

# Differential Action of Monohydroxylated Polycyclic Aromatic Hydrocarbons with Estrogen Receptors $\alpha$ and $\beta$

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Polycyclic aromatic hydrocarbons (PAHs) are a diverse group of widespread environmental pollutants, some of which have been found to be estrogenic or antiestrogenic. Recent data have shown that hydroxylated PAH metabolites may be responsible for the estrogenic effects of some PAHs. The purpose of this study was to investigate the effects of several PAHs, as well as their monohydroxylated metabolites, on estrogen receptors (ERs), ER $\alpha$  and ER $\beta$ . Three parent PAHs and their monohydroxylated metabolites were each evaluated using transcriptional reporter assays in isogenic stable cell lines to measure receptor activation, competitive binding assays to determine ligand binding, and bioluminescence resonance energy transfer assays to assess dimerization. Finally, the estrogenic effects of the hydroxylated metabolites were confirmed by quantitative real-time PCR of estrogen-responsive target genes. Although the parent PAHs did not induce ER $\alpha$  or ER $\beta$  transcriptional activity, all of the monohydroxylated PAHs (1-OH naphthanol, 9-OH phenanthrene, 1-OH pyrene) selectively induced ER $\beta$  transcriptional activity at the concentrations tested, while not activating ER $\alpha$ . Additionally, the monohydroxylated PAHs were able to competitively bind ER $\beta$ , induce ER $\beta$  homodimers, and regulate ER $\beta$  target genes. Although monohydroxylated PAHs appeared to have weak agonist activity to ER $\beta$ , our results showed that they can elicit a biologically active response from ER $\beta$  in human breast cancer cells and potentially interfere with ER $\beta$  signaling pathways.

**Key Words:** polycyclic aromatic hydrocarbons; estrogen receptors; monohydroxylated metabolites; dimerization; transcription; ligand binding.

Polycyclic aromatic hydrocarbons (PAHs) have been of increasing concern in the human health field due to their widespread dispersion in the environment and the adverse health effects associated with PAH exposure (Baird *et al.*, 2005). Formed through the incomplete combustion of organic compounds, PAHs can be found in charbroiled foods, cigarette smoke, contaminated soil, vehicle exhaust, and in the atmosphere

from the by-products of industrial processes. PAH exposure can have several adverse effects, including carcinogenesis and endocrine disruption.

Although PAHs are a diverse group of chemicals, most are metabolized by cytochrome P450s, a superfamily of enzymes that mediate the oxidation of lipophilic substrates (Anzenbacher, 2001; Bauer *et al.*, 1995; Kim *et al.*, 1998). The diol epoxide PAH metabolites are capable of inducing DNA damage (Baird *et al.*, 2005), and many PAHs have been shown to be carcinogenic (Bauer *et al.*, 1995; Kim *et al.*, 1998). PAHs can also act as endocrine disrupting chemicals by interfering with normal estrogen signaling. Upon monohydroxylation, PAHs can induce estrogenic effects by directly interacting with estrogen receptors (ERs) (Arcaro *et al.*, 1999; Fertuck *et al.*, 2001a,b). These data suggest that the estrogenic effects of PAHs are primarily mediated by the monohydroxylated PAH metabolites.

ERs, members of the nuclear receptor superfamily of transcription factors, exist in two distinct isoforms,  $\alpha$  and  $\beta$ . Encoded by separate genes on different chromosomes, ER $\alpha$  and ER $\beta$  have both overlapping and unique biological functions. The DNA-binding domains share 96% homology, and ERs bind similar estrogen response elements (EREs) to regulate transcription of target genes. The ligand-binding domains (LBDs), containing the hormone-dependent activation function (AF-2) (Tora *et al.*, 1989), have 55% identity and have similar, but not identical, ligand-binding pockets (Pike *et al.*, 1999). Upon ligand binding, the receptors dimerize and bind DNA to initiate transcription of target genes that mediate distinct biological effects. In the presence of estrogen, ER $\alpha$  is a known driver of cell proliferation, especially in breast cancer cells, whereas ER $\beta$  has been shown to inhibit ER $\alpha$ -mediated cell proliferation (Hartman *et al.*, 2006; Paruthiyil *et al.*, 2004; Trecek *et al.*, 2010).

Given the critical roles ERs play in regulating cell growth in response to estrogens, there has been significant effort put forth to understand and predict the impacts of xenoestrogens on ER

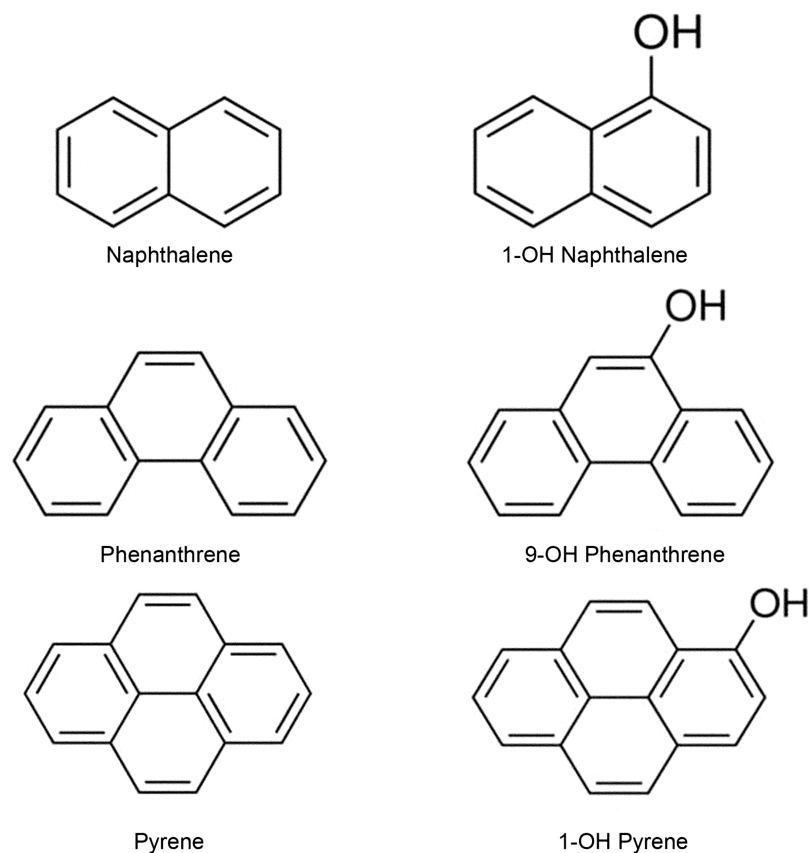


FIG. 1. Chemical structures of select polycyclic aromatic compounds and monohydroxylated metabolites studied.

singaling. However, most studies have been performed solely in the context of ER $\alpha$ , with a limited number of PAHs tested. Here we utilize several *in vitro* assays to assess the effects of three PAHs and their monohydroxylated metabolites, shown in Figure 1, on the transcriptional activation, ligand binding, and dimerization of both ER $\alpha$  and ER $\beta$ . Compounds were initially screened for transcriptional activation using a previously characterized pair of isogenic breast cancer cell lines with inducible expression of either ER $\alpha$  or ER $\beta$  and a stably integrated estrogen-responsive reporter (Shanle *et al.*, 2011). These cell lines provide a sensitive tool to directly compare the transcriptional induction of ER $\alpha$  and ER $\beta$ . Next, bioluminescence resonance energy transfer (BRET) assays were performed to evaluate the dimerization status of ERs. BRET assays are able to monitor protein-protein interactions in a live, cell-based system (Powell and Xu, 2008; Tremblay *et al.*, 1999). Fluorescence polarization experiments were utilized to generate competitive binding curves and determine half maximal inhibitory concentration (IC<sub>50</sub>) values. This provided a simple, yet specific way to determine whether the tested compound can compete with estrogen for binding to ER. Finally, compounds were evaluated for their ability to upregulate ER $\beta$  target genes via quantitative real-time PCR (qPCR).

Naphthalene, phenanthrene, and pyrene were chosen as parent PAH compounds for study because they have been detected

at high levels in contaminated environments (Arcaro *et al.*, 1999), and they are considered by to be Priority Pollutants according to the U.S. Environmental Protection Agency. The hydroxylated metabolites were chosen due to their detection after metabolism of the parent compound (Cho *et al.*, 2006; Rossbach *et al.*, 2007). This is the first study to assess ER selective activity of these PAHs and their hydroxylated metabolites at the levels of transcriptional activity using isogenic reporter cell lines, ligand binding, and dimerization. The data demonstrate that monohydroxylated PAHs differentially interact with ER $\alpha$  and ER $\beta$  and exhibit stronger agonistic activity toward ER $\beta$  compared with ER $\alpha$ , suggesting that ER $\beta$ -mediated biological processes need to be evaluated to assess the outcomes of PAH exposure on humans.

## MATERIALS AND METHODS

**Chemicals.** All PAH compounds were purchased from Sigma-Aldrich (St Louis, MO). Doxycycline (Dox) was obtained from Clontech (Mountain View, CA). ICI 182,780 was obtained from Tocris Bioscience (Ellisville, MO).

**Cell culture and reporter assays.** Cell culture media were obtained from Invitrogen (Carlsbad, CA). HEK293T cells were maintained in Dulbecco's Modified Eagle's Medium (DMEM) supplemented with 10% Gibco Fetal Bovine Serum (FBS; Invitrogen) at 37°C and 5% CO<sub>2</sub>. Hs578T-ER $\alpha$ Luc and Hs578T-ER $\beta$ Luc cells were previously created by Shanle *et al.* (2011) and were

**TABLE 1**  
**Primer and Probe Sequences**

RPL13A	Primer 1	5'-TGT TTG ACG GCA TCC CAC-3'
	Primer 2	5'-CTG TCA CTG CCT GGT ACT TC-3'
	Probe	5'-CTT CAG ACG CAC GAC CTT GAG GG-3'
C3	Primer 1	5'-AAC TAC ATC ACA GAG CTG CG-3'
	Primer 2	5'-AAG TCC TCA ACG TTC CAC AG-3'
	Probe	5'-CGT TTC CCG AAG TGA GTT CCC AGA-3'
JAG1	Primer 1	5'-GGA CTA TGA GGG CAA GAA CTG-3'
	Primer 2	5'-AAA TAT ACC GCA CCC CTT CAG-3'
	Probe	5'-TCA CAC CTG AAA GAC CAC TGC CG-3'
NRIP1	Primer 1	5'-AGA TTC CCT GTC CTC CTT CA-3'
	Primer 2	5'-GGA AGT GTT TGG ATT GTG AGC-3'
	Probe	5'-TGT GCA TCT TCT GGC TGT GTT TCT CC-3'

maintained in DMEM/F12 supplemented with L-glutamine and 10% Tet-system approved FBS (Clontech) at 37°C and 5% CO<sub>2</sub>.

Reporter assays were performed as previously reported (Shanle *et al.*, 2011). Briefly, cells were seeded in triplicate at 10<sup>4</sup> cells/well on white 96-well tissue culture plates in phenol red-free DMEM/F12 supplemented with 5% charcoal-stripped FBS treated with 50 ng/ml Dox. After 24 h, media were removed and replaced with media treated with 50 ng/ml Dox and vehicle (0.15% dimethyl sulfoxide [DMSO]) or PAH compounds diluted in DMSO. After 24 h of treatment, the cells were washed with 30  $\mu$ l of 1 $\times$  PBS and lysed with 35  $\mu$ l lysis buffer (100mM K<sub>2</sub>HPO<sub>4</sub>, 0.2% Triton X-100, pH 7.8). Thirty microliters of lysate were mixed 1:1 with luciferase substrate (Promega, Madison, WI), and luminescence was measured with a 700-nm filter on a Victor X5 microplate reader (PerkinElmer, Waltham, MA). Total protein was measured using the Bradford Method (Bio-Rad), and raw luciferase data were normalized to total protein. Approximate EC<sub>50</sub> values were calculated using GraphPad Prism Software (Version 5.04; Graph-Pad Software Inc., San Diego, CA) and a three-parameter log versus response nonlinear regression.

**BRET assays.** The BRET assays were performed similarly to those previously reported (Powell and Xu, 2008). Briefly, HEK293T cells were transfected with BRET fusion plasmids (pCMX-ER $\alpha$ -RLuc and pCMX-ER $\alpha$ -YFP or pCMX-RLuc-ER $\beta$  and pCMX-YFP-ER $\beta$ ). Twenty-four hours after transfection, cells were trypsinized and resuspended in triplicate in PBS at approximately 50,000 cells per well in a white 96-well plate. Cells were then incubated with vehicle (0.6% DMSO), 10nM E2, or monohydroxylated PAH compound for 1 h at room temperature. Coelenterazine h (Promega) was added to PBS at a final concentration of 5 $\mu$ M. Emission measurements at 460 nm and 535 nm were immediately taken on a Victor X5 microplate reader (PerkinElmer). BRET ratios were calculated as previously described (Koterba and Rowan, 2006; Powell and Xu, 2008).

**Competitive binding assays.** Competitive binding assays were performed using the PolarScreen ER $\beta$  Competitive Binding Assay Kit, Green (Invitrogen) according to the manufacturer's protocol. Recombinant human ER $\beta$  (20nM) and fluorescein-labeled estradiol were incubated for 4 h with the monohydroxylated PAH compounds. Fluorescence polarization was measured using a Victor X5 microplate reader (PerkinElmer). Approximate IC<sub>50</sub> values were determined by GraphPad Prism Software (Graph-Pad Software Inc.) from competitive binding curves.

**Western blot analysis.** Western blots were performed similarly to those previously reported (Shanle *et al.*, 2011) with cells treated for 48 h with vehicle (DMSO) or 10 $\mu$ M monohydroxylated PAH compound. Total protein was quantified using Bio-Rad Protein Assay (Bio-Rad), 35  $\mu$ g of protein was resolved by SDS-PAGE, and membranes were incubated with 1:1000 anti-FLAG-M2 antibody (Sigma) overnight at 4°C. Membranes were then incubated with goat anti-rabbit HRP secondary antibody (Licor Biosciences, Lincoln, NE) for 1 h at room temperature and visualized using SuperSignal West Pico Chemiluminescent Substrate (ThermoScientific, Waltham, MA)

on autoradiography film. Membranes were then washed and incubated with 1:5000 anti- $\beta$ -Actin (Sigma) for 1 h at room temperature, then incubated with goat anti-mouse HRP secondary antibody (Licor Biosciences) for 1 h at room temperature and visualized using SuperSignal West Pico Chemiluminescent Substrate (ThermoScientific) on autoradiography film.

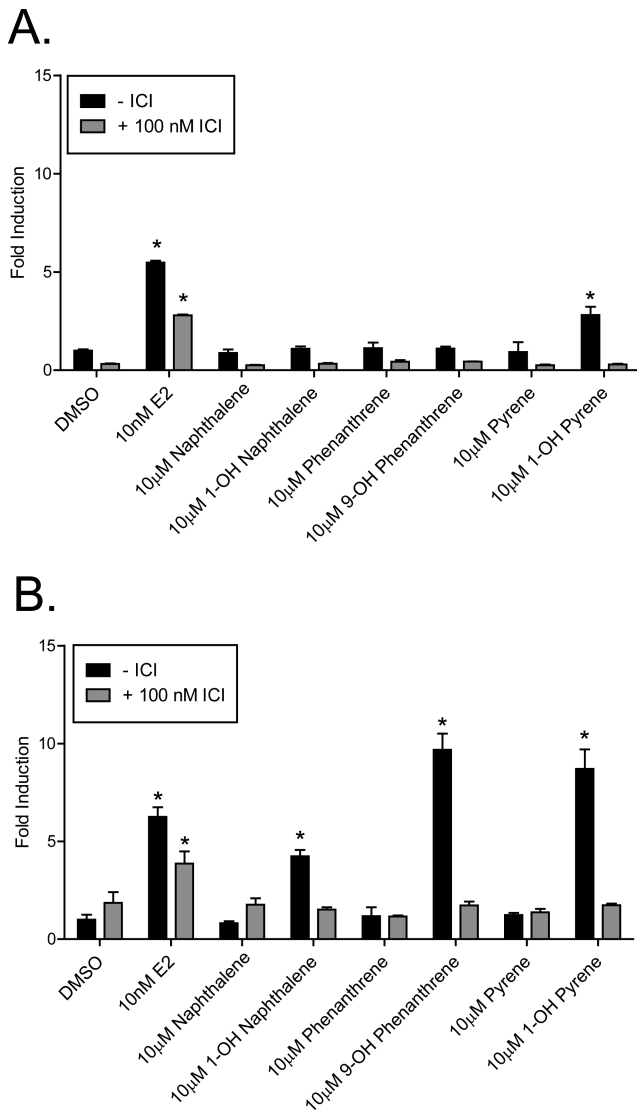
**qPCR analysis.** Hs578T-ER $\beta$ Luc cells were cultured in phenol red-free DMEM/F12 supplemented with 10% charcoal-stripped FBS for 3 days prior to experiment to remove any residual estrogens. Cells were seeded into 10-cm tissue culture plates in phenol red-free DMEM/F12 supplemented with 5% stripped serum and treated with 50 ng/ml of Dox 24 h prior to PAH treatment. Cells were then treated with 50 ng/ml Dox plus 0.1% DMSO control, 10nM E2, 10 $\mu$ M 1-OH-naphthalene, 5 $\mu$ M 9-OH phenanthrene, or 5 $\mu$ M 1-OH pyrene for 24 h. Total RNA was extracted using HP Total RNA Kit (VWR Scientific, West Chester, PA) according to the manufacturer's protocol. One microgram of RNA was reverse transcribed using Superscript II RT according to the manufacturer's protocol (Invitrogen), and qPCR was performed using TaqMan Prime Time custom designed assays (IDT, Coralville, IA), FastStart Universal Probe Master Mix (Roche Scientific, Basel, Switzerland), and a CFX96 instrument (Bio-Rad). Primer and probe sequences are shown in Table 1.

**Statistical analyses.** Two-tailed Student's *t*-tests were performed using GraphPad Prism version 5.04 for Windows, GraphPad Software ([www.graphpad.com](http://www.graphpad.com)).

## RESULTS

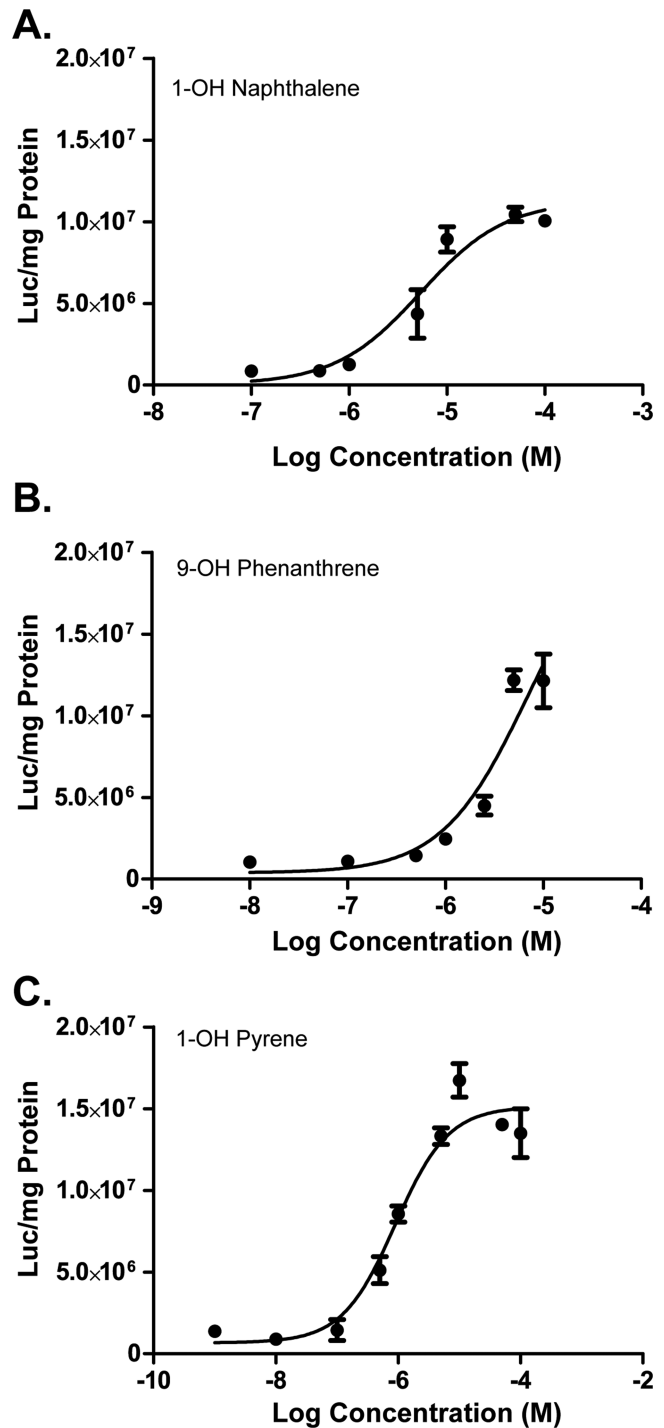
### *Monohydroxylated PAHs Selectively Activate ER $\beta$ in Reporter Cell Lines*

In order to test the hypothesis that hydroxylated PAHs may have estrogenic activity with differential effects on ER $\alpha$  and ER $\beta$ , we first utilized Hs578T-ER $\alpha$ Luc and Hs578T-ER $\beta$ Luc reporter cells (Shanle *et al.*, 2011). These cell lines have inducible expression of ER $\alpha$  or ER $\beta$ , respectively, and a stably integrated luciferase reporter just downstream of three tandem EREs. Previous work has shown that these cell lines are highly sensitive to estrogenic ligands and can be used to distinguish ER subtype selective ligands (Shanle *et al.*, 2011). In this system, cells are first treated with Dox to induce expression of the receptor, followed by treatment with the corresponding compounds. In our initial experiments comparing the activation of ER $\alpha$  and ER $\beta$ , we observed that only hydroxylated PAHs conferred estrogenic activity at 10 $\mu$ M (Fig. 2). The monohydroxylated PAH



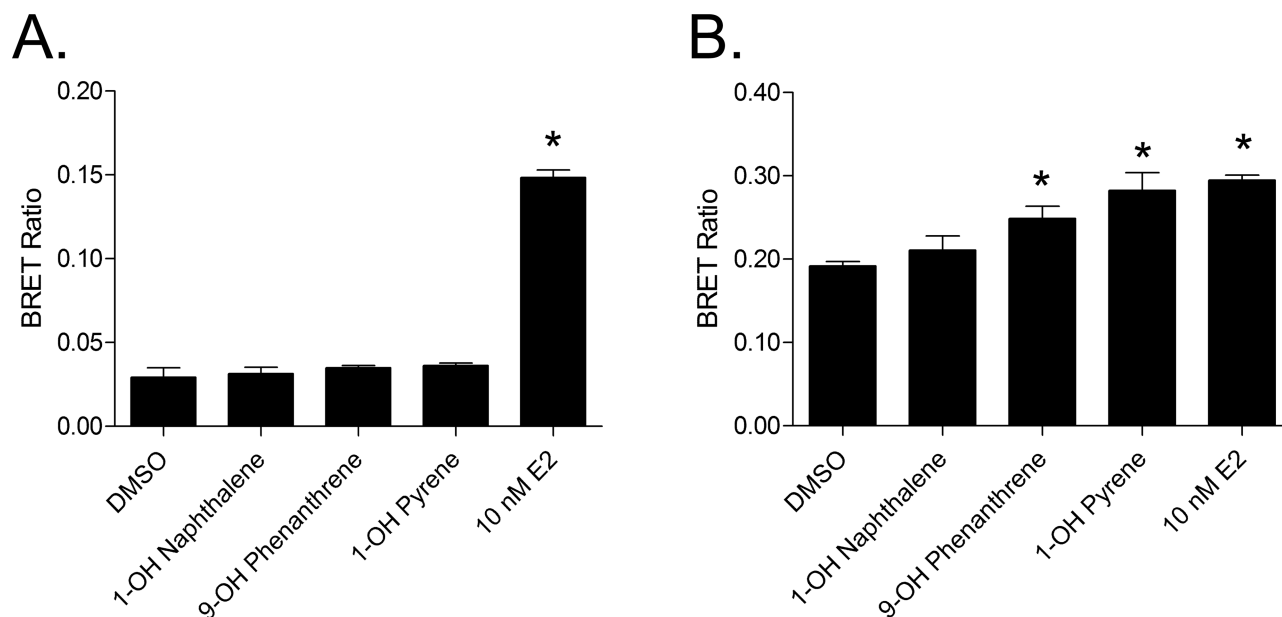
**FIG. 2.** Differential activation of ER $\alpha$  and ER $\beta$  by select monohydroxylated PAH compounds. (A) Hs578T-ER $\alpha$ Luc and (B) Hs578T-ER $\beta$ Luc stable cell lines were treated in triplicate with 10 $\mu$ M of PAH compound in the presence or absence of 100nM ICI 182,780 for 24h. Data are expressed as fold induction of raw luciferase units per mg protein over the DMSO control  $\pm$  SD. Experiments were repeated at least twice. \* $p$  < 0.01 compared with DMSO control.

compounds were able to induce the ERE-luciferase reporter activity primarily in the Hs578T-ER $\beta$ Luc cells (Fig. 2B). In these cells, 1-OH naphthalene, 9-OH phenanthrene, and 1-OH pyrene induced a 4.2-, 9.7-, and 8.7-fold change over DMSO vehicle control, respectively ( $p$  < 0.01 in all cases). In contrast, only 1-OH pyrene induced the ERE-luciferase reporter activity in the Hs578T-ER $\alpha$ Luc cell line ( $p$  < 0.01), but not nearly to the same degree as that of 17 $\beta$ -estradiol (E2) (Fig. 2A). The ER antagonist ICI 182,780 blocked PAH-induced expression in all cases, reducing the luciferase signal to that of vehicle-treated cells. Reporter expression induced by 10nM E2 was not fully blocked by ICI 182,780 cotreatment because of the high



**FIG. 3.** Monohydroxylated PAHs activate ER $\beta$  in a dose-dependent manner. Hs578T-ER $\beta$ Luc cells were treated with Dox for 24h followed by treatment with a range of concentrations of (A) 1-OH naphthalene, (B) 9-OH phenanthrene, or (C) 1-OH pyrene. The mean and SD shown are from triplicates of one representative experiment repeated twice.

concentration and potency of E2. No induction of reporter gene activity was seen in control experiments in which cells were not treated with Dox (Supplementary fig. 1), further confirming ER-mediated induction of the luciferase reporter.



**FIG. 4.** Monohydroxylated PAH compounds selectively induce ER $\beta$ / $\beta$  homodimers. (A) BRET data for 293T cells transfected with CMX-ER $\alpha$ -RLuc and CMX-ER $\alpha$ -YFP, showing no ER $\alpha$ / $\alpha$  dimerization upon treatment with PAH compounds in triplicate. (B) BRET data for 293T cells transfected with CMX-RLuc-ER $\beta$  and CMX-YFP-ER $\beta$ , showing ER $\beta$ / $\beta$  dimerization when treated with PAH compounds in triplicate. The experiment was performed three times. Error bars represent SEM. \* $p < 0.05$  compared with DMSO control.

We next determined the dose-dependent effects of the hydroxylated PAHs in the Hs578T-ER $\beta$ Luc cells (Fig. 3). The half maximal effective concentration (EC<sub>50</sub>) values for 1-OH naphthalene and 1-OH pyrene were found to be approximately 5.38 and 0.89 $\mu$ M, respectively. 9-OH Phenanthrene proved to be cytotoxic at concentrations greater than 10 $\mu$ M, and the dose-response curve did not adequately saturate; however, an approximate EC<sub>50</sub> value was estimated to be  $\geq 6.8\mu$ M.

#### Monohydroxylated PAHs Induce ER $\beta$ Dimers and Directly Bind the Receptor

To further dissect the mechanism through which the monohydroxylated PAHs activate ER $\beta$  and confirm the selectivity of the compounds, ER dimerization induced by the compounds was assessed using BRET assays. BRET assays allow the determination of dimer formation in live cells by transfecting cells with an energy donor (ER fused to Renilla luciferase) and acceptor (ER fused to yellow fluorescent protein) (see Powell and Xu, 2008). Upon transfecting the cells with the fusion constructs for ER $\alpha$  or ER $\beta$ , 9-OH phenanthrene and 1-OH pyrene were shown to significantly induce ER $\beta$  homodimerization ( $p = 0.02$  and  $0.01$ , respectively) (Fig. 4B). In contrast, 1-OH naphthalene did not significantly induce ER $\beta$  dimerization as determined by the BRET assay ( $p = 0.35$ ). Following the trend seen in the ER $\alpha$  ERE-reporter assay, the monohydroxylated PAH compounds were unable to induce ER $\alpha$  homodimers (Fig. 4A).

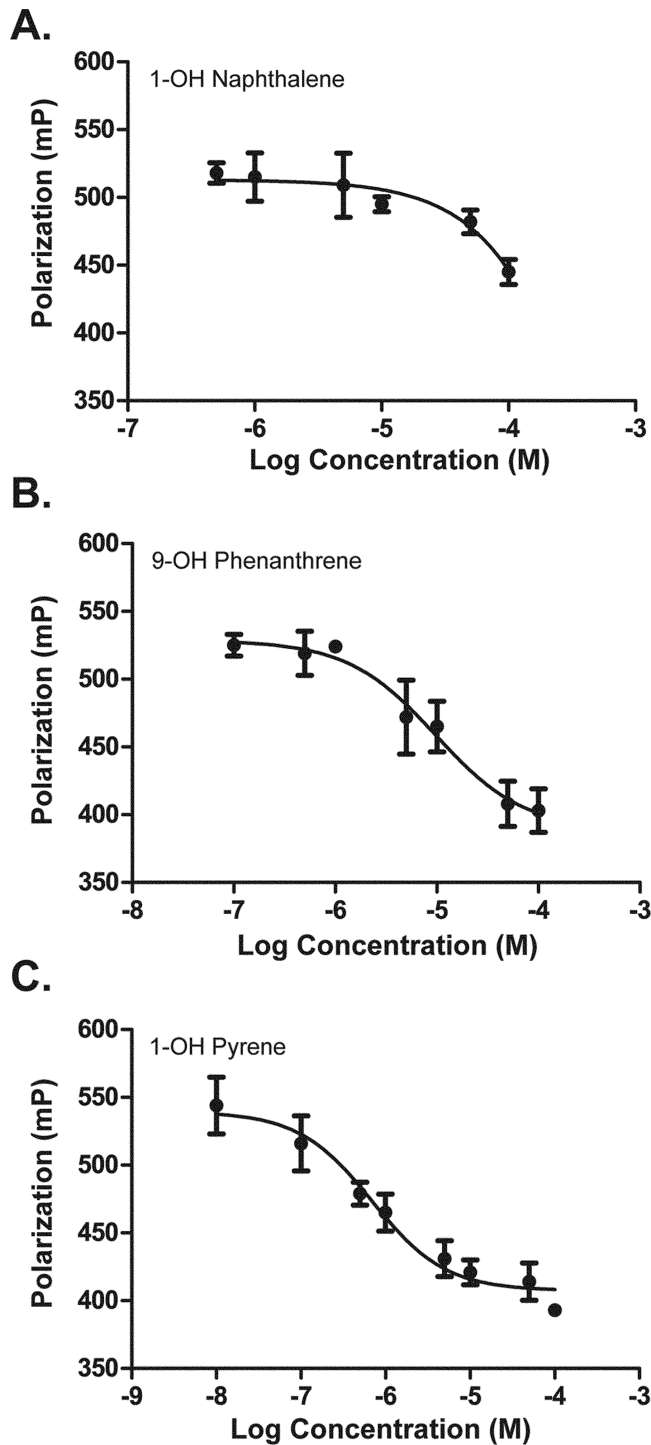
In order to confirm that ER $\beta$  dimerization and ERE-luciferase activity were directly induced by ligand binding, the ability of the monohydroxylated PAH compounds to displace

fluorescein-labeled estradiol from human ER $\beta$  was assessed in a competitive binding assay (Fig. 5). The competition with E2 indicates that compounds directly bind to ER $\beta$  in the same ligand-binding pocket as E2. These competitive binding data yielded half maximal inhibitory concentration (IC<sub>50</sub>) values for 9-OH phenanthrene and 1-OH pyrene at 9.75 and 0.69 $\mu$ M, respectively. In support of the BRET results, 1-OH naphthalene showed a much lower affinity for ER $\beta$  as evidenced by Figure 5A, but it was still able to displace E2 at higher concentrations. The approximate IC<sub>50</sub> value for 1-OH naphthalene was estimated at or greater than 0.48 $\mu$ M.

After determining that monohydroxylated PAHs bind ER $\beta$ , Western blots with FLAG antibody were used to determine the degradation status of the receptor (Supplementary fig. 2), as some ER ligands cause degradation of the receptor upon binding. These Western blots confirmed that ER $\beta$  was not degraded by the monohydroxylated PAHs within 48 h of treatment.

#### Monohydroxylated PAHs Exhibit Estrogenic Activity on ER $\beta$ Target Genes

To further validate the reporter assay and BRET assay results, the regulation of endogenous ER $\beta$  target genes was assessed. Estrogen responsive target genes of ER $\beta$  were previously identified in Hs578T-ER $\beta$  cells (Secreto *et al.*, 2007). Two upregulated target genes (*CC3* and *NR1P1*) and one downregulated target gene (*JAG1*) were selected for analysis by qPCR (Fig. 6). At 10 $\mu$ M, 1-OH naphthalene was able to induce *CC3* and *NR1P1* expression 2.1- and 2.2-fold over DMSO, respectively. Although the increased expression of *CC3* did not reach statistical significance ( $p = 0.06$ ),



**FIG. 5.** Monohydroxylated PAHs can bind ER $\beta$  *in vitro*. Competitive binding curves for monohydroxylated PAH compounds displacing fluorescein-labeled estradiol from human ER $\beta$ . Purified hER $\beta$  and fluorescein-labeled estradiol were incubated for 4h with serial dilutions in triplicate of (A) 1-OH naphthanol, (B) 9-OH phenanthrene, and (C) 1-OH pyrene. Error bars represent SD.

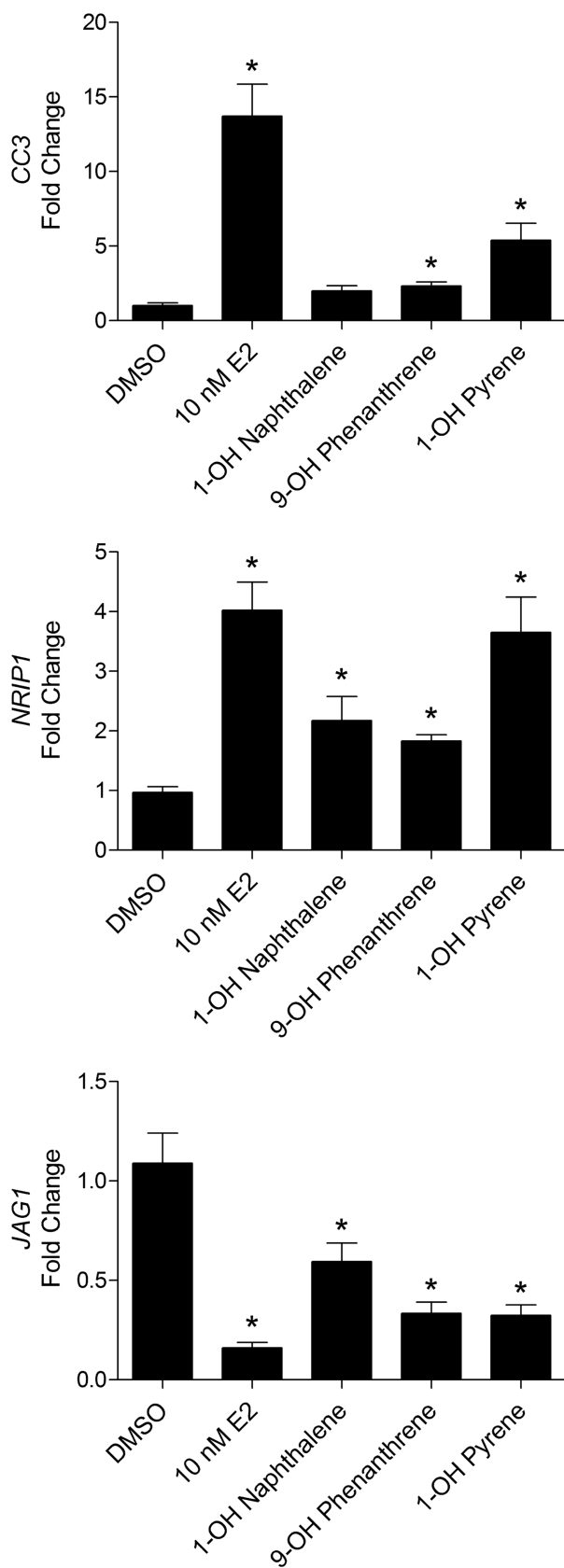
*NR1P1* was significantly upregulated by 1-OH naphthalene ( $p = 0.02$ ). Treatment with 5 $\mu$ M 9-OH phenanthrene was able to significantly induce *CC3* and *NR1P1* expression 2.4-fold ( $p = 0.02$ ) and 1.9-fold ( $p < 0.01$ ) over DMSO, respectively. Similarly, 5 $\mu$ M 1-OH pyrene was able to significantly induce *CC3* and *NR1P1* expression 5.6-fold ( $p = 0.02$ ) and 3.8-fold ( $p < 0.01$ ) over DMSO, respectively. Additionally, all three monohydroxylated PAH compounds were able to downregulate the expression of *JAG1*, generating mean fold changes of 0.64 ( $p = 0.02$ ), 0.36 ( $p < 0.01$ ), and 0.32 ( $p < 0.01$ ) over the DMSO control. It is important to note that although all compounds displayed some estrogenic activity on the target genes tested, the estrogenic response was not as robust as that of E2.

## DISCUSSION

Numerous studies have investigated the relationship between PAHs, their hydroxylated metabolites, and potential interactions with the ERs, yet most have focused on ER $\alpha$  (reviewed by Santodonato, 1997). Hayakawa *et al.* (2007) reported estrogenic and antiestrogenic activity for multiple monohydroxylated derivatives of common PAHs in a yeast two-hybrid assay expressing ER $\alpha$ . Similar to our findings, they also reported that the parent PAH compounds lacked any estrogenic or antiestrogenic activity. Charles *et al.* (2000) also reported estrogenic activity for hydroxylated metabolites of the carcinogen benzo[a]pyrene (B[a]P) in MCF-7 cells, which primarily express ER $\alpha$ . Despite these previous findings, there have been relatively few studies comparing the effects of monohydroxylated PAHs on the differential activation and dimerization of ER $\alpha$  and ER $\beta$ .

Our results, consistent with prior studies, indicate that hydroxylated PAHs are the active estrogenic species and can differentially activate either ER $\alpha$  or ER $\beta$ . Although the compounds we tested exhibited no interaction with ER $\alpha$ , the interaction with ER $\beta$  is novel and significant. Inhibition of luciferase signal by the ER antagonist ICI 182,780, as well as the lack of luciferase signal in the absence of Dox, demonstrates that the results of the reporter assay are ER $\beta$  mediated. Competition with fluorescein-labeled estradiol indicates that these monohydroxylated PAH compounds directly bind to ER $\beta$  at the same ligand-binding pocket as E2. Fertuck *et al.* (2001a) investigated different parental PAH and metabolite compounds, and they similarly reported that hydroxylated PAHs were able to compete with estrogen and bind ERs with a slight preference for ER $\beta$ . Their data, consistent with our findings, suggest that hydroxylated PAHs may preferentially affect ER $\beta$  signaling. Given ER $\beta$ 's role in normal development and function in reproductive tissues as well as in the lungs, colon, prostate, and cardiovascular system, disruption of and interference with ER $\beta$  signaling could have implications in normal development, as well as in cancers and malfunctions of these tissues.

In addition to the reporter assay and competitive binding data, the BRET and qPCR data confirm that 9-OH phenanthrene and 1-OH pyrene induce a biologically active ER $\beta$  response in this



system. Given our data, 1-OH naphthalene may not necessarily induce ER $\beta$  homodimers even at the high concentration tested (10 $\mu$ M). In support of these data, ligand-binding assays with 1-OH naphthalene demonstrate a relatively low binding affinity for ER $\beta$ . Despite the negative BRET data, qPCR for endogenous ER $\beta$  target genes suggest that 1-OH naphthalene is capable of inducing a slight biologically active ER $\beta$  response for some ER $\beta$  target genes although not to the same extent as E2. Collectively, the data obtained for 1-OH naphthalene demonstrate an important consideration of the *in vitro* assays used in this study: different assays have different sensitivities for detecting estrogenic activity and ER subtype selectivity. The ER $\beta$  homodimerization BRET assay typically shows a 1.5- to 2-fold induction with E2 treatment because of high ligand-independent dimerization (Powell and Xu, 2008). In addition, the BRET ratios ultimately depend on the conformational changes within the receptor fusion proteins, which allow for efficient energy transfer, and different ligands will induce different conformational changes, thereby affecting the BRET ratio output. Despite the lower fold changes for the ER $\beta$  homodimerization assay, BRET has been successfully used in a high-throughput manner to identify ER dimer selective ligands (Powell *et al.*, 2010) and, in this study, demonstrated a significant induction of ER $\beta$  homodimerization by two other monohydroxylated PAHs, 1-OH pyrene and 9-OH phenanthrene.

Although each monohydroxylated PAH tested gave a similar pattern of results, the relative activity of each compound is quite different. Our data indicate that 1-OH naphthalene is the weakest ER $\beta$  agonist among the tested metabolites, as demonstrated by low reporter gene output, a lack of saturation in the dose-response reporter assays, low induction of ER $\beta$  dimerization, and a lower binding affinity for ER $\beta$ . In contrast, 1-OH pyrene and 9-OH phenanthrene appear to be fairly efficient ER $\beta$  agonists. Both ligands induced ERE-reporter gene activity similar to E2 and effectively displaced E2 from the ER $\beta$  ligand-binding pocket. Both compounds also significantly elicited ER $\beta$  homodimerization. 9-OH Phenanthrene generated data similar to 1-OH pyrene with the exception that it proved to be cytotoxic at concentrations greater than 10 $\mu$ M, resulting in difficulty to obtain accurate EC<sub>50</sub> values. Despite the cytotoxicity of 9-OH phenanthrene at high concentrations, treatment with lower concentrations of 9-OH phenanthrene (5 $\mu$ M) stimulated the regulation of endogenous ER $\beta$  target genes in Hs578T-ER $\beta$ Luc cells. These data suggest that some monohydroxylated PAHs can affect ER $\beta$ -mediated signaling prior to inducing general cytotoxicity.

**FIG. 6.** Monohydroxylated PAHs can regulate ER $\beta$  target genes similar to estradiol. Expression of ER $\beta$  target genes (*CC3*, *NRIP1*, and *JAG1*) was determined by measuring relative mRNA levels using qPCR. RNA was collected following treatment with 0.1% DMSO, 10nM E2, 10 $\mu$ M 1-OH naphthalene, 5 $\mu$ M 9-OH phenanthrene, or 5 $\mu$ M 1-OH pyrene for 24-h and 48-h treatment with 50 ng/ml Dox. Data are expressed as fold induction compared with DMSO control. Error bars represent SEM. \* $p$  < 0.05 compared with DMSO control.

Although our data did not indicate that any of the monohydroxylated PAHs tested had an effect on ER $\alpha$ , others have reported ER $\alpha$  estrogenic effects for these compounds. Hayakawa *et al.* (2007) reported that all three monohydroxylated PAHs exhibited little to no ER $\alpha$  estrogenic activity, but that 1-OH pyrene was able to compete with E2 for ER $\alpha$  binding. Additionally, Wiele *et al.* (2004) reported that 1-OH pyrene showed ER $\alpha$  estrogenic activity in colon extracts from a simulator of the human intestinal microbial ecosystem. Discrepancies across these studies may be due to the use of different assays and cell lines to assess the estrogenic activity.

Overall, these data suggest that common monohydroxylated PAHs can interact, positively or negatively, with ER signaling. We can conclude from our results and from other studies that hydroxylated PAHs are the active estrogenic species and can differentially bind ER $\alpha$  or ER $\beta$ , likely in a cell- and tissue-specific manner. Few studies assessing the physiological serum concentrations of monohydroxylated PAHs have been published, although monohydroxylated PAHs may be used as urine biomarkers to assess exposure to PAHs (Elovaara *et al.*, 2006). It is therefore difficult to predict the concentrations of monohydroxylated PAHs that reach tissues such as the mammary gland, and the concentrations shown to be estrogenic in these studies may or may not be reached in the serum. Some estrogenic compounds in the diet, such as genistein found in soy products, can reach serum concentrations near the micromolar range (Cassidy *et al.*, 2006). Ultimately, the physiological exposure to monohydroxylated PAHs will be a function of both exposure and metabolic activity, which will greatly vary among individuals. These *in vitro* studies, however, demonstrate the potential for monohydroxylated PAHs to impact ER $\beta$ -mediated signaling and provide a framework for assessing the impacts of other environmental chemicals on the dimerization and transcriptional activities of ER $\alpha$  and ER $\beta$ .

#### SUPPLEMENTARY DATA

Supplementary data are available online at <http://toxsci.oxfordjournals.org/>.

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