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Bisphenol A inhibits cultured mouse ovarian follicle growth partially via the aryl hydrocarbon receptor signaling pathway

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Abstract

Bisphenol A (BPA) is an endocrine disruptor that inhibits growth of mouse ovarian follicles and disrupts steroidogenesis at a dose of 438 μ M. However, the effects of lower doses of BPA and its mechanism of action in ovarian follicles are unknown. We hypothesized that low doses of BPA inhibit follicular growth and decrease estradiol levels through the aryl hydrocarbon receptor (AHR) pathway. Antral follicles from wild-type and *Ahr* knock-out (*Ahr*KO) mice were cultured for 96 hours. Follicle diameters and estradiol levels then were compared in wild-type and *Ahr*KO follicles \pm BPA (0.004 - 438 μ M). BPA inhibited follicle growth (110 - 438 μ M) and decreased estradiol levels (43.8 - 438 μ M) in wild-type and *Ahr*KO follicles. However, at BPA 110 μ M, inhibition of growth in *Ahr*KO follicle growth partially via the AHR pathway, whereas its effects on estradiol synthesis likely involve other mechanisms.

Keywords

BPA; AHR; antral follicle; ovary; low dose; estradiol

1. Introduction

Antral follicles are the most mature types of mammalian ovarian follicles. They are essential for reproduction and endocrine homeostasis because they are the only follicle type capable of ovulation and they are the main source of estradiol production in the female body. Alterations in proper growth or function of the antral follicles may result in infertility. While there are many causes of infertility, environmental chemical exposures may lead to impaired follicular growth and reduced estradiol levels, and in turn, lead to infertility.

Bisphenol A (BPA) is a chemical widely used by plastic and epoxy resin manufacturers. It is incorporated in numerous products such as plastic containers, toys, building materials, and the lining of metal cans such as those used for food and baby formulas [1]. Unfortunately,

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BPA can leach out of the products it is incorporated in and end up in food and beverages [1]. Several studies indicate that BPA is detectable in different human tissues and fluids, including ovarian follicular fluid [2]. These findings are of concern because BPA is known to have endocrine disrupting properties and to be a reproductive toxicant [3]. For example, BPA exposure has been associated with adverse reproductive outcomes such as infertility [4, 5] and polycystic ovarian syndrome in women [6]. Further, postnatal exposure to BPA (150 µg/pup of 1.5 gram body weight) causes polyovular follicles in mice [7].

The calculated lowest observable adverse effect level (LOAEL) dose according to the United States Environmental Protection Agency (EPA) was determined to be 50 mg/kg/day. Though there is not much known yet on the pharmacokinetics of BPA in rodents or humans and it is difficult to extrapolate LOAEL doses obtained by in vivo studies to in vitro studies, it is still important to study the effects of a wide range of BPA doses in vitro to get a better understanding of the potential mechanism of action of BPA.

To date, studies have shown that in vitro treatment with BPA (100 pM-100 μ M) decreases viability of mouse granulosa cells [8]. Recently, BPA also has been shown to have toxic effects on cultured mouse ovarian follicles [9-11]. Specifically, BPA at 438 μ M inhibits growth of mouse antral follicles and BPA at 43.8 or 438 μ M decreases estradiol synthesis compared to vehicle control treated follicles [9]. However, not many studies have examined the effects of BPA on ovarian follicles using a wide dose range. Thus, one goal of the current study was to examine the effects of a wide range of doses of BPA (0.004 - 438 μ M) on ovarian follicle function by focusing on follicle growth and estradiol levels. Since it has been suggested that BPA may act in a non-monotonic manner [12], we specifically tested the hypothesis that a wide range of BPA doses causes U-shaped dose-response effects on follicle growth and estradiol levels.

Though BPA has been shown to act on various tissues and cell types such as MCF-7 cells [13], adipose tissue [14], and pituitary cell lines [15], the receptor types that mediate the effects of BPA on ovarian follicles have not been determined yet. A few studies have focused on whether BPA exerts its action through nuclear receptors such as estrogen [13], androgen [16], and thyroid [17, 18] receptors. Further, Peretz *et al.* [10] have shown that BPA does not exert its toxic effects via the genomic estrogenic pathway in mouse ovarian follicles. Thus, another goal of this study was to determine if BPA exerts its toxicity in mouse antral follicles via the aryl hydrocarbon receptor (AHR) pathway.

We focused on the AHR pathway because it mediates the effects of various environmental chemicals, including its most potent ligand 2,3,7,8,- tetrachlorodibenzo-p-dioxin (TCDD) [19]. Upon ligand binding, cytoplasmic AHR enters the nucleus, partners with its translocator (AHR nuclear translocator; ARNT), and binds to specific DNA response elements, leading to transcription of several genes including cytochrome P450, family 1, subfamily b, polypeptide 1(*Cyp1b1*). Further, the AHR repressor (AHRR) can heterodimerize with ARNT to terminate the activation of the AHR signaling pathway [20-22].

We also focused on the AHR because it plays a central role in the ovary [23]. For example, the AHR is involved in regulating the growth of pre-antral and antral follicles by increasing granulosa cell proliferation [24]. It is also involved in regulation of estradiol production by the ovary [25]. However, previous studies that have examined the potential role of the AHR in mediating the toxic effects of BPA have been equivocal. For example, Nishizawa *et al.* [26] reported that following in utero exposure to BPA, *Ahr* expression was elevated in the gonads harvested from embryos, whereas Bonefeld-Jorgensen *et al.* [27] reported weak

antagonistic effects of BPA with the AHR (AHR-CALUX system; mouse hepatoma cell line).

Nevertheless, no study has examined the direct role of the AHR in mediating the toxic effects of BPA on ovarian follicles. Thus, another goal of the current study was to test the hypothesis that the AHR mediates the toxic effects of BPA on the growth of cultured mouse antral follicles. Specifically, we hypothesized that BPA exposure leads to changes in the expression of main factors in the AHR signaling pathway in antral follicles isolated from wild-type (WT) mice, and that the effects of BPA exposure on the growth of antral follicles are attenuated in follicles lacking the AHR (i.e., follicles isolated from AHR global knock-out mice; *Ahr*KO) compared to WT follicles.

Lastly, activation of the AHR signaling pathway can decrease the levels of estradiol synthesizing enzymes and increase the levels of estradiol metabolizing enzymes (e.g., CYP1B1) [28]. For example, Karman *et al.* [29] found that exposure of cultured mouse antral follicles to various doses of the potent AHR ligand TCDD decreased estradiol levels compared to control follicles. Interestingly, Peretz *et al.* [9] reported that antral follicles cultured with BPA (43.8, 438 μ M) also synthesized lower estradiol levels compared to control follicles. Thus, it is possible that the AHR mediates the toxic effects of BPA on mouse antral follicles by decreasing estradiol levels. Therefore, a final goal of this study was to test the hypothesis that treatment of cultured mouse antral follicles with BPA leads to lower estradiol levels via mechanisms involving the AHR signaling pathway.

2. Materials and methods

2.1. Chemicals

BPA powder (99%) was purchased from Sigma-Aldrich (St. Louis, MO). Stock solution of BPA was prepared using dimethyl sulfoxide (DMSO) (Sigma-Aldrich) as a solvent to achieve a BPA dose of 438 μ M (133 mg/ml). Further dilutions of the stock solution with DMSO generated working dilutions of 0.004, 0.04, 0.44, 4.38, 43.8, 110, and 219 μ M respectively. These serial dilutions allowed us to add an equal volume of BPA to each individual well in which a follicle was placed, and to control for solvent concentration.

The selected dose range is based on previous findings and environmentally relevant levels [30, 31]. Previous studies indicate that BPA at 4.38, 43.8, and 438 µM doses exhibited a selective inhibition of growth of follicles isolated from Friend leukemia virus B (FVB) mice (30-32 days old) [9]. Specifically, BPA at 438 µM, but not at the 4.38 and 43.8 µM doses inhibited follicle growth starting at 72 hours and continuing for 96 hours of culture. In our current study, however, we used a different strain of mice and age of mice than the ones that were used in the study by Peretz et al. [9, 10]. Therefore, BPA at 4.38, 43.8, and 438 µM were included in the current study as well. Additionally, a wide range of other doses (0.004 -219 µM) were included to fully understand the dose response effects of BPA on ovarian follicles. BPA at 0.004 µM represents BPA levels that have been measured in women's follicular fluid (i.e., 0.004- 0.008 µM) [2]. BPA at 219 µM was used as a transition dose between the relatively low doses of BPA (up to 43.8 µM) to the higher doses of BPA (greater than 43.8 µM). Moreover, according to the U.S. Environmental Protection Agency the LOAEL of BPA is 50 mg/kg/day. Therefore, treatment with BPA at 219 µM and lower doses is important to get a better understanding of the in vitro effects of BPA at doses that may simulate the LOAEL doses and lower doses.

2.2. Animals

Cycling wild-type (WT) female C57BL/6 mice were obtained from Charles River Laboratories (Charles River, CA). *Ahr*KO mice (official symbol *Ahr^{tm1Bra}*) were in a

C57BL/6 background and were originally generated by Schmidt *et al.* [32]. The *Ahr*KO (*Ahr*^{-/-}) mice used in each experiment were generated by intercrossing either heterozygous (*Ahr*^{+/-}) female and male mice or heterozygous *Ahr*^{+/-} female mice with *Ahr*KO (*Ahr*^{-/-}) male mice. Mouse genetic screening was performed using ear tissue punches as previously described [33]. All mice were housed and bred in the core animal facility located at College of Veterinary Medicine, University of Illinois. Mice were maintained in polystyrene cages. Additionally, mice were kept under a 12L:12D photoperiod, at temperature of $22\pm1^{\circ}$ C, with $35\% \pm 4\%$ relative humidity. Food (Harlan Teklad 8626) and reverse-osmosis purified water were provided *ad libitum*. Mice were euthanized at 50-54 days of age by carbon dioxide inhalation followed by cervical dislocation. This age range was selected because previous studies indicate that at this age, WT antral follicles and *Ahr*KO antral follicles have similar growth patterns, estradiol levels, and expression levels of a major cell cycle regulator *Ccnd2* [34]. All procedures and experimental methods involving animals were approved by the University of Illinois Institutional Animal Care and Use Committee.

2.3. Follicle Culture

Ovaries were removed and antral follicles were mechanically isolated using fine watchmaker forceps and a dissecting microscope. We used 2-3 mice per replicate, and repeated the cultures at least three separate times. We obtained 15-40 antral follicles with relative sizes of 200-400 µm from each mouse. Following isolation, follicles were placed in individual wells of 96-well culture plates. Specifically, follicles from each mouse were equally distributed in all treatment groups and each treatment group included overall a minimum of 8 follicles. Next, follicles were covered with 75 µl of unsupplemented alpha minimal essential media (-MEM; Invitrogen, Carlsbad, CA) prior to treatment as previously described [35]. After plating, follicles were treated with 150 µl of supplemented media with vehicle control (DMSO) or BPA (0.004, 0.04, 0.4, 4.38, 43.8, 110, 219, and 438 µM). Supplemented media contained unsupplemented -MEM, 1% ITS (10 ng/ml insulin, 5.5 ng/ml transferrin, and 5.5 ng/ml selenium; Sigma-Aldrich, St. Louis, MO), 5% fetal bovine serum (Atlanta biological, Lawrenceville, GA), 100 U/ml penicillin (Sigma-Aldrich), 100 mg/ml streptomycin (Sigma-Aldrich), and 5 IU/ml human recombinant follicle stimulating hormone (Dr A. F. Parlow, National Hormone and Peptide Program, Harbor-UCLA Medical Center, Torrance, CA). An equal volume of BPA or DMSO (0.75 µl of treatment per 1 ml of media) was added to supplemented -MEM to keep vehicle concentration at a constant of 0.075% for each treatment. Follicles then were incubated for up to 96 hours, while supplying 5% CO_2 at 37°C.

2.4. Analysis of follicular growth

As a measurement of growth, follicle diameter was measured on perpendicular axes every 24 hours. Measurements were done with an inverted microscope equipped with a calibrated ocular micrometer. Follicle diameters were averaged and converted to percent change relative to baseline (0 hour, set as 100%). Data were plotted and statistically analyzed to compare the effects of BPA treatments on growth compared to DMSO per time point. At least three separate cultures were conducted per treatment group.

2.5. Gene expression analyses

At the end of each culture (either 24 hours or 96 hours), follicles were immediately snap frozen and stored at -80° C until further processed for quantitative real-time polymerase chain reactions (qPCR). Total RNA was extracted from at least 8 pooled follicles per treatment group per replicate (n 3) using RNeasy Micro Kit (Qiagen Inc., CA) following the manufacturer's protocol. The RNA concentration of each sample was determined at 260 nm using a Nanodrop ND1000 UV-Vis spectrophotometer (Nanodrop Technologies, Wilmington, DE). Unfortunately, we could not isolate enough RNA from follicles treated

with 219 and 438 μ M BPA following 96 hours of culture or from follicles treated with 438 μ M BPA following 24 hours of cultures. RNA Total RNA was reversed transcribed using an iScript cDNA synthesis kit (Bio-Rad, Hercules, CA) according to the manufacturer's instructions. Each cDNA sample was diluted 1:4 with nuclease-free water prior to further analyses.

Primer sets were either based on previous publications: *Ahr* [29], beta actin (*Actb*) [29], *Cyp1b1* [29], *Arnt* [36], B cell leukemia/lymphoma 2 (*Bcl2*) [37], and BCL2-associated X protein (*Bax*) [37] or originally designed (*Ahrr*, NM_009644.2) using a BLASTN search that was performed in GenBank to ensure that primers were unique to the gene of interest and that they spanned a large intron-exon junction. The sequences for the originally designed *Ahrr* primers were: forward 5 -AGA GCT GTG TCC CCA GGG AAG T - 3 and reverse 5 - CAG TGA GGA AAG ATG GCT TGT AGG - 3.

All gene expression analyses were performed using the CFX96 Real-time PCR Detection System C1000 Thermal Cycler (Bio-Rad). All qPCR reactions were done in triplicate in a total volume of 10 µl per sample accordingly: 1µl of diluted cDNA, 1µl of gene specific primer mix (500 nM; Integrated DNA Technologies, INC., Coralville, IA), 3 µl of nuclease free water, and 5 µl of SsoFast EvaGreen Supermix (Bio-Rad). The protocol for the qPCR runs included the following steps: initial denaturation of the cDNA and enzyme activation at 95°C for 1 min, followed by 40 cycles of 10 sec at 95°C, 10 sec at 60°C, and a fluorescent absorbance reading and one final annealing/elongation step for 5 min at 72°C. A heat dissociation curve (from 65°C to 95°C with a fluorescent absorbance reading after each 0.5° C increment) was performed at the end of every run to confirm specificity of each primer pair for the chosen transcript of interest. A standard curve was generated from six serial dilutions of a sample representing the treatment groups to calculate the amplification efficiencies of each primer set. Expression data were generated using the mathematical model developed by Pfaffl [38]. The housekeeping gene Actb was used as a reference gene for each sample because there were no statistical differences in the expression of this gene between any of the treatment groups, including DMSO (data not shown). Mean relative mRNA expression ratios from 3-4 separate follicle culture experiments were reported.

2.6. Measurement of estradiol levels

The media from follicles in each treatment group were collected at the end of the cultures, and stored at -80° C until performing measurements of estradiol levels. Randomly selected media samples per treatment group were subjected to enzyme linked immunosorbent assays (ELISA) for estradiol levels according to the manufacturer's guidelines (DRG International, Mountainside, NJ). The minimum detection limits were 9.71 pg/ml, and intra-assay and inter-assay coefficients of variation were 4.7%, and 7.8%, respectively. Individual follicle estradiol level was normalized to follicle diameter. Averaged estradiol levels represent normalized values from 12 individual follicles per treatment group obtained from three separate culture experiments.

2.7. Statistical analyses

Data were expressed as the mean \pm SEM from at least three separate experiments. Differences between DMSO and treatment groups were statistically analyzed using SPSS software (SPSS Inc., Chicago, IL). For all comparisons, statistical significance was assigned at p 0.05. Treatment group data were statistically compared to DMSO control data. When data were normally distributed and the homogeneity of variance assumption was met, we used the one-way analysis of variance (ANOVA) test followed by a Dunnett's post hoc test. When data were not normally distributed and/or when the homogeneity of variance assumption was not met, we used Kruskal-Wallis non-parametric test, followed by a Mann-Whitney test.

3. Results

3.1. Effect of BPA exposure on antral follicle growth in WT follicles (C57BL/6)

The results indicate that BPA selectively inhibited the growth of cultured WT antral follicles (Figure 1). Specifically, doses of BPA up to 43.8 μ M did not alter follicle growth compared to DMSO, whereas doses of BPA from 110 - 438 μ M significantly inhibited follicle growth, starting at 24 hours (219, 438 μ M) or 48 hours (110 μ M) and continuing throughout the culture (Figure 1).

3.2. Effect of BPA exposure on estradiol levels in WT follicles

Previous studies indicate that BPA exposure at 43.8 and 438 μ M decreases estradiol levels in mouse antral follicle cultures [9]. Therefore, we compared estradiol levels in the media from vehicle and BPA-treated follicles to expand our knowledge about potential effects of a wide range of BPA treatments on estradiol levels. The results indicate that exposure to 43.8, 110, 219, and 438 μ M BPA for 96 hours significantly decreased estradiol levels synthesized by WT antral follicles compared to WT follicles treated with DMSO (Figure 2). In contrast, exposure to 0.004, 0.04, 0.4, and 4.38 μ M BPA did not alter estradiol levels synthesized by WT antral follicles compared to WT follicles treated with DMSO (Figure 2).

3.3. Effect of BPA exposure on expression levels of Bcl2 and Bax in WT follicles

Inhibition of follicle growth and low estradiol levels can be a result of underlying atresia (programmed cell death of the ovarian cells). Further, previous studies indicate that BPA at 438 μ M causes atresia and increases expression of apoptotic factors [10]. Thus, we measured selected apoptotic factors levels at 96 hours in the current study. The results indicate that at 96 hours, 43.8 and 110 μ M BPA significantly up-regulated expression levels of the anti-apoptotic factor *Bcl2* in WT follicles compared to DMSO treated WT follicles (Figure 3). In contrast, BPA did not alter the expression of the pro-apoptotic factor *Bax* at any of the examined doses compared to DMSO (Figure 3).

3.4. Effect of BPA exposure on expression of factors in the AHR signaling pathway in WT follicles

To determine the potential role of the AHR in mediating the toxic effects of BPA, we compared relative mRNA levels of selected genes in the AHR signaling pathway (*Ahr, Ahrr, Arnt,* and *Cyp1b1*) in DMSO and BPA treated WT follicles. Because we observed inhibition of follicle growth beginning at 24 hours with BPA 219 and 438 μ M compared to control and because this inhibition continued for 96 hours, we compared expression levels at 24 and 96 hours per treatment group. After 24 hours of culture, BPA at 219 μ M significantly down-regulated *Cyp1b1* levels in WT follicles and BPA 110 μ M exposure resulted in a trend toward reducing *Cyp1b1* levels in WT follicles compared to DMSO treated WT follicles (Figure 4a). BPA at 0.4 μ M significantly down-regulated *Arnt* levels in WT follicles (Figure 4c). However, none of the selected doses of BPA altered the expression of *Ahr* or *Ahrr* in WT follicles compared to DMSO treated WT follicles (Figure 4 e and g). Similarly, at 96 hours, none of the selected doses of BPA altered the expression of *Cyp1b1*, *Ahr, Ahrr*, and *Arnt* in WT follicles compared to DMSO treated WT follicles (Figure 4 b, d, f, and h).

3.5. Effect of BPA exposure on antral follicle growth in AhrKO follicles

As a direct way to investigate the role of the AHR as a potential mediator of the toxic effects of BPA, we cultured follicles from *Ahr*KO mice with a wide range of doses of BPA and measured their growth over time. The data indicate that BPA at doses between 0.04 - 43.8 µM did not alter the growth of *AhrKO* follicles compared to DMSO treated *AhrKO* follicles (Figure 5). This growth pattern was similar to the growth of WT follicles following treatment with BPA at doses between 0.04 - 43.8 µM significantly inhibited the growth of *AhrKO* follicles compared to *AhrKO* follicles treated with DMSO starting at 24 hours, and continuing throughout the culture time (Figure 5). Again, these results are similar to what was observed with the WT follicles at the same doses (Figure 1). Lastly, BPA at 110 µM inhibited the growth of *AhrKO* follicles compared to *AhrKO* follicles treated with DMSO at 48 hours only (Figure 5). Interestingly, BPA at 110 µM did not significantly alter the growth of *AhrKO* follicles compared to *AhrKO* follicles treated with DMSO at any of the other time points (Figure 5). These results are different than what we observed with WT follicles treated with BPA 110 µM, in which inhibition of growth started at 48 hours, but lasted throughout the culture time (Figure 1).

3.6. Effect of BPA exposure on estradiol levels in AhrKO follicles

Previous studies indicate that BPA exposure at 43.8 and 438 μ M decreases estradiol levels in mouse antral follicle cultures [9], but the mechanism by which BPA leads to decreased estradiol synthesis is unknown. Therefore, we measured estradiol levels in the media from vehicle and BPA-treated *Ahr*KO follicles to expand our knowledge about the potential role of the AHR signaling pathway in mediating BPA toxic effects on estradiol synthesis. The results indicate that 96 hours exposure to BPA at 43.8, 110, 219, and 438 μ M significantly decreased estradiol levels synthesized by *Ahr*KO antral follicles compared to *Ahr*KO follicles treated with DMSO (Figure 6). In contrast, exposure to BPA at 4.38 μ M did not alter estradiol levels synthesized by *Ahr*KO antral follicles compared to *Ahr*KO follicles treated with DMSO (Figure 6). These results are similar to the results of WT follicles treated with these doses of BPA (Figure 2).

4. Discussion

The current study expands on the limited published data regarding the direct toxic effects of a wide dose range of BPA on cultured mouse antral follicles. It also provides information on the potential mechanism by which BPA may exert its toxic effects in antral follicles. In this study, we utilized a follicle culture system to closely examine the direct effects of exposure to a wide range of doses of BPA, including environmentally relevant doses, on antral follicles [2, 10, 11, 34]. We also compared the growth of antral follicles isolated from WT and AhrKO mice to examine the role of the AHR in mediating the toxic effects of BPA. The results suggest that BPA inhibits the growth and decreases estradiol levels of WT follicles in a dose dependent manner. Furthermore, the data suggest that BPA at 110 µM may selectively act through the AHR signaling pathway to exert its toxic effects. The maximum tolerated dose for BPA is considered to be 1,000 mg/kg/body weight/day and the calculated reference dose according to the U.S. EPA was determined to be 50 µg/kg/day based on the LOAEL. Though there is not much known yet on the pharmacokinetics of BPA in mice or humans and it is difficult to extrapolate from LOAELs determined in vivo to in vitro studies, our study adds valuable data on the effects of BPA exposure at levels lower than the LOAEL and the reference dose.

In our study, BPA inhibited the growth of WT follicles at doses of 110, 219, and 438 μ M, but not at doses of 43.8 μ M or less. These findings indicate that in our antral follicle bioassay, BPA has a linear dose-response effect on follicular growth. This finding is in

contrast to our original hypothesis that BPA exhibits non-monotonic effects on follicles as it does in other cell lines/tissues such as pituitary cell lines [39] and adipose explants [14]. The reasons for different dose-response effects with BPA are unknown, but studies have shown that BPA can act differently in different tissues and this appears to be the case with ovarian follicle cultures versus pituitary cell lines and adipose explants [14, 39]. Yet, our findings are in agreement with other studies that examined the effects of BPA on mouse ovarian follicles. Xu et al. [8] observed that BPA in a dose dependent manner decreases viability of mouse cultured granulosa cells. Specifically, treatment with doses of 0.1 nM, 0.1 µM, and 100 µM of BPA resulted in cytotoxic effects [8]. Moreover, the toxic effects of BPA at 100 µM (which is close to our BPA 110 µM dose) were observed as early as 24 hours [8]. Further, according to Lenie et al. [11], BPA exposure at 3 nM - 30 µM (doses similar to our doses of BPA 0.004 - 43.8 µM) did not affect early preantral follicle growth compared to control after 4 days of culture. Peretz et al. [9] reported that BPA at 438 µM inhibited follicle growth as it did in our current study. Interestingly, BPA at 438 µM inhibited follicle growth starting at 72 hours of culture in the Peretz et al. study [9], while it inhibited follicle growth beginning at 24 hours in the current study. This difference in the timing of BPA effects could be due to differences in strains of mice and animal ages in Peretz et al. [9] compared to our study. Peretz et al. [9] used FVB mice, whereas we used C57BL/6 mice. Additionally, Peretz et al. [9] used 32 day old mice, whereas we used 50-54 day old mice.

Despite differences in the timing of inhibition of growth between our current study and the study by Peretz et al. [9], both studies showed that BPA treatment decreases estradiol levels. Estradiol is one of the main factors needed to stimulate follicle growth and to protect follicles from atresia [40]. Thus, decreased estradiol levels are another indication of impaired function of the follicles that can be observed even when follicle growth is not inhibited. In the current study, after 96 hours of culture, BPA at 43.8 - 438 µM significantly decreased estradiol levels in a dose dependent manner (Figure 2). Specifically, BPA at 43.8 µM significantly decreased estradiol levels, even though it did not affect follicle growth compared to controls, whereas BPA at 110, 219, and 438 µM decreased both follicle growth and estradiol levels (Figures 1,2). These findings are in agreement with the results by Peretz et al. [9], who reported that BPA at 43.8 and 438 µM significantly decreased estradiol levels compared to controls after 120 hours of culture. These data may suggest that exposure to BPA at 43.8, 110, 219, and 438 µM pushes the follicles into cellular stress that results with impaired steroidogenesis followed by inhibition of growth with higher doses of BPA (110, 219, and 438 µM). BPA at 43.8 µM still allows the follicles to recover their growth, but not estradiol synthesis.

Estradiol can protect follicles from atresia [40] by suppressing the transcription of proapoptotic factors such as Bax [41]. Previous studies also indicate that overexpression of the anti-apoptotic factor Bcl2 enhances follicular growth and decreases apoptosis [42]. Studies also indicate that the balance between pro- and anti-apoptotic factors is a main determinant in whether the process of atresia will be initiated in follicles [43]. In our current study, Bcl2 levels (Figure 3) were significantly increased at BPA 43.8 and 110 µM, whereas the proapoptotic factor Bax levels were not altered (Figure 3) with any of the examined BPA treatments compared to controls. Therefore, it is likely that BPA-induced changes in the ratio of pro- and anti-apoptotic factors to favor an anti-apoptotic state may be an attempt of the follicles to restore normal growth following treatment with BPA 43.8 and 110 µM. This attempt is not completely successful because follicles were only able to retain normal growth at BPA 43.8 µM, but not at the higher tested doses. At higher doses, it may be that the follicular damage caused by BPA (110 - 438 μ M) was too profound to allow the follicles to be rescued or even partially protected by increasing expression of anti-apoptotic factors such as Bcl2. Xu et al. [8] have shown that BPA can induce apoptosis in isolated granulosa cells at concentration as low as 100 pM [8]. Therefore the difference in doses required to

cause apoptosis in our study versus that by Xu *et al.* [8] can be due to the fact that we cultured whole follicles, whereas Xu *et al.* [8] cultured isolated granulosa cells. Interestingly, Peretz *et al.* [10] measured expression levels of *Bax* and *Bcl2* in isolated antral follicles cultured with doses of BPA at 4.38, 43.8, and 438 μ M from 32-35 day old FVB mice. Their results indicate that at 96 hours, BPA at 438 μ M treatment significantly increased expression of both *Bax* and *Bcl2*. Again, this difference between the study by Peretz *et al.* [10] and the current study could be due to difference in the age and strain of mice that were used in both studies.

The AHR signaling pathway plays a major role in regulating follicle growth and estradiol levels in the ovary and in mediating the toxic effects of other environmental chemicals (e.g., dioxins, polycyclic aromatic hydrocarbons) [20-22]. A few studies have examined the AHR as a potential mediator of the toxic effects of BPA [26, 27, 44-46]. However, only Nishizawa *et al.* [26] have examined the AHR-BPA relationship in the gonads. The researchers observed elevated gonadal *Ahr* expression following in utero exposure to BPA. Therefore, we investigated expression levels of major factors in the AHR signaling pathway following BPA exposure in our isolated follicle culture system. Interestingly, there were no striking differences in the expression of major factors in the AHR signaling pathway in WT follicles treated with BPA versus WT follicles treated with DMSO. These results are similar to what has been observed following treatment of cultured mouse antral follicles even with the potent ligand of the AHR TCDD, in which the *Ahr* expression in TCDD treated follicles was not different than controls [47].

In the absence of the *Ahr*, however, antral follicles may have been rescued from BPA (110 µM) induced inhibition of growth. This recovery occurs only at BPA 110 µM and only starts at 72 hours. Interestingly, the rescue of follicular growth was not observed at BPA 219 or 438 µM. This may be because different doses of BPA may act through different signaling pathways to exert its toxic effects. It may also be that higher doses of BPA are cytotoxic to the follicles and that the follicles cannot be rescued from BPA exposure at higher doses even when removing the Ahr. These results are in agreement with other studies that have shown that disruption of the normal activity of the AHR reduces the toxic effects of other chemicals in the ovary. For example, -napthoflavone, a known AHR inhibitor, reduces the toxic effects of benzo(a)pyrene, 3-methylcholanthrene, and 7,12-dimethylbenz(a)anthracene on the ovary [48-50], whereas TCDD, a potent AHR agonist, decreases tumorgenic cell proliferation [51]. Further, Basavarajappa et al. [47], who used a similar bioassay to the one that was used in our current study reported that AhrKO follicles treated with the pesticide methoxychlor are also partially protected from inhibition of growth. These findings suggest that the AHR differentially modulates the toxic effects on follicular growth caused by BPA as well as other chemicals such as methoxychlor.

Lastly, unlike the case with follicular growth, BPA treatment of *Ahr*KO follicles resulted in similar estradiol levels to what we observed with WT follicles treated with BPA. These data are highly suggestive that the AHR signaling pathway is not a major route through which BPA exerts its toxic effects on the steroidogenic capacity of the follicles. These results add to a number of publications on the other potential targets through which BPA exerts its toxic effects on steroidogenesis. For example, Peretz *et al.* [9] concluded that BPA treatment of antral follicles may interfere with steroidogenesis by inhibiting cholesterol uptake and metabolism by decreasing the expression of the rate limiting factors steroidogenic acute regulatory protein (*StAR*) and cytochrome P450 cholesterol side chain cleavage (*Cyp11a1*). Similar conclusions were also published by other researchers in other species [52]. Our results indicate that it is unlikely that any of these effects of BPA on steroidogenesis are mediated by the AHR.

Overall findings from the current study expand our knowledge regarding the toxic effects of a wide range of BPA doses in ovarian follicles. The examined range of doses represents both human exposure and environmentally relevant levels. Results from our study indicate that BPA in a linear dose-response manner decreases in vitro antral follicle growth and decreases synthesis of estradiol. Interestingly, BPA (0.004 and 0.04 μ M) at doses similar to the levels measured in women's ovarian follicular fluid undergoing in vitro fertilization treatments [2] did not affect follicular growth or estradiol synthesis compared to the control group. BPA also alters expression levels of *Bcl2*, a factor that is involved in the process of follicular atresia. However, BPA treatment did not dramatically alter genes in the AHR signaling pathway. BPA-induced inhibition of follicle growth, however, was partially rescued in the absence of the *Ahr*, indicating that the AHR signaling pathway may be partially involved in mediating the toxic effects of BPA on the growth of cultured mouse antral follicles. Further studies, however, are needed to better understand the mechanism by which the AHR signaling pathway mediates the toxic effects of BPA (110 μ M).

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References

- Cao XL, Perez-Locas C, Dufresne G, Clement G, Popovic S, Beraldin F, et al. Concentrations of bisphenol A in the composite food samples from the 2008 Canadian total diet study in Quebec City and dietary intake estimates. Food Addit Contam Part A Chem Anal Control Expo Risk Assess. 2011; 28:791–8. [PubMed: 21623504]
- [2]. Ikezuki Y, Tsutsumi O, Takai Y, Kamei Y, Taketani Y. Determination of bisphenol A concentrations in human biological fluids reveals significant early prenatal exposure. Hum Reprod. 2002; 17:2839–41. [PubMed: 12407035]
- [3]. Craig ZR, Wang W, Flaws JA. Endocrine-disrupting chemicals in ovarian function: effects on steroidogenesis, metabolism and nuclear receptor signaling. Reproduction. 2011; 142:633–46. [PubMed: 21862696]
- [4]. Ehrlich S, Williams PL, Missmer SA, Flaws JA, Berry KF, Calafat AM, et al. Urinary bisphenol A concentrations and implantation failure among women undergoing in vitro fertilization. Environ Health Perspect. 2012; 120:978–83. [PubMed: 22484414]
- [5]. Ehrlich S, Williams PL, Missmer SA, Flaws JA, Ye X, Calafat AM, et al. Urinary bisphenol A concentrations and early reproductive health outcomes among women undergoing IVF. Hum Reprod. 2012; 27:3583–92. [PubMed: 23014629]
- [6]. Takeuchi T, Tsutsumi O, Ikezuki Y, Takai Y, Taketani Y. Positive relationship between androgen and the endocrine disruptor, bisphenol A, in normal women and women with ovarian dysfunction. Endocr J. 2004; 51:165–9. [PubMed: 15118266]
- [7]. Suzuki A, Sugihara A, Uchida K, Sato T, Ohta Y, Katsu Y, et al. Developmental effects of perinatal exposure to bisphenol-A and diethylstilbestrol on reproductive organs in female mice. Reprod Toxicol. 2002; 16:107–16. [PubMed: 11955941]
- [8]. Xu J, Osuga Y, Yano T, Morita Y, Tang X, Fujiwara T, et al. Bisphenol A induces apoptosis and G2-to-M arrest of ovarian granulosa cells. Biochem Biophys Res Commun. 2002; 292:456–62. [PubMed: 11906184]
- [9]. Peretz J, Gupta RK, Singh J, Hernandez-Ochoa I, Flaws JA. Bisphenol A impairs follicle growth, inhibits steroidogenesis, and downregulates rate-limiting enzymes in the estradiol biosynthesis pathway. Toxicol Sci. 2011; 119:209–17. [PubMed: 20956811]

- [10]. Peretz J, Craig ZR, Flaws JA. Bisphenol A inhibits follicle growth and induces atresia in cultured mouse antral follicles independently of the genomic estrogenic pathway. Biol Reprod. 2012; 87:63. [PubMed: 22743301]
- [11]. Lenie S, Cortvrindt R, Eichenlaub-Ritter U, Smitz J. Continuous exposure to bisphenol A during in vitro follicular development induces meiotic abnormalities. Mutat Res. 2008; 651:71–81.
 [PubMed: 18093867]
- [12]. Vandenberg LN, Maffini MV, Sonnenschein C, Rubin BS, Soto AM. Bisphenol-A and the great divide: a review of controversies in the field of endocrine disruption. Endocr Rev. 2009; 30:75– 95. [PubMed: 19074586]
- [13]. Sengupta S, Obiorah I, Maximov P, Curpan R, Jordan V. Molecular mechanism of action of bisphenol and bisphenol A mediated by oestrogen receptor alpha in growth and apoptosis of breast cancer cells. Br J Pharmacol. 2013; 169:167–78. [PubMed: 23373633]
- [14]. Hugo ER, Brandebourg TD, Woo JG, Loftus J, Alexander JW, Ben Jonathan N. Bisphenol A at environmentally relevant doses inhibits adiponectin release from human adipose tissue explants and adipocytes. Environ Health Perspect. 2008; 116:1642–7. [PubMed: 19079714]
- [15]. Vinas R, Watson CS. Bisphenol S disrupts estradiol-induced nongenomic signaling in a rat pituitary cell line: effects on cell functions. Environ Health Perspect. 2013; 121:352–8. [PubMed: 23458715]
- [16]. Teng C, Goodwin B, Shockley K, Xia M, Huang R, Norris J, et al. Bisphenol A affects androgen receptor function via multiple mechanisms. Chem Biol Interact. 2013
- [17]. Moriyama K, Tagami T, Akamizu T, Usui T, Saijo M, Kanamoto N, et al. Thyroid hormone action is disrupted by bisphenol A as an antagonist. J Clin Endocrinol Metab. 2002; 87:5185–90.
 [PubMed: 12414890]
- [18]. Rubin BS. Bisphenol A: an endocrine disruptor with widespread exposure and multiple effects. J Steroid Biochem Mol Biol. 2011; 127:27–34. [PubMed: 21605673]
- [19]. Poland A, Knutson JC. 2,3,7,8-tetrachlorodibenzo-p-dioxin and related halogenated aromatic hydrocarbons: examination of the mechanism of toxicity. Annu Rev Pharmacol Toxicol. 1982; 22:517–54. [PubMed: 6282188]
- [20]. Puga A, Ma C, Marlowe JL. The aryl hydrocarbon receptor cross-talks with multiple signal transduction pathways. Biochem Pharmacol. 2009; 77:713–22. [PubMed: 18817753]
- [21]. Abel J, Haarmann-Stemmann T. An introduction to the molecular basics of aryl hydrocarbon receptor biology. Biol Chem. 2010; 391:1235–48. [PubMed: 20868221]
- [22]. Pollenz RS. The mechanism of AH receptor protein down-regulation (degradation) and its impact on AH receptor-mediated gene regulation. Chem Biol Interact. 2002; 141:41–61. [PubMed: 12213384]
- [23]. Hernandez-Ochoa I, Karman BN, Flaws JA. The role of the aryl hydrocarbon receptor in the female reproductive system. Biochem Pharmacol. 2009; 77:547–59. [PubMed: 18977336]
- [24]. Bussmann UA, Bussmann LE, Baranao JL. An aryl hydrocarbon receptor agonist amplifies the mitogenic actions of estradiol in granulosa cells: evidence of involvement of the cognate receptors. Biol Reprod. 2006; 74:417–26. [PubMed: 16237154]
- [25]. Safe S, Wormke M. Inhibitory aryl hydrocarbon receptor-estrogen receptor alpha cross-talk and mechanisms of action. Chem Res Toxicol. 2003; 16:807–16. [PubMed: 12870882]
- [26]. Nishizawa H, Morita M, Sugimoto M, Imanishi S, Manabe N. Effects of in utero exposure to bisphenol A on mRNA expression of arylhydrocarbon and retinoid receptors in murine embryos. J Reprod Dev. 2005; 51:315–24. [PubMed: 15738621]
- [27]. Bonefeld-Jorgensen EC, Long M, Hofmeister MV, Vinggaard AM. Endocrine-disrupting potential of bisphenol A, bisphenol A dimethacrylate, 4-n-nonylphenol, and 4-n-octylphenol in vitro: new data and a brief review. Environ Health Perspect. 2007; 115(Suppl 1):69–76. [PubMed: 18174953]
- [28]. Swedenborg E, Pongratz I. AhR and ARNT modulate ER signaling. Toxicology. 2010; 268:132– 8. [PubMed: 19778576]
- [29]. Karman BN, Basavarajappa MS, Craig ZR, Flaws JA. 2,3,7,8-Tetrachlorodibenzo-p-dioxin activates the aryl hydrocarbon receptor and alters sex steroid hormone secretion without affecting

growth of mouse antral follicles in vitro. Toxicol Appl Pharmacol. 2012; 261:88–96. [PubMed: 22483799]

- [30]. Flint S, Markle T, Thompson S, Wallace E. Bisphenol A exposure, effects, and policy: a wildlife perspective. J Environ Manage. 2012; 104:19–34. [PubMed: 22481365]
- [31]. Vandenberg LN, Chahoud I, Heindel JJ, Padmanabhan V, Paumgartten FJ, Schoenfelder G. Urinary, circulating, and tissue biomonitoring studies indicate widespread exposure to bisphenol A. Cien Saude Colet. 2012; 17:407–34. [PubMed: 22267036]
- [32]. Schmidt JV, Su GH, Reddy JK, Simon MC, Bradfield CA. Characterization of a murine Ahr null allele: involvement of the Ah receptor in hepatic growth and development. Proc Natl Acad Sci U S A. 1996; 93:6731–6. [PubMed: 8692887]
- [33]. Benedict JC, Lin TM, Loeffler IK, Peterson RE, Flaws JA. Physiological role of the aryl hydrocarbon receptor in mouse ovary development. Toxicol Sci. 2000; 56:382–8. [PubMed: 10910997]
- [34]. Hernandez-Ochoa I, Barnett-Ringgold KR, Dehlinger SL, Gupta RK, Leslie TC, Roby KF, et al. The ability of the aryl hydrocarbon receptor to regulate ovarian follicle growth and estradiol biosynthesis in mice depends on stage of sexual maturity. Biol Reprod. 2010; 83:698–706. [PubMed: 20631400]
- [35]. Miller KP, Gupta RK, Greenfeld CR, Babus JK, Flaws JA. Methoxychlor directly affects ovarian antral follicle growth and atresia through Bcl-2- and Bax-mediated pathways. Toxicol Sci. 2005; 88:213–21. [PubMed: 16081520]
- [36]. Abbott BD, Schmid JE, Brown JG, Wood CR, White RD, Buckalew AR, et al. RT-PCR quantification of AHR, ARNT, GR, and CYP1A1 mRNA in craniofacial tissues of embryonic mice exposed to 2,3,7,8-tetrachlorodibenzo-p-dioxin and hydrocortisone. Toxicol Sci. 1999; 47:76–85. [PubMed: 10048155]
- [37]. Craig ZR, Hannon PR, Wang W, Ziv-Gal A, Flaws JA. Di-n-butyl phthalate disrupts the expression of genes involved in cell cycle and apoptotic pathways in mouse ovarian antral follicles. Biol Reprod. 2013; 88:23. [PubMed: 23242528]
- [38]. Pfaffl MW. A new mathematical model for relative quantification in real-time RT-PCR. Nucleic Acids Res. 2001; 29:e45. [PubMed: 11328886]
- [39]. Wozniak AL, Bulayeva NN, Watson CS. Xenoestrogens at picomolar to nanomolar concentrations trigger membrane estrogen receptor-alpha-mediated Ca2+ fluxes and prolactin release in GH3/B6 pituitary tumor cells. Environ Health Perspect. 2005; 113:431–9. [PubMed: 15811834]
- [40]. Quirk SM, Cowan RG, Harman RM, Hu CL, Porter DA. Ovarian follicular growth and atresia: the relationship between cell proliferation and survival. J Anim Sci. 2004; 82(E-Suppl):E40–E52. [PubMed: 15471814]
- [41]. Matsuda F, Inoue N, Manabe N, Ohkura S. Follicular growth and atresia in mammalian ovaries: regulation by survival and death of granulosa cells. J Reprod Dev. 2012; 58:44–50. [PubMed: 22450284]
- [42]. Hsu SY, Lai RJ, Finegold M, Hsueh AJ. Targeted overexpression of Bcl-2 in ovaries of transgenic mice leads to decreased follicle apoptosis, enhanced folliculogenesis, and increased germ cell tumorigenesis. Endocrinology. 1996; 137:4837–43. [PubMed: 8895354]
- [43]. Riou C, Tonoli H, Bernier-Valentin F, Rabilloud R, Fonlupt P, Rousset B. Susceptibility of differentiated thyrocytes in primary culture to undergo apoptosis after exposure to hydrogen peroxide: relation with the level of expression of apoptosis regulatory proteins, Bcl-2 and Bax. Endocrinology. 1999; 140:1990–7. [PubMed: 10218946]
- [44]. Nishizawa H, Imanishi S, Manabe N. Effects of exposure in utero to bisphenol a on the expression of aryl hydrocarbon receptor, related factors, and xenobiotic metabolizing enzymes in murine embryos. J Reprod Dev. 2005; 51:593–605. [PubMed: 16284450]
- [45]. Kruger T, Long M, Bonefeld-Jorgensen EC. Plastic components affect the activation of the aryl hydrocarbon and the androgen receptor. Toxicology. 2008; 246:112–23. [PubMed: 18294747]
- [46]. Imanishi S, Manabe N, Nishizawa H, Morita M, Sugimoto M, Iwahori M, et al. Effects of oral exposure of bisphenol A on mRNA expression of nuclear receptors in murine placentae assessed by DNA microarray. J Reprod Dev. 2003; 49:329–36. [PubMed: 14967926]

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- [47]. Basavarajappa MS, Hernandez-Ochoa I, Wang W, Flaws JA. Methoxychlor inhibits growth and induces atresia through the aryl hydrocarbon receptor pathway in mouse ovarian antral follicles. Reprod Toxicol. 2012; 34:16–21. [PubMed: 22484361]
- [48]. Cooper AR, Moley KH. Maternal tobacco use and its preimplantation effects on fertility: more reasons to stop smoking. Semin Reprod Med. 2008; 26:204–12. [PubMed: 18302112]
- [49]. Shiromizu K, Mattison DR. Murine oocyte destruction following intraovarian treatment with 3methylcholanthrene or 7,12-dimethylbenz(a)anthracene: protection by alpha-naphthoflavone. Teratog Carcinog Mutagen. 1985; 5:463–72. [PubMed: 2874631]
- [50]. Shiromizu K, Mattison DR. The effect of intraovarian injection of benzo(a)pyrene on primordial oocyte number and ovarian aryl hydrocarbon [benzo(a)pyrene] hydroxylase activity. Toxicol Appl Pharmacol. 1984; 76:18–25. [PubMed: 6484988]
- [51]. Holcomb M, Safe S. Inhibition of 7,12-dimethylbenzanthracene-induced rat mammary tumor growth by 2,3,7,8-tetrachlorodibenzo-p-dioxin. Cancer Lett. 1994; 82:43–7. [PubMed: 8033067]
- [52]. Mlynarcikova A, Kolena J, Fickova M, Scsukova S. Alterations in steroid hormone production by porcine ovarian granulosa cells caused by bisphenol A and bisphenol A dimethacrylate. Mol Cell Endocrinol. 2005; 244:57–62. [PubMed: 16225985]

HIGHLIGHTS

BPA (110-438 μ M) inhibits mouse ovarian antral follicle growth.

BPA (43-438 μ M) decreases mouse ovarian antral follicle estradiol levels.

BPA may inhibit ovarian follicle growth partially via the AHR signaling pathway.

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Figure 1.

Effect of in vitro BPA exposure on wild-type mouse antral follicle growth. Antral follicles mechanically isolated from wild-type mice were individually cultured with vehicle control (DMSO) or BPA (0.004 - 438 μ M) for 96 hours. During the culture, follicle diameters were measured and then were converted to percent change from baseline (time 0). Graph represents means ± SEM from 3-4 separate experiments (at least five follicles per treatment per replicate). Asterisks represent significant differences from DMSO per time point; *p* 0.05.

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Figure 2.

Effect of in vitro BPA exposure on estradiol synthesis by wild-type mouse antral follicles. Antral follicles mechanically isolated from wild-type mice were individually cultured with vehicle control (DMSO) or BPA (0.004 - 438 μ M). After 96 hours, the media were collected and analyzed for estradiol levels (at least three follicles per treatment, per experiment). Graph represents means ± SEM from three separate experiments. Asterisks represent significant differences from DMSO; *p* 0.05.

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Treatment

Figure 3.

Effect of 96 hours in vitro BPA exposure on expression of *Bcl2* and *Bax*. Antral follicles isolated from wild-type mice were treated with vehicle control (DMSO) or BPA (0.004 - 438 μ M) and cultured for 96 hours. Expression levels of *Bcl2* and *Bax* were measured using qPCR. Data represent means ± SEM from at least three separate experiments. Asterisks represent significant differences from DMSO; p 0.05.

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Figure 4.

Effect of in vitro BPA exposure on expression of major factors in the AHR signaling pathway, following 24 and 96 hours of culture. Antral follicles isolated from wild-type mice were treated with vehicle control (DMSO) or BPA (0.004 - 438 μ M) and cultured for 24 or 96 hours. Expression levels of *Cyp1b1* (a ,b), *Arnt* (c, d), *Ahr* (e, f), and *Ahrr* (g, h) were measured using qPCR. Data represent means ± SEM from at least three separate experiments. Asterisks represent significant differences from DMSO; *p* 0.05.

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Figure 5.

Effect of in vitro BPA exposure on *Ahr*KO mouse antral follicle growth. Antral follicles mechanically isolated from *Ahr*KO mice were individually cultured with vehicle control (DMSO) or BPA (0.004 - 438 μ M) for 96 hours. During the culture, follicle diameters were measured and then were converted to percent change from baseline (time 0). Graph represents means ± SEM from 3-4 separate experiments (at least five follicles per treatment per experiment). Asterisks represent significant differences from DMSO per time point; *p* 0.05.

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Figure 6.

Effect of in vitro BPA exposure on estradiol synthesis by *Ahr*KO mouse antral follicles. Antral follicles mechanically isolated from *Ahr*KO mice were individually cultured with vehicle control (DMSO) or BPA (4.38 - 438 μ M). After 96 hours, the media were collected and analyzed for estradiol levels. Graph represents means ± SEM from three separate experiments (four follicles per treatment, per experiment). Asterisks represent significant differences from DMSO; *p* 0.05.