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Metals and Breast Cancer

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Abstract

Metalloestrogens are metals that activate the estrogen receptor in the absence of estradiol. The metalloestrogens fall into two subclasses: metal/metalloid anions and bivalent cationic metals. The metal/metalloid anions include compounds such as arsenite, nitrite, selenite, and vanadate while the bivalent cations include metals such as cadmium, calcium, cobalt, copper, nickel, chromium, lead, mercury, and tin. The best studied metalloestrogen is cadmium. It is a heavy metal and a prevalent environmental contaminant with no known physiological function. This review addresses our current understanding of the mechanism by which cadmium and the bivalent cationic metals activate estrogen receptor- α . The review also summarizes the *in vitro* and *in vivo* evidence that cadmium functions as an estrogen and the potential role of cadmium in breast cancer.

Keywords

Estrogens; Estrogen receptor; Metalloestrogens; Metals; Endocrine disruptors; Breast cancer

Introduction

It has been suggested that the high incidence of hormone related diseases, such as breast cancer, is due, in part, to the presence of environmental estrogens. In fact, a number of chemicals present in the environment demonstrate estrogenlike activity [1]. Phytoestrogens, such as coumestrol and the isoflavone genistein, are naturally occurring nonsteroidal compounds that are derived from plants, while xenoestrogens, such as bisphenol A, and the polychlorinated biphenyls dichlorodiphenyltrichloroethane and its metabolite p, p'-dichlorodiphenyl dichloroethylene, are synthetic chemicals that activate the estrogen receptor (ER). Many of the phytoestrogens are flavonoids whereas the xenoestrogens are a structurally diverse group of compounds that have no common structural motif. Many have one or two aromatic rings and may be chlorinated giving them a negative charge. In contrast to phytoestrogens and xenoestrogens, metalloestrogens are small ionic metals and metalloids that also activate the estrogen receptor. The metalloestrogens fall into two separate subclasses, oxyanions that include arsenite, antimony, nitrite, selenite, and vanadate and bivalent cations that include cadmium, calcium, cobalt, copper, nickel, chromium, lead, mercury, and tin [2-10]. This review focuses on the bivalent cationic metalloestrogens and addresses our current understanding of their mechanism of action and role in breast cancer.

Estrogen Receptor-Alpha Structure and Function

Many of the actions of estrogens in the breast are mediated by two isoforms of the estrogen receptor, ER α and ER β . The mitogenic actions of the hormone are mediated by ER α while the antimitogenic actions are mediated by ER β [reviewed in 11]. Although ER α is a ligand activated transcription factor that belongs to the superfamily of nuclear receptors [12], the receptor has both genomic and nongenomic functions. Similar to other nuclear receptors, ER α is divided into regions A through F (Fig. 1) [13]. The N-terminal A/B region contains the transactivation function-1 (AF-1) domain that is involved in protein-protein interactions and plays an important role in ligand-dependent and - independent activation of the receptor. Region C is the DNA binding domain and consists of two zinc finger motifs that are responsible for binding to estrogen response elements (ERE) in target genes. Region D is the hinge region and plays a role in dimerization of the receptor. Region E is the hormone, or ligand, binding domain (LBD) and contains the transactivation function-2 (AF-2) domain. The LBD is the most structurally and functionally complex region of the receptor that is responsible for ligand dependent activation, dimerization, and recruitment of cofactors to the receptor. In the absence of hormone, ER α is associated with a complex containing heat shock proteins and immunophilins that maintains the LBD in a high affinity, ligand binding conformation and prevents the receptor from dimerizing and binding to DNA and cofactors [14, 15] (Fig. 2). Following the binding of estradiol, the receptor is phosphorylated on serines in the A/B region that increases the activity of the AF-1 domain [16, 17] and a conformational change occurs in the LBD leading to the dissociation of the heat shock complex and the formation of the AF-2 domain, the coactivator binding site [18]. In the classical genomic pathway, the activated receptor then localizes in the nucleus, dimerizes, binds to an ERE, and recruits coactivators and RNA polymerase II to the promoters of target genes. In the nongenomic pathway, ER α activates the ERK1/2 and PI3-K/Akt signal transduction pathways [reviewed in 19, 20].

The ligand binding domain of ER α is also structurally similar to other nuclear receptors [18, 21-26] (Fig. 3) and contains 11 alpha helices (H1, H3-H12) folded into a three layered antiparallel α -helical sandwich. The central core of the LBD contains the ligand binding pocket that is formed by helices H5/6, H9, and H10 inserted between two layers of helices composed of H1-4, H7, H8, and H11 with helix H12 flanking the ligand binding pocket [24]. Based upon the crystal structure of RXR- α in the absence and presence of its ligand [23, 27], several major conformational changes are thought to occur in the LBD as a result of estradiol binding in the pocket (Fig. 3). In the absence of estradiol, helices H10 and H11 are separated by a short loop with helix H11 positioned at an angle to helix H10 and helix H12 positioned to the side of the pocket. As a result of hormone binding in the pocket, helix H12 is repositioned over the ligand binding pocket and helix H11 is repositioned adjacent to helix H10 forming a continuous helix. In the activated receptor, the repositioning of helix H12 over the central core creates the AF-2 domain and the newly formed helix H10/H11, together with helices H8 and H9, constitutes the dimerization domain.

Cross Talk Between Signal Transduction Pathways and ER α

In addition to estradiol, ER α is activated by growth factors and cytokines [28], however, the mechanisms by which their downstream signaling pathways cross talk with ER α are not fully understood. Growth factors and cytokines are thought to activate the receptor, in part, through the phosphorylation of the N-terminal AF-1 domain [29] and through the posttranslational modification of coactivators [30]. Growth factors, such as EGF and IGF-1, activate the Ras-Raf-MEK-MAPK kinase and PI3K-AKT pathways resulting in the phosphorylation of serines in the N-terminal A/B region of the receptor [reviewed in 31]. Although phosphorylation of the A/B region is responsible for the activation of the AF-1 domain, it does not account for the conformational changes necessary for the activation of the AF-2 domain suggesting that additional intracellular events are required for the activation of the C-terminal LBD. Calcium is a bivalent cationic metal and a second messenger in signal transduction pathways. Recent studies show that calcium mediates the cross talk between growth factor/cytokine signaling pathways and the LBD of ER α and more importantly, that calcium is a ligand of ER α that activates it in the absence of estradiol [8]. Activation of ER α by calcium in breast cancer cells results in the induction of estrogen responsive genes and in hormone independent proliferation.

In contrast to estradiol that binds in the ligand binding pocket, calcium binds and activates ER α through sites on the surface of the LBD [8]. The calcium interaction sites are located at the ends of helices that are repositioned upon activation of the receptor suggesting that the interaction of calcium with these sites induces a conformational change in the LBD similar to the conformational change induced by estradiol (Fig. 3). By interacting with different amino acids in a protein, metals, such as calcium, can alter secondary structure, promote local folding, and assemble different regions into one domain. In the lipase enzyme of *Pseudomonas* for example, the active site of the enzyme is covered by a lid that is formed by two helices separated by a turn. The interaction of calcium with amino acids on the two helices repositions the helices and opens the lid [32]. In the case of ER α , there are four potential calcium interaction sites in the LBD [8]. One of the calcium interaction sites is located at the interface of helices H10 and H11 suggesting that the interaction of the metal at

the interface of the helices converts the helix-loop-helix structure of the inactive conformation into a continuous helical structure, helix H10/H11, in the active conformation (Fig. 3). Two of the calcium interaction sites are located at the ends of helices; one site is on the C-terminal end of helix H11 and one site is on the N-terminal end of helix H12 suggesting that the interaction of the metal with these sites repositions helix H12 over the ligand binding pocket. The fourth calcium interaction site is located on helix 4 in close proximity to helix H12 suggesting that the fourth site may contribute to the repositioning and closing of helix H12 over the ligand binding pocket and the formation of the coactivator binding site. The proposed model, that calcium activates ER α by inducing conformational changes in the ligand binding domain that mimic the conformational changes induced by estradiol, remains to be tested.

The ability of calcium to mediate the cross talk between growth factors/cytokines and the LBD of ER α and to activate the receptor in the absence of estradiol suggests that the bivalent cationic metalloestrogens activate ER α by mimicking calcium. In support of this hypothesis, cadmium replaces calcium in many biological systems and assays [33-35]. In the case of ER α , cadmium and the other bivalent cationic metalloestrogens activate it through the LBD and require the same amino acids (cys381, glu523, and asp538) as calcium [3, 5]. Cadmium also competes with calcium for binding to the LBD of the receptor [8]. An ionic charge of +2 appears to be important as chromium(II), a bivalent cation, activates ER α whereas, chromium(III), a trivalent cation, does not activate the receptor [3]. Although an ionic charge of +2 appears to be an important attribute of this subclass of metalloestrogens, not all bivalent cations activate ER α , e.g., zinc, a bivalent cation, does not activate ER α [3, 5, 36]. Many of the bivalent cationic metalloestrogens also have an effective ionic radius that is similar to calcium. In addition to having a charge and ionic radius similar to calcium, the bivalent cationic metals bind with high affinity to the LBD and noncompetitively block the binding of estradiol [3, 5, 37]. Several studies demonstrate a high affinity interaction of cadmium with the LBD that blocks the binding of estradiol, whereas other studies fail to show that the metal interacts with human ER α and blocks the binding of hormone [3, 38]. The differences between the studies that demonstrate specific binding of cadmium to the receptor and studies that fail to demonstrate binding may be attributed to differences in experimental conditions. In whole cell binding assays with endogenously [3] or exogenously expressed ER α [3, 38], the intracellular concentrations of naturally occurring metal chelators, such as glutathione, can influence the binding of cadmium to the receptor. In cell-free binding assays with either purified full length receptor [37] or purified recombinant LBD [3], the concentrations of preservatives, such as dithiothreitol, and metal chelators, such as EDTA, can also influence binding of the metal. In addition, temperature [38], order of addition, and incubation time [3] can alter the binding of cadmium to ER α as well as the ability of the metal to block the binding of estradiol to the receptor. Although less well studied, other bivalent cationic metals mimic the binding of calcium to ER α . Chromium(II), copper, cobalt, nickel, lead, mercury, and tin bind with high affinity to the receptor, block the binding of estradiol, and interact with the same amino acids in the ligand binding domain [5]. The ability of cadmium and the other metals to mimic the ability of calcium to bind to the LBD and to require the same amino acids to activate the receptor suggests that bivalent cationic metalloestrogens activate ER α by mimicking calcium.

Estrogen-Like Effects of Metals in Vitro

There is a growing body of evidence that metalloestrogens activate ER α in vitro. Most of the published studies address the ability of cadmium to activate the genomic and nongenomic pathways of ER α and show that, similar to estradiol, cadmium induces the proliferation of estrogen dependent breast cancer cells [2, 36, 39, 40], increases the transcription and expression of estrogen regulated genes such as the progesterone receptor (PR) [2, 41], activates ER α in transfection assays [2, 3, 36, 39, 42], and increases signaling through the ERK1/2 and Akt pathways [41, 43, 44]. In addition to activating ER α , there is some evidence that cadmium activates the membrane estrogen receptor GPR30 [43]. Although the majority of studies find that cadmium activates ER α , some studies failed to demonstrate an estrogen-like effect of the metal. One study failed to find an estrogen-like effect of cadmium in breast cancer cells or in yeast that expressed the human ER α [45]; a second study also failed to find a genomic effect but demonstrated a nongenomic effect of the metal in breast cancer cells [44]. Cell culture conditions may explain, in part, the inability of cadmium to activate ER α in vitro. Our experience has been that the presence and amount of sulfates, phosphates, lipoic acid, and the polyanion putrescine in the culture media and the presence of sulfates in sulfatase-treated stripped serum interferes with the ability of the metal to elicit an estrogenlike response in vitro (unpublished data). Sulfates and phosphates form weak and insoluble cadmium salts while lipoic acid and putrescine chelate metals. Although less well studied, there is also evidence that the other bivalent metals activate ER α in vitro [5, 36, 46]. Copper, cobalt, nickel, lead, mercury, tin, and chromium(II) induce the proliferation of estrogen dependent breast cancer cells [5, 36, 46], increase the transcription and expression of estrogen regulated genes [5], and activate ER α in transfection assays [5, 36] supporting the estrogen-like effects of these bivalent cationic metals in vitro.

Estrogen-Like Effects of Metals in Vivo

There is also increasing evidence that cadmium activates the genomic and nongenomic pathways of ER α in vivo [4, 47-55]. The rodent uterotrophic response assay is the most commonly used laboratory model to test for the estrogenicity of compounds in vivo. Traditionally, ovariectomized pubertal and adult rats are employed in the assay but ovariectomized pubertal and adult mice are also used [56]. In more recent studies, immature animals rather than ovariectomized adult animals were employed because developing tissues are thought to be more susceptible to perturbation by hormones [57]. In the uterotrophic assay, the estrogenicity of a compound is determined by measuring one or more endpoints including the ability of the compound to increase uterine wet weight, increase epithelial cell height and number, increase the number of glands, and induce the expression of estrogen responsive genes such as progesterone receptor and complement C3. In addition to measuring endpoints in the uterus, the estrogenicity of a compound can be measured in the mammary gland by its ability to enhance development of the gland and to induce the expression of estrogen responsive genes. In the case of cadmium, the estrogenicity of the metal has been measured in the uterus and mammary glands of immature and ovariectomized rats and mice. In Sprague–Dawley rats ovariectomized at 28 days of age and treated 3 weeks later with a single intraperitoneal (ip) dose of cadmium of 5 ug/kg body

weight (bw), treatment with the metal resulted in a significant increase in weight and expression of the estrogen regulated genes PR and complement C3 in the uterus and a significant increase in epithelial density and expression of PR and complement C3 in the mammary glands [4]. More importantly, the estrogen-like activity of cadmium was blocked by an antiestrogen [4]. In contrast to animals ovariectomized at 28 days of age, a higher dose of cadmium of 800 ug/kgbw administered intraperitoneally for 3 days was needed to obtain a significant increase in uterine weight in Sprague–Dawley rats ovariectomized at 23 days of age and treated 3 weeks later [49] suggesting that the age at the time of ovariectomy influences the estrogenic activity of cadmium. The route of exposure also influences the estrogenic activity of the metal. In Wistar rats ovariectomized at 28 days of age and treated 2 weeks later, a single intraperitoneal dose of cadmium ranging from 50 ug/kgbw to 2 mg/kgbw resulted in a dose dependent increase in uterine weight and increased in expression of complement C3 at the highest dose [48]. In the small intestine, the single intraperitoneal dose of cadmium also resulted in an estrogen-like effect on gene expression [52]. In contrast to intraperitoneal administration, administration of cadmium by gavage for 3 days at doses ranging from 50 ug/kgbw to 4 mg/kgbw or through drinking water for 28 days at doses ranging from 400 ug/kgbw to 9 mg/kgbw had no effect on uterine weight [48, 52]. However, there was an estrogen-like effect on gene expression in both the uterus and small intestines [48, 52]. A similar dose effect of cadmium on uterine weight following intraperitoneal administration was also observed in Wistar rats ovariectomized at 28 days of age and treated 3 weeks later with either 120 ug/kgbw or 1.2 mg/kgbw for 3 days [53], or in Wistar animals ovariectomized at 35 to 42 days of age and treated 2 weeks later with a single dose of either 50 ug/kgbw or 2 mg/kg bw [55]. In the former study, the increase in uterine weight was accompanied by an increase in phosphorylation of ERK [53] and in the latter study, the increase in uterine weight was accompanied by an increase in PR and complement C3 expression [55] suggesting that exposure to cadmium activates both the nongenomic and genomic pathways in vivo. The estrogen-like effects of the metal have also been studied in mice. In immature mice, subcutaneous administration of cadmium at a dose of 5, 50, or 500 ug/kgbw for 3 days had no effect on uterine weight but increased the height of the uterine luminal epithelium in a dose dependent manner [50]. In contrast to low doses of cadmium, intraperitoneal administration of a much higher dose of cadmium (3 mg/kg bw for 5 days per week for 2 weeks) to immature CD-1 mice resulted in a decrease in uterine weight [47]. However, as the animals aged (4 months old), intraperitoneal administration of cadmium (2 mg/kgbw given 5 days per week for 7 weeks) increased uterine weight and mammary gland development as measured by an increase in the development of lobuloalveolar structures [47]. Interestingly, when the mice were ovariectomized prior to treatment with cadmium, an estrogen-like effect was observed in the mammary glands but not in the uterus [47]. Together, these studies suggest that the age and hormonal status of the mice influences the estrogenic response to cadmium. Diet also influences the estrogenic activity of the metal. C57BL/6 mice that express the luciferase gene under the control of an estrogen response element were ovariectomized at 4 to 6 months of age and two weeks later were exposed to cadmium for 21 days by gavage (1 ug/kgbw), in white bread (17.57 ug/kgbw), or in flaxseed supplemented bread (49.2 ug/kg) [54]. In the animals exposed to cadmium by gavage, there was an increase in uterine weight, an increase in luciferase in the chest, and an increase in progesterone receptor expression in white adipose tissue. However in the animals exposed to

cadmium in white bread, there was no increase in uterine weight but an increase in luciferase in the chest, thymus, and white adipose tissue and an increase in expression of prothymosin in the thymus and uterus. Similarly, in animals exposed to the metal in flaxseed supplemented bread, there was no increase in uterine weight but an increase in luciferase in the chest, thymus, and liver and an increase in prothymosin in the thymus and uterus. Taken together, the published studies provide compelling evidence that cadmium activates ER α in vivo and demonstrate that the response depends on the species and strain of rodents, age and hormonal status of the animals, the dose and route of exposure, and the target tissue.

Environmental Exposure to Metals and Breast Cancer

Metals have diverse biological functions from being essential to toxic and carcinogenic. Metals, such as chromium, cobalt, copper, and nickel, are essential metals required in trace amounts. Essential metals play an important role in metabolism and respiration, in membrane integrity and permeability, and in cell proliferation and death [58-61] where alterations in their concentration may result in disease or toxicity [60, 62, 63]. For example, essential metals at low concentrations function as components of enzymes but at high concentrations can inhibit enzyme activity [61, 62]. Metals, such as cadmium, lead, mercury, and tin, are nonessential metals and exert their toxic effects by mimicking or blocking the function of essential metals [58, 64].

Many metals are also carcinogens. Cadmium, chromium, and nickel are established human and animal carcinogens, while copper, lead, and mercury are probable carcinogens or co-carcinogens [65-72]. Occupational exposure to metals including cadmium, chromium, nickel, copper, cobalt, lead, and mercury is associated with an increased risk of lung cancer. Exposure to chromium is also linked to an increased risk of liver, larynx, esophagus, and gastrointestinal cancer. Occupational exposure to cadmium and nickel are linked to renal and prostate cancer and copper exposure is linked to non-Hodgkin's lymphoma and skin cancer. Lead and mercury exposures are associated with glioma and stomach cancers and prostate and bladder cancers, respectively. Women working in dentistry also have an increased frequency of precancerous lesions of the cervix that correlates with the length of employment [73]. In animal studies, nickel, cobalt, mercury, lead, and chromium (VI) induce sarcomas and carcinomas in the breasts, kidneys, lungs, liver, and pancreas and sarcomas at the site of injection [65, 74-77]. While some copper salts produce sarcomas in chickens and mice, other copper salts and chelates suppress the tumorigenicity of chemical carcinogens. There is no experimental data to implicate tin as a carcinogen.

The general population is exposed to metals primarily through the environment. Environmental exposure to cadmium occurs primarily through dietary sources, cigarette smoking, and, to a lesser degree, drinking water [78, 79]. In the United States, dietary studies in the late 1970s and early 1980s found that potato, grain, and cereal products accounted for the largest portion of cadmium intake by the adult male, contributing 24 and 36 %, respectively. Fluids, which include drinking water, accounted for 3.2 % of cadmium intake. In the United States, the amount of cadmium exposure from the diet is estimated to range from 0.12 to 0.331 ug/kgbw/day with the highest exposure in children 1 to 6 years of age [78-80]. In the Ruhr district of Germany, exposure to cadmium from the diet ranges

from 0.37 ug/kg bw/day in adults, 0.49 ug/kgbw/day in children (mean age 3.8 years), to 0.17 ug/kgbw/day in young children (mean age 1.8 years) while in Amrum, an island in the German North Sea, cadmium exposure from the diet is 0.39 ug/kg bw/day in adults and 0.35 ug/kgbw/day in children [65, 81, 82]. Similar exposures are seen in the United Kingdom and Sweden, i.e., 0.2 and 0.25 ug/kg bw/day, respectively [83, 84]. Cigarette smoke is also an important source of human exposure to cadmium, reflecting the high efficiency of pulmonary absorption of the inhaled metal [69, 85]. Cadmium intake from one pack of cigarettes per day is estimated to be from 2 to 4 ug. The concentration of cadmium in the kidneys of nonsmokers is approximately 15 to 20 ug/gm tissue, while in smokers, the concentration doubles to 30 to 40 ug/gm tissue. High concentrations of cadmium are also present in the breasts of healthy women (20 to 30 ug/gm tissue) [86]. Animal studies show that cadmium can be transferred through the placenta to the developing fetus [87] and that transfer to the fetus is dose related and increases with advancing gestation [88]. Low but detectable amounts of cadmium are found in the gastrointestinal tract, liver, kidneys, and blood of the newborn [87] but by age 30, the body burden may reach 30 mg [89]. The estimated half-life of cadmium in the body ranges from 10 to 30 years [69] which may account for the significant accumulation of the metal in the body.

Cadmium is widely distributed in the earth and is mined for use principally in galvanizing and electroplating, in batteries, in electrical conductors, and in the manufacture of pigments, plastics stabilizers, and phosphate fertilizers [90]. The primary source of cadmium in the environment is due to industrial contamination and most contamination is a byproduct of smelters [91]. In the 1980s, the total atmospheric emissions of cadmium were estimated to be about 635,000 kg annually. The level of cadmium in streams and rivers generally reflects the level of cadmium contamination in the air and soil. Cadmium has been detected in surface water and ground water samples taken at about 70 % of hazardous waste sites. It has been detected in water samples collected from all of the Great Lakes. In a survey in New Jersey, cadmium was detected in 100 % of surface water and ground water samples, where concentrations as high as 405 ug/l (3.6 uM) were detected. California, Colorado, Idaho, and Maine also had high concentrations of cadmium in surface and ground water samples (340 to 2,000 ug/l, i.e., 3 to 18 uM). Although cadmium levels are high in many water sources, most drinking water in the United States probably does not contain more than 1 ug/l (9 nM). Although most environmental cadmium is due to industrial contamination, water may also contain cadmium as a result of leaching from the soil and the dissolution of cadmium from underlying geologic formations, especially in areas where soft, acidic waters are common. Stratiform deposits with some of the most important cadmium deposits are located in several states, including Missouri, Tennessee, the Missouri/Kansas/Oklahoma border area, the Wisconsin/Illinois border area, and Pennsylvania [90].

Environmental exposure to other metalloestrogens is also significant. In water and soil, the concentrations of chromium, mercury, and copper are 1 to 800 ug/l and 40 to 459 mg/kg, respectively [92-94]. The amounts of nickel, chromium, mercury, lead, and copper in fish range from 81 ng/gm to 328 mg/gm and in grain, the amount of copper ranges from 1 to 14 ug/gm [95]. Exposure of humans to these metals occurs primarily through dietary sources of food and water [96, 97], air, cigarette smoke [59], and occupational exposure [59, 96, 98] and can lead to significant accumulation in the body. The average daily intake of chromium,

mercury, and nickel is estimated to be from 0.28 to 25 ug/day while the daily intake of copper ranges from 1.46 to 1.63 mg/day [99]. Lead exposure results in significant accumulation in hair and toenails, 3.8 to 10.1 ug/gm [100] and in breast milk, 36 mg/l [95].

There is increasing epidemiological evidence linking exposure to cadmium with an increased risk of developing breast cancer. The first study was a hypothesis-generating case-control study that examined the death certificates of over 33,000 deaths attributed to breast cancer and over 117,000 non-cancer deaths between 1984 and 1989 [101]. The death certificates were coded for occupation and industry. The study found that occupational exposure to cadmium was associated with an approximate 8 to 20 % increase in breast cancer risk among white women and a 50 to 130 % increase in risk among African-American women. As acknowledged by the authors, the method of determining exposure (e.g., a single occupation listed on the death certificate, based on information from a proxy) may have subjected the study to substantial nondifferential misclassification which can markedly attenuate the observed odds ratios leading to an underestimation of risk. In addition, there was no information on other breast cancer risk factors that could have further distorted the results. A second epidemiological study in a retrospective cohort of working Swedish women also suggests a link between occupational exposure to cadmium and an increased risk of breast cancer [102]. In this study, women employed as metal platers and coaters had the highest standardized incidence ratio (relative risk: 2.38). Metal plating and coating exposes workers to cadmium, hexavalent chromium, and organic solvents. A population based case-control study of nonoccupationally exposed women in Wisconsin, that measured cadmium in urine, found that women in the highest cadmium quartile had more than a two-fold increased risk of breast cancer (odds ratio [OR]: 2.29; 95 % confidence interval [CI]; 1.3–4.2) compared to women in the lowest cadmium quartile and estimated that approximately 36 % of breast cancer may be attributed to exposure to the metal [103]. A second population based case-control sample of women living in Long Island also found a similar association [104]. Women in the highest cadmium quartile had more than a two-fold increased risk of breast cancer (OR: 2.69; 95 % CI; 1.07–6.78) compared to women in the lowest cadmium quartile. The latter study also investigated a cross-sectional U.S. probability sample from the National Health and Nutrition Examination Survey (NHANES 1999–2008). In the NHANES sample, the odds for breast cancer were significant and elevated for women in the third cadmium quartile (OR: 2.50; 95 % CI; 1.11–5.63) and marginally significant for the women in the fourth cadmium quartile. Similar to the study of Wisconsin women, the Long Island study estimated that approximately 35 % of breast cancer in the U.S. may be attributed to cadmium exposure. Since the case-control studies were conducted in breast cancer patients, the studies do not clearly establish whether cadmium is associated with the risk of developing breast cancer or is a consequence of the disease. A recent population-based prospective cohort study shows that long term dietary intake of cadmium is associated with an increased risk of breast cancer in postmenopausal women [105] suggesting a causal effect of cadmium in the development of the disease. Endometrial cancer is also an estrogen related cancer. A population-based prospective cohort study also showed that long term dietary intake of cadmium was associated with a 2.9-fold increased risk of endometrial cancer in postmenopausal women (95 % CI; 1.05–7.79) [106] providing additional evidence linking exposure to cadmium with an increased risk of hormone dependent cancer. Although

these epidemiological studies suggest a link between cadmium and breast cancer, more experimental and epidemiological studies are required to establish a cause and effect association between exposure to the metal and the development of the disease.

Although not as well studied, there is some evidence linking other metals with estrogen-like activity to breast cancer. The concentrations of copper, cobalt, and tin are significantly elevated in the serum of breast cancer patients and vary with the stage of the disease, the highest concentrations being observed in advanced stages [107-113]. Serum levels of copper are also higher in premenopausal than in postmenopausal breast cancer patients. A marginally significant association is also observed between toenail levels of chromium and breast cancer risk in postmenopausal women but an inverse association is found among premenopausal women [114].

Summary

Breast cancer is an epidemic, yet the underlying causes of the disease are largely unknown. The prominence of estrogens in the etiology of breast cancer has led to the suggestion that exposure to environmental estrogens may increase the risk of developing the disease. Metalloestrogens are small ionic metals and metalloids that include metal/metalloid anions and bivalent cations, such as cadmium, calcium, cobalt, copper, nickel, chromium, lead, mercury, and tin. Because metalloestrogens activate the estrogen receptor in the absence of estradiol, exposure to these metals may increase the risk of developing breast cancer. In support of this hypothesis, environmental exposure to many of the metalloestrogens is widespread and has increased significantly over the last 50 to 60 years. Many of the metalloestrogens also have a long biological half life (e.g., cadmium has a half life of 10 to 30 years) and accumulate in the body and in the breast. There is also credible experimental evidence that cadmium activates ER α in vitro and in vivo as well as increasing epidemiological evidence linking cadmium to breast cancer. Although there is evidence linking exposure to the metal with breast cancer, the role of cadmium and other metalloestrogens as causative agents in the etiology of the disease remains to be established.

Abbreviations

AF-1	transactivation function-1
AF-2	transactivation function-2
Akt	serine/threonine specific kinase
bw	body weight
EDTA	ethylenediaminetetraacetic acid
ERα	estrogen receptor-alpha
ERE	estrogen response element
ERK	extracellular signal-regulated kinases
GPR30	G protein-coupled receptor 30

LBD	ligand binding domain
PR	progesterone receptor

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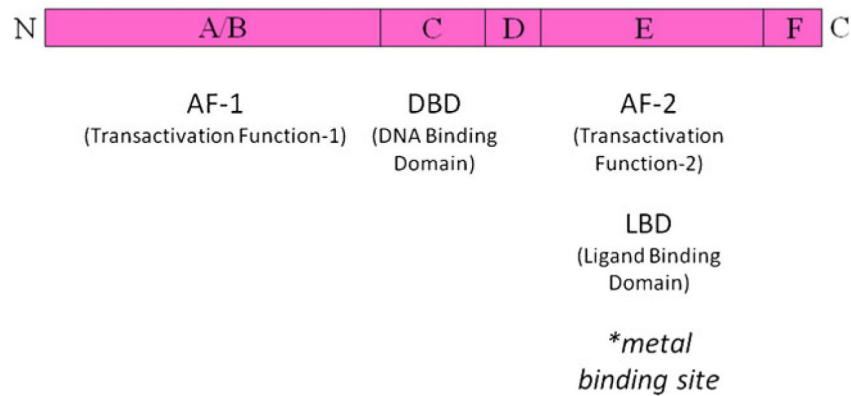
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Structure and Function of Estrogen Receptor- α **Fig. 1.**

Structure and function of estrogen receptor-alpha. Based on similarities to other nuclear receptors, ER α is divided into regions A through F. The N-terminal A/B region contains the transactivation function-1 (AF-1) domain that plays an important role in ligand dependent and independent activation of the receptor. Region C is the DNA binding domain (DBD) that is responsible for binding to target genes. Region D is the hinge region and plays a role in dimerization of the receptor. Region E is the ligand binding domain (LBD) that binds hormone and metals. Region E contains the transactivation function-2 (AF-2) domain that is responsible for ligand dependent activation, dimerization, and recruitment of cofactors to the receptor

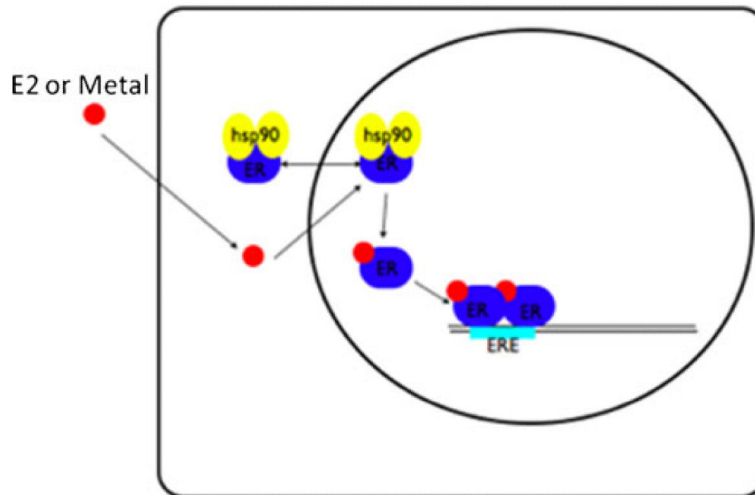


Fig. 2.

Classical genomic pathway of estrogen receptor- α : proposed role of metals. In the absence of ligand, ER α is associated with a complex containing heat shock proteins and immunophilins that maintains the LBD in a high affinity, ligand binding conformation and prevents the receptor from dimerizing and binding to DNA and cofactors. Following the binding of metals, it is proposed that, similar to estradiol, the receptor is phosphorylated on serines in the A/B region that increases the activity of the AF-1 domain and a conformational change occurs in the LBD that results in the dissociation of the heat shock complex and the formation of the AF-2 domain, the coactivator binding site. The activated receptor then localizes in the nucleus, dimerizes, binds to an ERE, and recruits coactivators and RNA polymerase II to the promoters of target genes

Conformational Changes in the Ligand Binding Domain - Model

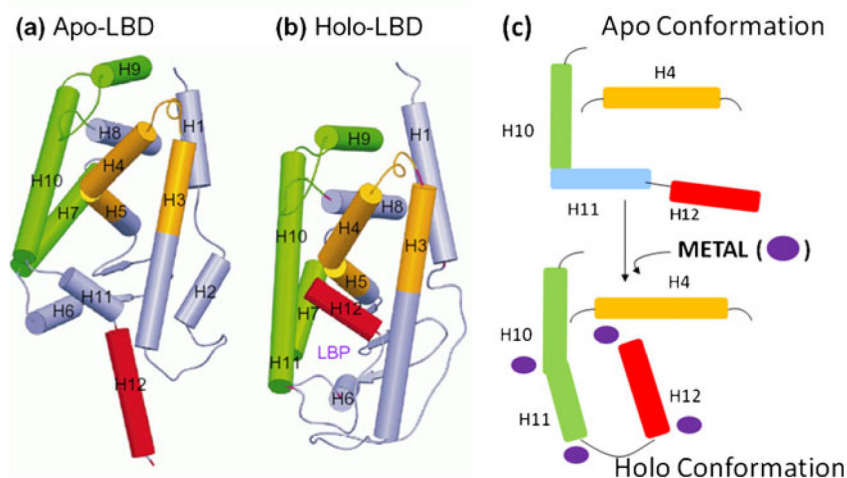


Fig. 3. Model of conformational changes in the ligand binding domain induced by metals. Schematic drawing of the unliganded (apo) ligand binding domain of retinoid X receptor **(a)** and the agonist bound (holo) ligand binding domain of retinoid receptor **(b)**. Comparison of helices H4, H10, H11, and H12 in the liganded (apo) and unliganded (holo) conformation **(c)**. In the absence of ligand, helices H10 and H11 are separated by a short loop with helix H11 positioned at an angle to helix H10 and helix H12 is positioned to the side of the ligand binding pocket (LBP). In the presence of ligand, helix H12 is repositioned over the ligand binding pocket and helix H11 is repositioned adjacent to helix H10 forming a continuous helix. In the proposed model, interaction of the metal (shown as *purple circles*) with ER α induces conformational changes that mimic the conformational changes induced by the ligand estradiol. Adapted with permission from Bourguet et al. Trends Pharmacol Sci 2000;21:381-8