

# Low dose ionizing radiation produces too few reactive oxygen species to directly affect antioxidant concentrations in cells

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**It has been hypothesized that radiation-induced oxidative stress is the mechanism for a wide range of negative impacts on biota living in radioactively contaminated areas around Chernobyl. The present study tests this hypothesis mechanistically, for the first time, by modelling the impacts of radiolysis products within the cell resulting from radiations (low linear energy transfer  $\beta$  and  $\gamma$ ), and dose rates appropriate to current contamination types and densities in the Chernobyl exclusion zone and at Fukushima. At  $417 \mu\text{Gy h}^{-1}$  (illustrative of the most contaminated areas at Chernobyl), generation of radiolysis products did not significantly impact cellular concentrations of reactive oxygen species, or cellular redox potential. This study does not support the hypothesis that direct oxidizing stress is a mechanism for damage to organisms exposed to chronic radiation at dose rates typical of contaminated environments.**

**Keywords:** Chernobyl; Fukushima; radiation; oxidative stress; biota; cell

## 1. INTRODUCTION

Oxidative stress results from ‘a mismatch between the production of damaging reactive oxygen species (ROS) and the organisms’ capacity to mitigate their damaging effects’ [1, p. 75]. At high dose rates, cellular oxidizing stress plays an important role in cell damage from ionizing radiation, and antioxidants may have a protective effect at such dose rates. For example, injection of vitamin E ( $\alpha$ -tocopherol) increased 30-day survival rates of mice exposed to acute high dose rate ( $1.2 \times 10^7 \mu\text{Gy h}^{-1}$ ) of low linear energy transfer (LET) radiation [2].

Significant radiation-induced oxidative stress of flora and fauna at lower dose rates (here defined as up to *ca*  $400 \mu\text{Gy h}^{-1}$  from internal and external sources) has also been hypothesized. Plant responses have been linked to oxidizing stress and antioxidant capacity in field and experimental studies [3]. Significantly, lower antioxidant concentrations have been observed in birds (barn swallow, *Hirundo rustica*;

great tit, *Parus major*) inhabiting areas contaminated by Chernobyl [4]—an effect attributed to radiation-induced oxidative stress. This hypothesis is supported by observations of decreased levels of the antioxidants retinol,  $\alpha$ -tocopherol and carotenoids in blood plasma, liver and egg yolk of barn swallows living near Chernobyl [4].

However, the relationship between antioxidant concentrations and oxidative stress is complex [1]. For example, studies in plants have found increased concentrations of antioxidant enzymes with radiation exposure [5] at low dose rates. Other low dose rate studies have found no changes in antioxidant concentrations either in plants [6] or birds [7], although the latter did observe a significant difference in metabolites produced by reactive oxygen (ROM). Recently, Bonisoli-Alquati *et al.* [8] found that ‘oxidative damage of sperm was negatively related to sperm motility’ in birds exposed to radiation at Chernobyl, but that ‘the highest values [of high sperm motility] were associated with relatively high radiation levels’ (p. 105).

The low radiation dose rate oxidizing stress hypothesis has not, to our knowledge, yet been tested at a mechanistic level. The present study tests this hypothesis (using previously published data on oxidizing stress in birds at Chernobyl) by modelling, for the first time, the capacity of selected antioxidants to reduce radiolysis products at radiation dose rates appropriate to current contamination densities pertaining at Chernobyl and Fukushima.

## 2. MATERIAL AND METHODS

For a given dose rate,  $D$  ( $\text{Gy s}^{-1}$ ), we can calculate the rate of production of ion pairs per unit mass of an organism by considering the radiolysis products of water when exposed to ionizing radiation. These include  $\text{H}_2$ ;  $\text{H}_2\text{O}_2$ ;  $\text{e}_{\text{aq}}^-$ ;  $\text{H}^\cdot$ ;  $\text{OH}^\cdot$ ;  $\text{HO}_2^\cdot$  (see electronic supplementary material). For low LET radiation,  $G$ -values giving radiolysis products per Joule of absorbed radiation energy are given in electronic supplementary material, table S1 (from Choppin *et al.* [9]).

The rate of production of each radiolysis product,  $r_p$ , per unit mass of tissue ( $\text{mol kg}^{-1} \text{s}^{-1}$ ) is

$$r_p = \theta DG, \quad (2.1)$$

where  $\theta$  is the fractional water content of the tissue.

Here, we investigate the potential impact of radiolysis products on cellular antioxidant concentrations. Assuming that the rate of replenishment of antioxidant molecules occurs at a rate proportional to the difference between the current concentration,  $C_T$  ( $\text{mol l}^{-1}$ ) and a ‘target’ equilibrium concentration,  $C_{TE}$  ( $\text{mol l}^{-1}$ ), the following equation describes the rate of change of  $C_T$ :

$$\frac{dC_T}{dt} = r_f(C_{TE} - C_T) - r_p, \quad (2.2)$$

where  $r_f$  is the fractional replenishment rate of antioxidant molecules ( $\text{s}^{-1}$ ). For boundary condition  $C_T = C_{TE}$  at  $t = 0$ , this has solution:

$$C_T = C_{TE} - \frac{r_p}{r_f}(1 - e^{-r_f t}). \quad (2.3)$$

The impact of radiolysis on the redox status of the cell can be quantified by considering the impact on redox potential,  $E_h$  (volts) of decreased glutathione (GSH) owing to oxidation to oxidized glutathione (GSSG) concentrations [10,11]:

$$\left. \begin{aligned} &\text{GSSG} + 2\text{H}^+ + 2\text{e}^- \rightarrow 2\text{GSH} \\ &\text{and} \\ &E_h = E_m - \frac{RT}{nF} 2.3 \log \left( \frac{[\text{GSH}]^2}{[\text{GSSG}]} \right), \end{aligned} \right\} \quad (2.4)$$

where  $E_m$  is the mid-point potential ( $-0.240 \text{ V}$  at pH 7; [11]),  $R$  is the gas constant ( $8.314 \text{ J K}^{-1} \text{ mol}^{-1}$ ),  $T$  is temperature (Kelvin),  $F$  is the Faraday constant ( $9.6485 \times 10^4 \text{ C mol}^{-1}$ ),  $n$  is the number of electrons involved in the redox of the couple, in this case, two.

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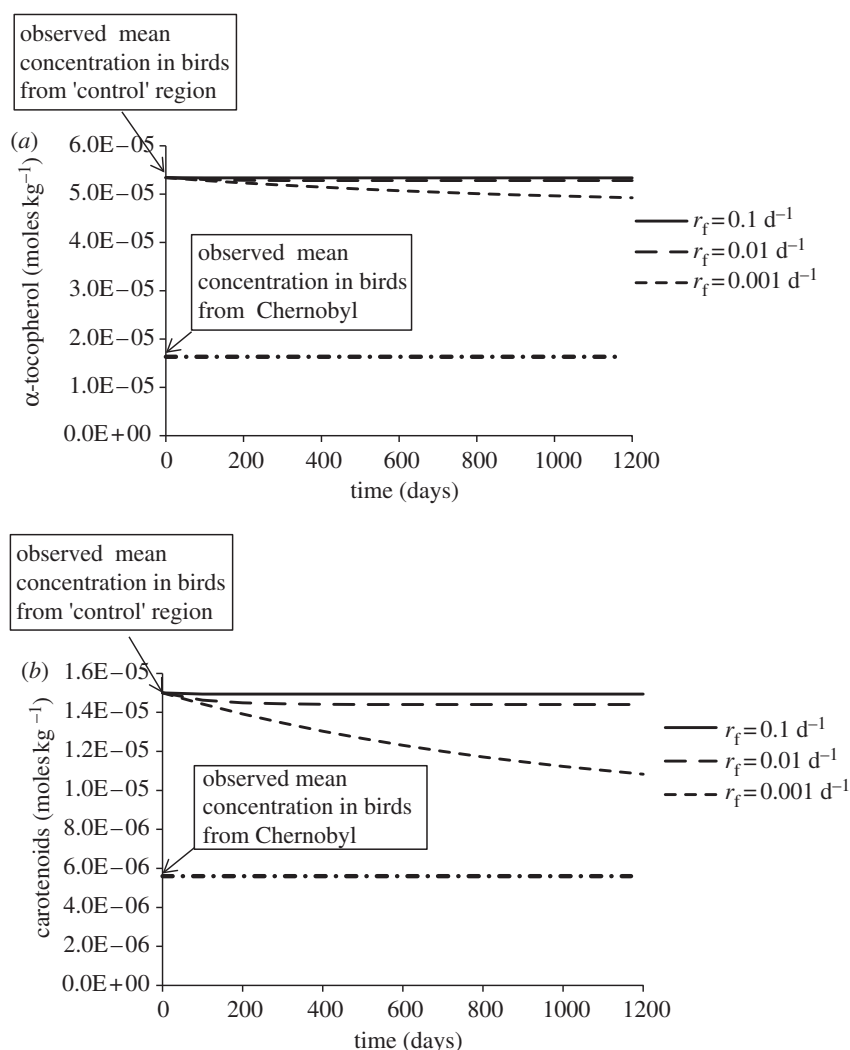


Figure 1. Changes in (a)  $\alpha$ -tocopherol and (b) carotenoids in birds' liver as a result of  $417 \mu\text{Gy h}^{-1}$  ionizing radiation (many times higher than the mean at the Chernobyl study sites; see electronic supplementary material). (Anti-oxidant concentrations were estimated from Møller *et al.* [4].)

### 3. RESULTS

#### (a) Effect of radiolysis on cellular antioxidant concentrations

A study by Møller *et al.* [4] observed that barn swallow liver cell  $\alpha$ -tocopherol and carotenoid concentrations decreased at sites near Chernobyl (contaminated with approx.  $3.9 \mu\text{Gy h}^{-1}$  external dose) compared with a control site. Equation (2.3) was used to determine whether radiolysis could account for these changes. Setting the equilibrium antioxidant concentration  $C_{\text{TE}}$  to that of the control site, and assuming that the cell has minimal antioxidant capacity (i.e. no other enzymatic or non-enzymatic antioxidants operate in the cell), the change in concentration,  $C_{\text{T}}$  can be calculated. Figure 1 shows the effect of ionizing radiation of  $417 \mu\text{Gy h}^{-1}$  (much higher than the birds were exposed to; see electronic supplementary material) on cellular  $\alpha$ -tocopherol and carotenoid concentrations for three different fractional rates of replenishment: 0.001, 0.01 and  $0.1 \text{ d}^{-1}$ .

#### (b) Effect of radiolysis on cellular redox potential

The effect of radiolysis on cellular redox potential was investigated by again assuming the cell has minimal antioxidant capacity, i.e. just GSH and no

other enzymatic or non-enzymatic antioxidants. Equation (2.3) was used to determine reduction in GSH concentration as a function of time following chronic exposure to  $417 \mu\text{Gy h}^{-1}$  at low LET radiation. Equation (2.4) was used to calculate the impact of this change in the GSH–GSSG balance on the redox potential of the cell. We have here assumed a hypothetical GSH concentration of 1 mM and have calculated the GSH/GSSG concentration starting with 1 per cent conversion of GSH to GSSG. Glutathione concentrations in cells typically range from 1 to 11 mM [11–13]. Figure 2 shows the change in GSH concentration and cellular redox potential for 1200 days exposure to  $417 \mu\text{Gy h}^{-1}$  low LET radiation assuming an unrealistically low ( $0.001 \text{ d}^{-1}$ ) replenishment rate of GSH.

#### (c) Comparison of rate of production of reactive oxygen species with reactive oxygen metabolite concentrations

The rate of production of ROS from radiolysis (table 1) is compared with measurements from barn swallows at a contaminated (up to  $2.9 \mu\text{Gy h}^{-1}$ ) site at Chernobyl using data presented in Bonisoli-Alquati *et al.* [7].

Table 1. Rate of production ( $\text{mol kg}^{-1} \text{s}^{-1}$ ) of radiolysis products of water at different exposures to  $\gamma$ - and high-energy  $\beta$ -radiation. (A typical cellular water content of  $\theta = 0.8$  is assumed.)

dose rate ( $\text{mGy d}^{-1}$ )	dose rate ( $\mu\text{Gy h}^{-1}$ )	$\text{H}_2$	$\text{H}_2\text{O}_2$	$e_{\text{aq}}^-$	H	OH	$\text{HO}_2$	$\Sigma_{\text{ROS}}$
1	41.7	4.35E-16	6.76E-16	2.59E-15	5.74E-16	2.59E-15	2.5E-17	6.9E-15
10	417	4.35E-15	6.76E-15	2.59E-14	5.74E-15	2.59E-14	2.5E-16	6.9E-14
100	4170	4.35E-14	6.76E-14	2.59E-13	5.74E-14	2.59E-13	2.5E-15	6.9E-13

Table 2. Daily production of ROS by radiolysis at  $10 \text{ mGy d}^{-1}$  ( $417 \mu\text{Gy h}^{-1}$ ) compared to differences in ROM (concentration of hydroperoxides) in the plasma of barn swallows between contaminated and control sites (data from [7]).

ROM $\text{mM H}_2\text{O}_2$ equivalents				rate of production of ROS	
sex	contaminated ( <i>ca</i> $3 \mu\text{Gy h}^{-1}$ )	control	difference ( $\Delta\text{ROM}$ )	rate of production of ROS ( $\text{mM d}^{-1}$ )	$\Delta\text{ROM}$ ( $\text{d}^{-1}$ )
males	2.45	2.03	0.42	$5.96 \times 10^{-6}$	$1.42 \times 10^{-5}$
females	2.92	1.97	0.95	$5.96 \times 10^{-6}$	$6.27 \times 10^{-6}$

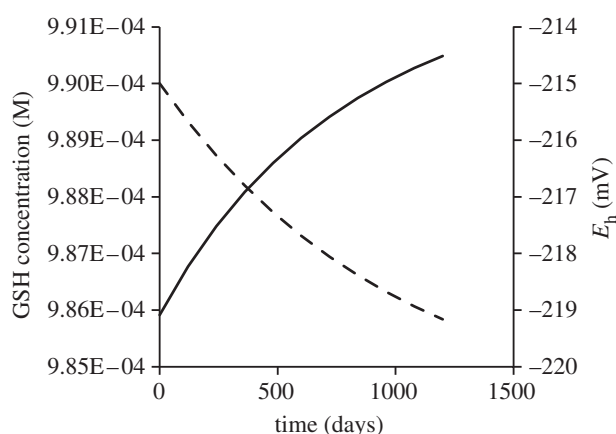
Figure 2. Predicted changes in glutathione (GSH) concentration (dashed line) and cellular redox potential ( $E_h$ ) following 1200 day exposure to  $10 \text{ mGy d}^{-1}$  ( $417 \mu\text{Gy h}^{-1}$ ) ionizing radiation assuming initial  $1 \text{ mM}$  total (GSH + GSSG) and an unrealistically slow glutathione replenishment rate of  $0.001 \text{ d}^{-1}$ . Solid line,  $E_h$  (mV).

Table 2 compares the daily rate of production of ROS by  $417 \mu\text{Gy h}^{-1}$  radiation with the difference in ROM between a contaminated and a control site. It can be seen (table 2) that daily rate of production of ROS by ionizing radiation represents a minuscule fraction (*ca*  $10^{-5}$ ) of the difference in ROM between contaminated (external dose rate of  $2.9 \mu\text{Gy h}^{-1}$ ) and control sites observed by Bonisoli-Alquati *et al.* [7]. As this study [7] measured only hydroperoxides, the difference between radiation-induced ROS production and ROM would in reality be greater.

#### 4. DISCUSSION

We calculated the direct effects of radiolysis on antioxidant concentrations at a total (external + internal) radiation dose rate of  $417 \mu\text{Gy h}^{-1}$  low LET radiation,

representative of the highest doses to organisms in the Chernobyl zone (see electronic supplementary material) and also relevant to current contamination densities and dose rates at Fukushima [14]. No significant changes in antioxidant concentrations or cellular redox potential were calculated. Assuming that only single antioxidants were used to reduce radiolysis products, and that fractional replenishment rates were as (unrealistically) low as  $0.001 \text{ d}^{-1}$ , antioxidant concentrations observed in 'control' birds are not reduced to those of exposed birds over 1200 days (figure 1). Differences in ROM between contaminated and control sites [7] cannot be explained by direct effects of radiolysis because the observed differences are orders of magnitude larger than the rate of production of ROS by radiolysis. The functional replenishment rate of glutathione in a range of animal tissues under different dietary conditions ranges from 10 per cent to 100 per cent per day [15] highlighting how conservative our assumptions are. Furthermore, Schafer & Buettner [11] suggest that changes in redox potential that cause cell changes including, sequentially, proliferation, differentiation, apoptosis and necrosis need to be of order of  $60 \text{ mV}$ , whereas changes calculated here are less than  $5 \text{ mV}$  over a 1200 day period.

We note that despite the minor direct impact of radiation on redox status of the cell and on antioxidant concentrations, it is well known that even low dose ionizing radiation can cause negative effects via DNA damage. Such damage is direct—caused by strand breaks and deletions—or indirect, from the free-radical products of water radiolysis in the immediate vicinity of nucleotides. At dose rates of order of  $417 \mu\text{Gy h}^{-1}$  (representing the most contaminated parts of the Chernobyl exclusion zone), radiation effects on organisms would be expected, and have indeed been observed [16,17]. The present study shows that observed effects are unlikely to be due to radiolysis products directly causing oxidative stress, significantly clarifying discussions about low-level

radiation and oxidative stress. Thus, while some radiation effects on organisms are likely (although see, for example, Wickliffe *et al.* [18]) at dose rates pertaining in the most contaminated sites at Chernobyl, the results of our study do not support the hypothesis [4,7,8] that direct oxidizing stress is the damage mechanism. It may also help us to explain the variety and inconsistency of radiation-induced antioxidant responses apparently observed at low doses. Although not directly tested against data from organisms other than birds, these results are also likely to apply to other organisms: direct generation of radiolysis products is not nearly high enough to affect oxidative stress even though there is variation in antioxidative capacity between different organisms. Some of the studies on birds take account of habitat differences between sites of different contamination level, but it seems more probable that differences in habitat, diet or ecosystem structure are associated with changed antioxidant concentrations rather than the direct effects of radiolysis products.

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