SCIENTIFIC REPORTS



SUBJECT AREAS: EARTH AND ENVIRONMENTAL SCIENCES ZOOLOGY PHYLOGENETICS BIODIVERSITY

Received 21 May 2012 Accepted 6 November 2012

Published 28 November 2012

Correspondence and requests for materials should be addressed to J.A.M. (joshuamackie@gmail. com)

* Current address: Department of Biological Science, San José State University, CA 95192-0100, USA

Ecology of cryptic invasions: latitudinal segregation among *Watersipora* (Bryozoa) species

Joshua A. Mackie¹*, John A. Darling² & Jonathan B. Geller¹

¹Moss Landing Marine Laboratories, 8272 Moss Landing Road, Moss Landing, CA 95039, USA, ²National Exposure Research Laboratory, US Environmental Protection Agency, 109 T. W. Alexander Drive, Research Triangle Park, NC 27711, USA.

Watersipora is an invasive genus of bryozoans, easily dispersed by fouled vessels. We examined Cytochrome c oxidase subunit I haplotypes from introduced populations on the US Pacific coastline to investigate geographic segregation of species and/or haplotypes. In California, the *W. subtorquata* group fell into three major sub-groups: *W. subtorquata* clades A and B, and W. "*new sp.*". *W. subtorquata* clades A and B were common in southern California south of Point Conception, a recognized biogeographic boundary, whereas further north, *W. subtorquata* clade A and *W. n.* sp. were frequent. The southern California region also had colonies of a morphologically distinct species, *W. arcuata*, also found in southern Australia and Hawaii; COI variation indicates a common ancestral source(s) in these introductions. The distribution of *Watersipora*-complex lineages on different coastlines is shown to be temperature correlated. Accordingly, pre-exisitng temperature-based adaptations may play a key role in determining invasion patterns.

enetic studies have widely supported the appearance of 'cryptic' evolutionary divergence at multiple levels of the taxonomic hierarchy, (e.g. multiple species existing within taxa traditionally considered to be single species¹⁻³). The non-native establishment of such diversity is typically referred to as 'cryptic invasion'⁴, and these phenomena often reflect unrecognized multiple introduction events. Given the unsettled taxonomic state and the high rate of transport by anthropogenic vectors for these taxa, it is not surprising that processes underlying marine invasions remain poorly understood⁴⁻⁶. For example, there are few cases in which one could confidently determine from vector history alone, in the absence of genetic information, whether the invasive spread of an organism is the result of propagules from one area or multiple areas⁴.

The existence of cryptic genetic diversity suggests the possibility that corresponding ecological differences could be important for patterns of invasion⁷. Limited examples suggest that invasion potential among such lineages can vary widely, presumably reflecting intrinsic ecological differences. For example, mussels in the genus *Mytilus* include three cryptic species with different environmental tolerances, of which the warmer water species *M. galloprovincialis* has proven an aggressive invader^{8,9}. Similarly, lineages of the barnacle *Balanus glandula* from warmer southern areas of the native range in North America invaded similarly warm coasts of Argentina, while lineages from colder northern areas were found in similarly cold areas in Japan¹⁰. However, for most marine cases where cryptic invasion has been demonstrated, little is known of the ecological differences among the cryptic lineages.

The bryozoan genus *Watersipora* Neviani, 1895 is a promising system to investigate the importance of cryptic invasions. The Bryozoa are a phylum of encrusting animals that have become common in fouling communities the world over due to a predisposition for human-mediated transport^{11,12}, either in ballast water¹³ or on ships' hulls¹⁴. Species of *Watersipora* are among the most invasive Bryozoa. Once released, populations of *Watersipora* grow explosively due to lateral growth of established colonies and settlement of short-lived larvae that may be retained near parents owing to short dispersal durations (generally <24 h)^{15,16}. Introduced *Watersipora* have become a major space occupier on natural and anthropogenic substrata in protected bays in many areas^{17,18}. In communities experimentally polluted with copper ions (an active agent in antifouling paint)^{19,20}, or subjected to heat-wave conditions²¹, invasive *Watersipora* species have settled and occupied space more successfully than most indigenous organisms, suggesting adaptations favoring the rapid spread of these species.

Cryptic diversity in *Watersipora* appears rampant and is incompletely resolved¹⁹ genetic analysis is therefore critical to resolving introduction patterns. As currently understood, the genus consists of *W. arcuata* Banta, 1969, a species with distally recurved tentacle apertures²² and supported by monophyly of mitochondrial Cytochrome c

oxidase subunit I (COI)²³ sequences, and a group of genetically distinct clades we refer to as the *W. subtorquata*-complex that share a proximally pointed sinus in the lophophore aperture (henceforth "sinusoidal"). This latter group has a tortured taxonomic history that has been described as a taxonomic "can of worms"²⁴. The taxonomic difficulty associated with this complex is largely because the genus lacks the spines, avicularia, and external ovicells used for taxonomic diagnosis in most other bryozoans.

The nominal species Watersipora subtorquata (d'Orbigny, 1852) is a widespread invader in cool-temperate areas globally. However, a divergent lineage (15% Kimura 2-parameter nucleotide divergence in COI) from the previously known W. subtorquata-complex was found as a single colony collected in California (this clade was referred to as W. "new sp."23) and in subsequent sampling in California²⁵. Thus, it appears at present that at least two cryptic species comprise the W. subtorquata-complex, and that genetic analysis is necessary to discover their patterns of distribution. While the type specimen locale of W. subtorquata is Rio De Janeiro, Brazil, native ranges of W. subtorquata-complex species are not known due to uncertainties of taxonomy and dispersal history, which may include undocumented movement by vessels. A third species, W. subovoidea (Ryland et al., 2009), which may be invasive in tropical areas, was recently re-described and formally separated from the W. subtorquata-complex following confirmation of genetic divergence in COI²⁶. Ryland et al.²⁶ postulate that W. subovoidea may be native to the Mediterranean, however origin is uncertain, and the IndoPacific presents another possible source for this species group according to suggested morphological affinity27.

Recognition of introductions of Watersipora began when it was realized that the species now referred to as W. arcuata and thought to be native to the tropical eastern Pacific²⁸, had invaded Australian coastline. In fact, by the mid 20th century, W. arcuata was extremely common around Australian harbors, indicating hull-fouling in spread^{23,29}, but the arrival time can only be approximated to within 1889–1940 due to a dearth of intervening surveys²⁹. An introduction of Watersipora arcuata to New Zealand occurred around 1957³⁰ from which time populations spread to become common on the north and south islands³¹. It is thought W. arcuata invaded in southern California, near Los Angeles around 196128. Colonies of a sinusoid species form, referred to generally as W. "subtorquata", were recognized as invading in Australia in the 1970s and in New Zealand and California in the 1980s^{31,32}. On the three landmasses, some regional displacement of W. arcuata was seen as colonies of sinusoid morph entered areas occupied by W. arcuata^{17,31}. Indicating that the rapid propagation of sinusoid Watersipora colonies occurring along California coastline did not originate from simply one introduction, an initial study using COI sequence comparison located a "subtorquata" haplotype occurring in Australia and elsewhere, along with an example of COI sequence that was divergent (in Monterey Bay California), this genetic group being referred to as the "new sp." COI phylogroup²³.

Our first objective was to determine the distribution of the cryptic species noted above in California, and whether intensive sampling of these communities would indicate a likelihood that the number of sources in introductions is greater still than recognized. The genetic (COI) analysis was used secondly to investigate whether *Watersipora* species were non-randomly distributed with respect to sea surface temperature (SST) in the California region and globally. By contrasting multiple invasions of multiple species, our analysis suggests that different temperature-related physiological mechanisms may be important drivers of the invasive distributional patterns of the *Watersipora* lineages.

Results

We generated 361 sequences (Table 1) with median length of 510 nucleotides and an average base content of A, 31.3%; T, 33.3%; C,

18.5%; G, 16.9%. Amino acid translations had no stop codons. Tajima's D statistic³³ was a mean of -0.1795, ranging generally between 0 and 2 (Table 2). Despite slight negativity (as is typical³⁴), few populations showed significant deviation from the simulated D null distribution, with the exception of some southern Californian W. subtorquata population samples, as discussed below.

In California we found divergent COI clades that correspond to previously recognized *Watersipora arcuata*, *W. n.* sp. and *W. subtorquata* phylogroups. Sequences were also included in the Bayesian tree (Figure 1) from colonies identified as *W. edmondsoni* (on coral, n=2, Kane'ohe Bay, Hawaii) and *W. subovoidea* colonies from southern Florida, Brazil, and tropical Australia (Table 1). Monophyletic groups differed by average net divergence of 18.5% (Kimura-2 parameter model); the lowest divergence was observed between *W. subtorquata* and *W. edmondsoni*, 12.3%; greatest divergence was observed between *W. arcuata* and *W. n.* sp. 24.0%). *Watersipora subtorquata* sequences differing by a net divergence of 2.8% formed clades we refer to as *W. subtorquata* clades A and B (Figure 1).

W. subtorquata clade A was abundant in southern California and in central California –Moss Landing, San Francisco and Tomales Bay (Figure 2). Clade B was common in southern California and was established at Humboldt Bay, northern California, where it was found in specimens collected in 2002 along with specimens of *n*. sp. clade. To increase our understanding of clade composition of Humboldt Bay, California we then determined COI phylogroup using a multiplex assay, in which five PCR primers each mismatched at its 3' end to one or two *A*, *B*, or *new* sp sequence populations, generating phylogroup-specific fragment lengths viewed on agarose gels³⁵. *W. n.* sp. was dominant at Humboldt Bay (92.67%), with clade A (3.84%) and clade B (3.49%) being both present at lower frequency (Figure 2).

W. subtorquata clade A has been the most common group sampled at a global scale to date (Figure 3). The most common clade A sequence in California (referred to as haplotype WSI^{23}) is also common in southern Australia, New Zealand, and Europe (Figure 3B). The second most frequent haplotype of clade A (*WS3*) occurs also in southern Australia and South Korea (Figure 1). Clade B variation consisted of one haplotype characterized previously²⁵ (GenBank accession: AY647167) and a second haplotype represented by two colonies collected at Long Beach (near Los Angeles), and a third unique sequence represented by a single colony sampled previously in China³⁴ (collected on seaweed at Qingdao Huiquan Beach, Qingdao, GenBank accession: EU365892) (Figure 3B).

The *W. n.* sp. haplotypes, consisting of one common sequence and several less frequent, related haplotypes (Figure 3C) were common at many California sites. The southernmost finding of *W. n.* sp. was Oxnard, central California, which is within the region defined by the 16° C long term mean SST isotherm (Figure 2). *W. n.* sp. was the one group found at Bodega Bay, Morro Bay, and Bremerton. Clade A and *n.* sp. occurred at similar frequencies in the area of San Francisco and Monterey Bay.

The five COI phylogroups that we recognized occur in statistically distinguishable SST regimes according to a partial Mantel test that included all five groups (P < 0.0001). From lower to higher SST, these were *W. n.* sp. (median=11.8°C, 95% confidence interval 10.9–15.1°C, n=111), *W. subtorquata* clade A (15.1°C, 14.0–17.3°C, n=212), *W. arcuata* (16.3°C, 15.1–25.7°C, n=82), *W. subtorquata* clade B (17.1°C, 11.8–18.2°C, n=29), and *W. subovoidea* (26.0°C, 25.0–26.3°C, n=23). Partial Mantel test comparisons of SST for pairs of phylogroups produced significant *r* coefficients at *P*-values between 0.020 and 0.001 (statistically significant at an unadjusted α value of 0.05); the weakest correlation (r = 0.1008) occurred in comparison of *W. arcuata* and *W. subtorquata* clade A distributions (P = 0.0127).

Population pairwise Φ_{ST} measures were generally higher in the *W*. subtorquata clade A, B and arcuata samples than among *n*. sp.



Table 1 | Specimens analyzed, listed according to phylogroup, and sample location along with estimated mean annual sea surface temperature (SST)

Region	Site or Area name	Coordinates ^A (Lat., Long.)	mean SST (°C)	COI phylogroup	N	Source (if another study)
w. US	Bremerton	47.5798, -122.6321	10.0	n. sp.	12	
w. US	Bodega Bay	38.00 , -123.00	11.3	n. sp.	1	25
w. US	Bodega Bay Harbor	38.3295, -123.0562	11.3	n. sp.	16	
w. US	Humboldt Harbor	40.8074, -124.1635	11.8	В	11	
w. US	Humboldt Harbor	40.8074, -124.1635	11.8	n. sp.	33	
Europe	Plymouth	50.30 , -4.14	13.0	A	1	26
Europe	Guernsey	49.50, -2.58	13.0	A	2	26
Europe	St Jacut	48.60, -2.15	13.0	A	4	20
w. US	Moss Landing Harbor	36.8051, -121.7852	13.0	A	25	
w. US	Moss Landing Harbor	36.8051, -121.7852	13.0	<i>n.</i> sp.	12	25
w. US	Morro Bay	35.37, -120.80	13.3	n. sp.	14	25
W. US	Morro Bdy	35.3707, -120.8585	13.3	n. sp.	14	25
w. US	San Francisco Day	37.91, -122.33	13.5	n. sp.	25	
w. US	San Francisco, Richmond	37.9130, -122.3303	13.5	A D	23	
w. US	San Francisco, Richmond	37.9130, -122.3303	13.5	D	1	
w. US	San Francisco, Richmond	37.9130, -122.3303	13.5	n. sp.	1	
w. US	San Francisco, Oakland	37.0102, -122.3230 37.7845, -122.3230	13.5	A	1	
w. 05 Australia	Hobart	-43.00, 147.28	14.0	^	4	23
Europo	Wellington	-43.00, 147.20 -41.00, 174.78	14.0	A A	2	23
	Tomalos Bay	38 1001 - 122 0220	14.2	A A	17	
	Ventura	34.17 - 110.23	14.5	R	1	25
Australia	Melbourne	-38.00, 144.82	15.1	Δ	7	23
w IIS	Channel Islands Harbor	34 1666 -119 2250	15.1	Δ	13	
w US	Channel Islands Harbor	34166 -1192250	15.1	n sn	1	
w US	Channel Islands Harbor	34 1666 -119 2250	15.1	п. эр. В	3	
w US	Port Hueneme	341532-1192095	15.1	arcuata	3	
w US	Port Hueneme	341532-1192095	15.1	Δ	2	
w US	Port Hueneme	341532-1192095	15.1	n sp	14	
w US	Marina Del Rev	33 9702 -118 4496	15.1	arcuata	8	
w. US	Marina Del Rev	33.9702, -118.4496	15.1	A	3	
w. US	Santa Barbara	34.4067, -119.6890	16.0	arcuata	19	
w. US	Santa Barbara	34.4067, -119.6890	16.0	n. sp.	1	
Australia	Adelaide	-34.50, 138.53	16.3	arcuata	12	23
Australia	Adelaide	-34.50, 138.53	16.3	A	1	23
w. US	Oceanside	33.21, -117.40	17.1	arcuata	2	25
w. US	San Diego, Shelter Island	32.71, -117.23	17.1	А	1	25
w. US	Oceanside	33.2121, -117.3954	17.1	arcuata	1	
w. US	Oceanside	33.2121, -117.3954	17.1	А	12	
w. US	Oceanside	33.2121, -117.3954	17.1	В	3	
w. US	Mission Bay	32.7671, -117.2362	17.1	A	18	
w. US	Mission Bay	32.7671, -117.2362	17.1	В	1	
w. US	Long Beach Harbor	33.7655, -118.2528	17.3	A	17	
w. US	Long Beach Harbor	33.7655, -118.2528	17.3	В	4	
w. US	Huntington Harbor	33.7175, -118.0658	17.3	A	17	
w. US	Dana Point Harbor	33.4591, -117.6992	17.4	A	8	
w. US	Dana Point Harbor	33.4591, -117.6992	17.4	В	5	
w. US	Dana Point, Tijuana Est.	33.4614, -117.7146	17.5	A	3	
w. US	Newport	33.6199, -117.8943	18.2	A	14	
w. US	Newport	33.6199,-117.8943	18.2	В		27
n. Asia	Korea, Namhae Sangju	34./1, 127.99	20.0	A	1	37
n. Asia	Korea, Namhae Sangju	34./1, 12/.99	20.0	n. sp.	17	37
Australia	Sydney	-33.87, 151.21	20.3	arcuata	17	23
Australia	Sydney	-33.87 , 151.21	20.3	A	0	23
Australia	Perth	-31.93,115.83	20.5	arcuata	10	23
		-31.73, 113.83 26.054 120.29	20.3	A D	3	36
	Singaao Elarida	30.034 , 120.38	23.U 25.0	D	1	23
e. US Europo		27.20, -80.22	25.0	supovolded	4	23
		21.00, -137.87 -20.66 116.71	20.7	arcuaid	12	23
Australia	Pio Do Janoiro	-20.00, 110./1	20.0	supovolded	4	
Australia		-23.01, -43.43 -16.88 145.80	20.0 24 5	subovoidea	14	23
	Cuirlis	-10.00, 143.00	20.5	supovolueu	1	
^A Coordinates given	to four decimal places refer to sampling si	te from present study. Coordinates to two pl	aces refer to region	al-level locality information	derived from other sou	irces.

Table 2 Summ	ary of p	oludor	ıtion div	/ersity an	d pairwise (Φ_{ST} measure	is for Water:	sipora arcuat	a, subtorquatc	а, and 'new	sp'. COI seq	Jences (lengt	h: 489 base p	airs)	
		Dive	ersity in	ndices					Φ_{ST}						
								A. Watersipa	ora arcuata						
		Ľ	× S	s /n	6	L	D		Santa Barbara						
Santa Barbara Marina Del Rae		ح 19	∞ 4 	8 0.947 4 2.000	74 0.0168≟ 00 0.0123≟	±0.0094 ±0.0079 -	0.5843 5 -1.3759 A	santa Barb. Asr. Del R. (
							æ	Watersipora	ı subtorquatc						
	n K	S	S/n		я	D	Aust-NZ	Tomales Bay	San Fran. Bay	Elkhorn SI.	Oxnard	Long B. H	unt. Harb. Danc	a Point Newpo	ort Ocean.
Australia-New Zealand ^A	23 2	-	0.0435	0.0002	- +0.0004	-1.1610	I								
Tomales Bay San Francisco	17 5 24 7	4 M	0.2353 0.2917	8 0.0018 0.0037	3±0.0015 - 7±0.0024 -	-0.7400 -0.1351	0.2039*** 0.3967***	_ 0.0785*	I						
Harbor Elkhorn SI. Oxnard Long Beach, LA Huntington	19 7 18 5 20 7 15 5	26 7 32 4	0.3684 1.4444 1.6000 0.4000	1 0.0025 0.0161 0.0175	5±0.0019 1±0.0088 5±0.0094 1±0.0025	-0.9214 0.1633 -0.2038 -0.1401	0.4535*** 0.1752** 0.1576*** 0.2290**	0.0475 0.1082 0.1030* 0.1713*	~ 0 0.1186* 0.1208** 0.1460**	_ 0.1041* 0.1045* 0.1649*	~ 0 0.0907	- 0.0714	I		
Harbor Dana Point Newport Oceanside Mission Bay	13 5 14 4 18 5 18 5	24 24 25	1.9231 1.7900 1.6000 1.3889	0.0255 0.0082 0.0171 0.0070	5±0.0139 2±0.0049 ±0.0094 1±0.0094	2.3880** -2.0674** 0.5516 -2.0950**	0.7302*** 0.0431* 0.2525** 0.0726**	0.5128** 0.0369** 0.1510* 0.0601*	0.5996** 0.0942** 0.1738* 0.0586*	0.5200** 0.0848* 0.1523* 0.0799**	0.0508 ~ 0 0.0035	~ 0.0712 ~ 0 ~ 0 0.0131	0.4669* 0.0262 0.2 0.1422* 0.0 0.003 0.2	- 158* - 103 0.005 916* ~0	1 – 0.0316
								C. Watersipc	ora new sp.						
		۲	×	S	S/n	μ		D	Bremerton	Humb. (2003)	Humb. (2006)	Bodega Bay	 Elkhorn SI. 	Morro I	ay
Bremerton (2010 Humboldt Bay (2(Humboldt Bay Bodega Bay Elkhorn Slough Morro Bay Port Hueneme	с)03) ^в	3110670102		00-0000	0 0.0588 0.0588 0 0.1818 0.1538	0 0 0.0002± 0 0.0010± 0.00010±	0.0004 0.0011 0.0008	0 0 0.0058 0 0 -0.7782 -1.4680	~ 0 - 2 0 - 2 0 - 2 0 - 2 0 - 2 0 0 - 2 0 0 0 0	~ 0 _ ~ 0 ~ 0 0.1264*	0.0166 ~	~ 0 ⁻ 0.1905*	* 0.1318**	- 000	ۍ د
$\begin{array}{l} \label{eq:period} \mbox{Permutation tests of D c} \\ \mbox{``P < 0.05.} \\ \mbox{``P < 0.001.} \\ \mbox{``P < 0.0001.} \\ ``P < $	nd Φ_{ST} wer es were cc Ref.23. 1 in Smiths	re condu illected ir onian (Ei Jensen, U	cted with 5 n Californic nvironmenty Jniversity c	5,000 replicat a Department tal Research C of Washingtor	les. tof Fish and Gam Center] surveys. n, 2010.	te (Introduced Spe	scies Surveys), 20	80							





Figure 1 Bayesian tree of COI sequences. Posterior probabilities (> 0.5) are shown at nodes. Squares at branch tips, within four phylogroups – *Watersipora arcuata, W. subtorquata* clades A and B, and *W. n.* sp. – indicate an introduced haplotype on the Californian coast. Haplotypes found in other areas are indicated: Washington state (Pacific US); AU, Australia; Fl, Florida (Atlantic, US); BR, Brazil; Ha, Hawaii (O'ahu); NZ, New Zealand (Wellington); UK (southern England and Channel Islands); FR, France. *Watersipora* colonies have uniform zooids. Typical dimensions were determined from multiple colonies in different COI groups, providing a stereotyped zooid appearance (see Figure 4).

populations (Table 2). Given the fact that the *W*. *n*. sp. had relatively low nucleotide diversity, COI is likely insufficiently sensitive for detecting post-introduction genetic isolation if present. However, there was significant differentiation (Φ_{ST}) occurring between the Morro Bay sample and other populations (pairwise comparisons of P < 0.05), but no other significant differentiation.

Ten *Watersipora arcuata* haplotypes were found in California (Figure 3A). Two (previously designated *h1* and *h8*, Mackie et al. 2006) were identical to haplotypes previously found in Australia, and three (*h5*, *h6*, and *h7*) identical to haplotypes known from both Australia and O'ahu, Hawaii. The COI variation of *W. arcuata* showed structuring at the scale of collection sites (Φ_{ST} , = 0.4459, *P* < 0.0001) and collection areas (i.e., within O'ahu, California, Perth or Adelaide; $\Phi_{SC} = 0.6285$, *P* < 0.0001). There was, however, no differentiation observed in comparing the regional sampling areas of O'ahu, California, Perth or Adelaide. In fact, the inter-area variance component was negative ($\Phi_{CT} = -0.4912$), indicating a spatially-dispersed but locally-structured pattern.

Spatial differentiation was evident within the *W. subtorquata* COI clade complex found on the California coastline. An AMOVA supported differentiation of northern and southern populations (from Santa Barbara southward) ($\Phi_{\rm CT} = 0.1068$, P = 0.0215); these populations were also structured in COI nucleotide variation at local scale ($\Phi_{\rm ST} = 0.1625$, P < 0.0001, $\Phi_{\rm SC} = 0.06229$, P = 0.0401). It is noted that southern Californian populations have *D* statistic deviations that may be due to the sampling of relatively long branches (clade A and B). Newport and Mission Bay (sites 14 and 16, Figure 2) had *D* statistic³³ values of < 2.0, which were statistically significant ($P_{D \text{ obs } > D \text{ exp}} < 0.05$); this appears to reflect unbalanced frequencies of haplotypes separated on divergent branches. In contrast, Dana Point (D = +2.388, $P_{D \text{ obs } > D \text{ exp}} = 0.9989$), and Oceanside (D = 0.5516, *ns*) had positive *D* measures, reflecting the more even spread of clades A and B in these samples.

The zooid area-to-operculum area ratio (Figure 4A) distinguished *Watersipora subovoidea* colonies as a homogeneous grouping from other sinusoidal colonies consisting of *subtorquata* and *n*. sp. COI clade colonies (ANCOVA: $F_{1.97} = 59.83$, P < 0.0001). An ANCOVA, comparing morphometric ratio slopes did not support a difference between *W. subtorquata* (clade A or B) and *W. n.* sp. populations ($F_{1.70} = 0.18$, P = 0.6727). In considering the invasive







Figure 3 | Parsimony networks describing relationships of COI haplotypes of *Watersipora*, introduced to California and other areas. Continuous straight lines represent connections with >95% confidence⁶⁸ and dashed lines represent connections with confidence below this limit. Circles indicating sampled haplotypes are scaled according to haplotype frequency. The frequency of the most common haplotype in each set is shown. *W. arcuata* haplotypes, h1–h8, were defined previously²³.

subtorquata–n sp. complex in California, there was no relationship between the COI-clade identity and zooid length (data not shown), however, zooid length scaled with the mean SST temperature in the

invaded area (Figure 4B). Deviating from the general trend, zooid lengths of *n*. sp. colonies collected from Bremerton, Washington (mean SST ~ 10°C, collected in 2010), and *W. subtorquata* clade A colonies from Humboldt Bay (mean SST ~ 10°C, collected in 2003) were short, for reasons unknown. Analysis of the general sample of sinusoidal *Watersipora* colony populations in California supported a negative relationship between temperature and zooid length ($R^2 = 0.245$, P < 0.001) (Figure 4B).

Discussion

On the basis of phylogenetic inference using the COI locus, the previously described Watersipora subtorquata-complex represents two cryptic species, W. subtorquata and W. n. sp., consistent with previous reports²³. According to the present sampling, W. subtorquata can be further divided into two genetically shallow groups clade A, which has haplotypes recognized in Europe and Australasia, and clade B, found to be common in southern California and present at much lower frequency in northern California. Other investigations show that these three known COI clades of the W. subtorquatacomplex also occur in the Asian western Pacific^{36,37}. The absence of reports of sinusoidal Watersipora in Californian waters prior to the 1980s probably indicates the true absence of such forms given the intensity of study there³². Since then, according to COI data, there have been multiple introductions from separate sources. The introduced Watersipora of the Californian region is more diverse than that of Australasia, and COI variation apparent in California reflects either different thermal tolerances of the W. n. sp. and W. subtorquata complex source populations, different sets of introductions to these areas, or both.

Watersipora species as a whole were absent from the US Pacific coastline until the 1960s and absent from the fossil record in that region²⁸. *Watersipora arcuata* was the first watersiporid to be recognized on the coastline, appearing in southern California around 1963²⁸. Soule and Soule³² then reported a sinusoid species as invading in Los Angeles Harbors and marinas in 1982–3, a period of unusually warm water due to El Niño. We have no reason to suspect that these specimens were not the sinusoidal *Watersipora subtorquata* complex described here. Examination of collections of the late Dorothy and John Soule at the Santa Barbara Natural History Museum unfortunately did not locate the material referred to in their report (Mackie, pers. obs., July, 2011). Banta²² suggested *W. arcuata* was native to the tropics or subtropics of the eastern Pacific. Evidence of the native source of *Watersipora subtorquata* clade A and clade B or *W. n.* sp. is lacking, as is a precise timing of arrival in California.

Early collections of colonies that best fit the morphological description of W. arcuata were made in the Galapagos Islands and the Pacific Mexican coast, including Baja California and Gulf of California, prior to the 1930s^{28,38}. By 1940, W. arcuata was a common fouling species in the Gulf of California^{39,40} but was not yet known on the US Pacific coastline. Two hypotheses have been proposed to explain the appearance of W. arcuata in California in the 1960s. Soule and Soule³² proposed that larval dispersal or dispersal of colonies on drift material from warmer Pacific areas into California could occur during El Niño events (such as 1956-57), when currents from southern areas extend further north than usual. Banta and Carlton each, however, favored a scenario of introduction of W. arcuata to California from Australasia through ship fouling, in part based on negative evidence, specifically the lack of W. arcuata in extant fouling community records or the fossil record of California²⁸. Positive circumstantial evidence includes the observation that Australasia has been a donor of a number of marine invasive species to the US Pacific region⁴¹.

Given the lack of segregation of COI genetic variation among widespread areas – Australia, California, and Hawaii – the arrival of *W. arcuata* colonizers through shipping from common sources is supported, as opposed to regionally independent introductions.



Figure 4 | (A) Averaged orifice area versus zooid area (log₁₀ versus log₁₀) for populations distinguishable as four COI clades. As shown previously²⁶ the plotted relationship discriminates the recognized species *W. subtorquata* and *W. subovoidea*, however as evident here, it does not discriminate the *subtorquata* (A or B) and *n*. sp. phylogroups. Inset: cartoon of zooids of averaged proportions, by COI phylogroup. (B) Plot showing means and SE (vertical bars) of zooid length at different sites on US west Coast, indicating a decrease in length toward warmer localities. Trend lines were calculated by least-squares regression.

Defining the routes of introductions around the globe based on COI is not possible though given the observed distribution of genetic variation at that locus. Inferring direction and chronology of these invasions genetically will likely rely on the use of multiple loci providing finer spatial resolution to the distribution of genetic variation.

Watersipora species ranges have undergone remarkable shifts with human assistance. Given such widespread global introductions and realizing that ranges are rapidly dynamic, source populations are perhaps traceable now only by searching for remnant phylogeographic patterns.

Collection records of *W. arcuata* and *W. subtorquata* indicate that species range boundaries have changed rapidly. In the 1980s *W. subtorquata* replaced *W. arcuata* in southeastern Australia and New Zealand, specifically in cooler areas of these landmasses^{17,31}. In the areas of *subtorquata* introduction, *W. arcuata* had been present since its introduction sometime between the late 1800s and 1940 in Australia²⁹ and around 1957 in New Zealand³⁰. These taxonomic records indicate that species of Watersipora may compete with one another for resources (with space presumably being one of the most important) in the human-modified fouling niche. Further population separation on the basis of COI supports widespread species-replacement interactions.

In the present study we confirmed that *W. subovoidea* (matching northern Australian and Brazilian populations morphologically) also occurs in Florida, US. COI sequences show $\leq 1\%$ divergence among *W. subovoidea* in Brazil, Florida and Australian populations,

supporting recent and widespread introductions of a species suited to tropical conditions. While genetic analysis of the type specimens or neotypes has not been conducted, a W. subtorquata-W. subovoidea delineation is supported on the basis of both morphometric and COI comparison²⁶. The W. subtorquata holotype (Gunabara Bay, Brazil) was described by d'Orbigny from material collected in 1837. Given this historical record, it was surprising that recent collections in the Rio De Janeiro and Sao Paulo regions have revealed only W. subovoidea. Similarly Ramalho et al.42 reported collections of Watersipora matching W. subovoidea morphologically at multiple localities spanning a wide range of surfaces and pollutant levels. Thus, there is reason to suspect W. subovoidea has displaced W. subtorquata in its native locale or at least an area where it was common in the 1800s, and there is clearly a need to relate collections from different points to the diversity patterns now suggested by genetics.

Prior studies of Bryozoa have suggested genetically divergent species are recognizable by morphological divergence^{43,44}, but our analysis suggests that this conclusion is not universal. Measurements of subsamples of *W. subtorquata* clades A and B and *W. n.* sp. in California showed these to be homogeneous in their zooid geometry (Figure 4A) and they were not definitively sorted by color (unpublished data). Colonies of *W. n.* sp. varied from flat, encrusting forms to large (30 cm) multi-colony ball-shaped forms. Although diagnostic morphological criteria may yet be discovered, currently none seem present that would be practical for rapid identification, and genetic analysis will be necessary for continued work on *Watersipora*.

The lack of morphological distinctiveness of *W. subtorquata* and *W. n.* sp. COI clades reflects an absence of obvious skeletal characters (a situation that is exacerbated by the relatively featureless zooid morphology of the genus), or wholesale hybridization leading to a continuous range of morphologies. Hybridization is a factor affecting identification and ecological responses in introductions of the *Mytilus* species complex, for example⁹. No introgression of mitochondrial DNA has been found in colonies of the *W. arcuata* and *W. subtorquata* morphologies (arcuate and sinusoidal orifice form respectively), which supports the standpoint that *Watersipora* COI clade groups are not randomly interbreeding. Assessment of genetic variation at microsatellite and other nuclear DNA loci is being used to test for admixture among COI clade populations.

Zooids were as much as 25% longer in northern California as compared to southern Californian colonies. The phylogroup itself showed no consistent relationship to zooid length, with rather, colony populations of multiple groups exhibiting latitude related zooid length (Figure 4). Size-latitude trends arguably still require extended documentation in invertebrates generally; in mollusks a recent metastudy has however indicated size trends to be common, with the direction of the relationship being variable⁴⁵. The direction of the zooid size to temperature relationship seen across these recently introduced Watersipora populations is consistent with studies of other Watersipora spanning the Galapagos archipelago³² and in other bryozoans where there is an inverse relationship between zooid size and temperature⁴⁶⁻⁴⁸. Phenotypic plasticity is a possible explanation (as seen in one study of a limpet in which shell variation - larger size in cold - was explainable by water temperature rather than genetic variance⁴⁹). Drosophila wing-traits⁵⁰ on the other hand provide an example of post-introduction variation responding in a clinal selective gradient. There are sharp differences in mean size are recognized following introductions in a number of marine metazoans⁵¹. Perhaps promisingly, Wateripora provide a useful system in which to assess the heritable/plastic components of zooid size, along with net overall growth, and reproductive characteristics, determining whether these variously influence or respond to observed range expansions.

The boundary presented in part by the cold California and warmer Davis current systems allows the California coast to be used as a sensitive test of the role of temperature in differentiating introduction processes. Examination of COI variation occurring in Watersipora revealed significant north-south separation of haplotype frequencies in California. This separation coincides with Point Conception, an area with a high turnover of ranges and phylogeographic breaks in native taxa⁵². All W. arcuata occurrence was to the south of Point Conception (and *n*. sp., conversely, has not been found far south of Point Conception). The range of W. arcuata however did expand briefly in a northward direction in 1982 and 1983, an El Niño period, such that the species was found in Monterey Bay, northern California³² where it has not been reported subsequently. This particular observation is notable for indicating the likely sensitivity of the ranges to temperature. Our study, and others examining genotypic variance (e.g^{10,53-55}), suggest genetically related invading propagules have temperature related fitness which determines organismal or genotype-level range limits at least in early stages of introduction. Average sea-surface temperatures of 14°C-20°C unite the southern Californian and some Australian localities, where W. arcuata and W. subtorquata were found together. The situation is analogous to invasions of two monophyletic Caulerpa (Chlorophyceae) groups^{56,57}, which also appear to be established in southern Australia, Mediterranean areas and southern Californian regions, but not northern California.

While the COI phylogroup-SST correlation is derived from relatively few global locales, it is possible to define widespread introductions of the five *Watersipora* COI groups by different temperature-zone envelopes, an indication that intrinsic differences in temperature-related fitness structure patterns of spread. With this background information, direct hypothesis testing can be used to determine whether phenotypic differences as opposed to vector transport patterns alone dictate introduction success or local densities of *Watersipora* species. The COI variation of California (which is greater than that observed in Australia) and haplotype distribution pattern provides a useful framework for common-environment experiments to test for physiological restrictions to ranges. The association of different species and COI clades of *Watersipora* with particular temperature zones suggests a global assortment of lineages into similar temperature zones, in other words, natural selection acting on existing variation in parallel in different areas.

While the invasion success of *Watersipora* populations is geographically limited by evolved ecological tolerances, the species in the genus as a whole have cumulatively extremely broad potential for global spread, with a collection of traits (including a high tolerance of copper-based antifouling paint that is apparent in larvae and colonies) which assists in colonization of painted hulls on ships that may further transport colonies. Perhaps because growth rate and reproductive potential are intrinsically connected in modular organisms, temperature modulated growth rate (temperature-related fitness) may prove more useful than other general hypotheses commonly put forward to explain the high invasive capacity of certain introduced species, such as the ability to escape specialist predators and pathogens⁵⁸⁻⁶⁰ or propagule pressure^{61,62}. Clearly, modular fouling organisms warrant attention as a group of organisms sensitively indicating community changes in response to environmental change.

Methods

Collections. Colonies were collected between 2005 and 2010 from docks and floats, predominantly, throughout California, and additional specimens were obtained from fouling panels in San Francisco Bay, and Humboldt Harbor (collections made in the period of 2002–2005), and field collections from Washington State (Bremerton), Florida, and Brazil (Tables 1 and 2). We generated COI sequence for 361 colonies. Sequences and collection information were lodged in GenBank (accession numbers: JQ715456–JQ715577). We included previously reported sequences^{23,25,26,36,37} in analyses.

PCR and sequencing. Colonies were preserved in 85–95% ethanol. Fragments (which were generally <2 cm across) were sorted into individual colonies with independent ancestrula. DNA was extracted by Qiagen DNeasy Tissue protocol. DNA for sequencing was obtained by amplification of 710 base pairs using LCO1490 and HCO2198 primers⁶³, followed by re-amplification of this product in a second PCR with LCO1490 and a bryozoan-specific primer BRY-HCOI-2161, effectively increasing product yield²³. PCR was carried out using GoTAQ[®] DNA polymerase and 2x Buffer, with 3.0 mM Mg²⁺ ion, at an annealing temperature of 40°C. Products were isolated using Qiagen Quickspin[®] columns and sequenced in both directions by BigDye[®] di-deoxy terminators.

Sequence analysis. Sequence chromatograms were read using Codon Code Aligner [®] software, aligned using MEGA4⁶⁴, and collated to haplotypes using DNA Collapser V. 1 (http://www.birc.au.dk/fabox). A Bayesian analysis (using a flat prior distribution and a General Time Reversible model of nucleotide substitution including a Gamma-distribution substitution rate parameter) was used to construct a tree. Nucleotide substitution model parameters were determined using ModelTest⁶⁵, and the tree constructed using MyBayes⁶⁶. Posterior probabilities at nodes were calculated using three parallel Metropolis Coupled Markov Chains, searching for 2 million generations. The Bayesian analysis produced a robust topology with posterior support for major clades of 0.94 or higher, at which level there was topological agreement with a parsimony tree found by heuristic search in PAUP*⁶⁷ (data not shown).

Haplotype relationships within *Watersipora arcuata, W. subtorquata* and the *W. n.* sp. clades were evaluated using median joining parsimony networks⁶⁸. Sequence lengths used in comparisons were determined by the minimum lengths of sequence data available in GenBank: a 388-nucleotide segment of *W. arcuata* COI sequences was compared, including samples from southern Australia (Perth, Adelaide, Sydney areas) and O'ahu, Hawaii²³ *Watersipora subtorquata* and *W. n.* sp. clade networks were constructed using a 489-nucleotide segment. AMOVA (Analysis of Molecular Variance⁶⁹) was used to quantify COI sequence variation partitioned among broad geographic regions for both *W. arcuata* (regions included Hawaii, California, and Australia) and *W. subtorquata* (including the Australasian region of southern

Australian and New Zealand and two California regions north and south of Point Conception). Permutation tests (5000 replicates) were used to evaluate AMOVA coefficient significance⁷⁰.

Mean sea surface temperature (SST) approximation. Local average SST was approximated using year-long measurements spanning 2002–2011. For most US sites, SST was obtained from a monitoring buoy located within 50-km of the sampling location, via NOAA National Oceanographic Data Center Coastal temperature tables (http://www.nodc.noaa.gov/dsdt/cvtg/cpac.html). Measurements were also obtained using the NASA satellite (Aqua) Moderate Resolution Imaging Spectroradiometer (MODIS) thermal map data archive²¹. Temperatures were resolved to 1°C unit of accuracy using the dominant pixel-record within 50×50 km squares positioned offshore to sampling areas. We verified, as elsewhere²², that MODIS and buoy-recorded mean SSTs were generally within 1°C.

Analysis of SST data. The 95% confidence interval of the median sea surface temperature experienced by major COI clades was estimated by bootstrapping (resampling populations of 20 individuals for 1000 replicates). To test the null expectation of no correlation between temperature and clade, Mantel tests were conducted correcting for spatial distance using a partial matrix⁷³. Pairwise SSTs and decimal grid coordinates were converted to Euclidean distances. Clades were encoded as presence or absence, and separate tests were run for all five clade groups and pairs of clades separately. Partial Mantel tests were conducted using the *R* Software Project package, ecodist⁷⁴. Data were ranked, which assists in linearizing relationships between dissimilarity matrices⁷⁵. The significance of the partial coefficient was determined using 10,000 matrix permutations.

Zooid-dimension comparisons for COI clades of sinusoidal Watersipora. The recognized species Watersipora subovoidea and the W. subtorquata-complex can be distinguished by zooid proportions: for a given frontal shield area, the tentacular orifice is smaller in W. subovoidea²⁶. In the current study, a subset of the sinusoid colonies analyzed by COI were photographed at 20X magnification using a dissecting scope, and analyzed using Image J imaging software⁷⁶. We recorded five zooid dimensions: zooid length (L_z), zooid width at maximum (W_z), orifice width (W_{or}), orifice length (L_{or}). We tested for a difference between W. n. sp. and other W. subtorquata-complex COI phylogroups, using zooid area and tentacular orifice area (as in²⁶) as covariates via ANCOVA. Log₁₀ transformations of areas were used, and regressions met the assumption of homogeneity of variances.

- 1. Knowlton, N. Sibling species in the sea. Annu. Rev. Ecol. Syst. 24, 189-216 (1993).
- 2. Carlton, J. T. Biological invasions of cryptic species. *Ecology* 77, 1653–1655 (1996).
- 3. Bickford, D. *et al.* Cryptic species as a window on diversity and conservation. *Trends Ecol. Evol.* **22**, 148–155 (2006).
- 4. Geller, J. B., Darling, J. A. & Mackie, J. A. Genetic perspectives on marine biological invasions. *Annu. Rev. Mar. Sci.* **2**, 367–393 (2010).
- Carlton, J. T. Pattern, process and prediction in marine invasion ecology. *Biol. Conserv.* 78, 97–106 (1996).
- Ruiz, G. M., Fofonoff, P. W., Carlton, J. T., Wonham, M. J. & Hines, A. H. Invasion of coastal marine communities in North America: Apparent Patterns, Processes, and Biases. *Annu. Rev. Ecol. Syst.* 31, 481–531 (2000).
- Wonham, M. J. & Carlton, J. T. Trends in marine biological invasions at local and regional scales: the Northeast Pacific Ocean as a model system. *Biol. Invasions* 7, 369–392 (2005).
- Wonham, M. J. Mini-review: Distribution of the Mediterranean mussel Mytilus galloprovincialis (Bivalvia : Mytilidae) and hybrids in the Northeast Pacific. J. Shellfish Res. 23, 535–543 (2004).
- Braby, C. E. & Somero, G. N. Ecological gradients and relative abundance of native (*Mytilus trossolus*) and invasive (*Mytilus galloprovoncialis*) blue mussels in the California hybrid zone. *Mar. Biol.* 148, 1249–1262 (2006).
- Geller, J. B., Sotka, E. E., Kado, R., Palumbi, S. R. & Schwindt, E. Sources of invasions of a northeastern Pacific acorn barnacle, *Balanus glandula*, in Japan and Argentina. *Mar. Ecol. Prog. Ser.* 358, 211–218 (2008).
- Watts, P. C., Thorpe, J. P. & Taylor, P. D. Natural and anthropogenic dispersal mechanisms in the marine environment: a study using cheilostome Bryozoa. *Philos. Trans. R. Soc. Lond. B* 353, 453–464 (1998).
- Barnes, D. Biodiversity: Invasions by marine life on plastic debris. Nature 416, 808–809 (2002).
- Carlton, J. T. & Geller, J. B. Ecological roulette: the global transport of nonindigenous marine organisms. *Science* 261, 78–82 (1993).
- Carlton, J. T. & Hodder, J. Biogeography and dispersal of coastal marine organisms: experimental studies on a replica of a 16th-century sailing vessel. *Mar. Biol.* 121, 721–730 (1995).
- Lynch, W. F. The behavior and metamorphosis of the larva of *Bugula neritina* (Linnaeus): experimental modification of the length of the free-swimming period and the responses of the larvae to light and gravity. *Biol. Bull.* **92**, 115–150 (1947).
- Wisely, B. The settling and some experimental reactions of a bryozoan larva Watersipora cucullata (Busk). Aust. J. Mar. Freshw. Res. 9, 362–371 (1958).
- Keough, M. J. & Ross, J. in *Marine Biological Invasions of Port Phillip Bay Victoria* (eds Hewitt, C. L., M. L. Campbell, R. E. Thresher, & R. B. Martin) 9–11 (CSIRO Marine Research, 1999).

- Rodriguez, L. F. & Ibarra-Obando, S. E. Cover and colonizarion of commercial oyster (*Crassostrea gigas*) shells by fouling organisms in San Quintin Bay, Mexico. *J. Shellfish Res.* 77, 337–343 (2008).
- Cohen, A. N. & Carlton, J. T. Nonindigenous aquatic species in a United States estuary: a case study of the biological invasions of the San Francisco bay and delta. (Aquatic Nuisance Species Taskforce, 1995).
- Dafforn, K. A., Glasby, T. M. & Johnston, E. L. Differential effects of tributyltin and copper antifoulants on recruitment of non-indigenous species. *Biofouling* 24, 23–33 (2008).
- Sorte, C. J. B., Williams, S. L. & Zerebecki, R. A. Ocean warming increases threat of invasive species in a marine fouling community. *Ecology* 91, 2198–2204 (2010).
- Banta, W. C. Watersipora arcuata, a new species in the subovoidea-cucullata-nigra complex (Bryozoa, Cheilstomata). Bull. South. Calif. Acad. Sci. 68, 96–102 (1969).
- Mackie, J. A., Keough, M. J. & Christidis, L. Invasion patterns inferred from cytochrome oxidase I sequences in three bryozoans, *Bugula neritina, Watersipora* subtorquata, and Watersipora arcuata. Mar. Biol. 149, 285–295 (2006).
- Gordon, D. P. The marine fauna of New Zealand: Bryozoa: Gymnolaemata (Cheilostomatida Ascophorina) from the western South Island continental shelf and slope. *Memoirs of the New Zealand Oceanographic Institute* 97, 1–158 (1989).
- Anderson, C. M. & Haygood, M. G. a-Proteobacterial symbionts of marine bryozoans in the genus *Watersipora*. *Appl. Environ. Microbiol.* **73**, 303–311 (2007).
- Ryland, J. S., De Blauwe, H., Lord, R. & Mackie, J. A. Recent discoveries of alien Watersipora (Bryozoa) in Western Europe, with redescriptions of species. Zootaxa 2093, 43–59 (2009).
- Soule, D. F. & Soule, J. D. in *Proceedings of the third Conference, International Bryozoology Association (1973)* Vol. 3 (ed S. Pouyet) 299–309 (Documents des Laboratoires de Gèologie de la Faculté des Sciences de Lyon, 1976).
- Banta, W. C. The recent introduction of *Watersipora arcuata* Banta (Bryozoa, Cheilostomata) as a fouling pest in Southern California. *Bull. South. Calif. Acad. Sci.* 68, 248–251 (1969).
- 29. Allen, F. E. Distribution of marine invertebrates by ships. Aust. J. Mar. Freshw. Res. 4, 303–316 (1953).
- Skerman, T. M. The recent establishment of the Polyzoan Watersipora cucullata (Busk) in Auckland Harbour, New Zealand. N. Z. J. Sci. 3, 615–619 (1960).
- Gordon, D. P. & Mawatari, S. H. Atlas of marine-fouling bryozoa of New Zealand ports and harbours. *Miscellaneous Publications N.Z. Oceanographic Institute* 107, 1–52 (1992).
- Soule, D. F. & Soule, J. D. in *Bryozoa: Ordovician to Recent* (eds C. Nielson & G. P. Larwood) 293–300 (Olsen & Olsen, 1985).
- Tajima, F. Statistical method for testing the neutral mutation hypothesis by DNA polymorphism. *Genetics* 123, 585–595 (1989).
- Wares, J. L. Natural distributions of mitochondrial sequence diversity support new null hypotheses. *Evolution* 64, 1136–1142 (2010).
- Láruson, Á. J., Craig, S. F., Messer, K. J. & Mackie, J. A. Rapid and reliable inference of mitochondrial phylogroups of *Watersipora*, an invasive group of ship-fouling species (Bryozoa, Cheilostomata). *Conserv. Genet. Res.* 4, 617–619 (2012).
- Sun, M. *et al.* The complete mitochondrial genome of *Watersipora subtorquata* (Bryozoa, Gymnolaemata, Ctenostomata) with phylogenetic consideration of Bryozoa. *Gene* 439(1–2), 17–24 (2009).
- Lee, H.-J. et al. DNA Barcode Examination of Bryozoa (Class: Gymnolaemata) in Korean Seawater. Korean J. Syst. Zool. 27, 159–163 (2011).
- Hastings, A. B. Cheilostomatous polyzoa from the vicinity of the Panama Canal collected by Dr. C. Crossland on the cruise of the S. Y. "St. George". *Proc. Zool. Soc. London* 1929, 687–740 (1930).
- Soule, J. D. Results of the Puritan-American Museum of Natural History expedition to western Mexico. Am. Mus. Novit. 2053, 1–66 (1961).
- Osburn, R. C. Bryozoa of the Pacific Coast of America. Part 2, Cheilostomata--Ascophora. Allan Hancock Pac. Expedit. 14(2), 271–612 (1952).
- Carlton, J. T. Patterns of transoceanic marine biological invasions in the Pacific Ocean. Bull. Mar. Sci. 41, 452–465 (1987).
- Ramalho, V. L., Muricy, G. & Taylor, P. D. Taxonomic revision of some leprailiomorph cheilostome bryozoans (Bryozoa: Lepraliomorpha) from Rio de Janeiro State, Brazil. J. of Nat. Hist. 45, 767–798 (2011).
- Jackson, J. B. C. & Cheetham, A. H. Evolutionary significance of morphospecies: a test with the Cheilostome bryozoa. *Science* 248, 579–583 (1990).
- Mackie, J. A., Keough, M. J., Norman, J. A. & Christidis, L. in *Bryozoan studies* 2001, *Proceedings of the Twelfth International Bryozoology Association Conference* (eds P. N. Wyse Jackson, C. J. Buttler, & M. E. Spencer-Jones) 199–206 (Balkema, 2002).
- 45. Berke, S. K., Jablonski, D., Krug, A. Z., Kaustuv, R. & Tomasovych, A. Beyond Bergmann's rule: size-latitude relationships in marine Bivalvia world-wide. *Global Ecol. Biogeogr.* 10.1111/j.1466-8238.2012.00775.x (2012, in press).
- O'Dea, A. Zooid size parallels contemporaneous oxygen isotopes in a large colony of *Pentapora foliacea* (Bryozoa). *Mar. Biol.* 146, 1075–1081 (2005).
- Atkinson, D., Morley, S. A. & Hughes, R. N. From cells to colonies: at what levels of body organization does the 'temperature-size' rule apply. *Evol. Dev.* 8, 202–214 (2006).
- Amuivedel, A., Hayward, P. & Porter, J. Zooid size and growth rate of the bryozoan *Cryptosula pallasiana* Moll in relation to temperature, in culture and in its natural environment. J. Exp. Mar. Biol. Ecol. 353, 12 (2007).

- Teske, P. R., Barker, N. P. & MacQuaid, C. D. Lack of genetic differentiation among four sympatric southeast African intertidal limpets (Siphonariidae): phenotypic plasticity in a single species? J. Molluscan Stud. 73, 223–228.
- doi:210.1093/mollus/eym1012 (2007).
 50. Gilchrist, G. W., Huey, R. B. & Serra, L. Rapid evolution of wing size clines in Drosophila subobscura. Genetica 112-113, 273-286 (2001).
- Grosholz, E. D. & Ruiz, G. M. Biological invasions drive size increases in marine and estuarine invertebrates. *Ecol. Lett.* 6, 700–705 (2003).
- Dawson, M. N. Phylogeography in coastal marine animals: a solution from California? J. Biogeogr. 28, 723–736 (2001).
- Zardi, G. I., McQuaid, C. D., Teske, P. R. & Barker, N. P. Unexpected genetic structure of mussel populations in South Africa: indigenous *Perna perna* and invasive *Mytilus galloprovincialis. Mar. Ecol. Prog. Ser.* 337, 135–144, doi:10.3354/ meps337135 (2007).
- 54. Asif, J. H. & Krug, P. J. Lineage distribution and barriers to gene flow among populations of the globally invasive marine mussel Musculista senhousia. *Biol. Invasions* 14, 1431–1444, doi:DOI 10.1007/s10530-011-0169-6 (2012).
- Roman, J. Diluting the founder effect: cryptic invasions expand a marine invader's range. Proc. R. Soc. B. 273, 2453–2459, doi:DOI 10.1098/rspb.2006.3597 (2006).
- 56. Klein, J. & Verlaque, M. The Caulerpa racemosa invasion: A critical review. Mar. Pollut. Bull. 56, 205–225 (2008).
- 57. Jousson, O. *et al.* Invasive alga reaches California. *Nature* **408**, 157–158 (2000).
- Keane, R. M. & Crawley, M. J. Exotic plant invasions and the enemy release hypothesis. *Trends Ecol. Evol.* 17, 164–170 (2002).
- Torchin, M. E., Lafferty, K. D., Dobson, A. P., McKenzie, V. J. & Kuris, A. M. Introduced species and their missing parasites. *Nature* 421, 628–630 (2003).
 Callaway, R. M. & Ridenour, W. M. Novel weapons, investive success and the
- 60. Callaway, R. M. & Ridenour, W. M. Novel weapons: invasive success and the evolution of increased competitive ability. *Front. Ecol. Environ.* 2, 419–426 (2004).
 61. Verling, E. et al. Supply side invasion scalars, characterizing memory.
- Verling, E. *et al.* Supply-side invasion ecology: characterizing propagule pressure in coastal ecosystems. *Proc. R. Soc. B.* 272, 1249–1256 (2005).
 Sinch ach ff, D. The rule of rule.
- Simberloff, D. The role of propagule pressure in biological invasions. Ann. Rev. Ecol. Evol. Syst 40, 81–102 (2009).
- Folmer, O., Black, M., Hoeh, W., Lutz, R. & Vrijenhoek, R. DNA primers for the amplification of mitochondrial cytochrome c oxidase subunit I from diverse metazoan invertebrates. *Mol. Mar. Biol. Biotechnol.* 3, 294–299 (1994).
- Kumar, S., Tamura, K. & Nei, M. MEGA3: Integrated software for Molecular Evolutionary Genetics Analysis and sequence alignment. *Brief Bioinform.* 5, 150–163 (2004).
- Posada, D. & Crandall, K. A. MODELTEST: testing the model of DNA substitution. *Bioinformatics* 14, 817–818 (1998).
- Ronquist, F. & Huelsenbeck, J. P. MRBAYES 3:Bayesian phylogenetic inference under mixed models. *Bioinformatics* 19, 1572–1574 (2003).
- 67. PAUP* Phylogenetic analysis using parsimony (*and other methods) v. 4 (Sinauer, Sunderland, 2000).
- Clement, M., Posada, D. & Crandall, K. A. TCS: a computer program to estimate gene genealogies. *Mol. Ecol.* 9, 1657–1660 (2000).
- Excoffier, L., Smouse, P. & Quattro, J. Analysis of molecular variance inferred from metric distances among DNA haplotypes: Applications to human mitochondrial DNA restriction data. *Genetics* 131, 479–491 (1992).

- Excoffier, L., Laval, G. & Schneider, S. Arlequin ver. 3.0: An integrated software package for population genetics data analysis. *Evol. Bioinform. Online* 1, 47–50 (2005).
- 71. Feldman, G. C. & McClain, C. R. Aqua MODIS sea surface temperature (11 μ nightime), Ocean Color Web, eds. Kuring, N., Bailey, S. W., Franz, B. F., Meister, G., Werdell, P. J., Eplee, R. E. NASA Goddard Space Flight Center. http://oceancolor.gsfc.nasa.gov/cgi/l3 (April 30 2012).
- Haines, S. L., Jedlovec, G. J. & Lazarus, S. M. A MODIS Sea Surface Temperature Composite for Regional Applications. *IEEE T. GeoSci. Remote* 45, 2919–2927 (2007).
- Smouse, P. E., Long, J. C. & Sokal, R. R. Multiple regression and correlation extension of the Mantel test of matrix correspondence. *Syst. Zool.* 35, 627–632 (1986).
- 74. Goslee, S. C. & Urban, D. L. The ecodist package for dissimilarity-based analysis of ecological data. J. Stat. Software 22, 1–19 (2007).
- Legendre, P. & Legendre, L. Numerical ecology, 2nd English Edition. (Elsevier Science BV, 1998).
- Abramoff, M. D., Magelhaes, P. J. & Ram, S. J. Image processing with ImageJ. Biophotonics Int. 11, 36–42 (2004).

Acknowledgements

This work was supported by the California Department of Fish and Game (grant to Geller). J. Mackie was assisted by the California State University Council on Ocean Affairs, Science and Technology (COAST) consortium and National Science Foundation (NSF award #1061695) support. We thank M. Ashe, S. Foss, and M. Sowby at California Department of Fish and Game for encouragement and project guidance. R. Fairey and his group at the Marine Pollution Studies Laboratory at Moss Landing Marine Laboratories provided the majority of specimens used in this project. Several individuals kindly provided additional samples: A. Cohen, G. Jenkins, L. McCann, A. Miggoto, G. Ruiz, and J. Winston, or sequence information, Yong-Jin Won; or provided general assistance: R. Blackwell, S. Bros, S. Craig, D. Gerhinger, E. Jensen, A. Láruson, K. Messer, G. Schroeder. Though this work has been reviewed by the U.S. Environmental Protection Agency and cleared for publication, it may not necessarily reflect official Agency policy.

Author contributions

All authors contributed to experiment design, experiments, manuscript text, and reviewed the manuscript. Mackie prepared figures.

Additional information

Competing financial interests: The authors declare no competing financial interests.

License: This work is licensed under a Creative Commons Attribution 3.0 Unported License. To view a copy of this license, visit http://creativecommons.org/licenses/by/3.0/

How to cite this article: Mackie, J.A., Darling, J.A. & Geller, J.B. Ecology of cryptic invasions: latitudinal segregation among *Watersipora* (Bryozoa) species. *Sci. Rep.* **2**, 871; DOI:10.1038/srep00871 (2012).