



Bacteria and Genes Involved in Arsenic Speciation in Sediment Impacted by Long-Term Gold Mining

Patrícia S. Costa¹, Larissa L. S. Scholte², Mariana P. Reis¹, Anderson V. Chaves¹, Pollyanna L. Oliveira¹, Luiza B. Itabayana¹, Maria Luiza S. Suhadolnik¹, Francisco A. R. Barbosa¹, Edmar Chartone-Souza¹, Andréa M. A. Nascimento^{1*}

1 Departamento de Biologia Geral, Instituto de Ciências Biológicas, Universidade Federal de Minas Gerais; Belo Horizonte, Brazil, **2** Grupo de Genômica e Biologia Computacional, Centro de Pesquisas René Rachou (CPqRR), Fundação Oswaldo Cruz (FIOCRUZ), Belo Horizonte, Brazil

Abstract

The bacterial community and genes involved in geobiocycling of arsenic (As) from sediment impacted by long-term gold mining were characterized through culture-based analysis of As-transforming bacteria and metagenomic studies of the *arsC*, *arrA*, and *aioA* genes. Sediment was collected from the historically gold mining impacted Mina stream, located in one of the world's largest mining regions known as the "Iron Quadrangle". A total of 123 As-resistant bacteria were recovered from the enrichment cultures, which were phenotypically and genotypically characterized for As-transformation. A diverse As-resistant bacteria community was found through phylogenetic analyses of the 16S rRNA gene. Bacterial isolates were affiliated with *Proteobacteria*, *Firmicutes*, and *Actinobacteria* and were represented by 20 genera. Most were AsV-reducing (72%), whereas AsIII-oxidizing accounted for 20%. Bacteria harboring the *arsC* gene predominated (85%), followed by *aioA* (20%) and *arrA* (7%). Additionally, we identified two novel As-transforming genera, *Thermomonas* and *Pannonibacter*. Metagenomic analysis of *arsC*, *aioA*, and *arrA* sequences confirmed the presence of these genes, with *arrA* sequences being more closely related to uncultured organisms. Evolutionary analyses revealed high genetic similarity between some *arsC* and *aioA* sequences obtained from isolates and clone libraries, suggesting that those isolates may represent environmentally important bacteria acting in As speciation. In addition, our findings show that the diversity of *arrA* genes is wider than earlier described, once none *arrA*-OTUs were affiliated with known reference strains. Therefore, the molecular diversity of *arrA* genes is far from being fully explored deserving further attention.

Citation: Costa PS, Scholte LLS, Reis MP, Chaves AV, Oliveira PL, et al. (2014) Bacteria and Genes Involved in Arsenic Speciation in Sediment Impacted by Long-Term Gold Mining. PLoS ONE 9(4): e95655. doi:10.1371/journal.pone.0095655

Editor: Celine Brochier-Armanet, Université Claude Bernard - Lyon 1, France

Received: June 3, 2013; **Accepted:** March 31, 2014; **Published:** April 22, 2014

Copyright: © 2014 Costa et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Funding: Sources of funding: FAPEMIG APQ 00801/12, CNPq n°472411/2012-8, CNPq/INCT no 15206-7. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* E-mail: amaral@ufmg.br

Introduction

Arsenic occurs naturally in the earth's crust and is widely distributed in the environment [1,2]. Natural mineralization and microorganisms enhance arsenic mobilization in the environment, but human interventions, such as gold mining, have aggravated the environmental arsenic contamination arousing health concerns. Water pollution by arsenic is one of the major challenges for public health, primarily due to its carcinogenic potential at low doses [3,4,5,6]. According to Nordstrom [7] over 50 million people in the world are at risk from drinking arsenic-contaminated water. Moreover, given that arsenic has a variety of valence states (+V, +III, 0, -III) with different physicochemical properties, the removal of arsenic from contaminated water bodies is yet a challenge.

In nature, microorganisms have developed different response mechanisms to metabolize As, mainly via reduction and oxidation reactions, leading to its speciation [8]. Previous studies have regarded As speciation as a result of microbial activity in the environment, including some derived from gold-mining activities [1,2]. However, few bacterial genera involved in As transformation have been found at any of the sites studied [9–13]. Thus, a

more comprehensive knowledge on the structure of the bacterial community involved in As-transformation in gold-mining sites remains warranted.

The arsenate (AsV) reducing pathways known are the detoxification (*arsC* gene) and the dissimilatory respiration (*arrA/B* genes). The organization of *ars* operons varies greatly between taxa, and the core genes include *arsR*, *arsB* and *arsC*, whereas *arsD* and *arsA* genes can eventually be found [1]. The *arsC* gene encodes the enzyme AsV reductase, which is located in the cytoplasm and is responsible for the biotransformation of AsV to AsIII. This enzyme together with a transmembrane efflux pump, encoded by *arsA* and *arsB* genes, is the most common As transformation mechanism in the environment [2,14–16]. Moreover, *arrA/B* genes encode a periplasmic AsV reductase that works during anaerobic respiration using AsV as the final electron acceptor for energy generation [17]. The AsV dissimilatory respiration reduction has already been described for many bacterial phyla, including obligatory and facultative anaerobic bacteria and some archaea [1].

The microbial oxidation of AsIII was first reported in 1918 and can be mediated by two distinct enzymes: AioBA, hardly studied, and ArxAB, recently described by Zargar et al. [18]. Both enzymes

have been found in several heterotrophic and chemolithoautotrophic bacterial species [19–22]. Aerobic AsIII oxidation is catalyzed by arsenite oxidase, which uses O₂ as terminal electron acceptor, and is encoded by *aioB/A* genes, formerly referred to as *aoxA/B*, *aroB/A* and *asoB/A* genes [20,23]. ArxAB detected in AsIII oxidizing bacteria in anoxic conditions, in which nitrate or chlorate reduction is coupled to AsIII oxidation in the chemolithotrophs [24,25]. Interestingly, members of the genus *Ectothiorhodospira* are able to use AsIII as electron donor for anoxygenic phototrophic growth [26]. According to Zargar et al. [18] the *arxA* gene is more closely related to *arrA* than to *aioA* genes.

In this research, we bioprospected As-resistant bacteria from As-enrichment culture of sediments collected from a stream located at the Brazilian gold mining area known as the Iron Quadrangle (IQ, Minas Gerais state), one of the world's largest mining regions. Much concern exists about As-contamination of gold-mining sites in this area because it is estimated that at least 390,000 tons of As have been released into this area since the beginning of gold-mining activity in the 17th century [27]. We also investigated the diversity of As-transforming genes using metagenomic strategies. This included the genes for arsenite oxidase (*aioA*) and arsenate reductases (*arsC* and *arrA*).

Materials and Methods

Ethics Statement

For sampling in Mina stream, no specific permit was required for the described field study. The study location is not privately owned or protected in any way and we confirm that the field study did not involve endangered or protected species.

Study Area and Sampling

Mina stream (19°58'46.80"S–43°49'17.07"W) is a natural body of water located at the Velhas River Basin (IQ, Minas Gerais state, Brazil) and characterized as backwater (Figure S1). This stream was chosen because is located near a historically impacted gold-mining area. Moreover, previous investigations [28] reported As concentrations superior to those permitted by Brazilian law (Conselho Nacional do Meio Ambiente – CONAMA) and by Canadian Environmental Quality Guidelines (Canadian Council of Ministers of the Environment– CCME).

Bulk water and superficial sediment samples (up to 1.0 cm depth) were collected on 13 July 2011, during the dry season. The typical sediment core can be divided into three zones: oxic, suboxic and anoxic [29]. According to literature the thick oxic zone can extend from several mm up to 10 cm [30,31]. In this work the sampling site was shallow (20 cm) and therefore highly influenced by the nutrients and oxygen concentrations of the water body. The analyzed sediment was taken from the upper part, representing the oxic zone. Samples were collected aseptically at three points at 1m distance from each other, subsequently pooled in a single sample, and stored at 4°C for bacterial analysis or at –20°C for chemical and molecular analyses.

To assess the bulk water conditions physicochemical characteristics such as temperature, pH, and dissolved oxygen (DO) concentration were measured *in situ* with a multiprobe (Horiba, model U-22) [30]. Concentrations of total nitrogen (TN), total phosphorus (TP), ammonium (NH₄⁺-N), nitrite (NO₂-N), nitrate (NO₃-N), and soluble reactive phosphorus (PO₄-P) were measured as previously described [32,33]. Metal and metalloid concentrations of water and sediment samples were determined by using an inductively coupled plasma-optical emission spectrometer (ICP-OES, Optima 7300 DV, PerkinElmer).

Arsenic Enrichment and Isolation

Sediment (10 g) samples were added to Erlenmeyer flasks containing 100 mL of CDM medium (0.012 mM Fe₂SO₄, 7 mM Na₂SO₄, 0.0574 mM K₂HPO₄, 9.5 mM NaHCO₃, 18.7 mM NH₄Cl, 8.12 mM MgSO₄, 0.457 mM CaCl₂ and 44.6 mM sodium lactate as organic carbon source, pH 7.2) with either 2 mM sodium arsenite or 10 mM sodium arsenate and incubated at 28°C for seven days. Then, serial 10-fold dilutions of the enrichment cultures were plated onto CDM agar media (1.5% agar) amended with 2 mM sodium arsenite or 10 mM sodium arsenate to selectively enrich and isolate AsIII- and AsV-resistant bacteria. Plates were incubated at 28°C for five days. The resulting colonies were repeatedly streaked on the same medium to accomplish their purification. The bacterial isolates from AsIII- and AsV-resistant bacteria (named MS-AsIII and MS-AsV, respectively) were stored at –20°C in 25% glycerol.

DNA Extraction from the Cultures and Sediment

Genomic DNA was extracted and purified from each MS-AsIII and MS-AsV isolate using a protocol previously described [34]. Additionally, metagenomic DNA was extracted from 10 g (wet weight) of sediment using the PowerSoil DNA Extraction Kit (MO BIO Laboratories, USA) according to the manufacturer's instructions. Total DNA from the MS-AsIII and MS-AsV isolates and sediment were quantified by absorbance at 260 nm using a NanoDrop Spectrophotometer (NanoDrop Technologies). DNA purity was assessed using the A260/A280 and A260/A230 ratios. DNA was stored at –20°C until further processing.

PCR Amplification and Construction of Clone Libraries

Briefly, touchdown PCR was carried out by amplifying bacterial MS-AsIII and MS-AsV isolates 16S rRNA gene fragments using the conditions previously described by Freitas *et al.* [35]. The reactions were performed using the bacterial-targeted primer set 8F (5'-AGAGTTTGATYMTGGCTCAG-3') and 907R (5'-CCGTCAATTCMTTTRAGT-3') [36]. Taq DNA polymerase and dNTPs were purchased from Fermentas (Canada) and used in all the PCR reactions.

Metagenomic and genomic DNA were used as template for PCR employing the *arsC*, *arrA* and *aioA* genes for construction of clone libraries and genotypic characterization of the bacterial MS-AsIII and MS-AsV isolates. PCR reactions targeting the *arsC*, *arrA* and *aioA* genes were carried out using primers and conditions as previously described by Sun *et al.* [37], Malasarn *et al.* [17] and Hamamura *et al.* [20], respectively. The *arsC* gene examined was of the glutaredoxin-dependent arsenate reductase enzyme, ArsC, from *Escherichia coli* R773 plasmid. The primer chosen has been successfully applied in several investigations of a variety of environmental samples [37,38].

The amplicons of *arsC*, *arrA*, and *aioA* genes were gel-purified using the Silica Bead DNA Gel Extraction Kit (Fermentas, Canada). PCR products were cloned into the vector pJET1.2/blunt (Fermentas, Canada), and propagated with *Escherichia coli* XL1-Blue electrocompetent cells according to the manufacturer's instructions.

Sequencing and Phylogenetic Analysis

Partial 16S rRNA, *arsC*, *arrA*, and *aioA* gene sequences were obtained using BigDye Terminator Cycle Sequencing kit (Life Technologies, USA) according to the manufacturer's instructions. The nucleotide sequences were quality checked and submitted to GenBank with accession numbers from KC577613 to KC577798. The 16S rRNA gene sequences were analyzed through blastn

(<http://www.ncbi.nlm.nih.gov>) and Classifier search tool (<http://rdp.cme.msu.edu>) to determine their phylogenetic affiliation. The *arsC*, *arrA*, and *aioA* gene sequences were compared with those available at the GenBank databases using *blastn* and *blastx* tools (<http://www.ncbi.nlm.nih.gov>) to retrieve potential homologs. Operational taxonomic units (OTUs) from As gene clone libraries were defined with DOTUR software [39] using a cut-off threshold of $\geq 97\%$ identity. Coverage of the clone libraries was calculated using the equation $C = 1 - (n/N) \times 100$, where *n* is the number of unique OTUs and *N* is the number of sequences analyzed in the library [40].

In total, five fasta files were obtained containing *arsC*, *arrA*, *aioA*, MS-AsIII 16S rRNA, and MS-AsV 16S rRNA gene sequences. Due to the short length of *arsC* and *arrA* amino acid sequences obtained in this study, added to the high similarity of some OTUs and isolates, we decided to reconstruct the phylogenetic relationships of As metabolism genes using nucleotide sequences to increase the phylogenetic signal and avoid overparameterization.

Sets of nucleotide sequences were independently aligned using MAFFT 7 with iterative refinement by the G-INS-i strategy [41]. Multiple sequence alignments were manually refined using Jalview [42]. To optimize the datasets for evolutionary analyses we removed redundancy and sequences too distantly related using the Decrease Redundancy tool available as a resource at ExPaSy (www.expasy.org). The Decrease Redundancy parameters were set as 99 for “% max similarity” and 30 for “% min similarity”. Identical sequences were clustered as single OTUs and filtered alignments were further used in phylogenetic analyses. Identifiers of filtered sequences were later included into the phylogenetic tree. To reconstruct phylogenetic trees we used the maximum likelihood method (ML) as implemented in PhyML [43]. For the phylogenetic reconstruction we tested seven different evolutionary models (HKY85, JC69, K80, F81, F84, TN93, and GTR) using the jModelTest 2 software [44]. The evolutionary model best fitting the data was determined by comparing the likelihood of tested models according to the Akaike Information Criterion (AIC). Statistical support value for each node was computed by approximate likelihood ratio test (aLRT). Trees were visualized and edited using the FigTree software (tree.bio.ed.ac.uk/software/figtree).

Susceptibility and Arsenic Transformation Tests

Minimum inhibitory concentrations (MIC) were established, in triplicate, by the agar dilution method in CDM with 1×10^5 CFU ml^{-1} as standard inoculums. CDM plates were supplemented with increasing concentrations (from 2 mM to 1024 mM) of AsIII or AsV and incubated at 28°C for seven days. MIC was defined as the lowest AsIII or AsV concentration that completely inhibited bacterial growth.

The ability to oxidize AsIII and reduce AsV was investigated using a qualitative screening according to [45]. To achieve that, bacterial MS-AsIII and MS-AsV isolates were grown in CDM broth with 100 mg l^{-1} 2 mM sodium arsenite or 100 mg l^{-1} sodium arsenate until an optical density of 0.4 at 595 nm was reached. After that, 20 μl of 0.01 mol l^{-1} of potassium permanganate solution were added in 1 ml of bacterial culture. The data were interpreted according to the change in medium color, i.e., a pink color indicated a positive oxidation of AsIII and a yellow color indicated a positive reduction of AsV.

Results

Environmental Parameters

The physicochemical characteristics of the water and sediment samples from the Mina stream are presented in Tables 1 and 2. Data displayed on Table 1 revealed that metal concentrations in the Mina stream exceeded the maximum allowable concentrations established by Brazilian and Canadian environmental regulations [46,47] for sediment and water. Al, Mn, Fe, Cu, As and Zn were the metals present in the highest concentrations in the sediment sample analyzed.

The physicochemical analysis revealed that the Mina stream can be characterized as a mesothermal oxidized environment with highly oxygenated and circum-neutral waters (Table 2). Nitrogen and phosphorus ratio was greater than nine (Table 2). According to Salas & Martino [48], this ratio indicates that the phosphorus was the most limiting nutrient and that the stream can be classified as eutrophic.

Phylogenetic Affiliation

In total, 123 bacterial isolates were recovered from the enrichment cultures (68 and 55 from the MS-AsIII and MS-AsV, respectively). Partial 16S rRNA gene sequences used for phylogenetic analysis were approximately 600 bp long and spanned the V2 to V4 variable regions. MS-AsIII and MS-AsV isolates were categorized into three phyla: *Proteobacteria* (56% and 59%, respectively, includes *alpha*, *beta*, and *gamma-Proteobacteria*), *Firmicutes* (36% in both enrichment cultures), and *Actinobacteria* (8% and 5%). Twenty genera represented these phyla in the Mina stream sample analyzed. Differences in the bacterial composition between the MS-AsIII and MS-AsV enrichment cultures were detected (Table 3 and Figure 1). The resulting Venn diagram shows that a higher bacterial diversity was observed in the MS-AsIII than in the MS-AsV enrichment cultures. Eight genera were specifically found in MS-AsIII and seven were shared between the culture systems (Figure 1).

Dominant genera in MS-AsV were *Bacillus* (26%), *Pseudoxanthomonas* (18%), and *Brevundimonas* (16%). The predominant population in MS-AsIII was *Bacillus* (30%), followed by *Pseudoxanthomonas* (25%). The other bacteria related to MS-AsIII and MS-AsV are listed in Table 3. Although the *Proteobacteria* phylum was the most diverse and dominant, the *Bacillus* (29%) genus was the most abundant and diverse among the genera because it harbored eight identified species.

Characterization of As-reducing and Oxidizing Isolates and Identification of their Genes Involved in As Metabolism

The MICs for the MS-AsIII and MS-AsV isolates were determined. The highest MIC was found for AsV in which 94% of the isolates exhibited values ≥ 256 mM, whereas 90% of the isolates displayed MICs ranging from 32 mM to 64 mM for the most toxic AsIII.

The As-transformation ability of the isolates was determined with a qualitative test that revealed that 72% of the isolates were AsV-reducing, whereas 20% were AsIII-oxidizing. Of those, 8% presented AsV-reducing as well as AsIII-oxidizing activities. Among the 20 genera identified in both MS-AsIII and MS-AsV enrichment cultures, *Acidovorax* and *Achromobacter* presented only AsIII-oxidizing activity. No As-transformation activity was found in 8% of the total of MS-AsIII and MS-AsV isolates (123) (Table 3).

The molecular analysis of the MS-AsIII and MS-AsV isolates unveiled that the *arsC* gene was the most frequent (85%), followed by *aioA* (20%) and *arrA* (7%) (Table 3). Of those, *Bacillus* was the

Table 1. Metal concentration from sediment and water of Mina Stream and limits permitted by law.

Metals	Sediment (mg kg ⁻¹)	CONAMA* (mg kg ⁻¹)	Water (mg l ⁻¹)	CONAMA** (mg l ⁻¹)
Fe	492.8	NE	0.52	15
Ni	9.0	18	<0.1	2
Mn	1284.5	NE	1.45	1
Cu	387.7	35.7	0.19	1
Pb	8.7	35	NE	0.5
Cd	<2.5	0.6	<0.1	0.2
Zn	180.9	123	0.2	5
Al	2343.2	NE	<0.5	NE
As	297.1	5.9	<0.1	0.5
Cr	17.3	37.3	<0.1	1
Hg	<2.5	0.17	<0.1	0.01

NE – Not established.

*CONAMA resolution 344/04.

**CONAMA resolution 430/11.

doi:10.1371/journal.pone.0095655.t001

only genus harboring all three genes, and *Shewanella* was the only genus which did not harbor the most common gene (*arsC*) in the isolate analyzed (MS-AsIII-61). *Achromobacter* and *Acidovorax* both harbored the *aioA* gene, confirming the phenotypic data. *Thermomonas* and *Pannonibacter* also harbored As resistance genes.

General Features of Clone Libraries

To unveil the molecular diversity of genes involved in As metabolism in the Mina stream sediment, three clone libraries for *arsC*, *arrA*, and *aioA* genes were constructed. One hundred sixty-four sequences were analyzed after quality control and the removal of chimeric sequences. The coverage values of the three libraries (80%, 70% and 63%, respectively for *arsC*, *arrA*, and *aioA*) indicated that most of the diversity of these genes was detected.

Blastx analysis of *arsC*, *aioA*, and *arrA*-OTUs revealed high similarity with sequences from glutaredoxin-glutathione arsenate reductase (from 76 to 100%), molybdopterin-binding arsenite oxidase (from 71 to 96%), and respiratory arsenate reductase (from 64 to 98%) (Tables S1, S2, and S3 in Tables S1). The sequences

corresponding to *arsC* were associated with *arsC* harboring different bacterial taxa from a variety of environments. The *aioA*-OTUs were closely related to uncultured and cultured clones from As contaminated environments. Furthermore, all *arrA*-OTUs were closely related to uncultured clones from rock biofilms of an ancient gold mine and Cache Valley Land Fill sediments, both arsenic-contaminated environments.

Phylogenetic Analyses of 16S rRNA, *arsC*, *aioA*, and *arrA* Genes Sequences

In this study we have amplified, sequenced and reconstructed the evolutionary relationships of 16S rRNA and As metabolism genes encoded by As-resistant bacteria retrieved from a stream located at the Brazilian gold mining area and cultivated on As-enrichment sediment's culture, as well as As metabolism genes of clone libraries. The phylogeny of the AsIII-resistant bacteria (MS-AsIII) 16S rRNA gene sequences was reconstructed from an alignment containing 57 operational taxonomic units and 719 sites, which represent 99 sequences (Fig. 2). Therefore, 42 sequences were considered redundant by the Decrease Redundancy tool (www.expasy.org). The phylogenetic tree reconstructed by using the maximum likelihood method as implemented in PhyML [43], shows sequence's clear separation into three strongly supported clades, which have representatives of the *Firmicutes*, *Actinobacteria*, and *Proteobacteria* phyla (Fig. 2). Similar results were obtained for the AsV-resistant bacteria (MS-AsV) 16S rRNA phylogenetic analysis (Fig. 3). The evolutionary history was based on an alignment containing 40 OTUs and 721 sites, representing 79 sequences (Fig. 3). The Decrease Redundancy tool filtered about 50% of the initially selected sequences. The resulting phylogeny also exhibits the presence of three well-supported clades containing *Firmicutes*, *Actinobacteria*, and *Proteobacteria* phyla representatives (Fig. 3).

Concerning evolutionary histories of As metabolism genes, the phylogenetic tree of *arsC* sequences was reconstructed with 48 nucleotide sequences and 352 sites, which represent 142 sequences (Fig. 4). TrN+I+G+F was selected as best fit model. The resulting phylogeny supports the hypothesis that horizontal gene transfer (HGT) seems to have played a role in the widespread distribution of *arsC* coding gene in *Actinobacteria* and *Proteobacteria*. Similar findings were retrieved on the phylogeny reconstructed for *arrA*

Table 2. Physicochemical parameters from water of Mina Stream.

Parameters	Water
pH	6.2
Conductivity (μs cm ⁻¹)	2151
Temperature (°C)	18.0
Dissolved Oxygen (mg l ⁻¹)	9.1
Redox (mV)	215
NO ₃ ⁻ -N (μg l ⁻¹)	3103.8
NO ₂ ⁻ -N (μg l ⁻¹)	161.3
NH ₄ ⁺ -N (μg l ⁻¹)	829.5
PO ₄ ³⁻ -P (μg l ⁻¹)	2.3
Total P (μg l ⁻¹)	77.6
Total N (μg l ⁻¹)	2916.8

doi:10.1371/journal.pone.0095655.t002

Table 3. Phylogenetic distribution of the bacterial isolates and their As-metabolism phenotype and genotype.

Enriched culture	Phylum	Genus	N° of isolates*	Phenotype**	Genotype***	
MS-AsV	Proteobacteria	<i>Acinetobacter</i>	1	reducer	<i>arsC aioA</i>	
		<i>Brevundimonas</i>	9	reducer (3)	<i>arsC</i> [9] <i>aioA</i> [1]	
		<i>Diaphorobacter</i>	1	reducer	<i>arsC</i>	
		<i>Pannonibacter</i>	2	reducer (2)	<i>arsC</i> [2]	
		<i>Pseudomonas</i>	6	reducer (2)	<i>arsC</i> [6] <i>aioA</i> [2]	
		<i>Pseudoxanthomonas</i>	10	reducer (6)	<i>arsC</i> [7] <i>aioA</i> [1]	
		<i>Stenotrophomonas</i>	5	reducer (5)	<i>arsC</i> [5] <i>aioA</i> [2]	
		<i>Thermomonas</i>	1	reducer	<i>arsC</i>	
		Firmicutes	<i>Bacillus</i>	14	reducer (9), oxidizer (8)	<i>arsC</i> [12] <i>aioA</i> [6] <i>arrA</i> [1]
			<i>Exiguobacterium</i>	1	reducer	<i>arsC aioA</i>
		Actinobacteria	<i>Microbacterium</i>	4	reducer (4) oxidizer (1)	<i>arsC</i> [3] <i>aioA</i> [1]
<i>Rhodococcus</i>	1		reducer	<i>arsC</i>		
MS-AsIII	Proteobacteria	<i>Achromobacter</i>	1	oxidizer	<i>arsC aioA</i>	
		<i>Acidovorax</i>	2	oxidizer(2)	<i>arsC</i> [2] <i>aioA</i> [1]	
		<i>Acinetobacter</i>	4	reducer (3)	<i>arsC</i> [2] <i>aioA</i> [1]	
		<i>Comamonas</i>	5	reducer (5)	<i>arsC</i> [5] <i>arrA</i> [2]	
		<i>Ochrobactrum</i>	1	reducer	<i>arsC</i>	
		<i>Pseudomonas</i>	3	reducer (3)	<i>arsC</i> [1] <i>aioA</i> [1]	
		<i>Pseudoxanthomonas</i>	17	reducer (9) oxidizer (5)	<i>arsC</i> [14] <i>aioA</i> [2] <i>arrA</i> [1]	
		<i>Shewanella</i>	1	reducer	<i>arrA</i>	
		<i>Stenotrophomonas</i>	2	reducer (2) oxidizer (1)	<i>arsC</i> [1] <i>aioA</i> [1]	
		<i>Thermomonas</i>	2	reducer (1)	<i>arsC</i> [2] <i>arrA</i> [1]	
		Firmicutes	<i>Bacillus</i>	21	reducer (19) oxidizer (5)	<i>arsC</i> [21] <i>aioA</i> [2] <i>arrA</i> [3]
			<i>Paenibacillus</i>	1	reducer oxidizer (1)	<i>arsC</i>
			<i>Staphylococcus</i>	4	reducer (4) oxidizer (1)	<i>arsC</i> [3]
		Actinobacteria	<i>Micrococcus</i>	3	reducer (2) oxidizer (1)	<i>arsC</i> [1] <i>aioA</i> [1]
			<i>Microbacterium</i>	1	reducer	-

*The number represents the total of bacterial isolates identified.
 **Values in parentheses indicate the number of As-redox isolates.
 ***Values in bracket indicate the number of isolates harboring As-metabolism genes.
 doi:10.1371/journal.pone.0095655.t003

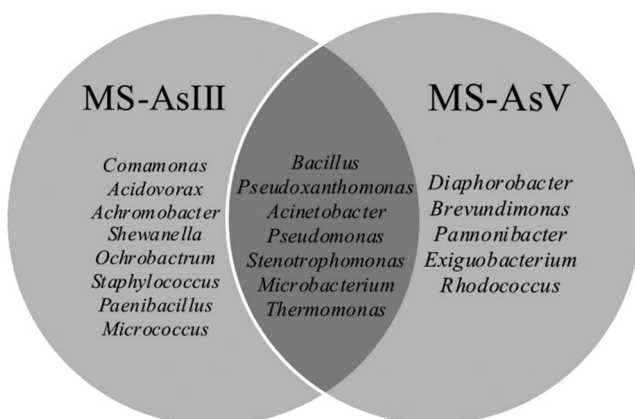


Figure 1. Venn diagram showing the exclusive and shared bacterial genera retrieved from MS-AsIII and MS-AsV enrichment cultures.
 doi:10.1371/journal.pone.0095655.g001

sequences based on 47 nucleotide sequences and 242 sites where GTR+I+G+F was selected as best fit model (Fig. 5). On the other hand, the phylogenetic analysis of *aioA* sequences based on 72 nucleotide sequences and 543 sites shows two clades strongly supported: *alpha*- and *beta*-*proteobacteria* (Fig. 6) without clear evidence of HGT. For this analysis GTR+I+G+F was selected as best fit model. Interestingly, all putative *arrA* sequences obtained in this study (*arrA*- OTU) were more closely related to themselves or to sequences from uncultured bacteria, showing that more studies involving *arrA* sequences will be relevant to better understand the molecular diversity of those genes (Fig. 5).

Discussion

The environmental impact of gold mining is presently a major concern because its processes release toxic metals such as As in both soil and groundwater. Considering the relevance of bacteria in the speciation of As in aquatic environments, we bioprospected As-resistant bacteria and As-transforming genes originated from sediments impacted by long-term gold mining. Although some studies have focused in the identification of As-resistant bacterial communities in a long-term As-contaminated environment [9–

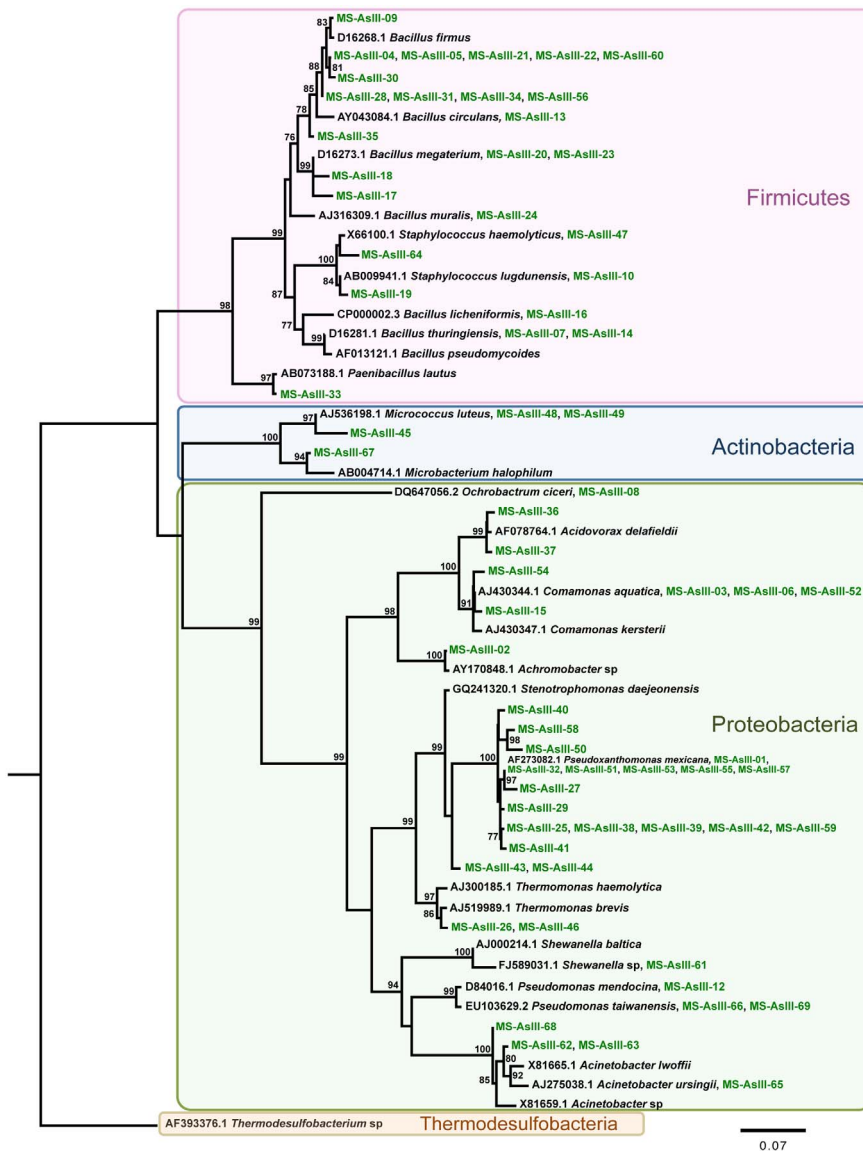


Figure 2. Evolutionary relationships of AsIII-resistant bacteria (MS-AsIII) 16S rRNA sequences. A total of 57 nucleotide sequences and 719 sites were analyzed. The phylogeny was reconstructed by maximum likelihood and TIM3+H+G+F was selected as best fit model. Support values for each node were estimated using the Akaike Likelihood Ratio Test (aLRT). Only support values higher than 70% are shown. Reference sequences retrieved from the non-redundant database of the NCBI are shown in black, bacterial isolates (MS-AsIII and MS-AsV) in green. Different background colors highlight three well-supported clades: *Firmicutes*, *Actinobacteria*, and *Proteobacteria*. *Thermodesulfobacteria* was used as outgroup. doi:10.1371/journal.pone.0095655.g002

13,49–52], the employment of combination of culture-based physiological and genomic approaches with metagenomic analysis in sediments collected from these areas are scarce [53,54]. In this study, we reveal a large number of phylogenetically distinct As-resistant bacterial genera retrieved from sediment collected from a stream in a long-term gold-mining area.

We found *Bacillus* as the dominant genus in both MS-AsIII and MS-AsV enrichment cultures. Members of *Bacillus* are often found in As-contaminated environments [9,11,16,54] being related to As-reduction and -oxidation, indicating that they are an essential component of As speciation in nature [54,55]. The observed abundance of *Bacillus* isolates harboring the *arsC* and *aiiA* genes confirmed its ubiquity and high As-resistance in As-rich environments, as it is the case of Mina stream sediment. This suggests an important role for *Bacillus* in As speciation. It also points to a

possible use of these natural isolates in future bioremediation projects.

A recent study of our group [56], using culture-independent approach to assess the prokaryotic diversity in Mina stream sediment, revealed the presence of the *Thermomonas*, *Acidovorax*, *Acinetobacter* and *Ochobacterium* genera also detected in the present study. Moreover, Bandyopadhyay *et al.* [57] have proposed a novel species of the *Pannonibacter* genus, *Pannonibacterindica*, which is able to grow in high concentrations of AsV. However, it should be noted that *Thermomonas* and *Pannonibacter* were not previously reported in the literature as As-transforming genera.

The phenotypic and genotypic characterization of the MS-AsIII and MS-AsV bacterial isolates revealed their ability to reduce and oxidize As. Most bacteria (85%) were AsV-resistant bacteria (ARB) harboring the *arsC* gene, responsible for the reduction of AsV to

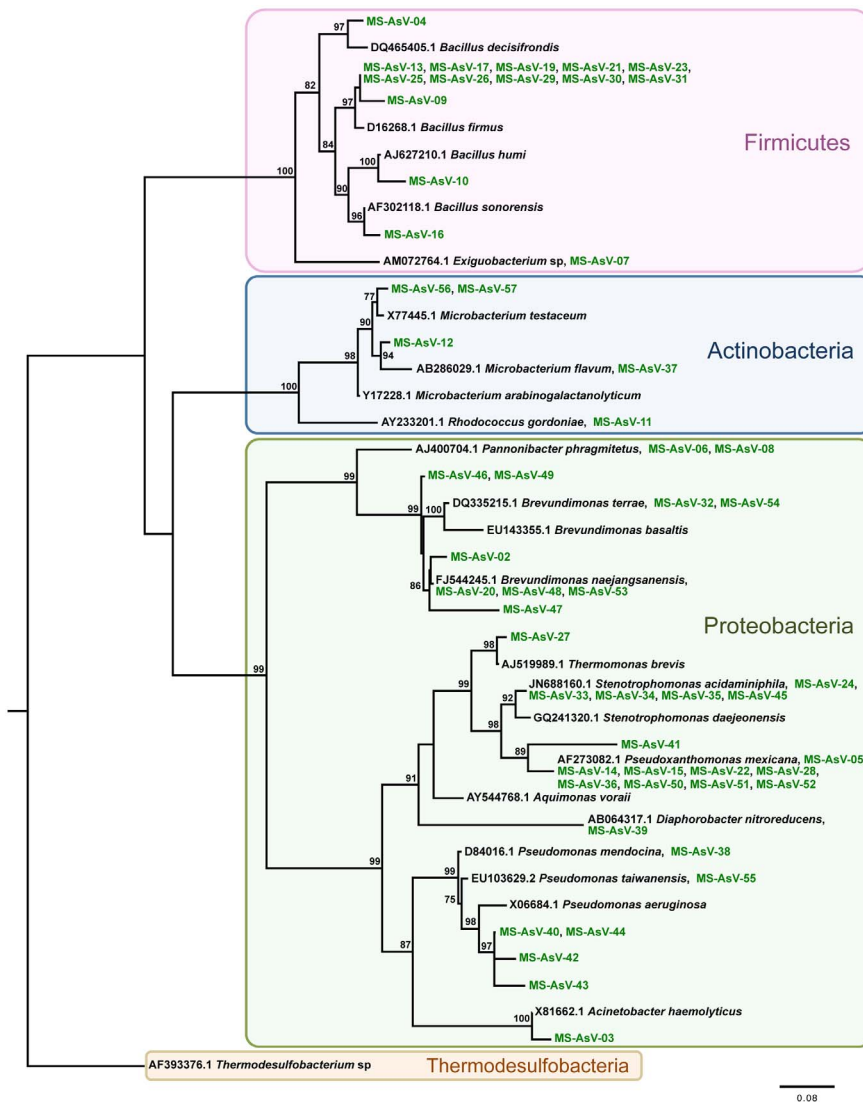


Figure 3. Evolutionary relationships of AsV-resistant bacteria (MS-AsV) 16S rRNA sequences. A total of 40 nucleotide sequences and 721 sites were analyzed. The phylogeny was reconstructed by maximum likelihood and TrN+G+F was selected as best fit model. Support values for each node were estimated using the Akaike Likelihood Ratio Test (aLRT). Only support values higher than 70% are shown. Reference sequences retrieved from the non-redundant database of the NCBI are shown in black, bacterial isolates (MS-AsIII and MS-AsV) in green. Different background colors highlight three well-supported clades: *Firmicutes*, *Actinobacteria*, and *Proteobacteria*. *Thermodesulfobacterium* was used as outgroup. doi:10.1371/journal.pone.0095655.g003

AsIII, which is the most frequent detoxification reaction among bacteria in the environment [8]. Although the aerobic enrichment culture condition employed in this study could inhibit the growth of dissimilatory arsenate-reducing bacteria (DARB), it is likely that these bacteria were present because the *arrA* gene was detected. Several reports have evidenced DARB bioremediation potential of As-contaminated samples [2,54,58,59].

AsIII-oxidizing bacterial isolates were minority (20%). This result is in agreement with those reported by Silver & Phung [14], who suggest that most isolates from natural environments lack AsIII-oxidizing ability. In this study, all AsIII-oxidizing isolates were classified as heterotrophic AsIII oxidizers (HAO) spanning 11 genera. However, only isolates belonging to *Bacillus*, *Pseudoxanthomonas*, *Stenotrophomonas*, *Micrococcus*, *Achromobacter*, and *Acidovorax* genera co-presented the oxidizing phenotype and genotype. From an ecological perspective, oxidizing bacteria are important

performers in As-contaminated environments because they promote transformation from AsIII into AsV.

In a few isolates (8%), oxidizing and reducing As-transformation activities were not observed in their phenotype and genotype. There are several possible explanations for this. First, the As-transformation gene expression observed in isolates grown in the laboratory is likely to be different from that encountered in these isolates in nature, because of the different conditions of these environments. Second, these differences may reflect mutations in the As-resistance genes studied. Third, alternative resistance genes may be expressed by these isolates [60].

The high diversity and adaptability of the bacterial community disclosed herein could be explained by the presence of multiple copies of As-resistance genes either on bacterial chromosomes or on plasmids as a consequence of pressure created by the long-term contamination that occurs in the Mina stream area. Nevertheless, further studies will be needed to establish this.

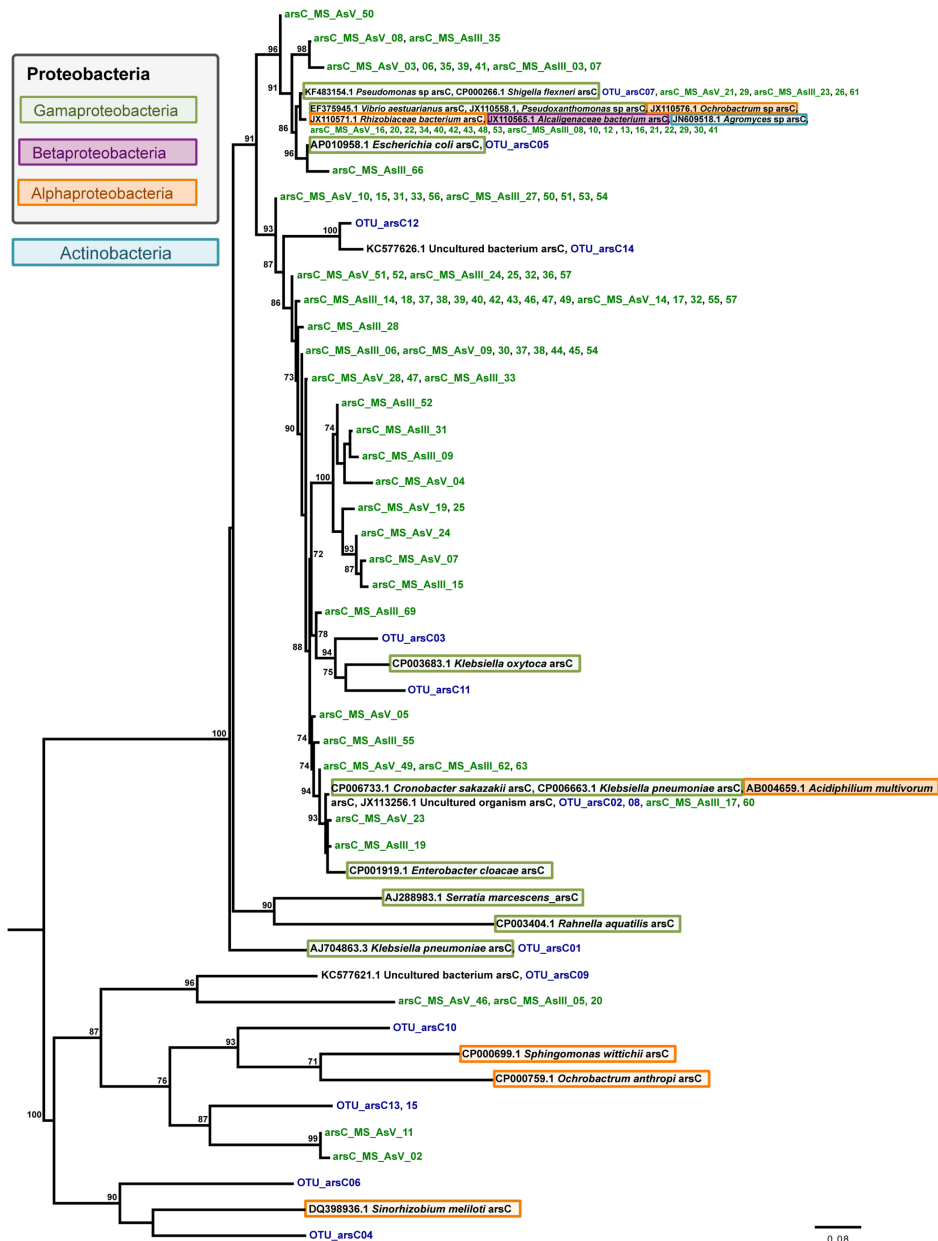


Figure 4. Evolutionary relationships of *arsC* sequences. A total of 48 nucleotide sequences and 352 sites were analyzed. The phylogeny was reconstructed by maximum likelihood and TrN+I+G+F was selected as best fit model. Support values for each node were estimated using the Akaike Likelihood Ratio Test (aLRT). Only support values higher than 70% are shown. Reference sequences retrieved from the non-redundant database of the NCBI are shown in black, bacterial isolates (MS-AsIII and MS-AsV) in green, and operational taxonomic unities (OTUs) from *As* gene clone libraries in blue. Different background colors highlight Actinobacteria and three Proteobacteria classes – Gamma-, Beta, and alpha-proteobacteria. doi:10.1371/journal.pone.0095655.g004

Previous studies on *As*-resistance genes are associated with *As*-resistant cultivable isolates [10,16,59,61]. Considering that the vast majority of bacteria are uncultivable, this traditional approach has limited our understanding of the extreme functional diversity in natural bacterial communities. Therefore, a metagenomic approach to investigate the functional genes associated with *As*-transformation in nature is essential to further our current knowledge on this matter. The analysis of *arrA* sequences revealed that all of them exhibited similarity with those from uncultured organisms. This predominance of uncultured organisms indicates that *arrA* gene present in Mina stream sediment is expressed by unidentified DARB. The *arsC* sequences detected in the sediment

were similar to those previously reported [37,62,63]. The primers used in this study amplified *aioA*-like sequences [20,64]. The *aioA* sequences were similar to several *aioA* genes of the Proteobacteria phylum. This finding is in agreement with Quéméneur *et al.* [65], who reported prevalence of *As*III-oxidizing Proteobacteria in mesophilic *As*-contaminated soils. However, it should be noted that *aioA* genes have been also detected in non-proteobacterial lineages [53,69].

Phylogenetic analyses' findings performed for MS-AsIII 16S rRNA, Ms-AsV 16S rRNA and *As* metabolism genes were consistent with findings obtained by similarity searches (blastx and blastn, respectively). Overall, the phylogenetic trees reconstructed

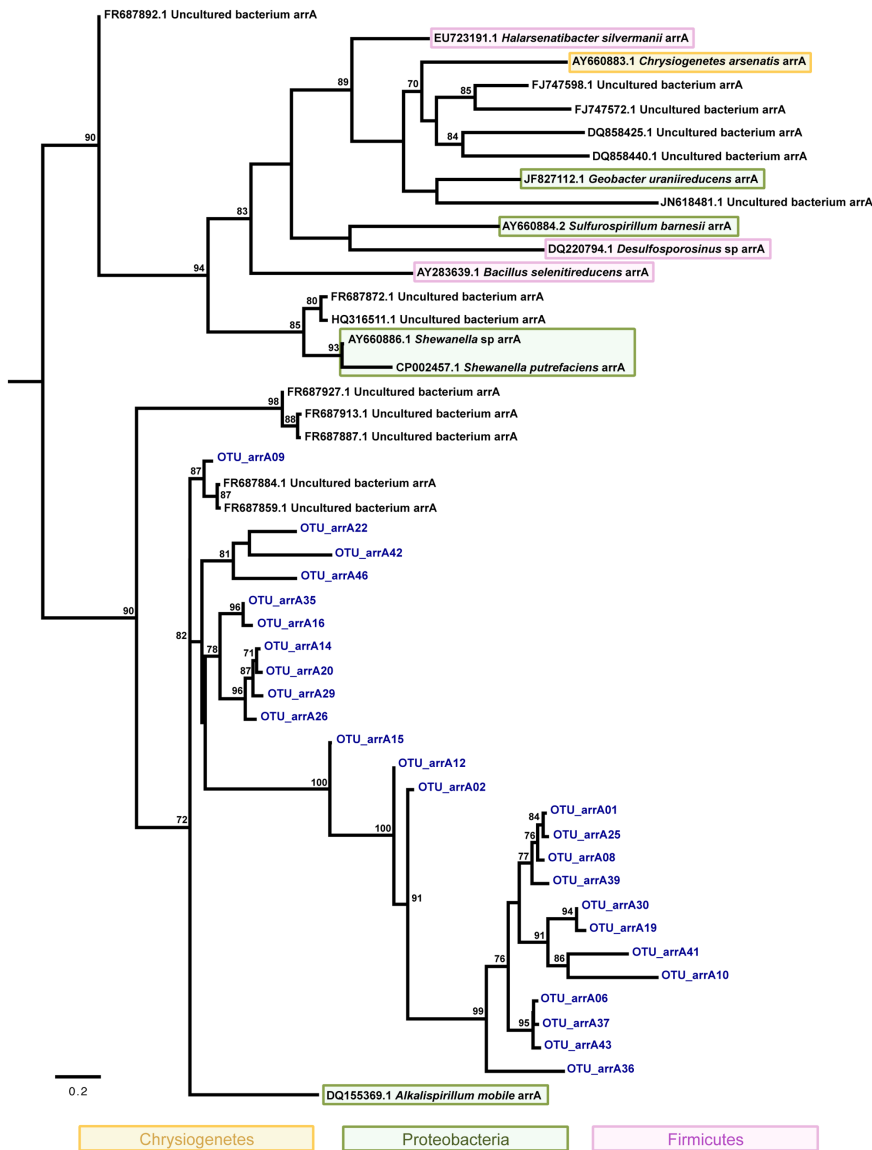


Figure 5. Evolutionary relationships of *arrA* sequences. A total of 47 nucleotide sequences and 242 sites were analyzed. The phylogeny was reconstructed by maximum likelihood and GTR+I+G+I was selected as best fit model. Support values for each node were estimated using the Akaike Likelihood Ratio Test (aLRT). Only support values higher than 70% are shown. Reference sequences retrieved from the non-redundant database of the NCBI are shown in black, bacterial isolates (MS-AsIII and MS-AsV) in green, and operational taxonomic units (OTUs) from As gene clone libraries in blue. Different background colors highlight three bacterial phyla - *Proteobacteria*, *Firmicutes*, and *Chrysiogenetes*. doi:10.1371/journal.pone.0095655.g005

for MS-AsIII and MS-AsV 16S rRNA sequences show very similar evolutionary histories where the relationships among *Firmicutes*, *Actinobacteria*, *Proteobacteria*, and *Thermodesulfobacteria* phyla members reflect the current knowledge regarding their evolution [66].

As previously described on the literature, the evolutionary relationships of *arsC* and *arrA* homologs (Figs. 4 and 5) support the role of horizontal gene transfer (HGT) on the evolution of arsenate oxidases e.g. [67,68]. The phylogeny reconstructed for *arsC* homologs (Fig. 4) clearly shows two *Ochrobactrum* sequences clustered in different well-supported clades suggesting that these two homologs were acquired by HGT from unrelated donors. Although it is known that due to HGT events *aiiA* sequences are not a suitable marker for microbial diversity studies [53], it was not observed on the *aiiA* phylogeny here presented (Fig. 6). Albeit *aiiA* sequences have been detected in non-proteobacterial lineages

[53,69], our findings show two strongly supported clades clustering *alpha*- and *beta*-*proteobacteria* homologs. Such results probably reflect the bias existing on GenBank databases where most *aiiA* sequences available are from proteobacterial lineages.

Overall, evolutionary analyses revealed high genetic similarity between some *arsC* and *aiiA* sequences obtained from isolates and clone libraries, suggesting that those isolates may represent environmentally important bacteria acting in As speciation. In addition, some *arsC*, *aiiA*, and *arrA* sequences were found to be closely related to homologs from uncultured bacteria. Thus, it may be hypothesized that these divergent sequences could represent novel variants of the As-resistance genes or other genes with related function. In addition, our findings show that the diversity of *arrA* genes is wider than earlier described, once none *arrA*-OTUs were affiliated with known reference strains. Therefore, the

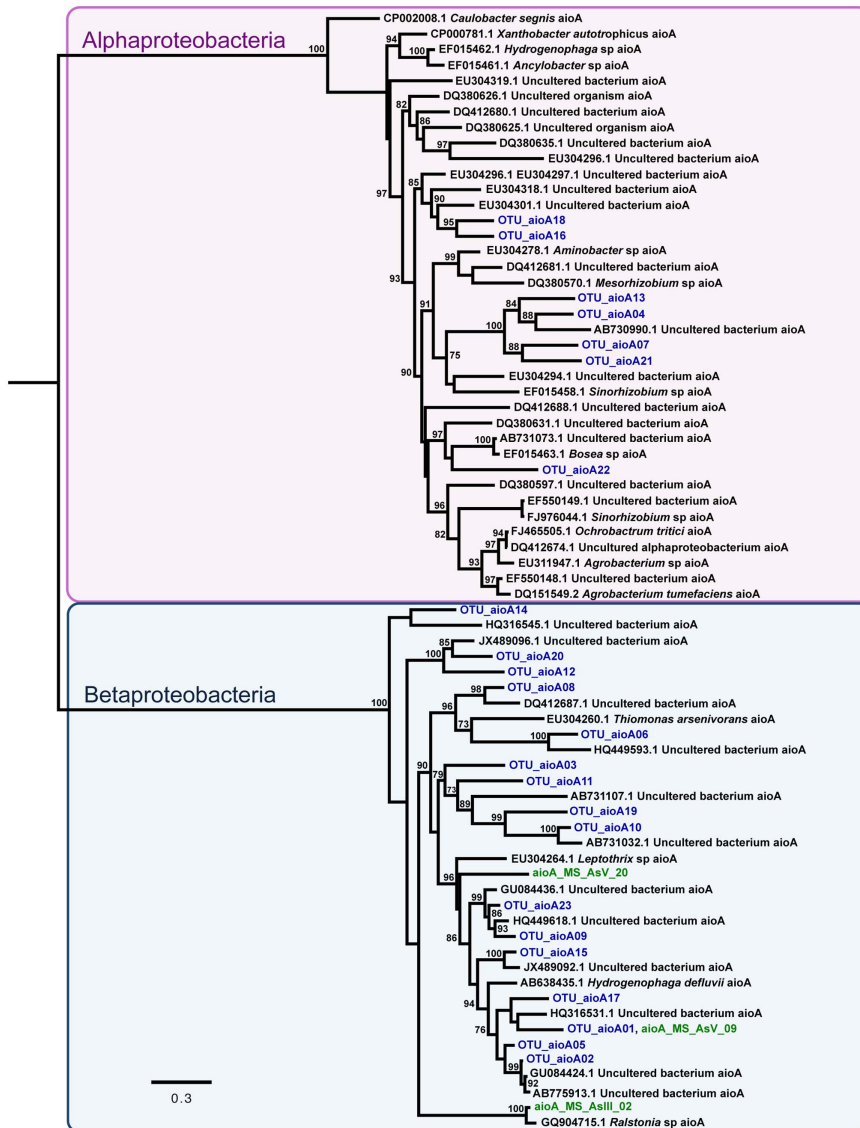


Figure 6. Evolutionary relationships of *aioA* sequences. A total of 72 nucleotide sequences and 543 sites were analyzed. The phylogeny was reconstructed by maximum likelihood and GTR+I+G+F was selected as best fit model. Support values for each node were estimated using the Akaike Likelihood Ratio Test (aLRT). Only support values higher than 70% are shown. Reference sequences retrieved from the non-redundant database of the NCBI are shown in black, bacterial isolates (MS-AsIII and MS-AsV) in green, and operational taxonomic unities (OTUs) from As gene clone libraries in blue. Different background colors highlight two Proteobacteria classes – *beta*- and *alpha*-proteobacteria.
doi:10.1371/journal.pone.0095655.g006

molecular diversity of *arrA* genes is far from being fully explored deserving further attention.

Altogether, this study is a bioprospection of AsIII-oxidizing and AsV-reducing bacteria and As-transforming genes in sediments impacted by long-term gold mining. Our culture efforts successfully identified a large number of phylogenetically distinct arsenic-resistant bacterial genera and revealed two novel As-transforming genera, *Thermomonas* and *Pannonibacter*. Our heterotrophic arsenite oxidizers and DARB isolates open new opportunities for their use in bioremediation of long-term gold-mining impacted areas. Furthermore, metagenomic analysis of As functional genes revealed a predominance of previously unidentified DARB.

Supporting Information

Figure S1 Map showing the sampling site. Crosshatch, red and yellow areas represent mining, urban, and sampling areas, respectively.

(TIF)

Tables S1 This file includes Table S1, S2 and S3. Table S1. Phylogenetic affiliation of *aioA* OTUs based on blastx protein database. Table S2. Phylogenetic affiliation of *arsC* OTUs based on blastx protein database. Table S3. Phylogenetic affiliation of *arrA* OTUs based on blastx protein database.

(DOCX)

Acknowledgments

We appreciate the technical support from Laboratório de Análises Químicas/DEMET/UFMG do Instituto Nacional de Ciência e Tecnologia em Recursos Minerais, Água e Biodiversidade – INCT –ACQUA in the chemical analyses. The authors acknowledge the use of the computing resources of the Center for Excellence in Bioinformatics (CEBio//CPqRR/Fiocruz, Brazil).

References

- Lièvreumont D, Bertin PN, Lett MC (2009) Arsenic in contaminated waters: Biogeochemical cycle, microbial metabolism and biotreatment processes. *Biochimie* 91: 1229–1237.
- Páez-Espino D, Tamames J, de Lorenzo V, Cánovas D (2009) Microbial responses to environmental arsenic. *Biomol* 22: 117–130.
- Neubauer O (1947) Arsenical cancer - a review. *Br J Cancer* 1: 192–251.
- Bhattacharya P, Welch AH, Stollenwerk KG, McLaughlin MJ, Bundschuh J, et al. (2007) Arsenic in the environment: Biology and Chemistry. *Sci Total Environ* 379: 109–20.
- McClintock TR, Chen Y, Bundschuh J, Oliver JT, Navoni J, et al. (2012) Arsenic exposure in Latin America: biomarkers, risk assessments and related health effects. *Sci Total Environ* 429: 76–91.
- Kruger MC, Bertin PN, Heipieper HJ, Arsène-Ploetz F (2013) Bacterial metabolism of environmental arsenic-mechanisms and biotechnological applications. *Appl Microbiol Biotechnol* 97: 3827–3841.
- Nordstrom DK (2002) Public health-worldwide occurrences of arsenic in ground water. *Science* 296: 2143–2145.
- Tsai SL, Singh S, Chen W (2009) Arsenic metabolism by microbes in nature and the impact on arsenic remediation. *Curr Opin Biotechnol* 20: 659–67.
- Anderson CR, Cook GM (2004) Isolation and characterization of arsenate-reducing bacteria from arsenic-contaminated sites in New Zealand. *Curr Microbiol* 48: 341–347.
- Chang JS, Yoon IH, Lee JH, Kim KR, An J, et al. (2008) Arsenic detoxification potential of *aox* genes in arsenite oxidizing bacteria isolated from natural and constructed wetlands in the Republic of Korea. *Environ Geochem Health* 32: 95–105.
- Drewniak L, Styczek A, Majder-Lopatka M, Sklodowska A (2008) Bacteria, hypertolerant to arsenic in the rocks of an ancient gold mine, and their potential role in dissemination of arsenic pollution. *Environ Pollut* 156: 1069–74.
- Cai L, Liu G, Rensing C, Wang G (2009) Genes involved in arsenic transformation and resistance associated with different levels of arsenic-contaminated soils. *BMC Microbiol* 9: 4.
- Cavalca L, Zanchi R, Corsini A, Colombo M, Romagnoli C, et al. (2010) Arsenic-resistant bacteria associated with roots of the wild *Cirsium arvense* (L.) plant from an arsenic polluted soil, and screening of potential plant growth-promoting characteristics. *Syst Appl Microbiol* 33: 154–64.
- Silver S, Phung LT (2005) Genes and enzymes involved in bacterial oxidation and reduction of inorganic arsenic. *Appl Environ Microbiol* 71: 599–608.
- Kaur S, Kamli MR, Ali A (2009) Diversity of arsenate reductase genes (*arsC* genes) from arsenic-resistant environmental isolates of *E. coli*. *Curr Microbiol* 59: 288–94.
- Liao VHC, Chu YJ, Su YC, Hsiao SY, Wei CC, et al. (2011) Arsenite-oxidizing and arsenate-reducing bacteria associated with arsenic-rich groundwater in Taiwan. *J Contam Hydrol* 123: 20–9.
- Malasarn D, Saltikov CW, Campbell KM, Santini JM, Hering JG, et al. (2004) *arsA* is a reliable marker for As(V) respiration. *Science* 306: 455.
- Zargar K, Conrad A, Bernick DL, Lowe TM, Stolc V, et al. (2012) ArxA, a new clade of arsenite oxidase within the DMSO reductase family of molybdenum oxidoreductases. *Environ Microbiol* 14: 1635–45.
- Muller D, Lièvreumont D, Simeonova DD, Hubert JC, Lett MC (2003) Arsenite oxidase *aox* genes from a metal-resistant *beta-proteobacterium*. *J Bacteriol* 185: 135–41.
- Hamamura N, Macur RE, Korf S, Ackerman G, Taylor WP, et al. (2009) Linking microbial oxidation of arsenic with detection and phylogenetic analysis of arsenite oxidase genes in diverse geothermal environments. *Environ Microbiol* 11: 421–31.
- Stolz JF, Basu P, Oremland RS (2010) Microbial arsenic metabolism: new twists on an old poison. *Microbe* 5: 53–59.
- Slyemi D, Bonnefoy V (2012) How prokaryotes deal with arsenic. *Environ Microbiol Reports* 4: 571–586.
- Lett MC, Muller D, Lièvreumont D, Silver S, Santini J (2012) Unified nomenclature for genes involved in prokaryotic aerobic arsenite oxidation. *J Bacteriol* 194: 207–8.
- Oremland RS, Saltikov CW, Wolfe-Simon F, Stolz JF (2009) Arsenic in the evolution of earth and extraterrestrial ecosystems. *Geomicrobiol J* 26: 522–536.
- Sun W, Sierra-Alvarez R, Milner L, Field JA (2010) Anaerobic oxidation of arsenite linked to chlorate reduction. *Appl Environ Microbiol* 76: 6804–6811.
- Kulp TR, Hoelt SE, Asao M, Madigan MT, Hollibaugh JT, et al. (2008) Arsenic(III) fuels anoxygenic photosynthesis in hot spring biofilms from Mono Lake, California. *Science* 321: 967–970.

Author Contributions

Conceived and designed the experiments: PSC ECS AMAN. Performed the experiments: PSC MPR PLO LBI MLSS FARB. Analyzed the data: PSC LLSS MPR AVC ECS AMAN. Contributed reagents/materials/analysis tools: FARB AMAN. Wrote the paper: PSC LLSS MPR ECS AMAN.

- Borba RP, Figueredo BR, Rawlins BG, Matchullat J (2000) Arsenic in water and sediment in the Iron Quadrangle, Minas Gerais state, Brasil. *Revista Brasileira de Geociências* 30: 554–557.
- Instituto Mineiro de Gestão das Águas (IGAM) (2004) Camargos LMM. Plano diretor de recursos hídricos da bacia hidrográfica do rio das Velhas: resumo executivo - Belo Horizonte, MG. Instituto Mineiro de Gestão das Águas, Comitê da Bacia Hidrográfica do Rio das Velhas.
- Salomons W, de Rooij NM, Kerdijk H, Bril J (1987) Sediments as a source for contaminants? *Hydrobiologia* 149: 13–30.
- Rasmussen H, Jørgensen BB (1992) Microelectrode studies of seasonal oxygen uptake in a coastal sediment: role of molecular diffusion. *Marine Ecology Progress Series* 81: 289–303.
- Marchand C, Lallier-Verges E, Allenbach M (2011) Redox conditions and heavy metals distribution in mangrove forests receiving shrimp farm effluents (Teremba bay, New Caledonia). *J Soils Sediments* 11: 529–541.
- Mackereth FJH, Heron J, Talling JF (1978) Water analysis: some revised methods for limnologists. Freshwater Biological Association Scientific Publication, Wareham.
- Golterman HL, Clymo RS, Ohnstad MAM (1978) Methods for chemical analysis of fresh waters. Blackwell Scientific Publications, Philadelphia, PA.
- Drams S, Biswas I, Maguin E, Braun L, Mastroeni P, et al. (1995) Entry of *Listeria monocytogenes* into hepatocytes requires expression of *inlB*, a surface protein of the internalin multigene family. *Mol Microbiol* 16: 251–261.
- Freitas DB, Lima-Bittencourt CI, Reis MP, Costa PS, Assis PS, et al. (2008) Molecular characterization of early colonizer bacteria from wastes in a steel plant. *Lett Appl Microbiol* 47: 241–249.
- Lane DJ (1991) 16S/23S rRNA sequencing. John Wiley and Sons, New York.
- Sun Y, Polishchuk EA, Radoja U, Cullen WR (2004) Identification and quantification of *arsC* genes in environmental samples by using real-time PCR. *J Microbiol Methods* 58: 335–49.
- Sarkar A, Kazy SK, Sar P (2013) Characterization of arsenic resistant bacteria from arsenic rich groundwater of West Bengal, India. *Ecotoxicology* 22: 363–376.
- Schloss PD, Handelsman J (2005) Introducing DOTUR, a computer program for defining operational taxonomic units and estimating species richness. *Appl Environ Microbiol* 71: 1501–1506.
- Good IJ (1953) The population frequencies of species and the estimation of population parameters. *Biometrika* 40: 237–262.
- Katoh K, Standley DM (2013) MAFFT multiple sequence alignment software version 7: improvements in performance and usability. *Mol Biol Evol* 30: 772–780.
- Versionhouse AM, Procter JB, Martin DM, Clamp M, Barton GJ (2009) Jalview Version 2—a multiple sequence alignment editor and analysis workbench. *Bioinformatics* 25: 1189–1191.
- Guindon S, Dufayard JF, Lefort V, Anisimova M, Hordijk W, et al. (2010) New algorithms and methods to estimate maximum-likelihood phylogenies: Assessing the performance of PhyML 3.0. *Syst Biol* 59: 307–21.
- Darriba D, Taboada GL, Doallo R, Posada D (2012) jModelTest 2: more models, new heuristics and parallel computing. *Nat Methods* 9: 772.
- Salmassi TM, Venkateswaren K, Satomi M, Nealon KH, Newman DK, et al. (2002) Oxidation of arsenite by *Agrobacterium albertinagni*, AOL15, sp. nov., isolated from Hot Creek, California. *Geomicrobiol J* 19: 53–66.
- Conselho Nacional do Meio Ambiente (CONAMA) (2011) Resolução N° 430, de 13 de maio de 2011. URL: <http://www.mma.gov.br/port/conama/estr.cfm>. Accessed 20 June 2011.
- Canadian Council of Ministers of the Environment (CCME) Canadian Environmental Quality Guidelines. URL: <http://www.ccme.ca/>. Accessed 02 April 2011.
- Salas HJ, Martino P (1991) A simplified phosphorus trophic state model for warm-water tropical lakes. *Water Res* 25: 341–350.
- Santini JM, Sly LI, Schnagl RD, Macy JM (2000) A new chemolithoautotrophic arsenite-oxidizing bacterium isolated from a gold mine: phylogenetic, physiological, and preliminary biochemical studies. *Appl Environ Microbiol* 66: 92–97.
- Santini JM, Sly LI, Wen A, Comrie D, Wulf-Durand P, et al. (2002) New arsenite-oxidizing bacteria isolated from Australian gold mining environments: phylogenetic relationships. *Geomicrobiol J* 19: 67–76.
- Santini JM, vanden Hoven RN (2004) Molybdenum-containing arsenite oxidase of the chemolithoautotrophic arsenite oxidizer NT-26. *J Bacteriol* 186: 1614–1619.

52. Oliveira A, Pampulha ME, Neto MM, Almeida AC (2009) Enumeration and characterization of arsenic-tolerant diazotrophic bacteria in a long-term heavy-metal-contaminated soil. *Water Air Soil Pollut* 200: 237–243.
53. Heinrich-Salmeron A, Cordi A, Brochier-Armanet C, Halter D, Pagnout C, et al. (2011) Unsuspected Diversity of Arsenite-Oxidizing Bacteria as Revealed by Widespread Distribution of the *aoxB* Gene in Prokaryotes. *Appl Environ Microbiol* 77: 4685–92.
54. Yamamura S, Ike M, Fujita M (2003) Dissimilatory arsenate reduction by a facultative anaerobe, *Bacillus* sp. strain SF-1. *J BiosciBioeng* 96: 454–60.
55. Chang JS, Kim IS (2010) Arsenite oxidation by *Bacillus* sp. strain SeaH-As22w isolated from coastal seawater in Yeosu Bay. *Environ Eng Res* 15: 15–21.
56. Reis MP, Barbosa FA, Chartone-Souza E, Nascimento AMA (2013) The prokaryotic community of a historically mining-impacted tropical stream sediment is as diverse as that from a pristine stream sediment. *Extremophiles* 17: 301–309.
57. Bandyopadhyay S, Schumann P, Das SK (2013) *Pannonibacter indica* sp. nov., a highly arsenate-tolerant bacterium isolated from a hot spring in India. *Arch Microbiol* 195: 1–8.
58. Dowdle PR, Laverman AM, Oremland RS (1996) Bacterial dissimilatory reduction of arsenic(V) to arsenic(III) in anoxic sediments. *Appl Environ Microbiol* 62: 1664–9.
59. Chang YC, Nawata A, Jung K, Kikuchi S (2012) Isolation and characterization of an arsenate-reducing bacterium and its application for arsenic extraction from contaminated soil. *J Ind Microbiol Biotechnol* 39: 37–44.
60. Achour AR, Bauda P, Billard P (2007) Diversity of arsenite transporter genes from arsenic-resistant soil bacteria. *Res Microbiol* 158: 128–37.
61. Sri LSM, Prashant S, Bramha CPV, Nageswara RS, Balaravi P, et al. (2012) Molecular identification of arsenic-resistant estuarine bacteria and characterization of their *ars* genotype. *Ecotoxicology* 21: 202–12.
62. Rasko DA, Rosovitz MJ, Myers GS, Mongodin EF, Fricke WF, et al. (2008) The pangenome structure of *Escherichiacoli*: comparative genomic analysis of *E. coli* commensal and pathogenic isolates. *J Bacteriol* 190: 6881–93.
63. Gootz TD, Lescoe MK, Dib-Hajj F, Dougherty BA, He W, et al. (2009) Genetic organization of transposase regions surrounding *blaKPC* carbapenemase genes on plasmids from *Klebsiella* strains isolated in a New York City hospital. *Antimicrob Agents Chemother* 53: 1998–2004.
64. Inskeep WP, Macur RE, Hamamura N, Warelou TP, Ward SA, et al. (2007) Detection, diversity and expression of aerobic bacterial arsenite oxidase genes. *Environ Microbiol* 9: 934–43.
65. Quéminéur M, Heinrich-Salmeron A, Muller D, Lièvreumont D, Jauzein M, et al. (2008) Diversity surveys and evolutionary relationships of *aoxB* genes in aerobic arsenite-oxidizing bacteria. *Appl Environ Microbiol* 74: 4567–73.
66. Olsen GJ, Woese CR, Overbeek R (1994) The winds of (Evolutionary) change: breathing new life into microbiology. *J Bacteriol* 176: 1–6.
67. Jackson CR, Dugas SL (2003) Phylogenetic analysis of bacterial and archaeal *arsC* gene sequences suggests an ancient, common origin for arsenate reductase. *BMC Evol Biol* 3: 18.
68. Duval S, Ducluzeau AL, Nitschke W, Schoepp-Cothenet B (2008) Enzyme phylogenies as markers for the oxidation state of the environment: the case of respiratory arsenate reductase and related enzymes. *BMC Evol Biol* 8: 206.
69. Andres J, Arsène-Ploetze F, Barbe V, Brochier-Armanet C, Cleiss-Arnold J, et al. (2013) Life in an arsenic-containing gold mine: genome and physiology of the autotrophic arsenite-oxidizing bacterium *Rhizobium* sp. NT-26. *Genome Biol Evol* 5: 934–953.