	Ricker (1995)

Table 2. Continued

Species	Traits examined	Location (period)	Putative factors	Evidence for potential evolutionary response	References
Pink salmon	Body length and weight	British Columbia, Canada (1951–1981)	E	Variable trends in size (mostly negative), depending on period	Healey (1986)
	Body weight	West coast N. America (1975–1993)	E	↓ In mean weight	Bigler et al. (1996)
	Spawn timing	Puget Sound, WA, USA (1946–2000)	H, E	Significant advances in spawn timing	Quinn et al. (2002)
	Body weight	British Columbia, Canada (1951–1975)	F	↓ In mean weight (even- and odd-year lines; all groups)	Ricker (1981)
	Body weight	British Columbia, Canada (1951–1991)	F, E	↓ In mean weight of all groups (especially southern odd-year)	Ricker (1995)
	Body weight	British Columbia, Canada (1953–1988)	F	↓ In mean weight	Ricker et al. (1978), Ricker (1981)
	Body length and weight	British Columbia, Canada (1951–1981)	E	Variable trends in size (mostly negative), depending on period	Healey (1986)
	Body weight	West coast N. America (1975–1993)	E	↓ In mean weight	Bigler et al. (1996)
	Allele frequency	Kamchatka, Russia (1979–1981)	F	↑ In heterozygosity at <i>PGM</i> & proportion of early-maturing fish	Altukhov and Salmenkova (1991); Altukhov et al. (1991); Thorpe (2007)
	Sockeye salmon	Body weight	West coast N. America (1975–1993)	E	↓ In mean weight
Body weight		British Columbia, Canada (1951–1991)	E	↓ In mean weight (27 of 37 groups)	Ricker (1981)
Body weight		British Columbia, Canada (1951–1991)	F, E	No sustained trend in mean weight	Ricker (1995)
Body length		West coast N. America (1975–1993)	E	↓ In mean length (selected groups)	Bigler et al. (1996)
Body length and weight		British Columbia, Canada (1951–1981)	E	Variable trends in size (mostly negative), depending on period	Healey (1986)
Body length at age		Fraser River, BC, Canada (1952–1993)	E	↓ In body size correlated with sea surface temperature	Hinch et al. (1995); Cox and Hinch (1997)
Body length at age		British Columbia, Canada; AK, USA (1967–1997)	E	↓ In body size correlated with ↑ abundance & SST	Pyper and Peterman (1999)
Age at maturation		West coast N. America (1975–1993)	E	↑ In mean age (selected groups)	Bigler et al. (1996)
Body girth		Bristol Bay, AK, USA (1994)	F	↓ In girth, scaling with harvest rate	Hamon et al. (2000)
Run timing		Bristol Bay, AK, USA (1969–2003)	F	Advance in river entry timing for two fishing districts	Quinn et al. (2007)
Allele frequency		Kamchatka, Russia (1930s–1980s)	F	↑ Proportion of heterozygous resident fish	Krogius (1979); Altukhov and Varnavskaya (1983); Altukhov and Salmenkova (1991)
Age, growth rate		Kamchatka, Russia (1968)	F	↑ proportion of early-maturing resident fish	Nikulin (1970)
Age, size, growth rate		Kamchatka, Russia (1935–1979)	F	↑ In proportion of early-maturing resident fish, ↓ in length	Krogius (1979); Varnavskaya and Varnavsky (1988); Altukhov (1994)
Allele frequency		Kamchatka, Russia (1979–1981)	F	↑ In heterozygosity at <i>PGM</i> & proportion of early-maturing fish	Altukhov and Varnavskaya (1983); Thorpe (1993)
Allele frequency	Kamchatka, Russia (1979–1981)	F	↑ In heterozygosity at <i>PGM</i> , <i>LDH</i> , <i>PX</i> & early-maturing males	Altukhov and Varnavskaya (1983)	

Table 2. Continued

Species	Traits examined	Location (period)	Putative factors	Evidence for potential evolutionary response	References
Brown trout	Body weight, age	Switzerland/France (1990s)	F	↑ Larger, older, Atlantic salmon and AB hybrids in catches	Mezzera and Largiadèr (2001)
Lake whitefish	Growth rate, age at maturation	Alberta, Canada (1941–1975); Lake Michigan (1932–1967); Germany (1947–1997)	F	↓ Growth rate, ↓ age at maturity	Handford et al. (1977); Taylor et al. (1992); Thomas and Eckmann (2007)
Grayling	Size at age, fecundity	NW Territories, Canada (1971–1978)	F	↑ Size at age and fecundity	Healey (1978, 1980)
	Age and size at maturation	Norway (1900s – most of 20th century)	F	↓ In weight, ↓ in age at maturation	Haugen and Vøllestad (2001)

E, environment (e.g. climate, ocean conditions); F, fishing selection; H, hatchery selection (e.g. domestication); MSW, multi sea winter; SST, sea surface temperature.

in many areas, and improving ocean conditions in the northern regions (Eggers et al. 2005). The recent declines in salmon numbers and concerns about loss of less productive populations have resulted in killing rates now more typically capped at 40–50%, although rates vary considerably among species and populations (Walters and Cahoon 1985; Walters and Martell 2004). Most Pacific salmon populations have experienced nearly a century of intensive fishing (Walters 1986; Walters and Martell 2004; Eggers et al. 2005; Hindar et al. 2007).

Salmon fisheries can be categorized by gear types such as hook and line (e.g. recreational fishing, commercial troll fishing), net (especially gillnet and purse seine), and trap technologies, and by the locations where gear intercepts fish on migration routes. These different gear types, and timing and location of use, exert different forms of selection. In general, hook and line salmon fisheries are size selective and timing is selective through regulation (Pacific Salmon Commission (PSC) 2004; Consuegra et al. 2005). Gillnet dimensions tend to be selective for body shape and migration timing (Todd and Larkin 1971; Hamley 1975; Millar and Fryer 1999; Hamon et al. 2000; Fujimori and Tokai 2001). Purse seines scoop up fish from aggregates and are thought to be less size selective (Pope et al. 1975; Ricker 1981) but could impose selection on migration timing and schooling behavior, particularly if the fishery employs specific time or area openings.

Traits under selection

Several salmonid traits are subject to direct or indirect effects of fishing. Two that have received considerable attention are body size and migration timing (Table 2). Fishing generally targets some aspect of body size, either through regulation or gear restriction. For example, gillnets target fish of particular girths but the degree of selectivity depends on population, sex, and state of maturation (Hamon et al. 2000; Fujimori and Tokai 2001; Quinn et al. 2001). Furthermore, size is correlated, genetically as well as phenotypically (Hard 2004), with other life history traits that influence salmon fitness. Even in the absence of direct selection on body size, changes in overall mortality level are driving selection on life history traits that involve trade-offs between performance in early and later life. This is most obvious for traits that relate to timing of major life history events such as smolting and maturation (Riddell 1986; Campbell et al. 2006; Thorpe 2007), but also applies to other traits such as growth and reproductive effort.

Although fishing mortality can account for only a fraction of total salmon mortality (Healey 1986; Riddell 1986; but see Heard 1991 for a counterexample), a sufficiently

high fishing mortality can result in selection that has a substantial impact on fitness variation. It is sometimes argued that because most salmon die during the early stages of life, fishing mortality cannot have a decisive effect in shaping salmon life history. However, salmon approaching maturity are those that are most likely to pass their genes to future generations, and selective mortality among them is capable of generating substantial selection differentials as well as influencing population growth rate, particularly when fishing mortality is high. The decrease in population size through fishing mortality can indirectly select against sexually selected morphologies on the spawning grounds, including investment in male kypes and humps for fighting for access to females, and female body size for fighting for quality nest sites and for increasing survival through parental care (van den Berghe and Gross 1986, 1989; Fleming and Gross 1989). It can also bias the selective advantage of alternative life histories, for example favoring 'jack' or early maturing precocial males at the expense of later maturing 'hooknose' males (Gross 1996). Fishing with nets can directly target sexually selected characters when males with larger kypes have higher probabilities of entanglement (Hamley 1975).

In addition to selective effects within populations, differential selection on mixtures of populations with distinct characteristics can alter stock composition in fisheries. For example, spawning populations often differ in their migration timing through the fishery (Quinn et al. 2007), which might affect patterns of fisheries-induced selection on size, age, or morphology among populations.

Approaches to detecting fisheries-induced evolution

Regression analyses and reaction norms

Two approaches have been used to try to disentangle genetic effects of fishing from other factors influencing phenotypes, but with mixed success for salmonids: regression-based analyses (e.g. Ricker 1981, 1995; Rijnsdorp 1993; Morita et al. 2001) and analyses using probabilistic maturation reaction norms (Heino et al. 2002; reviewed in Dieckmann and Heino 2007; see Thériault et al. in press for an application of reaction norm methodology to migratory tendency). Both approaches have considerable appeal but their limitations arise from how they deal with genetic and environmental influences on phenotypic expression of growth, size, and maturation. Maturation reaction norms may offer a powerful tool for specific situations, although there is some debate as to how cleanly they separate genetic and environmental effects acting on maturation (see below). Regression analysis is a generic but often weaker approach. How-

ever, incorporating elements of quantitative genetics (see below) to regression-based analysis can improve its power (Swain et al. 2007).

Analyses of changes in maturation likelihood as influenced by size and age (e.g. Morita and Morita 2002; Morita and Fukuwaka 2006, 2007) have tried to separate the influence of phenotypic plasticity from those of environmental variation in size and age on maturation using probabilistic maturation reaction norms (PMRN). A PMRN describes probability of maturation as a function of age and size, and potentially other explanatory variables (Heino et al. 2002). The analysis of PMRNs can help to distinguish the influences of genetic components of variation from those of phenotypic plasticity on maturation, and thereby characterize the relationship between age, size and likelihood of maturation for different levels of exploitation (Dieckmann and Heino 2007). Indeed, the PMRN approach allows removal of the influences of demography and a major source of phenotypic plasticity from analyses of trends in maturation. However, as a purely phenotypic approach, it cannot be used to unambiguously demonstrate genetic change (Dieckmann and Heino 2007; Marshall and McAdam 2007; Wright 2007); the method can also be confounded by violations of assumptions about genetic control of maturation and growth that are difficult to test.

Quantitative genetic models of response to selection

A more direct approach to determining the direction and rate of evolutionary change under fishing is through quantitative genetic analysis of phenotypic evolution (Lande 1979; McGuigan 2006). Selection requires phenotypic variation and differential reproduction or survival. With sufficient knowledge of the population's relatedness structure, observed (i.e., phenotypic) patterns of mean trait values together with their variances and covariances can be used to estimate the genetic parameters that determine its responses to selection in a population. The framework for relating selection and its response in a particular trait relies on a simple empirical function that relates a population's short-term evolutionary response to the selection intensity and to the amount of genetic variation present. For a single trait, the 'breeders' equation' is given as

$$R = h^2 S$$

where R is the single-generation response to selection, h^2 is the trait heritability, and S is the selection differential (McGuigan 2006). R represents the change in the population's phenotypic mean for the trait from generation to generation, h^2 is the trait's heritability (i.e. the proportion of phenotypic variation that results from variation in

expression of the trait's constituent genes), and S is the difference between the phenotypic mean before selection and that of potential breeders that survive selection within the same generation.

To fully characterize the evolutionary consequences of selection, a single-trait approach is insufficient because some traits are genetically linked and therefore can respond to selection even if not directly exposed to it. A multivariate, discrete-generation form of the breeders' equation takes these trait relationships into account (Lande 1979):

$$\Delta z = GP^{-1}s$$

where Δz is a vector of changes in the phenotypic means for all the traits under consideration, G is the genetic covariance matrix composed of the additive genetic covariances among the traits within an individual, P^{-1} is the inverse of the phenotypic covariance matrix, and s is the vector of selection differentials ($P^{-1}s$ is a vector describing the multivariate selection gradient β). Because this equation relates phenotypic changes to the selection applied through the genetic structure underlying those phenotypes, it (together with its age-structured analogs – see Law 1991a) provides a more complete characterization of short-term phenotypic response to selection imposed by fishing (Law 1991a; Policansky 1993a; Hard 2004; McGuigan 2006).

Fisheries-induced evolution in salmonids

The critical roles of growth and maturation

Most salmonids mature over a range of ages and sizes (Hendry and Stearns 2004; Quinn 2005; Table 1). Their propensity to mature depends on growth and physiological state at any of several potentially critical points in the life history, as dictated by their developmental programs. In anadromous salmon, reproductive investment appears to depend on energy availability; in coho salmon (*Oncorhynchus kisutch*), for example, ovary mass, egg size, and egg number are highly correlated with growth rate during the final spring and summer prior to ovulation (Campbell et al. 2006). A positive relationship between egg size and adult body size often varies with marine growth but not size at smoltification. Fish might be expected to grow at different rates when heavily fished, for behavioral, ecological or energetic reasons (such as a reduction in density resulting from fishing mortality, or an increase in relative predator abundance; e.g. Healey 1980; Trippel 1995; Salvanes and Baliño 1998), but changes in growth and maturation will also depend on their genetic architecture, as well as on how concurrent environmental changes affect the energetics of growth and the allocation of resources to reproductive effort. Thériault

et al. (in press) show that migratory and reproductive patterns in anadromous brook charr (*Salvelinus fontinalis*) are likely to be influenced by mortality experienced at key points in the life cycle across the marine life-history transition. Fishing may therefore alter the size or age at which allocation of resources to gonads versus somatic tissues begins to shift. This, in turn, will affect the productivity of the population as well as the biomass available for harvest. Selection for faster growth might also affect rates of natural mortality by increasing foraging intensity and risk-taking behaviors (Lee 1912; Ricker 1969; Kristiansen and Svåsand 1998; Walker et al. 1998; Mangel and Stamps 2001).

The maturation process of anadromous salmonids is complex and protracted. Salmon initiate maturation well in advance of its phenotypic expression, apparently in response to physiological state or growth rate at a particular size or developmental stage (e.g. Thorpe 2007; Wright 2007). The consequences of selective fishing for growth and maturation may affect the onset of underlying developmental processes. Analysis of these effects using a PMRN typically invokes an assumption that maturation probability can be described by age and body size and therefore by average immature growth rate, but this assumes that the actual growth trajectory leading to a particular combination of age and size is unimportant. However, this is biologically implausible for most salmonids. In chum salmon (*Oncorhynchus keta* Walbaum), Morita and Fukuwaka (2006) found that probability of maturing was more closely linked to recent growth history than to body size (this example also shows how the PMRN approach can be extended with additional data). If the relationship between size and age is itself heritable, then the evolutionary consequences of fishing on size and age at maturation will depend on the shape of that relationship (Kuparinen and Merilä 2007). For example, if the reaction norm describing propensity to mature as a function of age (x) and size (y) is relatively flat (approaching size-constrained maturation, wherein fish tend to mature at the same size regardless of age), then fishing is expected to lead to faster growth and earlier maturation (Fig. 1A). By contrast, if this function is relatively steep (approaching age-constrained maturation, wherein fish mature at the same age irrespective of size, e.g. pink salmon, *O. gorbuscha*, and coho salmon, *O. kisutch*), then fishing could lead to slower growth and delay maturation (Fig. 1B). For age-structured salmonids, this relationship would be relatively flat, leading to a prediction that size-selective fishing will favor faster growth and younger adults. A more complex function (Perrin and Rubin 1990; Ernande et al. 2004) would have less predictable consequences.

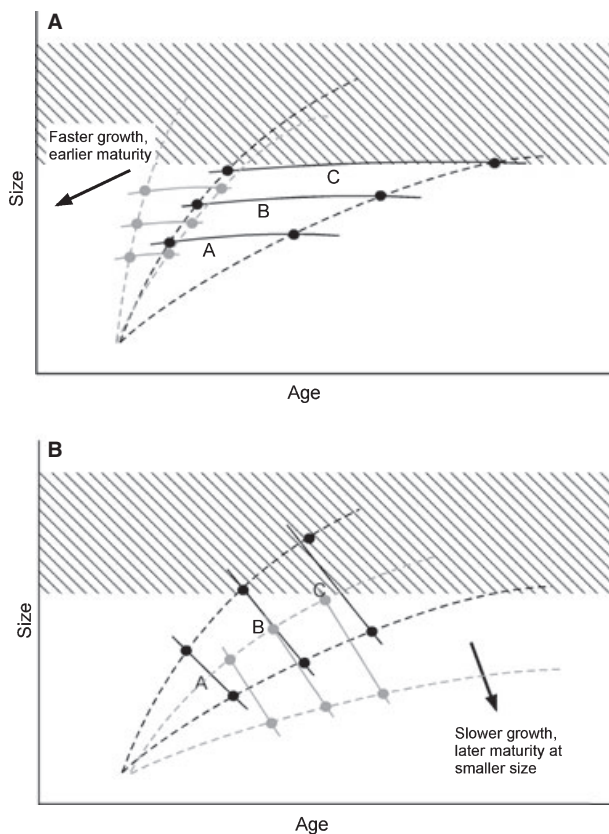


Figure 1 Hypothetical maturation reaction norms for size and age at maturation in salmonids under variable opportunities for growth. The dotted black curves depict hypothetical growth trajectories, from rapid (steep) to slow (shallow). In the strictest sense, reaction norms reflect phenotypic differences among distinct genotypes, although such functions are often used to evaluate patterns in other genetically differentiated groups. Here, A, B, C refer to distinct genotypes, families, or populations, with their maturation reaction norms indicated by the three solid curves in each pane. Solid black dots indicate the intersections of the growth trajectories and reaction norms for each group. (A) Maturation reaction norms corresponding to a primary influence of size on first maturation ('size-constrained maturation'). In this case the reaction norms are relatively flat, so that size selection imposed by fishing, indicated by the hatched area, is likely to increase growth rate and reduce size and age at first maturation in an exploited population. Possible responses in the reaction norms predicted by the arrow are given by the curves and dots in grey. This scenario appears consistent with the biology and phenotypic response of several species, such as Atlantic, Chinook, chum, and sockeye salmon, and steelhead and anadromous cutthroat trout (as well as some marine species such as cod and plaice). (B) Maturation reaction norms corresponding to a primary influence of age on first maturation ('age-constrained maturation'). In this case the reaction norms are more vertical, so that size selection imposed by fishing is likely to reduce growth rate, and perhaps increase age and reduce size at first maturation, in an exploited population. Possible responses in the reaction norms predicted by the arrow are given by the curves and dots in grey. This scenario is consistent with the biology of species with a constrained age structure, such as pink or coho salmon.

Ingredients of fisheries-induced evolution

Fishing as selection

The extent to which a population responds to fishery selection has some key prerequisites (Law 1991a; Hard 2004). First, fishing must be sufficiently strong to alter the distribution of phenotypes in the breeding population. Under constant fishing selectivity and genetic variability, higher fishing rates are more likely than lower rates to elicit an evolutionary response. If fishing selectivity is not sufficiently high to impose a detectable selection differential on size (or size at age), a short-term evolutionary response is less likely, although nonselective fishing mortality can still lead to evolution through changes in the maturation schedule (Policansky 1993a,b; Hard 2004). So too can accumulation of very small selection differentials that are repeated over the long time periods that fisheries can operate (tens or hundreds of years).

Fisheries that target maturing salmon concentrated near terminal areas are less likely to cause pronounced selection for age at maturation than those targeting immature fish migrating over ocean pathways, at least for semelparous populations or iteroparous populations with low rates of repeat spawning (Healey 1986). The primary reason for this is that fishing on semelparous individuals that have already made the physiological decision to mature will tend to have a reduced impact on age at maturation. Fisheries that target maturing fish expose all ages to the same mortality (subject to gear selectivity for size, etc.), while in fisheries that target immature fish, mortality is directly proportional to how long fish delay maturation once they become vulnerable to gear. Fisheries on immature individuals directly select for fish that mature earlier, or become vulnerable later, which might result in genetically based changes in reproductive output. Salmon fisheries in terminal areas, within rivers, or otherwise closely associated with aggregates of maturing fish are less apt to result in rapid evolutionary responses in age at maturation and correlated traits than those that are not (e.g. Kuparinen and Merilä 2007). Nevertheless, fishing on maturing individuals can alter other aspects of life history associated with size or age at maturation, including fecundity, egg size, redd size and depth, and nest defense (see van den Berghe and Gross 1989; Hamon et al. 2000; Hamon and Foote 2005).

Genetic variation in salmonid life history

Life history variation within and among populations of salmonids reflects both genetic and environmental sources of variation (Table 3; see also Carlson and Seamons in press). The genetic potential for key life history traits in salmon to respond to selection is high. However, few studies have examined specifically the genetic covariation

Table 3. Summary of heritability estimates for life history traits in anadromous salmonids likely to respond to fishing selection. With few exceptions, only studies involving narrow-sense estimates from correlation among relatives or response to selection in wild or hatchery-ranched, but not farmed, populations (i.e. considerable fraction of life cycle spent in wild and exposed to fishing mortality) are included. Data for only the species included in Table 2 are given here, and heritability estimates for disease resistance, juvenile behavior, and other traits are not included.

Species	Trait type	Description	Range of h^2	References	
Atlantic salmon	Body size/morphology	Juvenile length	0.04–0.79	Bailey and Loudenslager (1986); Garant et al. (2003); Refstie and Steine (1978)	
		Juvenile weight	0.10–0.89	Bailey and Loudenslager (1986); Jónasson et al. (1997)	
		Immature length	0.57–0.73	Bailey and Loudenslager (1986)	
		Immature weight	0.20–0.67	Bailey and Loudenslager (1986)	
		Mature weight	0.20–0.36	Jónasson (1993); Jónasson and Gjedrem (1997); Jónasson et al. (1997)	
	Survival	Marine survival	0.01–0.24	Jónasson et al. (1997)	
Chinook salmon	Body size/morphology	Juvenile length	~0.0–1.0	Hard et al. (1999); Bryden and Heath (2000)	
		Juvenile weight	0.99	Hard et al. (1999)	
	Growth rate	Development rate	0.05–0.23	Kinnison et al. (1998)	
		Age at maturation	0.30–0.57	Hankin et al. (1993); Hard (2004); Hard (1995)	
	Survival	Marine survival	~0.0–0.12	Unwin et al. (2003)	
	Migration or spawn timing	Maturation timing	0.23–1.0	Quinn et al. (2000); Hard (2004)	
	Egg number		~0.0–0.76	Kinnison et al. (2001)	
	Egg size		0.5–0.78	Kinnison et al. (2001)	
Chum salmon	Body size/morphology	Juvenile length	0.13–0.86	Beacham (1990); Kanno (1990)	
	Survival	Embryo/alevin survival	~0.0	Beacham (1988)	
Coho salmon	Body size/morphology	Juvenile length	~0.0–0.47	Murray et al. (1993)	
		Juvenile weight	~0.0–0.62	Withler and Evelyn (1990); Murray et al. (1993)	
		Immature length	0.32–0.69	Silverstein and Hershberger (1995)	
	Growth rate	Immature weight	0.07–0.85	Silverstein and Hershberger (1995)	
		Juvenile/immature	0.06–1.0	Sato (1980); Silverstein and Hershberger (1995); Vøllestad and Quinn (2003)	
	Age at maturation	Male precocity	0.05–0.13	Silverstein and Hershberger (1992)	
	Survival	Juvenile survival	~0.0–0.35	Beacham (1988); Murray et al. (1993)	
	Pink salmon	Body size/morphology	Mature length	~0.0–1.0	Smoker et al. (1994); Dickerson et al. (2005)
Mature weight			~0.0–0.66	Smoker et al. (1994)	
Survival		Embryo survival	~0.0–0.21	Beacham (1988)	
Migration or spawn timing		Return timing	~0.0–1.0	Smoker et al. (1998); Dickerson et al. (2005)	
		Spawn timing	0.06–0.54	Smoker et al. (1994)	
Egg number			~0.0	Funk et al. (2005)	
Egg size		0.22	Funk et al. (2005)		
Sockeye salmon	Body size/morphology	Gill raker count	0.57	Foote et al. (1999)	
Rainbow trout/steelhead	Body size/morphology	Immature length	0.11–0.58	McKay et al. (1986); Sylvén and Elvingson (1992); Thrower et al. (2004)	
		Immature weight		0.13–0.65	McKay et al. (1986); Sylvén and Elvingson (1992); Thrower et al. (2004)
				0.12–0.73	McIntyre and Blanc (1973); McKay et al. (1986); Thrower et al. (2004)
	Growth rate	Proportion smolting	0.45–0.73	Thrower et al. (2004)	
	Age at maturation	Early male maturation	0.02–1.0	Sylvén and Elvingson (1992); Thrower et al. (2004)	

among salmonid life history traits, which can constrain or augment selection response; virtually all studies of fisheries selection to date have focused on single characters. More recent studies that have focused on the genetic architecture of salmonid life history include analyses of growth, size and maturation (e.g. Smoker et al. 1994; Quinn et al. 2000; Kinnison et al. 2001; Hard 2004; Thrower et al. 2004), juvenile body size and shape (Kanno 1990; Hard et al. 1999), and pathogen resistance (Withler and Evelyn 1990; Fjalestad et al. 1996; Guy et al. 2006; Hard et al. 2006). Genetic correlations are difficult to estimate with precision, especially without adequate breeding designs, and such estimates are not available for most exploited populations. Nevertheless, in general these analyses suggest that the indirect responses of traits to selection depend critically on their genetic and phenotypic covariances and that these will be difficult to predict solely from phenotypic information on the trait subject to direct selection (McGuigan 2006; Law 2007).

Few studies have provided estimates of selection differential imposed by fishing. Some of the best known estimates have been derived for body length in Atlantic cod, which varied from -1 to $+2$ cm for North Sea cod (Law and Rowell 1993) and from -4 to $+4$ cm for cod from Canadian catches (Sinclair et al. 2002). For Atlantic salmon, Hindar et al. (2007) provided estimates of selection differential on body weight for one-sea winter (1SW) grilse ranging from -0.08 to -0.52 kg, depending on the population and year. For Pacific salmon, Ricker (1981) estimated that the selection differential imposed by fishing on British Columbia coho salmon body weight between 1951 and 1975 varied from -0.50 to -0.73 kg. Hamon et al. (2000) estimated that the Bristol Bay (Alaska) gillnet fishery imposed selection differentials on body girth in sockeye salmon that ranged from -0.6 to -3.6 mm for females and -3.6 to $+0.3$ mm for males. Analyses by Washington Department of Fish and Wildlife (WDFW) biologists of coho salmon caught in gillnets in Washington state in recent years indicate that selection differentials on body length varied from -3.3 to $+0.2$ cm for females and -5.8 to $+0.2$ cm for males (C. Knudsen and C. Busack, WDFW, personal communication). Unfortunately, these estimates were not standardized, so direct comparisons are difficult, but from available information most selection differentials estimated for fishing appear to be in the range of ~ 0 to ± 0.5 phenotypic standard deviations.

The combination of selection differentials with estimates of heritability for these traits indicates that responses in salmon size, growth, and maturation age to fishing-induced selection are likely to vary considerably among populations and over time. In most cases, these responses are expected to be modest over the short term

(ca. 10 or fewer generations), although they could potentially be as large as -1 cm for length and -100 g for weight on an annual basis under stable environmental conditions. That said, the estimates of selection differentials tend to be similar to, but perhaps usually lower than, estimates of selection intensity imposed by natural and sexual selection in naturally reproducing salmon populations, which can sometimes exceed 0.5 standard deviations (van den Berghe and Gross 1989; Hamon and Foote 2005).

Only a few investigations have explored the consequences of such trait architecture under selection for viability. Hankin and Healey (1986) found that selective fisheries can decrease the mean age of Chinook salmon populations and increase the probability of significant population decline. The results of simulations of fisheries-induced evolution by Hard (2004) suggest that the selective exploitation of large Chinook salmon could lead to modest reductions in size-at-age within approximately five generations; further exploratory modeling (Hard et al., unpublished data) has shown that such responses can reduce abundance and catch and produce some maladaptive changes in life history that are likely to increase risk to population viability.

Evidence for fisheries-induced evolution in salmonids

The selectivity of fishing on many fitness traits in salmonids, coupled with the ample evidence of underlying genetic variation in these traits, indicates that rapid evolutionary responses to fishing are possible. Several studies over the past quarter century have explored the potential evolutionary effects of fishing on salmon (e.g. Ricker 1981, 1995; Hankin and Healey 1986; Healey 1986; Riddell 1986; Altukhov 1994; Hard 2004; Morita et al. 2005; Quinn et al. 2007). In a recent perspective, Jørgensen et al. (2007) identified 46 studies involving six traits in 18 fish species that implied fishing-induced evolution and estimated appreciable rates of evolutionary change. For salmon, these studies involved five species, and provided evidence for evolutionary rates from less than 20 to more than 30% over 24 years (on the order of 1% change annually). However, since the design of the study by Jørgensen et al. (2007) excluded research which did not suggest evolutionary change, the overall effects of fisheries-induced evolution are likely to be less than this.

Ricker's (1981, 1995) pioneering analysis of changes in mean weight of several Canadian species of Pacific salmon *Oncorhynchus* spp. (and in mean age for Chinook salmon, *O. tshawytscha*) caught between 1950 and 1993 raised concerns about future fishery yields. Ricker (1995) concluded that the effects of size-selective fishing were

complex and difficult to disentangle from other factors affecting survival and growth in age-structured species, but that fisheries-induced genetic changes were likely, especially in pink and coho salmon (Ricker et al. 1978; Ricker and Wickett 1980). In a separate study, terminal fisheries with minimum size limits did not lead to changes in mean length in many populations of Pacific salmon in Canada, and observed changes were probably not genetic but due to environmental variation (Healey 1986; see also Riddell 1986). Summers (1995) and Friedland et al. (2000) observed changes in life history for Atlantic salmon consistent with temporal variation in marine environmental conditions. In an Asian fishery, chum salmon exhibited a decrease in the mean size at age of mature individuals, and an increase in the age at maturation, after the fishery switched from a high-seas gill-net fishery to a terminal set-net fishery (Morita et al. 2005).

Healey (1986) concluded that observed declines in the size of Pacific salmon previously attributed to selective fisheries probably also reflect changes in climate affecting marine growth and productivity of salmon, and he and Riddell (1986) identified several factors that tend to limit detection of an evolutionary response to fishing. First, the data may be inadequate or of low quality. The characteristics of many fisheries and of much of the associated catch data, such as those considered by Ricker (1981, 1995), do not lend themselves well to genetic analysis because of variable stock composition of the catch, because the data suitable for monitoring are limited, and because selection differentials are not easily quantified. Second, the environmental contribution to variation in size and age is likely to be large. Third, the genetic structure of size, age and correlated traits can constrain response to selection. Fourth, the consequences of tetraploid ancestry in salmonids for genetic variation and evolutionary dynamics are still not well understood. Fifth, response to selection can be complex for age-structured species due to variation in selection differentials for fish maturing at different ages, and specifically tailored life history models are required to adequately capture the evolutionary dynamics of salmonids (whether iteroparous or semelparous). Finally, countervailing selection in the wild (e.g. natural and sexual selection on spawners) might oppose fishing selection (Healey 1986; Riddell 1986; Carlson et al. 2007).

Hamon et al. (2000) found that selectivity in gillnet fisheries can impose strong selection on adult body morphology (girth). The magnitude and direction of this may vary as well (Miller and Kapuscinski 1994). In the Yukon River, Alaska, which historically produced appreciable numbers of large, old Chinook salmon, the numbers of very large (≥ 90 cm) fish have been declining in recent

decades (Hyer and Schleusner 2005). Declines in body size can affect fertility (Healey and Heard 1984), mate choice and breeding behavior (Quinn and Foote 1994; Esteve 2005), and redd construction and defense and subsequent fry survivorship (van den Berghe and Gross 1989; Steen and Quinn 1999).

Some authors have also argued that fishing can affect migration timing (Quinn et al. 2007). For example, Quinn et al. (2002) demonstrated that run timing of both Chinook and coho salmon from three hatcheries in Washington has shifted in recent decades as a result of selection of brood stock which has responded to fishing patterns. For Atlantic salmon in Ireland, Quinn et al. (2006) documented a long-term delay in run timing, as well as a decline in weight, changes which they argued probably resulted from patterns of angling pressure on returning adults.

These studies point to the importance of considering selective fishing as a factor in altering salmon life history. Unfortunately, most inferences about fishing selection are based on an evaluation of selectivity or fishing mortality rate and therefore focus on only one aspect of adaptive evolution: the opportunity for directional change through an apparent measure of selection intensity. Because evolution involves change in gene frequencies, an evolutionary response requires genetic variability, and inferring evolution in response to fishing pressure in the absence of this information is far from straightforward.

What we need to know about fisheries-induced evolution

The changes in life history observed in many exploited fish populations are fueling controversy among biologists and conservationists over whether these fisheries and the populations that support them can persist (e.g. Birkeland and Dayton 2005; Kuparinen and Merilä 2007). Our review of this body of work in salmon, summarized in Table 2, suggests one reason: none of these studies provides direct evidence for evolutionary responses to fishing, or whether such responses reduce viability. Nevertheless, the collective evidence across a variety of species and environmental conditions highlights trends in size, age, and other traits – traits that have large influences on productivity and fitness – that are consistent with evolutionary responses to size-selective fishing (International Council for the Exploration of the Sea (ICES) 2007; Jørgensen et al. 2007). As Law (2007) noted, such responses may often be modest over the short term and difficult to detect without evaluating longer trends.

A concerted empirical attempt to dissect effects of fishing from those of other factors is clearly warranted. This would include careful experiments to discriminate these

factors in a real-world – spatially or temporally structured but carefully monitored – evaluation of fishing effects on abundance, size, and life history of free-ranging salmon, such as the experiments suggested by McAllister and Peterman (1992) and McAllister et al. (1992). These authors emphasized the difficulty in evaluating fishing effects empirically but provided valuable guidance for how to structure the necessary experiments with adequate statistical power. Among their recommendations are to focus on species with simple life histories, such as pink salmon, and to employ adequate spatial replication; both of these recommendations can improve power considerably for relatively short-term (~5 generation) experiments.

Large-scale manipulative experiments and evaluation of management strategies in the context of conceivable responses in the fishery (see Walters 1994) are logistically and politically challenging to implement, and would be met with resistance from fishers without adequate compensation or a clear sense of a perceived longer-term benefit. Nevertheless, such approaches, when coupled with data on trends and knowledge of selectivity and genetics, would be more convincing to scientists and more compelling to managers. As suggested by Wright (2007), additional lines of inquiry that would likely prove profitable include comparisons of patterns of reproductive investment and allocation among populations varying in exploitation history, and contrasts of state-dependent thresholds for maturation among populations that differ in exploitation history.

Implications for fisheries management

Conceptually, the simplest way of reducing fisheries-induced selection pressures and consequences of excessive exploitation in general is to reduce overall fishing pressure. However, such overall reductions are hardly ever practical, and more specific measures are probably required. Considerable discussion in the literature has focused on the merits of minimum size limits, slot limits, and other fishing strategies as means to maintain current and preserve future yields. Some researchers have argued that minimum size limits tend to lead to 'recruitment overfishing,' whereas practices that increase catch of smaller, younger fish tend to lead to 'growth overfishing,' which is often thought to have less deleterious impacts on productivity and yield (Ricker 1976; Larkin 1978).

Other management options such as fisheries moratoria, time and area closures or catch limitations, and marine reserves also merit consideration to reduce long-term effects of fishing. Baskett et al. (2005) showed in a quantitative genetic model mimicking a cod life history that marine protected areas (MPAs) could help to reduce fish-

eries-induced selection for size at maturation in some long-lived species. MPAs may, however, have limited utility in mitigating for evolutionary change in highly migratory fish unless reserves are very large or are carefully networked. Protecting adults on spawning grounds could favor earlier maturation for some life histories, for example (Law 2007). In practice, the boundaries of MPAs that will be effective for anadromous salmon might not be difficult to identify but they will be difficult to implement and manage.

The socioeconomic factors that maintain exploitation are unlikely to ease until a clear biological threat is identified. Because resistance to reducing fishing rates will remain high in such circumstances, reducing fishing selectivity should become a tool for management, as reducing selectivity will preserve genetic and life history variability. Even so, as Policansky (1993b) pointed out, it must be recognized that a nonselective fishery will affect a population's evolutionary trajectory to the extent that it alters the mortality schedule. The key issue is where and how intense these pressure points are exerted by fishing on the mortality schedule relative to growth and maturation profiles.

Discussions that solely focus on productivity and yield often overlook the importance of standing genetic variation for size and associated life history traits to the resilience of an exploited population (Nelson and Soulé 1987). If exploited populations are to cope with the ecological and evolutionary pressures posed by fishing they must retain the adaptive capacity to respond. This capacity may be threatened by several of the characteristics of size-selective exploitation, especially selective removal of individuals with higher reproductive potential and elevation of the rate of stochastic genetic processes through reduction of genetic diversity (Smith et al. 1991; Harris et al. 2002).

The potential consequences of fisheries-induced evolution to viability of salmon remain poorly understood. Adaptation to fishing might reduce vulnerability of salmon to fisheries and thereby improve population viability compared to a hypothetical situation where evolution is not permitted. However, fishers may quickly adjust their capture strategies to changing fish characteristics, thereby engaging in a co-evolutionary 'arms race' and eradicating potential viability benefits (Heino 1998). Furthermore, when fish adapt to fishing they are likely to evolve away from configurations that natural and sexual selection alone would favor.

For example, a modest genetic influence on size at age ($h^2 \sim 0.3$) appears to permit adaptation of Chinook salmon to selection on size imposed by fishing (Hard et al., unpublished data); this adaptation is generally expressed as increased growth rate and earlier age at maturation,

which tends to reduce fishery vulnerability under a fixed minimum capture size threshold. Such adaptation can take considerable time, however – several to many generations. During this period, evolution is likely to entail a period of reduced fitness, which Walsh et al. (2006) referred to as repayment of a Darwinian ‘debt.’ Thus, under some circumstances, an evolutionary response to fishing selection can negatively affect population viability. Fisheries-induced evolution might compound the demographic risk posed by overfishing, and this evolutionary trend may be difficult to reverse. Yield might decline, and vulnerability to other threats to viability during this period is likely to remain high (Hard et al., unpublished data).

The consequences of environmental variation for fish growth, size, and age at maturation might – at least in some years – overwhelm the impacts of fishing. Indeed, what cannot be determined yet is whether the selection imposed by fishing, the evolutionary response to it, and any attendant effects on viability will be sufficiently large to precipitate a fishery collapse. It is unfortunate that our knowledge of the long-term consequences of fishing in salmon has not changed appreciably since the commentaries of Larkin (1978) and the reviews by Nelson and Soulé (1987) and Policansky (1993a,b). Whether fishing selection on salmon is in most cases intense enough to pose a problem for long-term management and conservation remains unclear, but it behooves managers, in the spirit of the precautionary principle, to work with scientists to incorporate the possibility in management planning.

Given sufficient genetic variability and a stable fishing regime, salmon populations will evolve in ways that reduce fishing mortality (and yield), primarily by increasing growth rate (and, in species with complex age structure, potentially accelerating the maturation schedule). Short-term adaptation will probably not be enough to compensate for the loss of aggregate yield due to size-selective fishing. Two assumptions that are critical to recovery of exploited populations suffering from changes caused by fisheries-induced evolution are that genetic variability in size and age is not eroded by fishing selection, and that productivity is not depleted by fishing-induced changes in size and age of spawners. Quantitative genetic models indicate that aggressive reduction of fishing mortality to a fraction of initial values within several generations might be sufficient to permit an exploited salmon population to show recovery of abundance, but achieving pre-fishing maturation schedules and size distributions after adaptation to fishing mortality can take a very long time (Hard et al., unpublished data).

Salmon are unique among exploited fishes in the scale on which cultured individuals are released from

hatcheries to the wild where they can be caught in fisheries or potentially spawn with naturally reproducing fish. Thus, for many stocks the hatchery and fishery regimes must be considered components of an integrated management system. To what degree selection in hatchery fish (i.e., domestication) and natural and sexual selection on spawning grounds might alter responses to fishing selection remains unclear. Understanding how hatchery and wild fish might differ in response to fishing and how domestication in hatcheries may degrade fitness of wild fish that interbreed with hatchery fish is critical to the development of sustainable hatchery production-fishing systems for salmon. For example, it is possible that fish spawned in hatcheries might have different short-term responses than fish spawning in the wild to fishing selection owing to the relaxation of natural and sexual selection on adult size in hatchery fish at time of spawning, but this issue remains unexplored (Hard 2004).

Conclusions

Do we know enough about the genetic effects of fishing on salmonids to justify reassessing current approaches to managing them? We believe so. Our survey of the literature indicates that the opportunity for fishing selection is amply demonstrated, even if it does not yet provide unambiguous evidence for rapid evolution. There are three critical uncertainties: whether trends in life history of exploited salmon are genetically based (Kuparinen and Merilä 2007), how quickly fisheries-induced evolution might occur, and whether such evolution is ‘reversible’ through management responses. Addressing these uncertainties is necessary to develop management regimes that are most effective in limiting evolutionary change caused by fishing. Meantime, a precautionary approach to fishery management that limits opportunity for adverse fishing selection is clearly warranted, and we recommend that this approach incorporate sufficient monitoring of key demographic parameters and life history traits such as run size and timing, escapement, size at age, and reproductive condition (Kuparinen and Merilä 2007). Fisheries management that promotes reduced gear selectivity with respect to size to allow sufficient larger, older individuals to breed, and focuses fishing activity on mature individuals in areas close to spawning grounds to reduce directional selection on maturation will provide some benefits to exploited populations. This is likely to be particularly important for species with restricted age structure, such as coho salmon, where variation in size of individuals vulnerable to size-selective fishing directly reflects variation in marine growth rate and high harvest rates could

impose substantial negative selection differentials on growth and size at maturation.

Several researchers (e.g. Law 1991b, 2007; Heino 1998; Jørgensen et al. 2007; Kuparinen and Merilä 2007; Hutchings and Fraser 2008) are urging managers and scientists to coordinate in developing management schemes that directly account for fisheries-induced evolutionary change. The weight of evidence from the large number of studies summarized in Table 2 and the estimates of heritability in Table 3 indicate that we can be confident that evolution is being caused by fishing even if none of the individual studies is entirely conclusive. It is time to incorporate evolutionary principles into the management of salmon fisheries.

Acknowledgements

We thank Robin Waples (Northwest Fisheries Science Center), Raymond Huey (University of Washington), and Joseph Travis (Florida State University) for organizing a workshop on 'Evolutionary consequences of anthropogenic changes on long-term viability of Pacific salmon and steelhead' in Seattle, Washington in December 2006, where the ideas for this paper were generated. We are also grateful to the workshop participants for their contributions to the discussion of the ideas presented here, and to two anonymous reviewers for their very helpful comments on an earlier draft.

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