

Mercury-Selenium Relationships in Liver of Guiana Dolphin: The Possible Role of Kupffer Cells in the Detoxification Process by Tiemannite Formation

José Lailson-Brito^{1,2*}, Renato Cruz^{3†}, Paulo Renato Dorneles^{1,2}, Leonardo Andrade⁴, Alexandre de Freitas Azevedo¹, Ana Bernadete Fragoso^{1,5}, Lara Gama Vidal¹, Marianna Badini Costa², Tatiana Lemos Bisi¹, Ronaldo Almeida⁶, Dario Pires Carvalho⁶, Wanderley Rodrigues Bastos⁶, Olaf Malm²

1 Laboratório de Mamíferos Aquáticos e Bioindicadores “Profa. Izabel Gurgel” (MAQUA), Faculdade de Oceanografia, Universidade do Estado do Rio de Janeiro (UERJ), Rio de Janeiro, Rio de Janeiro, Brazil, **2** Laboratório de Radioisótopos “Eduardo Penna Franca”, Instituto de Biofísica Carlos Chagas Filho, Universidade Federal do Rio de Janeiro (UFRJ), Rio de Janeiro, Rio de Janeiro, Brazil, **3** Departamento de Biologia Celular e Molecular, Universidade Federal Fluminense (UFF), Niterói, Rio de Janeiro, Rio de Janeiro, Brazil, **4** Instituto de Ciências Biomédicas, Universidade Federal do Rio de Janeiro (UFRJ), Rio de Janeiro, Rio de Janeiro, Brazil, **5** Instituto de Ciências Exatas e Naturais, Universidade do Estado do Rio Grande do Norte (UERN), Mossoró, Rio Grande do Norte, Brazil, **6** Laboratório de Biogeoquímica Ambiental, Universidade Federal de Rondônia (UNIR), Porto Velho, Rondônia, Brazil

Abstract

Top marine predators present high mercury concentrations in their tissues as consequence of biomagnification of the most toxic form of this metal, methylmercury (MeHg). The present study concerns mercury accumulation by Guiana dolphins (*Sotalia guianensis*), highlighting the selenium-mediated methylmercury detoxification process. Liver samples from 19 dolphins incidentally captured within Guanabara Bay (Rio de Janeiro State, Brazil) from 1994 to 2006 were analyzed for total mercury (THg), methylmercury (MeHg), total organic mercury (TOrgHg) and selenium (Se). X-ray microanalyses were also performed. The specimens, including from fetuses to 30-year-old dolphins, comprising 8 females and 11 males, presented high THg (0.53–132 µg/g wet wt.) and Se concentrations (0.17–74.8 µg/g wet wt.). Correlations between THg, MeHg, TOrgHg and Se were verified with age ($p < 0.05$), as well as a high and positive correlation was observed between molar concentrations of Hg and Se ($p < 0.05$). Negative correlations were observed between THg and the percentage of MeHg contribution to THg ($p < 0.05$), which represents a consequence of the selenium-mediated methylmercury detoxification process. Accumulation of Se-Hg amorphous crystals in Kupffer Cells was demonstrated through ultra-structural analysis, which shows that Guiana dolphin is capable of carrying out the demethylation process via mercury selenide formation.

Citation: Lailson-Brito J, Cruz R, Dorneles PR, Andrade L, Azevedo AdF, et al. (2012) Mercury-Selenium Relationships in Liver of Guiana Dolphin: The Possible Role of Kupffer Cells in the Detoxification Process by Tiemannite Formation. PLoS ONE 7(7): e42162. doi:10.1371/journal.pone.0042162

Editor: David L. McCormick, IIT Research Institute, United States of America

Received: January 25, 2012; **Accepted:** July 4, 2012; **Published:** July 31, 2012

Copyright: © 2012 Lailson-Brito et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Funding: This study was supported by Brazilian Research Council - CNPq (Edital Universal - grant #482938/2007-2), by Rio de Janeiro State Government Research Agency - FAPERJ (“Pensa Rio” Program, grant #E-26/110.371/2007), by the Ministry of Education of Brazil - CAPES (Ciências do Mar n° 09/2009) and by Society for Marine Mammalogy. LB-J had a scholarship from CAPES and has a research grant from FAPERJ/UERJ (“Prociência” Program and JCNE #E-26/103.247/2011) and from CNPq (grant #305303/2010-4). AFA has a research grant from CNPq (grant #304826/2008-1) and from FAPERJ (JCNE #101.449/2010). OM has a research grant from CNPq and from FAPERJ (CNE). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* E-mail: lailson@uerj.br

† Deceased.

Introduction

In general, apex predators present high mercury concentrations in their tissues due to the biomagnification of the most toxic form of the metal, methylmercury (MeHg) [1]. Mammals absorb mercury through skin contact and inhalation, as well as via placental or lactational transfer; however, the main exposure occurs through diet [2,3,4].

Studies have shown high mercury accumulation in aquatic mammal tissues [2,3,5,6]. This micropollutant is well known for its neuro- and immunotoxic effects on wildlife, and its toxicity to marine mammals was established through in vitro studies [7,8,9,10]. Additionally, some investigations have tried to link trace metal concentrations in free-ranging marine mammals and health status [11,12,13]. However, there are several studies that

have not found adverse effects related to mercury toxic action in marine mammals [5]. Such observation seems to be associated with the protective action of metallothioneins, as well as with mercury-selenium interactions, through the formation of mercuric selenide [14,15]. Stoichiometric relations between mercury and selenium have been reported for marine mammal liver since 1970s and these relations have been attributed to methylmercury detoxification processes [16,17,18].

The Guiana dolphin (*Sotalia guianensis*) can be useful as a model for the comprehension of mercury accumulation and methylmercury detoxification processes. The species inhabits coastal zones, which implies that Guiana dolphins are in constant interaction with anthropogenic activities. This is specially important in Guanabara Bay, the most degraded area along the species distribution [19,20,21,22,23]. This bay constitutes the marine

environment under the highest anthropogenic pressure along the Brazilian coast [24]. An assessed population of 14 million people lives in the 12 cities of its drainage basin and 35 rivers flow into the bay. These cities house 6 000 industrial plants. In addition, 500 tons of “in natura” sewage and 6.9 tons of oil are released daily into the bay [25,26]. Regarding heavy metal contamination of this body of water, the elements that raise the highest concern are Cr, Pb, Cu and Hg [27,28,29]. Other sources of metals, such as atmospheric deposition, also contribute significantly to the total metal burden that enters Guanabara Bay [27,29]. Recent studies have shown Guiana dolphin presents residence pattern in bays and estuaries along its distribution [30,31], which emphasizes its usefulness as sentinel of environmental pollution in coastal tropical ecosystems from West Atlantic.

The present study aimed to investigate the methylmercury (MeHg) detoxification in hepatic tissue of Guiana dolphins through the relationships among total mercury (THg), total organic mercury (TOrgHg), MeHg and selenium (Se). Additionally, intracellular sites of mercury accumulation were identified in hepatic tissue through ultrastructural histological investigations, revealing a possible route for MeHg detoxification.

Materials and Methods

Ethics Statement

Liver samples of Guiana dolphins were collected with appropriate permissions from Brazilian Environmental Agencies – IBAMA/MMA (permission number 11495-1) and ICMBio/MMA (permission number 11579-1).

Sampling

Hepatic tissue samples were obtained from Guiana dolphins, *Sotalia guianensis*, that were incidentally captured in gillnet fishery off Rio de Janeiro coast, from 1994 to 2006. All samples were stored in individual polyethylene bags and kept frozen (-20°C) until analysis. Nineteen dolphins, including 11 males and 8 females, were analyzed. Age was determined by counting the growth layer groups present in dentine and cementum of the teeth [32] and varied from 0 (fetus and neonate) up to 30 years. Total length (TL) ranged from 71 to 198 cm (Table 1).

Se Determination

Aliquots of approximately 200 mg of liver were digested with 2 mL of 65% HNO_3 in a screw-capped vessel, during 24 h. The solution was then heated to 60°C for 120 min in a water bath. Total selenium was determined by electrothermal atomic absorption spectrometry (ET-AAS), using an Analytik Jena spectrometer ZEE nit 60 equipped with Zeeman-effect background correction. Palladium nitrate was used as a matrix modifier and hence added to each solution to be analyzed. Standard solutions in the working range (20–100 ng/mL) were prepared from Se stock solution (1000 mg/L), using 13% v/v HNO_3 . Quality control was carried out using blanks, as well as, standard reference material DOLT-2 (NRC, Canada). Our analytical results (in $\mu\text{g/g}$ wet wt. \pm S.D.) for the determination of selenium in DOLT-2 (5.94 ± 0.07 $\mu\text{g/g}$) were in good agreement with certified value (6.06 ± 0.49 $\mu\text{g/g}$). The detection limit of the method was 0.00326 $\mu\text{g/g}$.

THg Determination

Aliquots of approximately 200 mg of liver were weighted in a vessel in which 1 mL of H_2O_2 was added. After this first step, the addition of 5 mL of an acid mixture of HNO_3 and H_2SO_4 (1:1) was performed and the solution was then heated to 60°C for

120 minutes in a water bath. After cooling during 15 minutes, the addition of 5 mL of KMnO_4 (5%) was executed. The samples went through a new step of heating in the water bath (60°C) for 15 minutes before overnight digestion stage. The day after, hydroxylamine hydrochloride (1 mL) was added to the samples. Total mercury determination was performed using an Atomic Absorption Spectrophotometer for Hg measurement with Flow Injection System (FIMS-400, Perkin Elmer). Blanks and standard reference material (DOLT-2 – NRC, Canada) were carried through the procedure in the same way as the sample. Our analytical results (in $\mu\text{g/g}$ wet wt. \pm S.D.) for mercury determination in DOLT-2 (2.16 ± 0.08 $\mu\text{g/g}$) were in good agreement with certified value (2.14 ± 0.28 $\mu\text{g/g}$). The detection limit of the method was 0.0017 $\mu\text{g/g}$.

MeHg Determination

Aliquots of approximately 50 mg of liver were put in a stove 68°C during 3–4 hours for samples lixiviation [33,34,35]. The sample was digested in an alkaline solution with 25% KOH/methanol degree HPLC and then made up to a known volume with high purity deionised water (18.2 M Ω cm) from a Milli-Q system. After extraction, 30 μL from the sample were transferred to an amber vessel and 200 μL of buffer solution ($\text{NaC}_2\text{H}_3\text{O}_2$ 2 M) were added to adjust to pH 4.9. The solution was then made up to 40 mL with Milli-Q water. After that, the samples were ethylated by the addition of 50 μL of NaBEt_4 . MeHg was quantified on a GC-AFS (MERXTM Automated Methyl Mercury Analytical System, Brooks Rand) [36]. Our analytical results for MeHg determination in DOLT-2 (0.671 $\mu\text{g/g}$) were in good agreement with certified value (0.693 $\mu\text{g/g}$). The detection limit of the method was 0.0005 $\mu\text{g/g}$.

TOrgHg Determination

TOrgHg determination was conducted according to Uthe et al. [37] and Wagemann et al. [38]. Samples were mixed with an aqueous solution of acid potassium bromide (30% in 4 N H_2SO_4) and cupric sulfate (2.5% in 4 N H_2SO_4) to release all the organic forms of mercury, and then extracted into dichloromethane (DMC) and N-hexane phase. An aliquot (1 mL) of the organic phase was removed and then heated with a mixture of acids (HNO_3 - H_2SO_4 ; 1:4 v/v) in water bath for 30 min at 60°C . The remaining phase was digested and analyzed for total mercury as explained above. Our analytical results for MeHg determination (used as reference to TOrgHg) in DORM-3 (NRCC) were (0.388 $\mu\text{g/g}$) in good agreement with certified value (0.355 $\mu\text{g/g}$).

Optical Microscopy

During dolphin necropsies, different hepatic tissue fragments (5 mm) were transferred to vials containing 10% paraformaldehyde (EMS) in 0.1 M sodium phosphate buffer (pH 7.2) for 2 hours. The tissue was then rinsed in the same buffer twice for 15 minutes, dehydrated in graded ethanol series (50, 70, 90 and 100%) for 15 minutes each bath. After dehydrating, the ethanol was replaced with xylene solutions and the tissue was embedded in paraffin. Semi-thin sections (6 μm) were obtained with histological razor blade (Ted Pella) in a Leica microtome and the sections were stained with hematoxylin & eosine (HE). The sections were observed in a Jenaval Universal (Carl Zeiss-Jena) equipped with a digital camera (Sony, Cybershot DSC-P93). The images (1024 \times 960 pixels) were acquired in bright field and under polarized light with crossed Polaroids.

Table 1. Biological information and analyte concentrations in liver of Guiana dolphins.

Specimen	Sex	TL ¹	Age ²	THg ³	THg ⁴	MeHg ⁵	%MeHg ⁶	TOrgHg ⁷	Se ⁸	Se ⁹	Se:THg ¹⁰
Sg #01	F	71.0	0	2.52	12.5	0.48	19.0	0.50	4.51	57.1	4.55
Sg #02	F	73.0	0	11.8	58.7	0.11	0.93	0.62	74.8	948	16.1
Sg #03	F	75.0	0	3.80	18.9	0.31	8.15	0.56	0.17	2.14	0.11
Sg #04	M	92.5	0	1.33	6.65	0.38	28.6	0.42	0.65	8.25	1.24
Sg #05	M	133.0	1	2.26	11.2	0.91	40.3	1.33	0.36	4.56	0.40
Sg #06	M	147.5	1	0.59	2.93	0.56	94.9	0.66	nd ¹¹	nd ¹¹	nd ¹¹
Sg #07	M	176.0	4	2.51	12.5	0.78	31.1	0.93	nd ¹¹	nd ¹¹	nd ¹¹
Sg #08	M	177.0	8	4.60	22.9	0.20	4.35	0.89	0.64	8.11	0.35
Sg #09	F	185.0	8	5.68	28.3	0.60	10.6	1.99	1.16	14.6	0.52
Sg #10	M	184.5	9	17.3	86.4	nd ¹¹	nd ¹¹	1.00	6.56	83.1	0.96
Sg #11	M	189.0	13	24.7	123	0.78	3.15	1.16	nd ¹¹	nd ¹¹	nd ¹¹
Sg #12	M	190.0	13	47.6	237	1.35	2.83	1.58	15.3	193	0.81
Sg #13	F	198.0	23	132	661	1.74	1.31	1.86	53.9	682	1.03
Sg #14	F	180.0	30	41.7	208	1.84	4.41	nd ¹¹	21.2	268	1.29
Sg #15	F	185.0	30	60.1	299.6	1.92	3.19	nd ¹¹	nd ¹¹	nd ¹¹	nd ¹¹
Sg #16	F	149.0	nd ¹¹	0.53	2.63	nd ¹¹	nd ¹¹	nd ¹¹	0.23	2.87	1.09
Sg #17	M	187.0	nd ¹¹	13.5	67.3	0.95	7.03	1.12	nd ¹¹	nd ¹¹	nd ¹¹
Sg #18	M	187.0	nd ¹¹	3.57	17.8	0.45	12.6	0.63	1.59	20.1	1.13
Sg #19	M	189.0	nd ¹¹	3.69	18.4	1.22	33.1	1.43	2.01	25.4	1.38

¹Total length in cm.²Age in years.³Total mercury concentration in µg/g, wet wt.⁴Total mercury concentration in nmol/g, wet wt.⁵Methylmercury concentration in µg/g, wet wt.⁶Percentage of methylmercury contribution to total mercury.⁷Total organic mercury concentration in µg/g, wet wt.⁸Selenium concentration in µg/g, wet wt.⁹Selenium concentration in nmol/g, wet wt.¹⁰Selenium: total mercury molar ratio.¹¹Not determined.

doi:10.1371/journal.pone.0042162.t001

Conventional and Analytical Transmission Electron Microcopy

For ultrastructural observations and X-ray microanalysis, samples were fixed in 2.5% glutaraldehyde (Sigma, EM grade), in 0.1 M sodium cacodylate buffer (pH 7.4), as well as in filtered seawater for 2 hours. Then, the samples were rinsed in the same buffer twice for 15 minutes. Samples were not post-fixed in osmium tetroxide to avoid undesirable X-ray peaks. The tissue fragments were dehydrated in graded acetone series (50, 70, 90, 100%) for 15 minutes each bath and embedded in Spurr resin (Ted Pella). Ultra-thin sections (90 nm) were obtained in an ultramicrotome (Sorval) and transferred to copper grids (300 mesh). The ultra-thin sections were observed in a Zeiss EM 906 and operated at 80 kV. X-ray microanalysis was performed in a Jeol 1200 EX equipped with a Noran-Voyager analytical system.

Statistical Treatment

Since Shapiro-Wilk's *W* test indicated that the data did not present normal distribution, non-parametric tests were used. The Spearman's (*R*_s) correlation test was used for investigating the occurrence of correlation between the concentrations of each element or compound and the variables total length and age; as well as for testing correlations between Se and THg, and between the percentage of MeHg in THg and THg concentration. All

statistical analyses were performed using the software Statistica (version 7, StatSoft, Inc.).

Results and Discussion

THg, TOrgHg, MeHg and Se concentrations (µg/g wet wt.) in livers of Guiana dolphins are presented in the Table 1.

Mercury Accumulation

Hepatic THg concentrations of Guiana dolphins varied from 0.53 to 132 µg/g wet wt. A similar range for THg (0.35 to 95.0 µg/g wet wt.) in liver samples of Guiana dolphins from southern São Paulo and northern Paraná State was reported [39]. However, Guiana dolphins from Ceará State (northeast Brazil) presented apparently lower hepatic Hg concentrations (0.10 to 29.5 µg/g wet wt.) [40]. Comparing to other small coastal odontocetes, hepatic THg concentrations found in the present study were higher than those observed in harbor porpoises (*Phocoena phocoena*) from North Atlantic (3.00 to 26.0 µg/g dry wt.) [41], but similar to those hepatic concentrations verified in Indo-Pacific humpback dolphins (*Souza chinensis*) from waters of Hong Kong (from <0.15 to 228 µg/g wet wt.) [42].

THg was significantly positively correlated with total length (Spearman correlations; *r*_s = 0.61, *p* < 0.05), which seems to be a result of Hg accumulation with the age of Guiana dolphins

(Spearman correlations; $r_s = 0.79$, $p < 0.05$) (Fig. 1A). An increasing THg concentration with age constitutes a pattern reported for marine mammals from all over the world, including odontocete and mysticete cetaceans [4,39,43]. This probably occur due to the high biological half-life of metals that present high affinity for the cysteine SH group (Cd, Hg, Pb and Ag) [44].

Maternal transfer of mercury is known to occur in marine mammals [4] and fetus/mother ratio of hepatic THg concentration may be used for comprehension on the placental transfer of this metal [45]. Fetus/mother ratios for two Guiana dolphin pairs were 0.19 and 0.03, demonstrating occurrence of restrict transplacental transfer of Hg in the species. These ratios were comparable with Honda et al. [45], who found ratios lower than 1.0. Hg transfer during pregnancy was estimated varying from 0.4 to 1.0% for striped dolphin, *Stenella coeruleoalba* [46]. The low values observed for this ratio both in the present study and in literature suggest the existence of a certain restriction in this transfer. Born et al. [47] reported similar concentrations in newborn walrus and their mothers (*Odobenus rosmarus rosmarus*), which would indicate a higher Hg transfer efficiency through milk than via placenta. However, it was recently suggested that, in northern elephant seals

(*Mirounga angustirostris*), the maternal transfer of mercury was larger during gestation than during lactation [48]. Similar result was previously reported for harp seal pups (*Phoca groenlandica*), which received much of the mercury as methylmercury [49].

One lactating (Sg#14) and two pregnant (Sg#13 and Sg#15) females presented the highest THg concentrations among the analyzed Guiana dolphins. Caurant et al. [14] also observed the highest THg concentrations in lactating females and the authors attributed this finding to a feeding habit alteration due to a higher energetic demand. However, it has been observed that pregnant females or iron-deficient organisms tend to undergo an increase in their non-essential elements gastrointestinal absorption rate, such as Cd [50]. Cd and Hg are from the group IIB metals, which could indicate the existence of a similar behavior for these elements. Hence, there could be an enrichment of inorganic Hg during pregnancy.

Selenium-mercury Relationships and Their Role in Methylmercury Detoxification Processes

Significant positive correlations were found between hepatic selenium concentrations and both age and total length when fetus

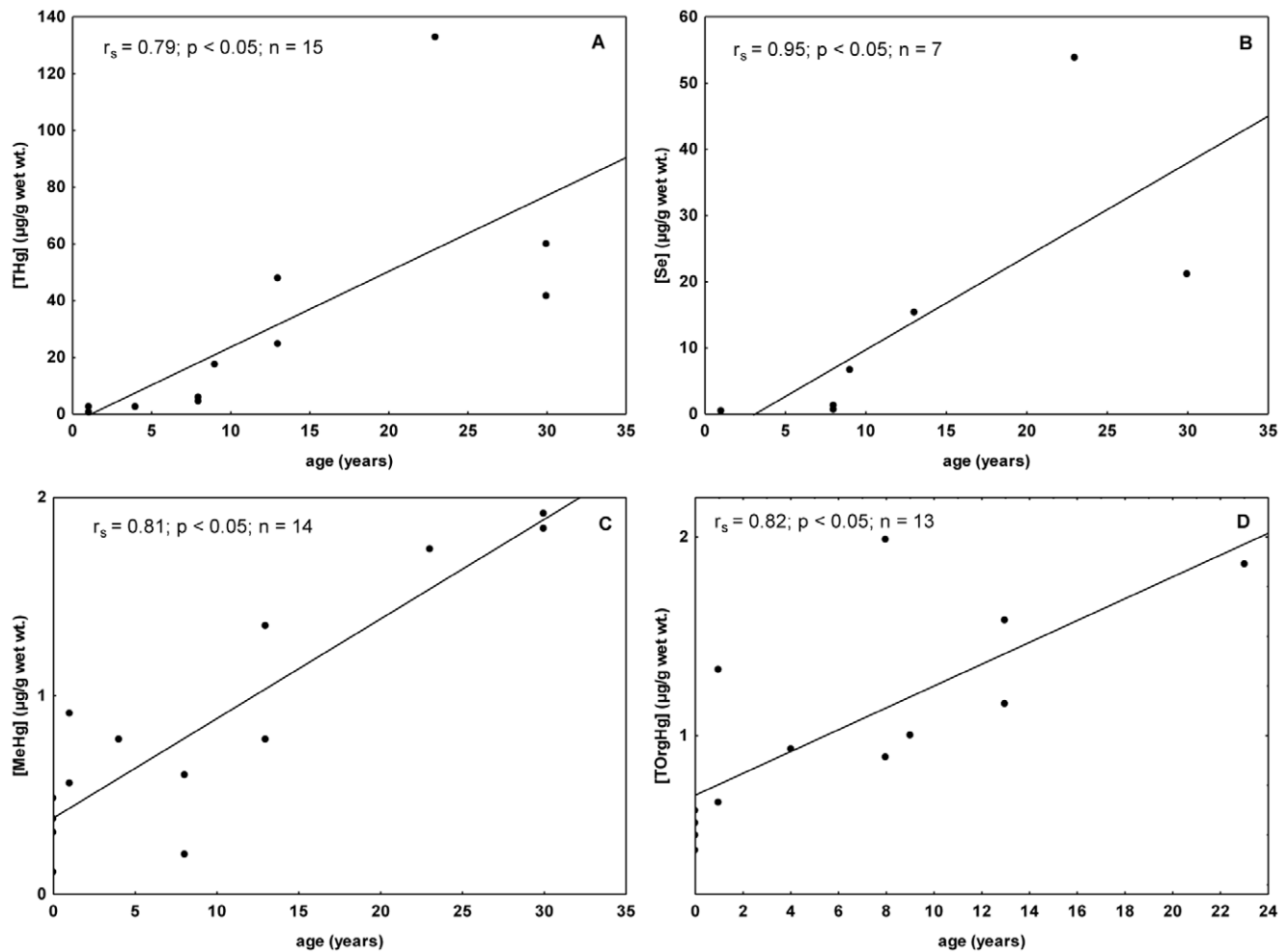


Figure 1. Spearman correlations between hepatic concentrations of THg, Se, MeHg, TORGHg and age of Guiana dolphins. The specimens are from Guanabara Bay, Rio de Janeiro State. Concentrations are in $\mu\text{g/g wet wt.}$ and age in years. (A) Relationships between total mercury concentration (THg) and age. (B) Relationships between selenium concentration (Se) and age - data from fetus and neonates were excluded from the analysis. (C) Relationships between methylmercury concentration (MeHg) and age. (D) Relationships between total organic mercury (TORGHg) and age.

doi:10.1371/journal.pone.0042162.g001

and neonates were excluded from the analysis (Spearman correlations, $p < 0.05$) (Fig. 1B). This pattern has been verified in a number of aquatic mammal species [14,15,51]. Fish and cephalopods may constitute important source of selenium for predator organisms, as it is the case of two teleost fish that comprise the diet of Guiana dolphins, mullet (*Mugil* sp.) and whitemouth croaker (*Micropogonias furnieri*) [52].

Aquatic mammals accumulate high Hg concentrations in their tissues, particularly in liver. Despite this common finding, investigations that describe deleterious effects of mercury on the health of marine mammals are scarce [11,53]. However, it was reported that mercury, selenium and zinc concentrations, as well as Hg:Se ratios, were significantly higher in harbor porpoises that died from infectious diseases than in those individuals that died as consequence of physical trauma [12]. In general, it has been argued that the molar ratio between the elements is more important than the concentration itself. It is clear that understanding the mercury-selenium relationship is of fundamental importance for the comprehension of the methylmercury detoxification processes, as well as for the understanding of the processes related to mercury immobilization under a non-toxic form [18,54,55]. Several studies have reported Hg-Se molar ratios close to 1.0 in marine mammal liver [17,56,57,58], including an investigation that dealt with samples from two Guiana dolphins from Suriname [56]. Koeman et al. [17,56] suggested a causal relationship between both elements, which would be involved in a detoxification process. Additionally, Lemes et al. [55] have detected methylmercuric glutathionate (CH_3HgSG) and Hg-Se complexes in liver and brain tissues of beluga whales from the western Canadian Arctic. The authors suggest that these compounds are associated to the protection against the toxic effect of the high concentrations of MeHg found in beluga whales [55].

In the present study, hepatic THg accumulation was followed by hepatic Se accumulation. The positive correlation coefficient of $r_s = 0.72$ (Spearman correlation, $p < 0.05$; Fig. 2A) turned it possible to verify the latter finding. An increase in the correlation coefficient was observed when the data from fetuses and neonates were removed from the test ($r_s = 0.89$; Fig. 2B). During both of these early life stages, mammals may not present well-developed MeHg detoxification mechanisms, which alters the selenium-mercury relationship in comparison to adults. Another possible explanation is that Hg and Se may have different transfer efficiency. As discussed above, Habran et al. [48] verified that maternal transfer of Se was extended during lactation, whereas Hg transfer was more important during gestation in the northern elephant seal. The mean Hg-Se molar ratio was 2.2 and fetuses were the individuals that presented both the highest and the lowest Hg-Se molar ratios (0.11 e 16.1) among all dolphins analyzed (Table 1). A high transfer of both Hg and Se from mother (Sg#15) to fetus (Sg#2) constitutes a possible explanation for this wide variation. The specimen Sg#2 showed the highest Hg concentration among fetuses and calves and the highest Se concentration among all individuals.

Palmisano et al. [15] suggest Hg-Se molar ratio to get near 1.0 only in hepatic concentrations close to 100 $\mu\text{g/g}$ in striped dolphins (*Stenella coeruleoalba*). However, it is important to highlight the mercury fraction used by the authors, since they neither used the protein-bound fraction nor they used MeHg (THg – (Hg^{++} bound to proteins + MeHg)). Considering that just few individuals presented hepatic Hg concentrations higher than 10.0 $\mu\text{g/g}$, our results may not be representative. However, our results indicated a trend for Hg-Se molar ratios close to 1.0 (Table 1). Palmisano et al. [15] did not find Hg concentrations between 10.0 and

100 $\mu\text{g/g}$. Therefore, it is possible that the Hg-Se molar ratio get near 1.0 even below the concentration suggested by the latter authors. This hypothesis agrees with the approximate limit of 50.0 $\mu\text{g/g}$ suggested for another delphinid species, the long-finned pilot whale (*Globicephala melas*) [14].

The occurrence of demethylation process seems to be responsible for the hepatic MeHg concentration. Low percentages of MeHg and TOrgHg contribution to THg were found in the present study (Table 1). This behavior had been described for Guiana dolphins from southern São Paulo and northern Paraná states regarding organic mercury [39], as well as, for other cetacean species [14,15,46,49]. Additionally, the percentage of MeHg to THg tend to decrease with the increase of THg in livers of Guiana dolphins (Spearman correlations; $r_s = -0.87$, $p < 0.05$) (Fig. 3). Apparently, as MeHg suffers demethylation, Hg is immobilized under an inorganic form. Therefore, the percentage of MeHg contribution to THg tends to reduce due to the high rates of inorganic Hg accumulation.

Young cetaceans and pinnipeds do not seem to be able to demethylate MeHg efficiently [14,46,49,59]. In general, marine mammal calves and fetuses present higher percentages of MeHg contribution to THg than their mothers. In the present study, the mother and the fetus (Sg#13 and Sg#3) presented the percentages of 1.3 and 8.2, respectively, which corroborates the general observation. Fetuses and one-year-old calves presented percentages of MeHg contribution to THg varying from 0.93 to 94.9%. Moreover, higher mean values were observed in fetuses and calves than in adults (32×13%), which constitute a pattern verified in other marine mammal species [14,46,49]. *In vitro* studies performed with hepatocytes from gray seal pups (*Halichoerus grypus*) did not evidence biochemical MeHg demethylation, whereas gray seal adults seem to demethylate MeHg [60]. It is interesting to note that the fetus (Sg#2) that showed the lowest percentage of MeHg contribution to THg was the specimen that presented the highest Hg-Se molar ratio (16.1). The mother (Sg#15) showed the highest THg concentration (299 $\mu\text{g/g}$ wet wt.) and the percentage of MeHg contribution to THg was relatively low (3.19%). As previously discussed, the percentage of MeHg contribution to THg tends to reduce due to the high rates of inorganic Hg accumulation as result of the MeHg demethylation. Consequently, a high percentage of the THg transferred from the mother to the fetus was probably inorganic mercury.

In general, positive correlations were found between TOrgHg and age [2,39] as observed to Guiana dolphins for both MeHg and TOrgHg concentrations (Spearman correlations; $p < 0.05$; Figure 1C and 1D). Dietz et al. [59] suggested that MeHg concentrations seldom exceed 2.00 $\mu\text{g/g}$ in health animals, which was corroborated by results obtained in other studies [3,61]. Since MeHg presents a high positive correlation with age, Wagemann et al. [3] suggested that the demethylation rate reduces during the aging process. Our results lead to the hypothesis that the demethylation continues in a high rate in Guiana dolphins, as well as to the supposition that the process does not allow occurrence of higher concentrations than the mentioned threshold (2.0 $\mu\text{g/g}$). A plausible explanation for that relies on the fact that the liver of marine mammals is normally capable of keeping MeHg in low levels [59].

Regarding the percentage of MeHg contribution to TOrgHg, the values varied from 17.8 to 96.0% (mean value = 70%). This finding showed the TOrgHg concentration included organomercurial species other than MeHg [38,62]. An arithmetic mean of 23% of MeHg contribution to TOrgHg was found in liver of ringed seals (*Phoca hispida*) [38], which was about three times lower than the MeHg percentage found in the present study.

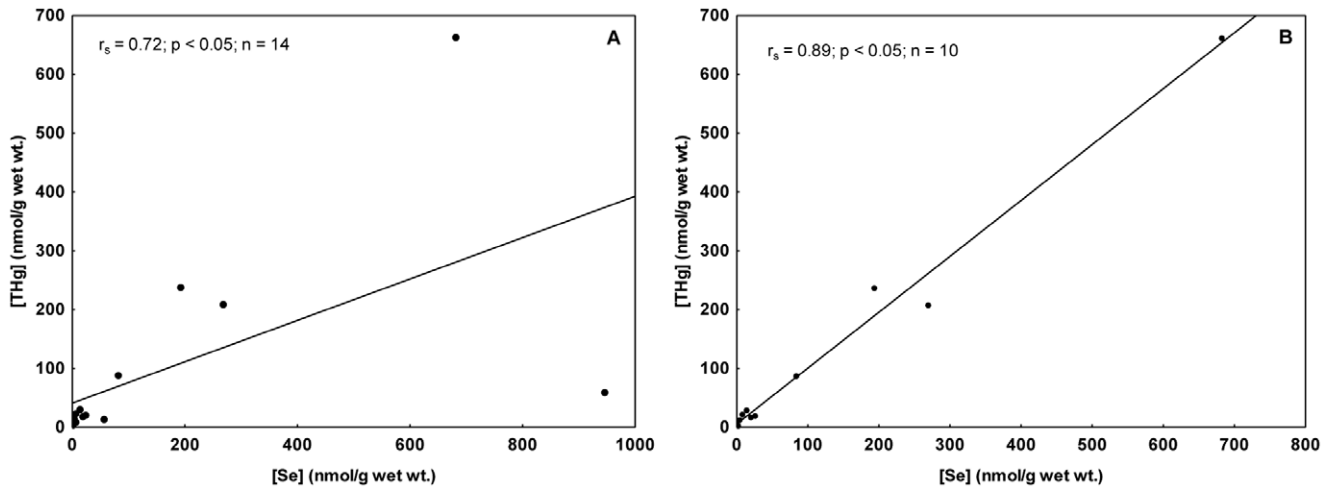


Figure 2. Spearman correlations between THg and Se molar concentrations in liver of Guiana dolphins. The specimens are from Guanabara Bay, Rio de Janeiro State. THg and Se concentrations are in nmol/g wet wt. (A) Relationship between total mercury and selenium considering all individuals. (B) Relationship between total mercury and selenium excluding data from fetus and neonates. doi:10.1371/journal.pone.0042162.g002

Kupffer Cells in the Detoxification Process by Tiemannite Formation

The presence of Hg under an insoluble form (mercury selenide granules - tiemannite) in marine mammal liver was suggested first by Koeman et al. [17]. Later, the accumulation of mercury selenide granules as amorphous crystals was demonstrated in hepatic tissue of Cuvier's beaked whales [63,64]. André et al. [65] suggested these particles do not suffer proteolytic enzyme attack and stay inert. The tiemannite accumulation was showed in liver,

muscle and brain [57], as well as these crystals were identified in lung [66,67].

In the present study, histological and ultrastructural investigations have been performed. The hepatic tissue originated from a 9-year-old dolphin (17.3 $\mu\text{gTHg/g}$) presented the basic appearance of a normal liver, with hepatic parenchyma formed by hepatic lobules in a prism form, surrounded by connective tissue. The lobules were formed by plates of hepatocytes separated by hepatic sinusoids that radiate out from a central vein. Hepatocytes are polyhedral cells that fill the most volume of hepatic parenchyma.

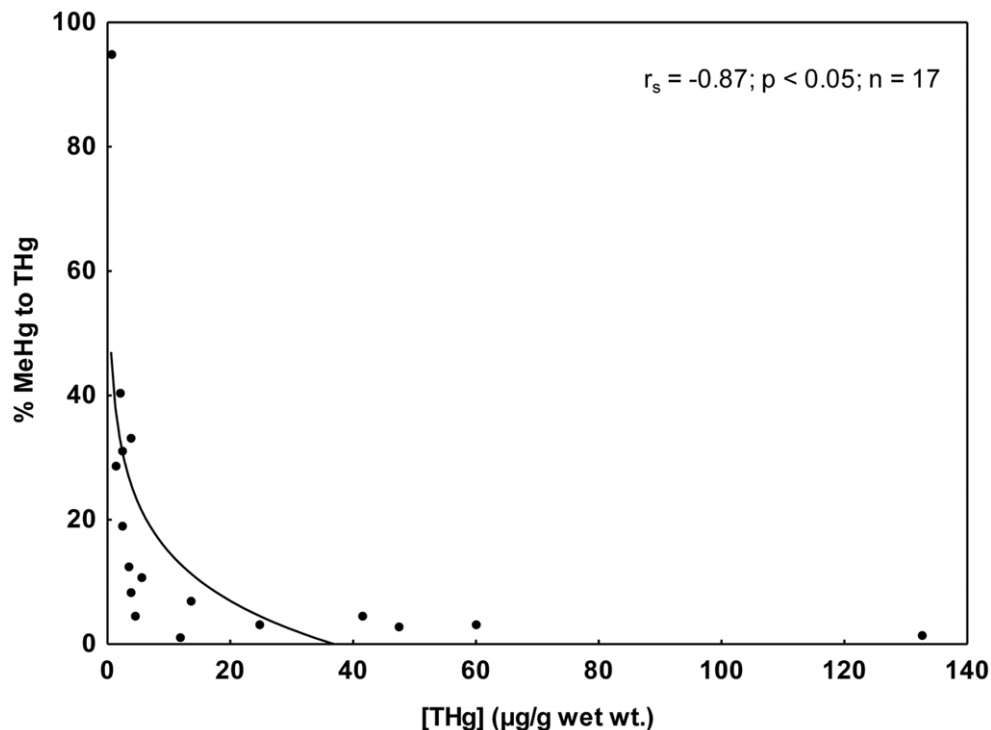


Figure 3. Spearman correlation between percentages of methylmercury (MeHg) to total mercury (THg) concentrations in liver samples from Guiana dolphin. The specimens are from Guanabara Bay, Rio de Janeiro State. THg concentration is in $\mu\text{g/g}$ wet wt. doi:10.1371/journal.pone.0042162.g003

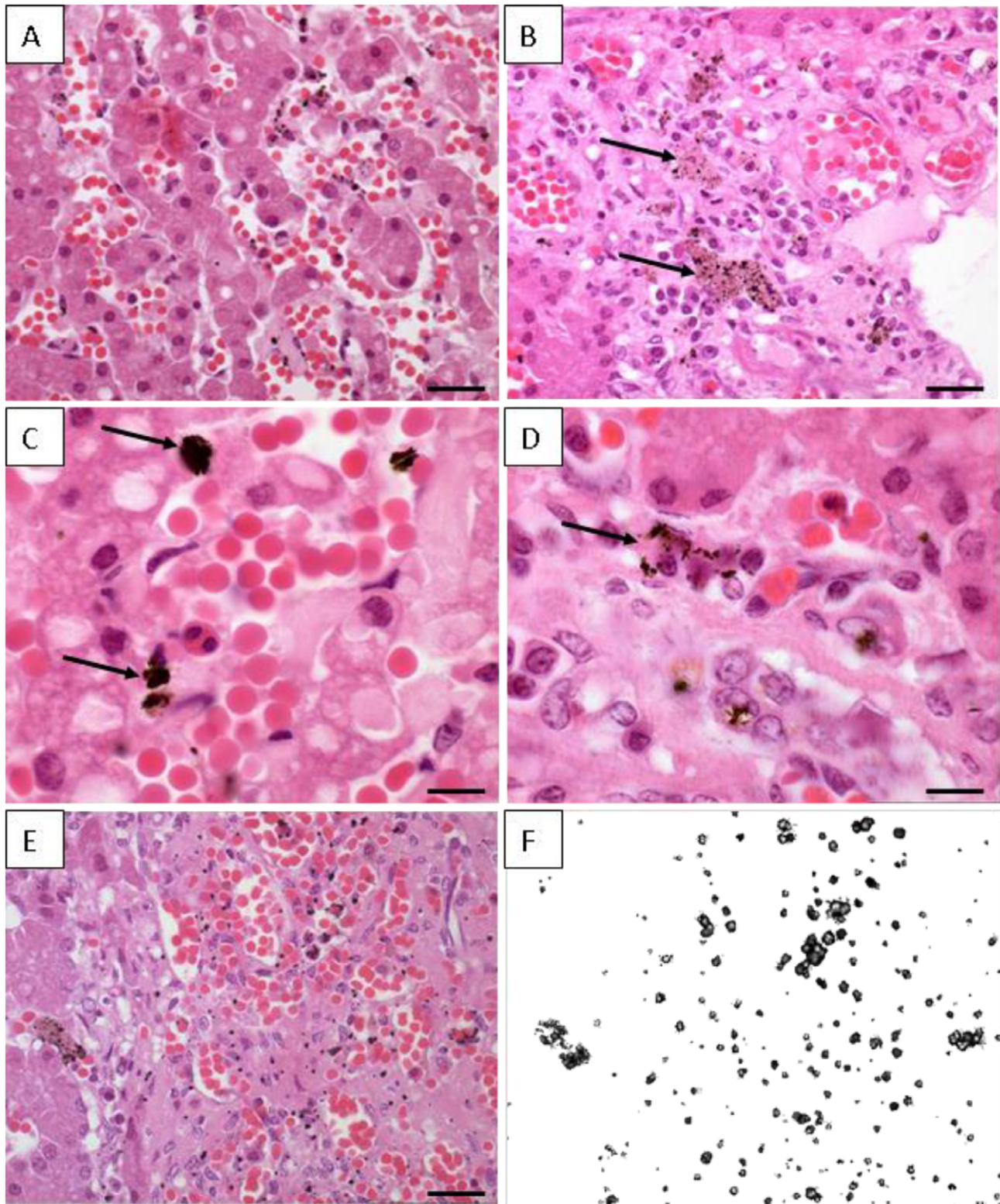


Figure 4. Histological sections from hepatic tissue of Guiana dolphin (*Sotalia guianensis*). The sample is from a 9-year-old male dolphin from Guanabara Bay, Rio de Janeiro State. (A) Semi-thin section of the hepatic tissue showing the hepatocytes arranged in layers or plates; Longitudinal blood vessel filled with numerous red cells can be seen; Barr = 80 μm . (B) Different field of the organ showing the hepatic parenchyma with many blood vessel and some dark deposits within the Kupffer cells (arrows); Barr = 100 μm . (C) Agglomerated dark deposits found in Kupffer cells (arrows); Barr = 20 μm . (D) Deposits distributed in small dots (arrows); Barr = 20 μm . (E) Semi-thin section of dolphin liver showing the cells with dark inclusions; Barr = 100 μm . (F) - Same area depicted in figure E observed under polarized light; all the dark dots are birefringent materials and corresponds to deposits within the Kupffer cells.
doi:10.1371/journal.pone.0042162.g004

Sinusoids walls are lined by flattened endothelial cells interspersed with plumper Kupffer cells that often presented dark deposits (Fig. 4A and 4B). These deposits had different sizes and they can be seen as agglomerated or small dots (Fig. 4C and 4D). These materials within Kupffer cells, when observed under polarized light, were birefringents, indicating a crystalline nature (Fig. 4E and 4F).

The hepatic sinusoids are located adjacent to the hepatocytes, allowing the exchange of soluble substances between blood and hepatocytes. The Kupffer cells are derived from bone marrow monocytes and contain numerous endocytic vesicles, lysosomes and usually fagocytosed particulate matter. The dark vesicles found in our sample resembled lysosomes typical from those cells. In addition, these cells act as macrophages, removing virus, bacteria, tumor cells and parasites from the blood stream. A possible explanation is that Hg and Se diffused through the sinusoids, were absorbed and formed a crystalline mineral that remained insoluble within the Kupffer cells. Therefore, it is plausible to assume that besides being produced from the hepatic demethylation process, the Se-Hg crystal produced in other tissues

is also accumulated in liver by the action of Kupffer cells, since it finds its way to the blood stream.

Transmission electron microscopy of Kupffer cells showed numerous circular vesicles, some electron-dense, other electron-lucent (Fig. 5A). In addition, stick-like shaped electron-dense deposits (approximately 2 μm) were also seen (Fig. 5B and 5C). These deposits were not found within the hepatocytes. They were analyzed by X-ray microanalysis and Hg ($L\alpha$ and $L\beta$ layers) and Se ($K\alpha$ layer) were detected co-localizing in the same studied region (Fig. 5D).

It is important to highlight that the low signal might have a straight relationship with the option for the tissue site to be analyzed. Priority was given to crystal analysis specifically rather than to crystal agglomerate, which certainly would produce an increase in the sign. However, the option for agglomerate would generate information not only on the crystal specifically but also on other structures that would mislead the interpretation of the results. Despite being close to the background noise, Hg and Se were the elements that could be highlighted, which did not happen by chance. Other abundant elements, such as copper, did not

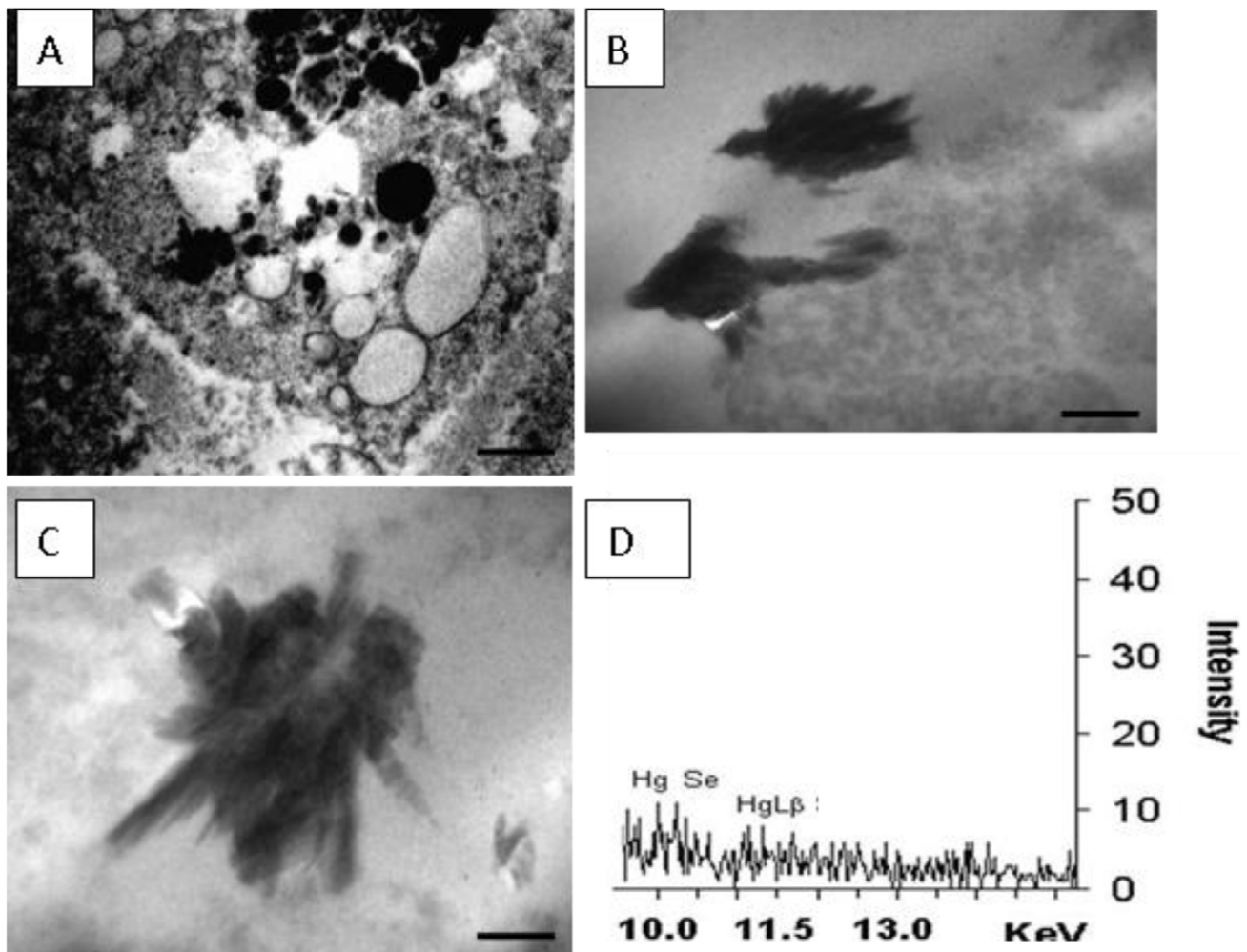


Figure 5. Ultrastructure of Guiana dolphin (*Sotalia guianensis*) liver. The sample is from a 9-year-old male dolphin from Guanabara Bay, Rio de Janeiro State. (A) Electron microcopy image showing a Kupffer cell and its organelles. Numerous circular vesicles, some electron-dense, other electron-lucent; Barr = 5 μm . (B) Image of stick-like shaped electron-dense deposits found in Kupffer cells; Barr = 1 μm . (C) A detail of other electron-dense deposit; Barr = 500 nm. (D) X-ray microanalysis spectra showing the elements Hg and Se as components of those dark deposits found within Kupffer cells.

doi:10.1371/journal.pone.0042162.g005

appear which clearly shows that the signal corresponds to the crystal rather than to the entire site. This finding suggests that Se-Hg granules are formed even in low concentrations.

Martoja and Berry [63] raised the hypothesis that tiemannite is a product of the biosynthesis process and that it represents the final stage of the MeHg degradation, suggesting that Se played a fundamental role in this process. The authors suggested also that tiemannite crystallization occurs in liver, since it is an organ in which the B12 vitamin is present in levels that allow the methyl group transfer from Hg to Se. This transfer would occur through dimethylselenide (CH_3SeCH_3) formation, which would be excreted by lungs. Nigro and Leonzio [57] corroborate such proposition suggesting that MeHg demethylation might occur in absence of biochemical process due to the high affinity between the two elements. The immobilization of Hg under the tiemannite form in cetacean liver, as well as in other top aquatic predators, represents the final stage of methylmercury biomagnification process. These insoluble deposits probably remain in liver for long periods in an integrative process, as proposed by Martoja and Berry [63].

In conclusion, through ultra-structural analyses, our study suggests the accumulation of Se-Hg amorphous crystals in Kupffer Cells. Our investigation shows a picture of a crystal that is

probably tiemannite, exposing it in a separated view rather than showing the entire accumulation site. Finally, our findings show that Guiana dolphin is capable of carrying out the demethylation process via mercury selenide formation.

Acknowledgments

Liver samples were collected under permits 11495-1 and 11579-1, issued by the Brazilian Ministry of the Environment (IBAMA/MMA and ICMBio/MMA, respectively). We thank to Radioisotopes Laboratory (UFRJ) and to Aquatic Mammal and Bioindicator Laboratory (MAQUA/UERJ) teams for invaluable assistance in sampling, as well as in sample preparation and analysis. We also thank three anonymous reviewers who made useful suggestions that have helped to improve the manuscript.

Author Contributions

Conceived and designed the experiments: JL-B RC OM. Performed the experiments: JL-B PRD RC LA AFA ABF LGV MBC RA DPC WRB. Analyzed the data: JL-B AFA TLB OM. Contributed reagents/materials/analysis tools: RC WRB OM. Wrote the paper: JL-B RC PRD LA AFA TLB. Photographed histological and ultrastructural sections: RC LA.

References

1. Renzoni A, Zino F, Franchi E (1998) Mercury Levels along the Food Chain and Risk for Exposed Populations. *Environmental Research* 77: 68–72.
2. Leonzio C, Focardi S, Fossi C (1992) Heavy metals and selenium in stranded dolphins of the northern Tyrrhenian (NW mediterranean). *Science of the Total Environment* 119: 77–84.
3. Wagemann R, Trebacz E, Boila G, Lockhart WL (1998) Methylmercury and total mercury in tissues of arctic marine mammals. *The Science of the Total Environment* 218: 19–31.
4. Das K, Debacker V, Pillet S, Bouquegneau JM (2003) Heavy metals in marine mammals. In: Vos JG, Bossart GD, Fournier M, O'Shea TJ, editors. *Toxicology of marine mammals*. London: Taylor and Francis Group. 135–167.
5. O'Shea TJ (1999) Environmental contaminants and marine mammals. In: Reynolds III JE, Rommel SA, editors. *Biology of marine mammals*. Washington: Smithsonian Institution Press. 485–563.
6. Cardellicchio N, Decataldo A, Di Leo A, Misino A (2002) Accumulation and tissue distribution of mercury and selenium in striped dolphins (*Stenella coeruleoalba*) from the Mediterranean Sea (southern Italy). *Environmental Pollution* 116: 265–271.
7. Betti C, Nigro M (1996) The comet assay for the evaluation of the genetic hazard of pollutants in cetaceans: preliminary results on the genotoxic effects of methylmercury on the bottle-nosed dolphin (*Tursiops truncatus*) lymphocytes in vitro. *Marine Pollution Bulletin* 32: 545–548.
8. De Guise S, Bernier J, Martineau D, Bédard P, Fournier M (1996) Effects of in vitro exposure of beluga whale splenocytes and thymocytes to heavy metals. *Environmental Toxicology and Chemistry* 15: 1357–1364.
9. Dufresne M, Frouin H, Pillet S, Hammill M, Lesage V, et al. (2010) Comparative sensitivity of harbour and grey seals to several environmental contaminants using in vitro exposure. 60: 344–349.
10. Frouin H, Loseto LL, Stern GA, Haulena M, Ross PS (2012) Mercury toxicity in beluga whale lymphocytes: Limited effects of selenium protection. *Aquatic Toxicology* 109: 185–193.
11. Rawson AJ, Patton GW, Hofmann S, Pietra GG, Johns L (1993) Liver abnormalities associated with chronic mercury accumulation in stranded Atlantic bottlenose dolphins. *Ecotoxicology and Environmental Safety* 25: 41–47.
12. Bennett PM, Jepson PD, Law RJ, Jones BR, Kuiken T, et al. (2001) Exposure to heavy metals and infectious disease mortality in harbour porpoises from England and Wales. *Environmental Pollution* 112: 33–40.
13. Siebert U, Joiris C, Holsbeek L, Benke H, Failing K, et al. (1999) Potential relation between mercury concentrations and necropsy findings in cetaceans from German waters of the North and Baltic Seas. *Marine Pollution Bulletin* 38: 285–295.
14. Caurant F, Navarro M, Amiard JC (1996) Mercury in pilot whales: Possible limits to the detoxification process. *Science of the Total Environment* 186: 95–104.
15. Palmisano F, Cardellicchio N, Zamboni PG (1995) Speciation of mercury in dolphin liver - a 2-stage mechanism for the demethylation accumulation process and role of selenium. *Marine Environmental Research* 40: 109–121.
16. Koeman JH, Van de Ven WSM, de Goeij JJM, Tijoe PS, Van Haften JL (1975) Mercury and selenium in marine mammals and birds. *The Science of the Total Environment* 3: 279–287.
17. Koeman JH, Peeters WHM, Koudstaal CH, Tijoe PS, Goeij J (1973) Mercury-selenium correlations in marine mammals. *Nature* 245: 385–386.
18. Cuvillier-Vincent M, Furness RW (1991) Mercury and selenium interaction: A review. *Ecotoxicology and Environmental Safety* 21: 348–364.
19. Azevedo AF, Lailson-Brito J, Dorneles PR, Van Sluys M, Cunha HA, et al. (2009) Human-induced injuries to marine tucuxis (*Sotalia guianensis*) (Cetacea: Delphinidae) in Brazil. *Marine Biodiversity Records* 2: e22.
20. Dorneles PR, Lailson-Brito J, Azevedo AF, Meyer JS, Vidal LG, et al. (2008) High Accumulation of Perfluorooctane Sulfonate (PFOS) in Marine Tucuxi Dolphins (*Sotalia guianensis*) from the Brazilian Coast. *Environmental Science & Technology* 42: 5368–5373.
21. Dorneles PR, Lailson-Brito J, Durtu AC, Wejls L, Azevedo AF, et al. (2010) Anthropogenic and naturally-produced organobrominated compounds in marine mammals from Brazil. *Environmental International* 36: 60–67.
22. Dorneles PR, Lailson-Brito J, Fernandez MAS, Vidal LG, Barbosa LA, et al. (2008) Evaluation of cetacean exposure to organotin compounds in Brazilian waters through hepatic total tin concentrations. *Environmental Pollution* 156: 1268–1276.
23. Lailson-Brito J, Dorneles PR, Azevedo-Silva CE, Azevedo AF, Vidal LG, et al. (2010) High organochlorine accumulation in blubber of Guiana dolphin, *Sotalia guianensis*, from Brazilian coast and its use to establish geographical differences among populations. *Environmental Pollution* 158: 1800–1808.
24. Rebello AL, Ponciano CR, Lges LH (1988) Avaliação da produtividade primária e da disponibilidade de nutrientes na Baía de Guanabara. *Academia Brasileira de Ciências* 60: 419–430.
25. Lacerda LD, Graça NM, Quintanilha MCP (1994) Bibliografia sobre a contaminação por metais pesados em ambientes costeiros do estado do Rio de Janeiro. Niterói: Programa Geoquímica/UFF. 57 p.
26. JICA (1992) The Study on Recuperation of the Guanabara Bay Ecosystem. Japan International Cooperation Agency.
27. Marins RV, Paula-Filho FJ, Maia SRR (2005) Distribuição de mercúrio total como indicador de poluição urbana e industrial na costa brasileira. *Química Nova* 27: 763–770.
28. Pfeiffer WC, Fiszman M, Malm O, Azcue JMP, Lacerda LD, et al. (1987) Heavy metals pollution studies through critical pathways analysis: The Rio de Janeiro State case, Brazil; New Orleans. Edinburgh: CEP Consultants.
29. Lacerda LD, Pfeiffer WC, Fiszman M (1983) Mineral distribution and ecological role of a recently formed halophyte in the Guanabara Bay, Rio de Janeiro. *Tropical Ecology* 24: 162–169.
30. Azevedo AF, Lailson-Brito J, Cunha HA, Van Sluys M (2004) A note on site fidelity of marine tucuxis (*Sotalia fluviatilis*) in Guanabara Bay, southeastern Brazil. *J Cetacean Res Manage* 6: 265–268.
31. Flores PAC, Da Silva VMF (2009) Tucuxi and Guiana Dolphin: *Sotalia fluviatilis* and *S. guianensis*. In: William FP, Bernd W, Thewissen JGM, editors. *Encyclopedia of Marine Mammals (Second Edition)*. London: Academic Press. 1188–1192.
32. Dietz R, Heide-Jørgensen MP, Teilmann J, Valentin N, Härkönen T (1991) Age determination in European harbor seals *Phoca vitulina*. *Sarsia* 76: 17–21.

33. Liang L, Bloom NS, Horvat M (1994) Simultaneous Determination of Mercury Speciation in Biological Materials by GC/CVAFS After Ethylation and Room-Temperature Pre-collection. *Clinical Chemistry* 40: 602–607.
34. EPA (2001) Method 1630: Methyl Mercury in Water by Distillation, Aqueous Ethylation, Purge and Trap, and CVAFS. Washington, D.C.: U.S. Environmental Protection Agency. 49 p.
35. Bloom NS, Fitzgerald WF (1988) Determination of volatile mercury species at the picogram level by low temperature gas chromatography with cold vapor atomic fluorescence detection. *Analytica Chimica Acta* 208: 151–161.
36. Taylor VF, Carter A, Davies C, Jackson BP (2011) Trace-Level Automated Mercury Speciation Analysis. *Analytical Methods* 3: 1143–1148.
37. Uthe JF, Solomon J, Grift B (1972) Rapid semimicro method for the determination of methyl mercury in fish tissue. *Journal Association of Official Analytical Chemists* 55: 583–589.
38. Wagemann R, Trebacz E, Boila G, Lockhart WL (2000) Mercury species in the liver of ringed seals. *The Science of The Total Environment* 261: 21–32.
39. Kunito T, Nakamura S, Ikemoto T, Anan Y, Kubota R, et al. (2004) Concentration and subcellular distribution of trace elements in liver of small cetaceans incidentally caught along the Brazilian coast. *Marine Pollution Bulletin* 49: 574–587.
40. Monteiro-Neto C, Itavo RV, Moraes LE (2003) Concentrations of heavy metals in *Sotalia fluviatilis* (Cetacea: Delphinidae) off the coast of Ceará, northeast Brazil. *Environmental Pollution* 123: 319–324.
41. Das K, Beans C, Holsbeck L, Mauger G, Berrow SD, et al. (2003) Marine mammals from northeast atlantic: relationship between their trophic status as determined by $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ measurements and their trace metal concentrations. *Marine Environmental Research* 56: 349–365.
42. Parsons ECM (1998) Trace metal pollution in Hong Kong: implications for the health of Hong Kong's Indo-Pacific hump-backed dolphins (*Sousa chinensis*). *Science of The Total Environment* 214: 175–184.
43. Wagemann R, Innes S, Richard PR (1996) Overview and regional and temporal differences of heavy metals in Arctic whales and ringed seals in the Canadian Arctic. *Science of The Total Environment* 186: 41–66.
44. Mason AZ, Jenkins KD (1995) Metal detoxication in aquatic organisms. In: Tessier A, Turner DR, editors. *Metal speciation and Bioavailability in Aquatic Systems*. Chichester: Wiley & Sons. 479–608.
45. Honda K, Tatsukawa R, Fujiyama T (1982) Distribution Characteristics of Heavy Metals in the Organs and Tissues of Striped Dolphin, *Stenella coeruleoalba*. *Agricultural and Biological Chemistry* 46: 3011–3021.
46. Itano K, Kawai S, Miyazaki N, Tatsukawa R, Fujiyama T (1984) Mercury and selenium levels in striped dolphins caught off the Pacific coast of Japan. *Agricultural and Biological Chemistry* 48: 1109–1111.
47. Born EW, Kraul I, Kristensen T (1981) Mercury, DDT and PCB in the Atlantic walrus (*Odobenus rosmarus rosmarus*) from the Thule District, North Greenland. *Arctic* 34: 255–260.
48. Habran S, Debier C, Crocker DE, Houser DS, Das K (2011) Blood dynamics of mercury and selenium in northern elephant seals during the lactation period. *Environmental Pollution* 159: 2523–2529.
49. Wagemann R, Stewart REA, Lockhart WL, Stewart BE, Povoledo D (1988) Trace metals and methyl mercury: associations and transfer in harp seal (*Phoca groenlandica*) mothers and their pups. *Marine Mammal Science* 4: 339–355.
50. Flanagan PR, McLellan JS, Haist J, Cherian MG, Chamberlain MJ, et al. (1978) Increased dietary cadmium absorption in mice and human subjects with iron deficiency. *Gastroenterology* 74.
51. Wagemann R, Muir DCG (1984) Concentrations of heavy metals and organochlorines in marine mammals of northern waters: overview and evaluation. Canadian Technical Report of Fisheries and Aquatic Sciences 1279: 1–97.
52. Seixas TG, Kehrig HA, Moreira I, Malm O (2006) Selênio em tecidos de organismos marinhos da Baía de Guanabara, Brasil. *Journal of the Brazilian Society of Ecotoxicology* 1: 21–25.
53. Ronald K, Tessaro JF, Uthe HC, Freeman HC, Frank R (1977) Methylmercury poisoning in the harp seal (*Pagophilus groenlandicus*). *Science of The Total Environment* 8: 1–11.
54. Khan MAK, Wang F (2010) Chemical Demethylation of Methylmercury by Selenoamino Acids. *Chemical Research in Toxicology* 23: 1202–1206.
55. Lemes M, Wang F, Stern GA, Ostertag SK, Chan HM (2011) Methylmercury and selenium speciation in different tissues of beluga whales (*Delphinapterus leucas*) from the Western Canadian Arctic. *Environmental Toxicology and Chemistry* 30: 2732–2738.
56. Koeman JH, Peters WHM, Smit CJ, Tjioe PS, Goeij JJM (1972) Persistent chemicals in marine mammals. *TNO Nieuws* 27: 570–578.
57. Nigro M, Leonzio C (1996) Intracellular storage of mercury and selenium in different marine vertebrates. *Marine Ecology Progress Series* 135: 137–143.
58. Lockhart WL, Stern GA, Wagemann R, Hunt RV, Metner DA, et al. (2005) Concentrations of mercury in tissues of beluga whales (*Delphinapterus leucas*) from several communities in the Canadian Arctic from 1981 to 2002. *Science of the Total Environment* 351–352: 391–412.
59. Dietz R, Nielsen CO, Hansen MM, Hansen CT (1990) Organic mercury in Greenland birds and mammals. *Science of the Total Environment* 95: 41–51.
60. Van de Ven WSM, Koeman JH, Svenson A (1979) Mercury and selenium in wild and experimental seals. *Chemosphere* 8: 539–555.
61. Julshamn K, Andersen A, Ringdal O, Mørkøre J (1987) Trace elements intake in the Faroe Islands I. Element levels in edible parts of pilot whales (*Globicephalus meleanus*). *Science of the Total Environment* 65: 53–62.
62. Wagemann R, Trebacz E, Hunt R, Boila G (1997) Percent methylmercury and organic mercury in tissues of marine mammals and fish by different experimental and calculation methods. *Environmental Toxicology and Chemistry* 16: 1859–1866.
63. Martoja R, Berry JP (1980) Identification of tiemanmtte as a probable product of demethylation of mercury by selenium in cetaceans. *Vie Milieu* 30: 7–10.
64. Martoja R, Viale D (1977) Accumulation de granules de seleniure mercurique dans le foie d'odontocetes (Mammiferes, Cetaces): un mecanisme possible de detoxication du methylmercure par le selenium. *Comptes Rendus de l'Academie des Sciences* 285: 109–112.
65. Andre JM, Ribeyre F, Boudou A (1990) Mercury contamination levels and distribution in tissues and organs of delphinids (*Stenella attenuata*) from the eastern tropical Pacific, in relation to biological and ecological factors. *Marine Environmental Research* 30: 43–72.
66. Rawson AJ, Bradley JP, Teetsov A, Rice SB, Haller EM, et al. (1995) A role for airborne particulate in high mercury levels of some cetaceans. *Ecotoxicology and Environmental Safety* 30: 309–314.
67. Augier H, Benkoel L, Chamlian A, Park WK, Ronneau C (1993) Mercury, zinc and selenium bioaccumulation in tissues and organs of Mediterranean striped dolphins *Stenella coeruleoalba meyen*. *Toxicological result of their interaction. Cellular and Molecular Biology* 39: 621–634.