

Quantification of Bacterial Indicators and Zoonotic Pathogens in Dairy Wastewater Ponds

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Zoonotic pathogens in land-applied dairy wastewaters are a potential health risk. The occurrence and abundance of 10 pathogens and 3 fecal indicators were determined by quantitative real-time PCR (qPCR) in samples from 30 dairy wastewaters from southern Idaho. Samples tested positive for *Campylobacter jejuni*, *stx*₁- and *aeaeA*-positive *Escherichia coli*, *Listeria monocytogenes*, *Mycobacterium avium* subsp. *paratuberculosis*, and *Salmonella enterica*, with mean recoveries of genomic DNA corresponding to 10² to 10⁴ cells ml⁻¹ wastewater. The most predominant organisms were *C. jejuni* and *M. avium*, being detected in samples from up to 21 and 29 of 30 wastewater ponds, respectively. The qPCR detection limits for the putative pathogens in the wastewaters ranged from 16 cells ml⁻¹ for *M. avium* to 1,689 oocysts ml⁻¹ for *Cryptosporidium*. *Cryptosporidium* and *Giardia* spp., *Yersinia pseudotuberculosis*, and pathogenic *Leptospira* spp. were not detected by qPCR.

In the United States, there are 9.3 million milk cows (33) producing an estimated 200 million metric tons of manure annually. Because cattle can be reservoirs of zoonotic pathogens, there is concern over the contamination of soil, water, air, and crops when the manure solids and liquids are land applied (6, 9, 44). At dairies, the solids are often removed from the manure slurries and the liquid fraction (i.e., urine, wash water) is then sent to a pond for storage or can be anaerobically digested first to produce biogas (43). During the crop-growing season, the pond wastewater is diluted with irrigation water and land applied through pressurized irrigation systems to improve the soil nutrient status. It is during spray irrigation that zoonotic pathogens could be aerosolized, increasing the risk of exposure to downwind receptors via inhalation or ingestion after deposition on fomites or food crops (8, 23). Once in the soil after manure addition, pathogens can be internalized by plants (49) and also reach recreational waters by overland flow transport during rainfall events (36, 53, 54), causing significant contamination.

Zoonotic pathogens of potential interest in cattle are *Campylobacter jejuni*, *Escherichia coli*, *Leptospira* spp., *Listeria monocytogenes*, *Mycobacterium avium* subsp. *paratuberculosis*, *Salmonella* spp., *Yersinia* spp., *Giardia lamblia*, and *Cryptosporidium parvum* (38, 39, 40). While numerous studies have measured the occurrence of these zoonotic pathogens in cattle manures and assessed their fate and transport in the environment (41, 48, 55), to our knowledge, no comprehensive studies have been conducted to quantify pathogens in dairy wastewaters. In addition, only a few studies to date have quantified a wide range of pathogens within cattle manures (20, 21, 26). Understanding the number of pathogens in any land-applied waste is particularly important when developing a quantitative microbial risk assessment (12, 17, 56). Estimation of the risk represented by pathogens in animal manures, however, has largely been based on the cultivation and enumeration of fecal indicator organisms (2, 10). It has been recently shown that molecular methods can be successfully applied with such difficult-to-analyze materials to complement culture-dependent approaches (26, 27).

In this study, we attempted to use quantitative real-time PCR (qPCR) to enumerate fecal indicator organisms and putative zoo-

notic pathogens in wastewaters from dairies in southern Idaho. qPCR is often used as a convenient alternative to culture-dependent techniques that has the added advantage of being able to detect viable but nonculturable (VBNC) cells of potential pathogens (7, 27, 59). While there are documented advantages and disadvantages with both molecular method- and culture-dependent approaches, it was our intent to provide the first quantitative survey of selected bacterial indicators and putative pathogens in dairy wastewaters. The bacterial indicators (enterococci, total coliforms, *E. coli*, and *Clostridium perfringens*) were also quantified using culture-dependent methodologies to gauge the effectiveness of qPCR as an enumeration technique for use with dairy wastewaters.

MATERIALS AND METHODS

Dairy operations and sample collection. Wastewater samples were collected from 30 storage ponds at dairy operations of various levels of stocking density located in southern Idaho. Eight 500-ml near-surface samples were collected from the perimeter of each storage pond and then composited in a sterile 4-liter Nalgene container. The composite samples were transferred to our laboratory in coolers with ice packs and immediately stored in the dark under conditions of refrigeration at 5°C upon receipt. All samples were processed within 24 h of collection. Three sets of samples were collected from the ponds during the summer and fall of 2011.

Culture-dependent quantification of fecal indicators. Prior to cultivation as described below, 10-fold serial dilutions of thoroughly mixed wastewater samples were prepared in room temperature phosphate-buffered saline (PBS). The most probable numbers (MPNs) for *Escherichia coli* (and total coliforms) and enterococci were determined using Colilert-18 and Enterolert detection kits, respectively, in conjunction with a Quanti-

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TABLE 1 Target genes, qPCR assays, and modified thermocycler conditions

Target organism	Target gene	No. of targets per genome	Primers	Primer annealing temp (°C); extension time (s)	Reference(s) ^a
<i>Campylobacter jejuni</i>	VS1	1	VS15-F, VS16-R	55; 30	58
<i>Clostridium perfringens</i>	Alpha toxin	1	cpa-F, cpa-R	55; 30	15
<i>Enterococcus faecalis</i>	23S rRNA	2	ECF, ECR	60; 30	18
<i>Escherichia coli</i>	Glucuronidase	1	Eco-F, Eco-R	60; 30	46
Enterohemorrhagic <i>E. coli</i>	Shiga toxin	1	stx1-F, stx1-R	55; 30	47
	Shiga toxin	1	stx2-F, stx2-R	55; 30	47
	Intimin	1	eae-F, eae-R	55; 30	47
<i>Leptospira</i> spp.	Lipoprotein, <i>lipL32</i>	1	45F, 286R	60; 30	51
<i>Listeria monocytogenes</i>	Listeriolysin O	1	F, R	60; 30	35
<i>Mycobacterium avium</i> subsp. <i>paratuberculosis</i>	IS900	20	F2, R2	65; 30	5, 25
<i>Salmonella enterica</i>	Invasin	1	139-F, 141-R	65; 30	19
<i>Yersinia pseudotuberculosis</i>	Invasin	1	inv-F, inv-R	60; 30	52
<i>Cryptosporidium</i> spp.	COWP ^b	4	P702-F, P702-R	60; 30	16
<i>Giardia</i> spp.	β-Giardin	16	P241-F, P241-R	60; 30	16

^a Probe sequences from the associated references were used for the PCRs, except SYBR green was used for *Y. pseudotuberculosis*.

^b COWP, *Cryptosporidium* oocyst wall protein gene.

Tray 2000 (IDEXX Laboratories, Inc., Westbrook, ME). Vegetative cells and spores of *C. perfringens* were assayed using membrane filtration and mCP media (Neogen, Lansing, MI) with mCP selective Supplement I (Sigma-Aldrich, St. Louis, MO) as described by Armon and Payment (1). The mCP plates were anaerobically incubated at 44.5°C for 1 day, with exposure to ammonia hydroxide vapors afterward to quantify presumptive *C. perfringens* colonies. Use of negative controls was implemented, while positive controls consisted of *E. coli* (ATCC 13706), *Enterococcus faecalis* (ATCC 700802), and *C. perfringens* (ATCC 13124). All trays and plates were manually counted, and their numbers were reported as MPN and CFU per milliliter of wastewater, respectively.

Isolation of microbial DNA from the wastewater. Three aliquots were removed from each 4-liter container of wastewater after thorough mixing. Depending on the solids content, aliquots ranging in volume from 5 to 70 ml were dispensed into 85-ml Oak Ridge tubes (Nalge Nunc International, Rochester, NY) and brought up to volume with cold PBS. After centrifugation at 10,000 × g for 10 min, the supernatant was discarded and the remaining pellet was washed twice with PBS. The pellet was then resuspended with 600 μl of PBS and then transferred to a bead-beating tube from a FastDNA Spin kit for feces (MP Biomedicals LLC, Solon, OH) and processed using a FastPrep FP120 instrument at a speed setting of 6 m s⁻¹ for 45 s according to the manufacturer's recommendations. However, all wash and centrifugation steps were carried out twice to further reduce possible humic acid contamination. The DNA was eluted during the final step with 100 μl of *N*-tris(hydroxymethyl)methyl-2-aminoethanesulfonic acid (TES) buffer and then stored at -20°C until analysis by qPCR.

Quantitative PCR. qPCR was performed with a Bio-Rad (Hercules, CA) iQ5 multicolor real-time PCR detection system using dually labeled Black Hole Quencher (BHQ) probes or iQ SYBR green Supermix (Bio-Rad). The probes were manufactured by Biosearch Technologies (Novato, CA), with FAM (6-carboxyfluorescein) and BHQ-1 as the fluorophore and quencher, respectively. Organisms enumerated in the wastewater samples are listed in Table 1, along with their target genes, individual cycling conditions, and references for primers and probes used. Quantification standards were prepared from pure cultures of *C. jejuni* (ATCC BAA-1062), *C. perfringens* (ATCC 13124), *E. faecalis* (ATCC 700802), *E. coli* O157:H7 strain 3032 (courtesy of Tom Casey, USDA-ARS), *Leptospira interrogans* serovar Copenhageni (ATCC BAA-1198), *L. monocytogenes* (ATCC BAA-679), *M. avium* subsp. *paratuberculosis* (ATCC BAA-968), *Salmonella enterica* serovar Typhimurium (courtesy of Pina Frantamico, USDA-ARS), and *Yersinia pseudotuberculosis* (ATCC 29833). Genomic DNA was extracted from the cultures using a Power-Microbial Maxi DNA isolation kit (MO BIO Laboratories, Inc., Carlsbad,

CA), while DNA for *C. parvum* (PRA-67D) and *Giardia intestinalis* (30888D) was directly purchased from the American Type Culture Collection (Manassas, VA).

One microliter of nondiluted DNA template was used in a PCR with 0.3 μM (each) primer, 0.1 μM probe, 12.5 μl of iQ Supermix (Bio-Rad), and molecular analysis-grade water to achieve a final volume of 25 μl. SYBR green was used in lieu of probe only for the detection of *Y. pseudotuberculosis*. After an initial denaturation step (95°C for 180 s), 40 cycles of 94°C for 20 s followed by an extension time of 30 s at primer-annealing temperatures were performed as indicated in Table 1. The number of putative pathogens in the samples was calculated using the DNA concentration of the standards, genome molecular weight, and target gene frequency (26).

Pathogen recovery and detection limits. To determine the performance of the DNA extraction technique and qPCR to detect pathogens in the dairy wastewaters, one wastewater sample was individually spiked with *C. jejuni*, enterohemorrhagic *E. coli*, *L. monocytogenes*, *M. avium* subsp. *paratuberculosis*, *S. enterica* serovar Typhimurium, and *Y. pseudotuberculosis*. The wastewater chosen had a pH of 7.5 and total solids of 6.7 g liter⁻¹. In brief, an overnight culture was serially diluted in sterile PBS and then 1 ml of the appropriate dilution was dispensed into triplicate 9-ml aliquots of wastewater. The titer of the inoculum was determined by plate counting after growth of the organisms on appropriate media. For each pathogen, a total of 4 to 5 spiked dilutions were prepared, with final wastewater titer levels ranging from log₁₀ 1.5 to 8.5. After spiking, the wastewater samples were processed in a vortex apparatus for 30 s and then a 5-ml aliquot was immediately processed as previously described for the DNA isolation.

For *Giardia* and *Cryptosporidium* spp., the same spiking and DNA isolation procedures were utilized as described above, except that the cysts and oocysts were spiked into two different wastewaters to determine a possible influence of levels of solids on qPCR performance. The low and high total solids contents of the wastewaters were 3.6 g liter⁻¹ and 12.7 g liter⁻¹, respectively. *Giardia lamblia* and *C. parvum* at a titer of 10⁷ cysts/oocysts ml⁻¹ were purchased from Waterborne, Inc. (New Orleans, LA).

The percent recovery was determined for each pathogen by dividing the qPCR result by the calculated wastewater titer level, which was based on the organism density of each spiking dilution. Detection limits were determined via a linear regression analysis of the wastewater titer levels versus the qPCR cycle threshold (C_T) values, where $r^2 > 0.96$. The minimum detection limits were determined at a C_T of no greater than 35, a value at which detection errors may become significant due to pipetting errors or potential traces of cross-contamination.

Statistical analysis. Our *a priori* hypothesis was that there would be no statistical difference between the titers as determined by the culture-dependent techniques and qPCR for the indicator organisms. The two-sample paired *t* tests for the means were performed on log-transformed data using SAS statistical software version 9.2 (45). Statements of statistical significance were based on an α value of 0.05.

RESULTS AND DISCUSSION

Dairy wastewater samples and chemical properties. Thirty freestall and open-lot dairies with <1,000 to 10,000 milk cows were targeted for this study. Most of the dairies were stocked with Holstein cows; however, in some cases, they were stocked with a combination of Holstein and Jersey cows. The freestall dairies used either a flush or a vacuum system to remove the manure from the alleys, while the open-lot dairies used a scrape system. The flushed or vacuumed manure is then commonly transferred to solid separation cells, followed, in some cases, by processing with a separator to reduce the solids content prior to discharge to the wastewater storage pond.

During the months of August and October, the pH of the wastewaters ranged from 6.6 to 8.8, with a median value of 7.7 (see Table S1 in the supplemental material). The total solids content ranged from a very low value of 0.28 g liter⁻¹ to as high as 57.3 g liter⁻¹, with a median value of 8.2 g liter⁻¹. The total Kjeldahl nitrogen value ranged from 0.001 to 1.5 g liter⁻¹, with a median value of 0.24 g liter⁻¹. The low values, which are much lower than anticipated for dairy wastewaters, were related to the fact that a few of the wastewater ponds were flooded with irrigation water. This practice was implemented at a few of the dairies as a means to blend the wastewaters prior to land application via pressurized spray irrigation. Because at some point more irrigation water than wastewater enters the pond, the pond takes on the chemical characteristics of the irrigation water.

Culture-dependent quantification of fecal indicators. The fecal indicator organisms were quantified in the dairy wastewaters using a MPN (enterococci, total coliforms, and *E. coli*) or plate-counting (*C. perfringens*) technique (see Table S2 in the supplemental material). Median titer values for enterococci and *C. perfringens* were approximately 10² MPN or CFU ml⁻¹, respectively, throughout the sampling period. Total coliform median values were 1 to 2 orders of magnitude greater, with a maximum titer of 10⁷ MPN ml⁻¹ occurring in August. As expected, the titer of generic *E. coli* was lower than that of total coliforms in all pond samples (data not shown). On average, the organism titer levels were greater in October than in June and August, except for the titers of total coliforms, which were greatest in August. In the United Kingdom, Hutchison and coworkers (21) found that *E. coli* O157, *Campylobacter* spp., and *Listeria* spp. were more prevalent in fresh cattle feces during the spring months and December. Additional studies have confirmed that the prevalences and levels of various zoonotic agents within fresh cattle manures were affected in a season-dependent manner (3, 50). While too few seasonal data points were collected for a statistical analysis, our results suggest that seasonal factors such as temperature or solar irradiation could have influenced the levels of fecal indicators in the dairy wastewaters. It should be noted that changes in the levels of fecal indicators in the wastewater ponds may not be seasonal in nature and could be related to the manure aging process or nutrient status of the pond. In general, it has been reported that levels of bacterial indicators and zoonotic agents in stored livestock wastes decline over time (22, 28, 34, 57). However, regrowth of indicator

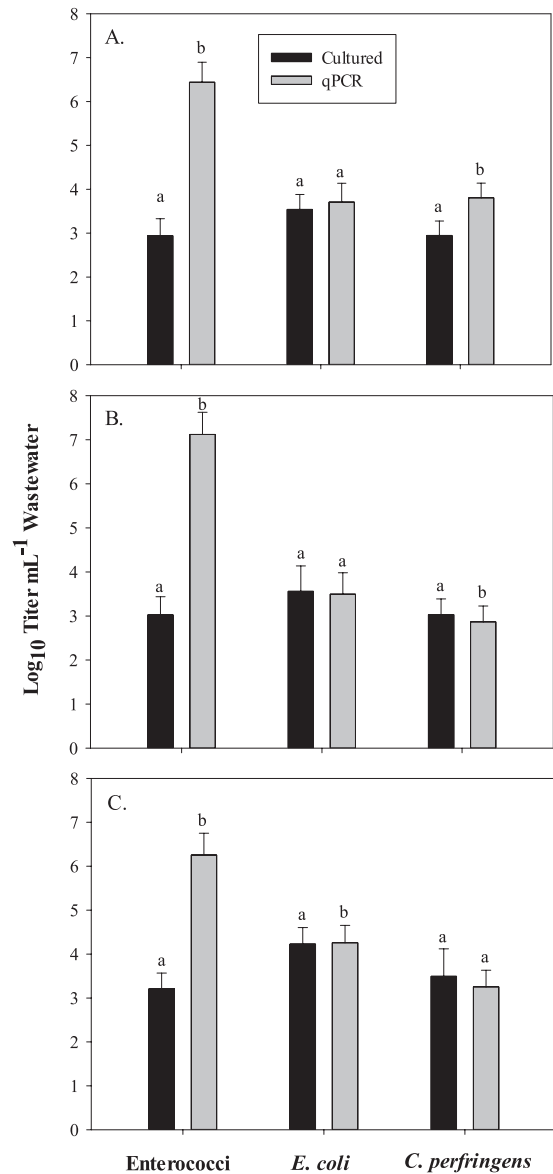


FIG 1 Comparison of culture-dependent and qPCR estimates of numbers of enterococci, *E. coli*, and *C. perfringens* in the dairy wastewaters from (A) June, (B) August, and (C) October. Quantitative PCR was based on positive amplification of target genes using EC, Eco, and *cpa* primers for enterococci, *E. coli*, and *C. perfringens*, respectively. Columns represent means \pm standard deviations ($n = 30$). Columns with different letters (a or b) indicate a significant difference between the two quantification techniques ($P < 0.05$).

bacteria to some extent, due to the nutrient-rich conditions, has also been reported (13, 28).

Quantification of fecal indicators by qPCR. In addition to quantification by culture-dependent techniques, qPCR of bacterial DNA was used to calculate the titers of enterococci, *E. coli*, and *C. perfringens* in the dairy wastewaters. The qPCR results are presented in Fig. 1 along with the MPN and CFU values for comparison. During each of the 3 months in which samples were collected, the qPCR values for enterococci corresponded to 10⁶ to 10⁷ cells ml⁻¹ wastewater. The qPCR method was specifically designed to detect *E. faecalis*, which is one of the most common enterococcus species found in the gastrointes-

TABLE 2 Recovery and detection limits for bacterial pathogens in a dairy wastewater determined by qPCR after spiking

Target organism	Wastewater titer range after spiking (log ₁₀ CFU ml ⁻¹ wastewater)	Range of recovery values (%)	Mean recovery (%) ^a	qPCR detection limit (no. of cells ml ⁻¹)
<i>C. jejuni</i>	3.4–8.4	41–836	326	768
<i>E. coli</i>				
<i>stx1</i>	3.2–8.2	38–120	72	1,229
<i>stx2</i>	3.2–8.2	54–238	121	232
<i>eaeA</i>	2.2–8.2	41–117	75	140
<i>L. monocytogenes</i>	3.5–8.5	45–127	81	410
<i>M. avium</i> subsp. <i>paratuberculosis</i>	1.5–6.5	25–52	37	16
<i>S. enterica</i>	3.3–8.3	13–133	56	325
<i>Y. pseudotuberculosis</i>	3.2–8.2	126–179	160	234

^a Data represent mean values as determined by spiking triplicate wastewater samples with pathogens at several titer levels.

tinal tract of cattle (14, 32). qPCR values were about 3 logs greater than those determined by the MPN technique (a statistically significant difference; $P < 0.0001$), which indicated that the majority of enterococci might have gone into a VBNC state such as was seen in earlier studies involving bovine feces and manure (28, 30). However, since the DNA could also have been extracted from damaged or dead cells, qPCR analysis may have resulted in overestimation of actual titer values in the samples (7, 31, 37).

In contrast to the results seen with enterococci, the average qPCR and culture-dependent values for both *E. coli* and *C. perfringens* were markedly but not statistically significantly similar in all cases (Fig. 1). In June and August, the qPCR and CFU values for *C. perfringens* were statistically significantly different ($P < 0.0001$), while the October titer values for *E. coli* were also different ($P = 0.02$). During the 3 months in which wastewater samples were collected, the titers of *E. coli* and *C. perfringens* (as determined by qPCR) ranged from about 10^3 to 10^4 cells ml⁻¹.

Quantification of zoonotic pathogens by qPCR. Prior to the enumeration of zoonotic bacterial pathogens in the dairy wastewaters, wastewater samples were spiked to determine the efficiency of the DNA extraction kit and precision of the qPCR technique. On average, recovery levels ranged from 56% for *S. enterica* to as high as 326% for *C. jejuni* (Table 2). Although these recovery levels are presumed to be representative of all

wastewaters examined in this study, spiking was not performed on all of the wastewaters due to the logistics of conducting this task. The qPCR detection limits for the bacterial pathogens were determined to range from 16 to 1,229 cells ml⁻¹ (Table 2). Due to the presence of multiple copies of the target gene within *M. avium* (i.e., 14 to 20 copies per genome), analysis performed using those sequences is more sensitive than analysis using the other targeted sequences (5), thus explaining the low detection limit of 16 cells ml⁻¹.

The putative bacterial pathogens detected in the dairy wastewater by qPCR were *C. jejuni*, *E. coli* (*stx*₁ and *eaeA* positive), *L. monocytogenes*, *M. avium*, and *S. enterica* (Table 3). All of these organisms were detected at least twice in samples from June, but *L. monocytogenes* and *S. enterica* were not detected in the August and October samples. The organisms with the greatest number of positive detections were *C. jejuni* and *M. avium*, with 21 and 29 of 30 ponds showing the presence of those species in samples collected in June, respectively. While the greatest number of detections for all organisms occurred in that first month of analysis, the number of positive pond results generally decreased during each subsequent month. The decrease in the number of detections may have been affected by the pathogen-shedding rate of the cattle, environmental variables (e.g., solar irradiation, temperature), pond characteristics (e.g., solids content), or competition from other indigenous microorganisms leading to a reduction of pathogen

TABLE 3 Enumeration of pathogens in dairy wastewater samples as determined by qPCR^a

Organism	June			August			October					
	No. of ponds	No. of cells (ml ⁻¹)		No. of ponds	No. of cells (ml ⁻¹)		No. of ponds	No. of cells (ml ⁻¹)				
		Min	Max	Mean		Min	Max	Mean		Min	Max	Mean
<i>C. jejuni</i>	21	<7.7 × 10 ²	2.7 × 10 ⁴	4.8 × 10 ³	17	<7.7 × 10 ²	1.8 × 10 ⁴	2.5 × 10 ³	13	<7.7 × 10 ²	2.5 × 10 ⁴	3.4 × 10 ³
<i>E. coli</i>												
<i>stx</i> ₁	13	<1.2 × 10 ³	6.3 × 10 ³	2.0 × 10 ³	2	3.6 × 10 ³	9.9 × 10 ³	6.8 × 10 ³	5	<1.2 × 10 ³	2.7 × 10 ³	1.0 × 10 ³
<i>eaeA</i>	10	1.5 × 10 ²	6.3 × 10 ³	2.0 × 10 ³	2	2.4 × 10 ³	4.2 × 10 ³	3.3 × 10 ³	1	1.4 × 10 ²	1.4 × 10 ²	1.4 × 10 ²
<i>L. monocytogenes</i>	2	1.4 × 10 ³	6.6 × 10 ³	4.0 × 10 ³	0	<4.1 × 10 ²			0	<4.1 × 10 ²		
<i>M. avium</i>	29	<1.6 × 10 ¹	7.1 × 10 ⁴	2.9 × 10 ³	24	<1.6 × 10 ¹	1.3 × 10 ³	1.5 × 10 ²	22	<1.6 × 10 ¹	1.4 × 10 ³	2.2 × 10 ²
<i>S. enterica</i>	5	3.0 × 10 ³	8.3 × 10 ⁴	2.1 × 10 ⁴	0	<3.2 × 10 ²			0	<3.2 × 10 ²		

^a No sample had detectable *L. interrogans*, *Y. pseudotuberculosis*, *Cryptosporidium* spp., *Giardia* spp., or *stx*₂ gene of *E. coli* O157:H7 during any month. No. of ponds, number of wastewater ponds containing the indicated pathogen ($n = 30$ ponds). Min, minimum; Max, maximum.

TABLE 4 Recovery of and detection limits for *Giardia* and *Cryptosporidium* spp. in dairy wastewaters as determined by qPCR with low and high total solids

Target organism	Total solids (g liter ⁻¹)	Wastewater titer range after spiking (log ₁₀ cysts/oocysts ml ⁻¹ wastewater)	Range of recovery values (%)	Mean recovery (%) ^a	qPCR detection limit (log ₁₀ cysts/oocysts ml ⁻¹)
<i>Giardia</i> spp.	3.6	3–6	8–20	15	130
	12.7		15–25	19	182
<i>Cryptosporidium</i> spp.	3.6	3–6	22–110	61	1,485
	12.7		71–426	210	1,689

^a Data represent mean values as determined by spiking triplicate wastewater samples with pathogens at several titer levels.

survival in the ponds (29, 41, 42). The fact that *Leptospira* spp. were not detected in the ponds could be related to the performance of the DNA extraction, as spiking studies were not conducted with these organisms. However, sufficient recovery levels of these organisms in cattle manure have been reported from studies using similar DNA isolation techniques (26), making it more likely that they were present at levels below the detection limits. *Yersinia pseudotuberculosis* was also not detected, but, based on the spiking results, it can be expected that the titer in the ponds was below the detection limit of 234 cells ml⁻¹.

To determine the effect of solids on the recovery and detection of pathogens, wastewaters with low and high total solids contents were spiked with *G. lamblia* cysts and *C. parvum* oocysts (Table 4). Regardless of the solids content, the mean levels of recovery of cysts from the low and high wastewater totals, at 15% and 19%, respectively, were very similar. In contrast, the recovery of oocysts from the high total content of wastewater solids was 210%, which is 3.4 times higher than in the low total content. Although this test was performed with only two pathogens, these results suggest that recovery of organisms might be slightly enhanced in some wastewaters with a higher solids content. Despite our efforts to quantify *Giardia* and *Cryptosporidium* spp. in the dairy wastewaters, no pond samples were found to contain them above their average detection limits of 156 cysts and 1,587 oocysts ml⁻¹. Our results, however, do not exclude the possible presence of *Giardia* and *Cryptosporidium* spp. in the wastewaters, since they are generally found at relatively low concentrations in cattle feces (20, 21).

The bacterial pathogen titers were generally consistent, with mean values of between 10² and 10⁴ cells ml⁻¹ (Table 3). The organisms detected at the highest titer levels were *M. avium*, *S. enterica*, and *C. jejuni* at 10⁴ cells ml⁻¹ in June, while *C. jejuni* was also detected at titer values of up to 10⁴ cells ml⁻¹ in August and October. The mean titer value for *L. monocytogenes* and *stx*₁- and *eaeA*-positive *E. coli* was 10³ cells ml⁻¹, except in October, where the titer for *eaeA*-positive *E. coli* was lower by 1 order of magnitude. The fact that the *stx*₁ and *eaeA* gene markers were detected suggests that some of the wastewaters were presumptive for enterohemorrhagic or enteropathogenic *E. coli*. While most studies do report the presence of virulence genes encoding Shiga toxin 2 in dairy manures (4, 11), the *stx*₂ gene was not detected in the wastewater ponds. Although our qPCR detection limit for *stx*₂ was 232 cells ml⁻¹, the lack of detections suggests that our methodology may not have been sensitive enough, as the majority of Shiga toxin-producing *E. coli* isolates from dairy cattle are known to possess both *stx*₁ and *stx*₂ genes (24).

Concluding remarks. Based on our results, the oligonucleotide primers and probes used in this study appeared to be suitable for quantitation of zoonotic bacterial pathogens in dairy wastewaters. While some of the key pathogens in the wastewaters were below their qPCR detection limits, this study successfully detected and quantified bacterial pathogens of public health concern at levels similar to those reported in the literature in related environments. In addition, the qPCR and culture-dependent results were in general agreement, further suggesting the suitability of using qPCR as an alternative to laborious culture-dependent techniques when analyzing dairy and (potentially) other livestock wastewaters. Once more, qPCR has turned out to be a reliable method to quantify pathogens in materials that are often difficult to analyze. Despite the fact that some of the values tended to be more conservative due to a possible codetection of irreversibly damaged or dead organisms, this approach allows the detection of putative pathogens in a dormant state which would not be detectable using culture techniques alone.

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Mention of trade names or commercial products in this publication is solely for the purpose of providing specific information and does not imply recommendation or endorsement by the USDA.

REFERENCES

- Armon R, Payment P. 1988. A modified m-CP medium for enumerating *Clostridium perfringens* from water samples. *Can. J. Microbiol.* 34:78–79.
- Budnick GE, Howard RT, Mayo DR. 1996. Evaluation of Enterolert for enumeration of enterococci in recreational waters. *Appl. Environ. Microbiol.* 62:3881–3884.
- Chapman PA, Siddons CA, Cerman Malo AT, Harkin MA. 1997. A 1-year study of *Escherichia coli* O157 in cattle, sheep, pigs and poultry. *Epidemiol. Infect.* 119:245–250.
- Cobbold RN, Rice DH, Szymanski M, Call DR, Hancock DD. 2004. Comparison of shiga-toxigenic *Escherichia coli* prevalences among dairy, feedlot, and cow-calf herds in Washington state. *Appl. Environ. Microbiol.* 70:4375–4378.
- Cook KL, Britt JS. 2007. Optimization of methods for detecting *Mycobacterium avium* subsp. *paratuberculosis* in environmental samples using quantitative, real-time PCR. *J. Microbiol. Methods* 69:154–160.
- Cook KL, Bolster CH, Ayers KA, Reynolds DN. 2011. *Escherichia coli* diversity in livestock manures and agriculturally impacted stream waters. *Curr. Microbiol.* 63:439–449.
- Dungan RS, Leytem AB. 2009. Qualitative and quantitative methodologies for determination of airborne microorganisms at concentrated animal-feeding operations. *W. J. Microbiol. Biotechnol.* 25:1505–1518.
- Dungan RS. 2010. Board-Invited Review: Fate and transport of bioaerosols associated with livestock operations and manures. *J. Anim. Sci.* 88:3693–3706.
- Dungan RS. 2012. Use of a culture-independent approach to characterize

- aerosolized bacteria near an open-freestall dairy operation. *Environ. Int.* 41:8–14.
10. Eckner KF. 1998. Comparison of membrane filtration and multiple-tube fermentation by the Colilert and Enterolert methods for detection of waterborne coliform bacteria, *Escherichia coli*, and enterococci used in drinking and bathing water quality monitoring in southern Sweden. *Appl. Environ. Microbiol.* 64:3079–3083.
 11. Franz E, Klerks MM, De Vos OJ, Termorshuizen AJ, van Bruggen AHC. 2007. Prevalence of shiga toxin-producing *Escherichia coli* *stx*₁, *stx*₂, *eaeA*, and *rfbE* genes and survival of *E. coli* O157:H7 in manure from organic and low-input conventional dairy farms. *Appl. Environ. Microbiol.* 73:2180–2190.
 12. Gale P. 2005. Land application of treated sewage sludge: quantifying pathogen risks from consumption of crops. *J. Appl. Microbiol.* 98:380–396.
 13. Gantzer C, et al. 2001. Monitoring of bacterial and parasitological contamination during various treatment of sludge. *Water Res.* 35:3763–3770.
 14. Gilmore MS. 2002. The enterococci: pathogenesis, molecular biology, and antibiotic resistance. ASM Press, Washington, DC.
 15. Gurjar AA, Hegde NV, Love BC, Jayarao BM. 2008. Real-time multiplex PCR assay for rapid detection and toxinotyping of *Clostridium perfringens* toxin producing strains in feces of dairy cattle. *Mol. Cell. Probes* 22:90–95.
 16. Guy RA, Payment P, Krull UJ, Horgen PA. 2003. Real-time PCR for quantification of *Giardia* and *Cryptosporidium* in environmental water samples and sewage. *Appl. Environ. Microbiol.* 69:5178–5185.
 17. Hardy R, Schilling K, Fromm J, Dai X, Cook M. 2006. Technical background document: microbial risk assessment and fate and transport modeling of aerosolized microorganisms at wastewater land application facilities in Idaho. Idaho Department of Environmental Quality, Boise, ID. http://www.deq.idaho.gov/media/529643-microbial_risk_assessment.pdf. Accessed 24 September 2012.
 18. He JW, Jiang S. 2005. Quantification of enterococci and human adenovirus in environmental samples by real-time PCR. *Appl. Environ. Microbiol.* 71:2250–2255.
 19. Hein I, Flekna G, Krassnig M, Wagner M. 2006. Real-time PCR for the detection of *Salmonella* spp. in food: An alternative approach to a conventional PCR system suggested by the FOOD-PCR project. *J. Microbiol. Methods* 66:538–547.
 20. Hutchison ML, Walters LD, Avery SM, Synge BA, Moore A. 2004. Levels of zoonotic agents in British livestock manures. *Lett. Appl. Microbiol.* 39:207–214.
 21. Hutchison ML, Walters LD, Avery SM, Munro F, Moore A. 2005. Analyses of livestock production, waste storage, and pathogen levels and prevalences in farm manures. *Appl. Environ. Microbiol.* 71:1231–1236.
 22. Hutchison ML, Waters LD, Moore A, Avery SM. 2005. Declines of zoonotic agents in liquid livestock wastes stored in batches on-farm. *J. Appl. Microbiol.* 99:58–65.
 23. Hutchison ML, Avery SM, Monaghan JM. 2008. The air-borne distribution of zoonotic agents from livestock waste spreading and microbiological risk to fresh produce from contaminated irrigation sources. *J. Appl. Microbiol.* 105:848–857.
 24. Irino K, et al. 2005. Serotypes and virulence markers of shiga toxin-producing *Escherichia coli* (STEC) isolated from dairy cattle in São Paulo State, Brazil. *Vet. Microbiol.* 105:29–36.
 25. Kim SG, et al. 2002. Development and application of quantitative polymerase chain reaction assays based on the ABI 7700 system (TaqMan) for detection and quantification of *Mycobacterium avium* subsp. *paratuberculosis*. *J. Vet. Diagn. Invest.* 14:126–131.
 26. Klein M, et al. 2010. Diversity and abundance of zoonotic pathogens and indicators in manures of feedlot cattle in Australia. *Appl. Environ. Microbiol.* 76:6947–6950.
 27. Klein M, et al. 2010. Monitoring bacterial indicators and pathogens in cattle feedlot waste by real-time PCR. *Water Res.* 44:1381–1388.
 28. Klein M, Brown L, Ashbolt NJ, Stuetz RM, Roser DJ. 2011. Inactivation of indicators and pathogens in cattle feedlot manures and compost as determined by molecular and culture assays. *FEMS Microbiol. Ecol.* 77:200–210.
 29. Kudva IT, Blanch K, Hovde CJ. 1998. Analysis of *Escherichia coli* O157:H7 survival in ovine or bovine and manure slurry. *Appl. Environ. Microbiol.* 64:3166–3174.
 30. Lebuhn M, Effenberger M, Gronauer A, Wilderer PA, Wuertz S. 2003. Using quantitative real-time PCR to determine the hygienic status of cattle manure. *Water Sci. Technol.* 48:97–103.
 31. Lleo MM, et al. 2005. Molecular vs culture methods for the detection of bacterial faecal indicators in groundwater for human use. *Lett. Appl. Microbiol.* 40:289–294.
 32. Mete E, Kaleli I. 2006. Species distribution and antibiotic resistance of enterococci isolated from cattle farmers and cattles. *Mikrobiyol. Bul.* 40:75–80. (In Turkish.)
 33. National Agricultural Statistics Service. 2009. 2007 census of agriculture. USDA-NASS, Washington, DC. http://www.agcensus.usda.gov/Publications/2007/Full_Report/index.asp. Accessed 24 September 2012.
 34. Nicholson FA, Groves SJ, Chambers BJ. 2005. Pathogen survival during livestock manure storage and following land application. *Bioresour. Technol.* 96:135–143.
 35. Nogva HK, Rudi K, Naterstad K, Holck A, Lillihaug D. 2000. Application of 5'-nuclease PCR for quantitative detection of *Listeria monocytogenes* in pure cultures, water, skim milk, and unpasteurized whole milk. *Appl. Environ. Microbiol.* 66:4266–4271.
 36. Oliver DM, Clegg CD, Haygarth PM, Heathwaite AL. 2005. Assessing the potential for pathogen transfer from grassland soils to surface waters. *Ad. Agron.* 85:125–180.
 37. Oliver JD. 2010. Recent findings on the viable but nonculturable state in pathogenic bacteria. *FEMS Microbiol. Rev.* 34:415–425.
 38. Pell AN. 1997. Manure and microbes: public and animal health problem? *J. Dairy Sci.* 80:2673–2681.
 39. Pelzer KD, Currin N. 2007. Zoonotic diseases of cattle. Virginia Cooperative Extension, publication 400-460. Virginia Tech, Blacksburg, VA. <http://pubs.ext.vt.edu/400/400-460/400-460.html>. Accessed 24 September 2012.
 40. Raizman EA, et al. 2004. The distribution of *Mycobacterium avium* ssp. *paratuberculosis* in the environment surrounding Minnesota dairy farms. *J. Dairy Sci.* 87:2959–2966.
 41. Ravva SV, Sarreal CZ, Duffy B, Stanker LH. 2006. Survival of *Escherichia coli* O157:H7 in wastewater from dairy lagoons. *J. Appl. Microbiol.* 101:891–902.
 42. Ravva SV, Korn A. 2007. Extractable organic components and nutrients in wastewater from dairy lagoons influence the growth and survival of *Escherichia coli* O157:H7. *Appl. Environ. Microbiol.* 73:2191–2198.
 43. Rico JL, Garcia H, Rico C, Tejero I. 2007. Characterisation of solid and liquid fractions of dairy manure with regard to their component distribution and methane production. *Bioresour. Technol.* 98:971–979.
 44. Salgado M, et al. 2011. Fate of *Mycobacterium avium* subsp. *paratuberculosis* after application of contaminated dairy cattle manure in agricultural soils. *Appl. Environ. Microbiol.* 77:2122–2129.
 45. SAS Institute. 2008. SAS/STAT 9.2 user's guide. SAS Institute, Cary, NC.
 46. Shannon KE, Lee DY, Trevors JT, Beaudette LA. 2007. Application of real-time quantitative PCR for the detection of selected bacterial pathogens during municipal wastewater treatment. *Sci. Total Environ.* 382:121–129.
 47. Sharma VK, Dean-Nystrom EA. 2003. Detection of enterohemorrhagic *Escherichia coli* O157:H7 by using a multiplex real-time PCR assay for genes encoding intimin and Shiga toxins. *Vet. Microbiol.* 93:247–260.
 48. Slana I, Pribylova R, Kralova A, Pavlik I. 2011. Persistence of *Mycobacterium avium* subsp. *paratuberculosis* at a farm-scale biogas plant supplied with manure from paratuberculosis-affected dairy cattle. *Appl. Environ. Microbiol.* 77:3115–3119.
 49. Solomon EB, Yaron S, Matthews KR. 2002. Transmission of *Escherichia coli* O157:H7 from contaminated manure and irrigation water to lettuce plant tissue and its subsequent internalization. *Appl. Environ. Microbiol.* 68:397–400.
 50. Stanley KN, Wallace JS, Currie JE, Diggle PJ, Jones K. 1998. The seasonal variation of thermophilic campylobacters in beef cattle, dairy cattle and calves. *J. Appl. Microbiol.* 85:472–480.
 51. Stoddard RA, Gee JE, Wilkins PP, McCaustland K, Hoffmaster AR. 2009. Detection of pathogenic *Leptospira* spp. through TaqMan polymerase chain reaction targeting the LipL32 gene. *Diagn. Microbiol. Infect. Dis.* 64:247–255.
 52. Thoerner P, et al. 2003. PCR detection of virulence genes in *Yersinia enterocolitica* and *Yersinia pseudotuberculosis* and investigation of virulence gene distribution. *Appl. Environ. Microbiol.* 69:1810–1816.
 53. Tyrrel SF, Quinton JN. 2003. Overland flow transport of pathogens from agricultural land receiving faecal wastes. *J. Appl. Microbiol.* 94:875–935.
 54. Unc A, Goss MJ. 2004. Transport of bacteria from manure and protection of water resources. *Appl. Soil Ecol.* 25:1–18.
 55. Venglovsky J, Saskova N, Placha I. 2009. Pathogens and antibiotic resi-

- dues in animal manures and hygienic and ecological risks related to subsequent land application. *Bioresour. Technol.* **100**:5386–5391.
56. Viau E, Bibby K, Paez-Rubio T, Peccia J. 2011. Toward a consensus view on the infectious risks associated with land application of sewage sludge. *Environ. Sci. Technol.* **45**:5459–5469.
57. Wang L, Mankin KR, Marchin GL. 2004. Survival of fecal bacteria in dairy cow manure. *Trans. ASAE* **47**:1239–1246.
58. Yang C, Jiang Y, Huang K, Changqing Z, Yin Y. 2003. Application of real-time PCR for quantitative detection of *Campylobacter jejuni* in poultry, milk and environmental water. *FEMS Immunol. Med. Microbiol.* **38**: 265–271.
59. Zhang G, Brown EW, González-Escalona N. 2011. Comparison of real-time PCR, reverse transcriptase real-time PCR, loop-mediated isothermal amplification, and the FDA conventional microbiological method for the detection of *Salmonella* spp. in produce. *Appl. Environ. Microbiol.* **77**: 6495–6501.