

## Abundance, diversity, and feeding behavior of coral reef butterflyfishes at Lord Howe Island

Morgan S. Pratchett<sup>1</sup>, Andrew S. Hoey<sup>1</sup>, Christopher Cvitanovic<sup>2,3</sup>, Jean-Paul A. Hobbs<sup>4</sup> & Christopher J. Fulton<sup>3</sup>

<sup>1</sup>ARC Centre of Excellence for Coral Reef Studies, James Cook University, Townsville, Queensland 4811, Australia

<sup>2</sup>Oceans and Atmosphere Flagship, CSIRO, Hobart, Tasmania 7000, Australia

<sup>3</sup>ARC Centre of Excellence for Coral Reef Studies, Research School of Biology, Australian National University, Canberra, Australian Capital Territory 0200, Australia

<sup>4</sup>Department of Environment and Agriculture, Curtin University, Perth, Western Australia 6845, Australia

### Keywords

Chaetodontidae, corallivore, disturbance, ecological function, endemism, schooling, selectivity, specialization.

### Correspondence

Morgan S. Pratchett, ARC Centre of Excellence for Coral Reef Studies, James Cook University, Townsville, Qld 4811, Australia.  
Tel: +61 7 47815747; Fax: +61 7 47816722;  
E-mail: morgan.pratchett@jcu.edu.au

### Funding Information

Funding was provided by ARC Centre of Excellence for Coral Reef Studies.

Received: 18 June 2014; Revised: 24 July 2014; Accepted: 25 July 2014

*Ecology and Evolution* 2014; 4(18): 3612–3625

doi: 10.1002/ece3.1208

### Abstract

Endemic species are assumed to have a high risk of extinction because their restricted geographic range is often associated with low abundance and high ecological specialization. This study examines the abundance of *Chaetodon* butterflyfishes at Lord Howe Island in the south-west Pacific, and compares interspecific differences in local abundance to the feeding behavior and geographic range of these species. Contrary to expected correlations between abundance and geographic range, the single most abundant species of butterflyfish was *Chaetodon tricinctus*, which is endemic to Lord Howe Island and adjacent reefs; densities of *C. tricinctus* ( $14.1 \pm 2.1$  SE fish per 200m<sup>2</sup>) were >3 times higher than the next most abundant butterflyfish (*Chaetodon melannotus*), and even more abundant than many other geographically widespread species. Dietary breadth for the five dominant butterflyfishes at Lord Howe Island was weakly and generally negative correlated with abundance. The endemic *C. tricinctus* was a distinct outlier in this relationship, though our extensive feeding observations suggest some issues with the measurements of dietary breadth for this species. Field observations revealed that all bites taken on benthic substrates by *C. tricinctus* were from scleractinian corals, but adults rarely, if ever, took bites from the benthos, suggesting that they may be feeding nocturnally and/or using mid-water prey, such as plankton. Alternatively, the energetic demands of *C. tricinctus* may be fundamentally different to other coral-feeding butterflyfishes. Neither dietary specialization nor geographic range accounts for interspecific variation in abundance of coral reef butterflyfishes at Lord Howe Island, while much more work on the foraging behavior and population dynamics of *C. tricinctus* will be required to understand its' abundance at this location.

### Introduction

Endemic species are an important component of biodiversity but are also considered to be disproportionately affected by disturbances, and more likely to go extinct because relatively moderate disturbances can simultaneously affect the entire population (McKinney 1997; Gaston 1998; Roberts et al. 2002). Moreover, geographic range is often correlated with abundance (e.g., Lawton 1993; Gaston 1994, 1996; Brown et al. 1995; McKinney 1997), further increasing the risk of extinction for restricted range species (Gaston et al. 1997; Gaston 1998).

This double jeopardy of extinction risk may also be further compounded if small range size is associated with other traits (e.g., ecologically specialization and low dispersal: Gaston et al. 1997; Malcolm et al. 2006; Pimm et al. 2014), making these species even more vulnerable to extinction (Davies et al. 2004; Brook et al. 2008; Olden et al. 2008).

Ecological specialization (the extent to which species specialize in their use of prey or habitat resources) is increasingly considered alongside population size and geographical range as a key determinant of extinction risk (e.g., McKinney 1997; Davies et al. 2004; Dulvy et al.

2004; Koh *et al.* 2004; Brook *et al.* 2008). Ecological theory (e.g., Brown 1984) suggests that specialized species should have narrower geographic ranges and be less abundant than generalist counterparts, but empirical data (e.g., Gaston *et al.* 1997; Manne and Pimm 2001; Päävönen *et al.* 2005; Reif *et al.* 2006; Hobbs *et al.* 2010, 2011; Berkström *et al.* 2012) does not always support the theory. An alternative explanation is that extinction filtering promotes persistence of species with compensatory relationships between range size, ecological specialization and population size that reduce the risk of extinction (e.g., Johnson 1998; Williams *et al.* 2006).

Despite the importance of ecological specialization for the biology, ecology and evolution of animals (e.g., Futuyma and Moreno 1988), ecological specialization is either rarely or poorly quantified (Devictor *et al.* 2010). Coral-feeding butterflyfishes (*Chaetodon*; Chaetodontidae) are an ideal group to study ecological specialization because their feeding behavior and dietary composition is easily measured, as is the differential availability of alternative prey (e.g., Berumen *et al.* 2005; Blowes *et al.* 2013; Noble *et al.* 2014). This enables direct estimates of dietary specialization across gradients of prey availability (e.g., Lawton *et al.* 2012), clearly distinguishing species that display distinct preferences regardless of prey availability (fundamental or obligate specialists) versus those that vary in their patterns of prey use simply to make use of locally abundant prey types (realized or facultative specialists). Moreover, sympatric butterflyfishes often exhibit significant variation in dietary selectivity, ranging from species that feed almost exclusively on just one coral species (e.g., *Chaetodon trifascialis*, Pratchett 2005; Pratchett *et al.* 2013a) to species that feed on >50 coral species, often in direct accordance with their relative abundance (e.g., *Chaetodon lunulatus*, Pratchett 2005).

Butterflyfishes are among the best-studied group of coral reef fishes (Pratchett 2014), owing partly to their inherent reliance on live coral for food and associated vulnerability to significant and widespread declines in live coral cover (e.g., Wilson *et al.* 2006, 2014). Pratchett *et al.* (2008, 2011) showed that interspecific differences in the vulnerability of butterflyfishes to coral loss are greatest among species for which corals represent >80% of total food intake (termed obligate corallivores, Cole *et al.* 2008). However, even among obligate coral-feeding fishes, responses to coral loss vary depending upon the extent to which species are more or less specialized in their use of different coral prey (Pratchett *et al.* 2008). There is, therefore, a definite need to better understand the specific foraging behavior and ecological specialization of coral reef butterflyfishes, especially among those species that are geographically restricted and exposed to local coral depletion (Lawton *et al.* 2012).

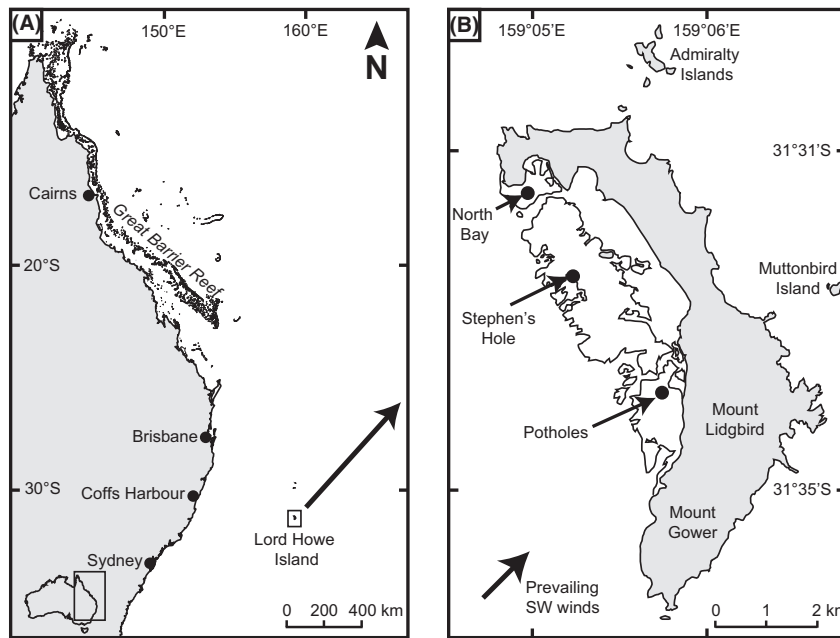
In this study, we explore the abundance, diversity and feeding behavior of *Chaetodon* butterflyfishes Lord Howe Island, and assess whether local abundance of individual species is related to their dietary specialization and/or geographic range. Lord Howe Island is the world's southernmost coral reef, with fish faunas comprising a mix of both tropical and temperate species (Zann 2000), and a relatively high number of endemics (Randall 1976). Previous studies conducted within tropical coral-dominated environments have revealed that specialist coral-feeding species tend to dominate butterflyfish assemblages (Emslie *et al.* 2010; Pratchett *et al.* 2013a), but coral-feeding fishes are under-represented at some marginal or peripheral coral reef locations (e.g., Pratchett *et al.* 2013b). Given high cover of corals across much of the reef habitat at Lord Howe Island (Hoey *et al.* 2011), we would expect to find a high abundance of coral-feeding butterflyfishes, though the isolation and extreme latitude may moderate the abundance of some species. In this study, direct feeding observations were used to quantify both feeding rates and diet (or feeding substrata) of dominant butterflyfishes. Notably, this is the first study on the feeding habits of the three-striped butterflyfish (*Chaetodon tricinctus*), which is endemic to Lord Howe Island and nearby reefs (Hobbs *et al.* 2009; van der Meer *et al.* 2013).

## Methods

### Field surveys

Lord Howe Island (31°32'S, 159°04'E) is located 630 km east of the Australian mainland in the Tasman Sea (Fig. 1A). The western side of the island is dominated by an extensive lagoon with a high cover (*ca.* 30%), but low diversity, of scleractinian corals (e.g., Hoey *et al.* 2011). Sampling for this study was undertaken at three sites (North Bay, Stephen's Hole and Potholes) equally spaced along the lagoon in areas of distinct platform reef <2 m depth, separated by deeper (4–6 m) sandy areas (Fig. 1B).

Butterflyfish abundance was quantified using underwater visual census (UVC) along haphazardly placed 50 × 4 m belt transects ( $n = 12$  replicates per site) in December 2011. Butterflyfishes were surveyed while simultaneously deploying a 50-m transect tape to delineate transect length. All butterflyfishes 2 m either side of the transect midline were then recorded to species, as well as estimating their total length (TL, to nearest cm) and recording group size. Coral cover and benthic composition were quantified using point-intercept transects (following Pratchett *et al.* 2004, 2011) to record the specific substratum type underlying uniformly spaced points (0.5 m apart) along the length of each 50 m transect. Scleractinian (hard) corals, alcyonacean (soft) corals, and



**Figure 1.** Map showing (A) geographic location of Lord Howe Island, and (B) the location of the three study sites on Lord Howe Island (North Bay, Stephen's Hole, and Potholes) used to quantify butterflyfish assemblages.

macroalgae (>5 mm) were identified to genus (and *Acropora* hard corals were further defined to tabulate or arborescent growth forms), with other substratum types categorized as sand/rubble or pavement.

### Feeding observations

To characterize and compare the feeding rates and diets of butterflyfishes at Lord Howe Island, the range of prey types, and the proportional use of different prey types by each species of butterflyfish (use was defined as an observed bite by the individual on a prey type), was quantified using replicate 3-min feeding observations following Pratchett (2005). Feeding observations were conducted during a similar time of year in each two consecutive years, May 2010 and June 2011. Feeding observations only commenced after the focal individual had taken their first bite, or 3-min after the observation started to allow fish to acclimate to observer presence. Observations were aborted if the focal individual fled or sought shelter from the observer. During each feeding observation, the total number of bites taken from different genera of hard coral, soft coral or any other noncoral macroinvertebrate was recorded. For the dominant coral genera, *Acropora*, we also distinguished between tabular (e.g., *Acropora glauca*), and arborescent (e.g., *Acropora yongei*) colonies. The number of bites taken from other reef substrata (i.e., consolidated reef pavement, coral rubble, or sand) that were not obviously occupied by corals or macroinvertebrates was also recorded. A minimum of 20 feeding observations were conducted for each of the

five most common butterflyfish species recorded at Lord Howe Island: *C. lunulatus*, *Chaetodon melannotus*, *Chaetodon plebeius*, *C. tricinctus*, and *C. trifascialis*. Increased sampling effort was applied to the endemic *C. tricinctus* (186 of 419 feeding observation) due to apparent size-based differences in feeding behavior (discussed below).

### Data analyses

Spatial variation in the abundance and composition of *Chaetodon* butterflyfishes and categories of reef substratum were examined across the three sample sites (North Bay, Stephen's Hole, and Potholes) using permutational multivariate analysis of variance (PERMANOVA). PERMANOVAs were conducted with 9999 permutations of the raw data constructed into resemblance matrices for the *Chaetodon* assemblages using a modified Gower Log<sub>10</sub> measure (Anderson *et al.* 2006), and for the reef substratum categories using a Bray-Curtis similarity measure on square-root transformed data for the 36 transects (Anderson *et al.* 2008). Ordinations were used to visualize structure within the reef substratum and *Chaetodon* assemblages via principal coordinates analysis (PCO) on the same resemblance matrices. Pairwise PERMANOVA was used to further explore differences between sites. PCOs were optimized with vector overlays of raw Pearson correlations (limited to  $r > 0.4$ ) and bubble plots to explore key *Chaetodon* species and substratum categories underlying spatial structure in this reef assemblage.

The extent to which spatial differences in *Chaetodon* assemblages could be explained by reef habitat composition

was explored by distance-based linear models (DISTLM), which were based on the same resemblance matrices above, and used Akaike Information Criteria for finite samples (AICc) to select the “best” models with a range of settings (models with either 1, 2, 3, or 4 substratum categories incorporated) from all of the possible combinations of habitat predictor variables (Anderson *et al.* 2008). As recommended by Anderson *et al.* (2008), we checked for multicollinearity among possible habitat predictor variables using draftsman plots. This led to exclusion of abiotic substratum categories (sand/rubble, pavement) from the DISTLM analysis, as they were strongly (negatively) correlated with biotic categories (chiefly scleractinian corals). All analyses and ordinations were performed in PRIMER (version 6.1.16) with PERMANOVA+ (version 1.0.6).

To compare dietary composition and feeding selectivity among *Chaetodon* butterflyfishes, forage ratios were calculated following Manly *et al.* (2002), which illustrate the use of each prey category (number of bites taken) relative to the availability of each prey type across the three study sites. Bonferroni-corrected 95% confidence limits were calculated for each prey category used by each butterflyfish species to establish the significance of prey selectivity. Selection ratios  $-95\%$  CI that were  $>1$  indicate that prey that were used significantly more than expected based on their availability (*i.e.*, preferred), while ratios  $+95\%$  CI that were  $<1$  indicate prey that were used disproportionately less than expected (*i.e.*, avoided).

Variation in both bite rates and diet breadth were analyzed using two-way ANOVAs to detect differences among species (*C. lunulatus*, *C. melannotus*, *C. plebeius*, *C. tricinctus*, and *C. trifascialis*) and among locations (North Bay, Stephen's Hole, and Potholes), and Tukey's post hoc test was used to reveal major differences among species. Raw data on the number of bites taken by each individual butterflyfish were square-root transformed prior to analyses to reduce the influence of occasional very large values. Replicate estimates of diet breadth were based on the number of distinct coral types that were consumed by each individual during the 3-min feeding observation; specialist species are expected to concentrate feeding on only 1–2 coral species, whereas generalists may feed on predominant or preferred prey while actively foraging across a range of different prey types (Pratchett 2014). One-way ANOVA was used to test for size-related differences in feeding rates for *C. tricinctus*, comparing among individuals with an estimated TL of  $<5$  cm, 5–10 cm, and  $>10$  cm. It was apparent during feeding observations that bite rates were highest among the smallest size classes and tended to decline with increasing size, so a minimum of 20 feeding observations were conducted for each size class. Similar analyses were not performed for other *Chaetodon* butterflyfishes, mainly because there

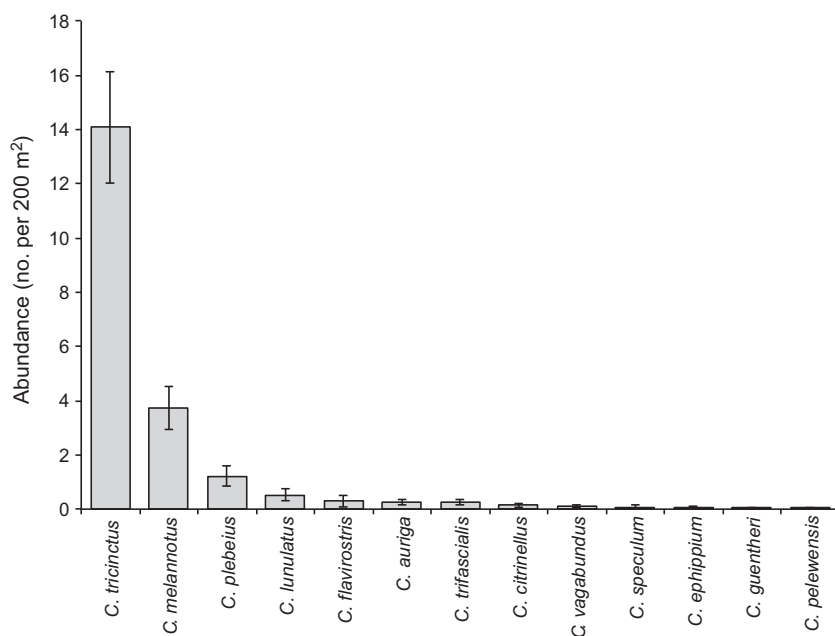
was much less variation in the size of fishes, and so most feeding observations were of larger (presumably adult) individuals.

After accounting for spatial variation in abundance of different butterflyfishes, overall abundance of each species was determined by averaging across all sites. This aggregate measure of individual abundance was then used to examine whether interspecific differences in local abundance are related to geographic range (across all species present) and diet breadth (for subset of species for which dietary composition was measured). To compare geographic range among butterflyfishes, we used published estimates of maximal area of occurrence (Jones *et al.* 2002). Diet breadth was calculated as described above.

## Results

A total of 13 species of *Chaetodon* butterflyfish were recorded across the three lagoonal reef sites at Lord Howe Island, although six of these species (*Chaetodon citrinellus*, *Chaetodon vagabundus*, *Chaetodon speculum*, *Chaetodon ephippium*, *Chaetodon guentheri*, and *Chaetodon pelewenisis*) were rare (Fig. 2). Butterflyfish assemblages were significantly different among sites (PERMANOVA: pseudo- $F_{2,33} = 2.98$ ,  $P = 0.003$ ), largely due to significant differences between North Bay and the other sites (pseudo- $t_{22} = 1.79$ ,  $P = 0.009$  and pseudo- $t_{22} = 2.30$ ,  $P = 0.001$  pairwise comparisons with Potholes and Stephen's Hole, respectively), with no significant difference between Potholes and Stephen's Hole (pseudo- $t_{22} = 0.73$ ,  $P = 0.760$ ). Ordination revealed that spatial variation in *Chaetodon* assemblages was largely due to variation in abundance of five abundant species: *C. tricinctus*, *C. melannotus*, *C. plebeius*, *C. lunulatus*, and *C. trifascialis* (Fig. 3A). Densities of both *C. tricinctus* and *C. melannotus* were 2–3 times higher at North Bay (average = 23.0 and 7.42 fishes per 200 m<sup>2</sup>, respectively) compared to Stephen's Hole and Potholes.

Similarly, reef substratum composition was significantly different among sites (pseudo- $F_{2,33} = 3.34$ ,  $P = 0.009$ ), particularly between North Bay and the other two sites (pseudo- $t_{22} = 2.33$ ,  $P = 0.004$  and pseudo- $t_{22} = 1.87$ ,  $P = 0.034$ ), but not between Potholes and Stephen's Hole (pseudo- $t_{22} = 1.27$ ,  $P = 0.176$ ). Spatial variation in reef habitat structure was largely attributable to seven benthic categories: sand/rubble, pavement, *Acropora* (arborescent), *Acropora* (tabular), *Pocillopora*, *Isopora* and *Porites* (Fig. 3B). Cover of scleractinian corals was much higher at North Bay (43.4%) compared to Stephen's Hole (38.7%) and Potholes (30.3%), mostly because of higher cover of arborescent *Acropora* (32.1%), which was the dominant coral at North Bay (comprised 73.9% of all



**Figure 2.** Mean ( $\pm$ SE) abundance of all *Chaetodon* butterflyfishes recorded at Lord Howe Island. Data are pooled across all sites to highlight relative abundance of different species.

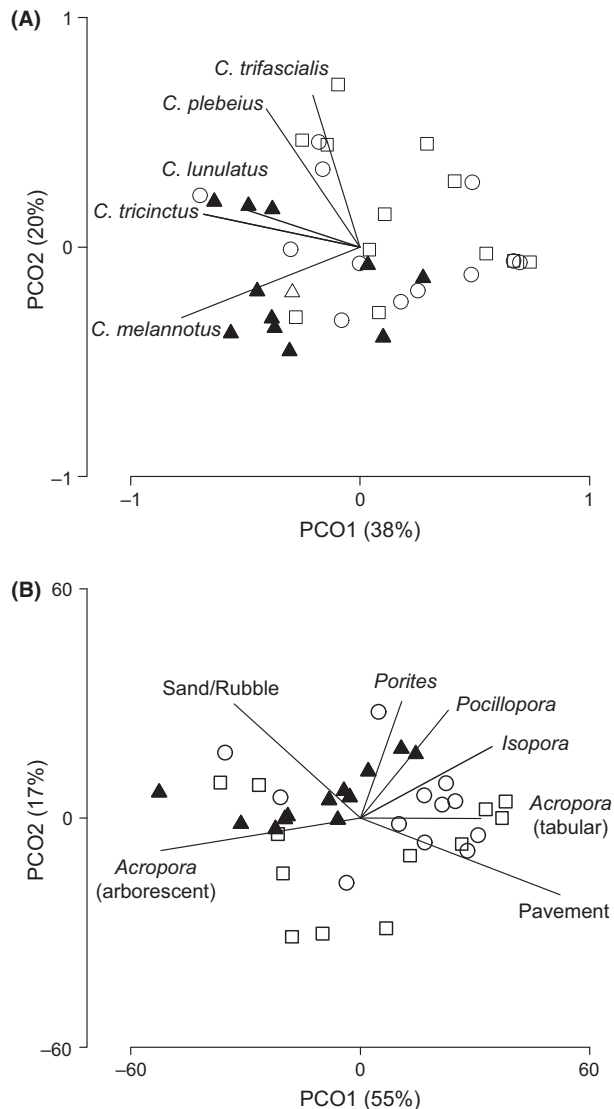
coral). DISTLM marginal tests indicated scleractinian corals accounted for 46.0% of variation in *Chaetodon* assemblages, followed by abiotic substratum types (sand/rubble and pavement, 20.4%), soft coral and macroalgae (<0.1% each, Table 1A). *Porites*, *Acropora* (arborescent), *Pocillopora*, and/or *Cyphastrea* appear to provide the best explanatory habitat variables in distance-based linear models of spatial variation in the Lord Howe Island *Chaetodon* assemblage (Table 1B). While proportional abundances for each of the above five *Chaetodon* species tended to be highest in areas characterized by some of these types of coral (Fig. 4), considerable variation remains unexplained in these habitat-based DISTLMs (i.e., all  $r^2 < 0.28$ , Table 1B).

The three stripe butterflyfish, *C. tricinctus* was by far the most abundant *Chaetodon* species at all locations, accounting for 67.7% of all individuals (Fig. 2). The mean abundance of *C. tricinctus* was  $14.08 \pm 2.05$  (SE) fish per  $200 \text{ m}^2$ , compared to  $3.72 \pm 0.78$  SE fish per  $200 \text{ m}^2$  for the next most abundant species, *C. melannotus* (Fig. 2). Most *C. tricinctus* (374 of 640 individuals) occurred in schools of up to 42 individuals, with only 16% of individuals ( $n = 142$ ) recorded in pairs, and 14% of individuals ( $n = 124$ ) observed on their own. Larger aggregations of *C. tricinctus* tended to be found in inter-reefal habitats (over sand), but in close proximity to colonies of arborescent *Acropora* (Fig. 4B). Abundance of *C. trifascialis* was also highest where there was high arborescent *Acropora* (Fig. 4F), while abundance of *C. lunulatus* was highest where there was high cover of *Pocillopora* (Fig. 4E).

## Feeding behavior

Feeding rates (number of bites taken per 3-min) varied greatly within and among the butterflyfishes considered during this study (*C. lunulatus*, *C. plebeius*, *C. tricinctus*, and *C. trifascialis*). Notably, a large proportion of *C. melannotus* (35 of 67) and *C. tricinctus* (43 of 141) were not seen to take any bites throughout an entire 6-min observation period (i.e., when including the 3-min acclimation period), in contrast to very few (0–2) instances of non-feeding in the other species. Accordingly, mean bites rates of *C. melannotus* and *C. tricinctus* were markedly lower than *C. lunulatus*, *C. plebeius*, and *C. trifascialis*, with mean bite rates (averaged across all sites) varying by a factor of six among these species (Table 2). Bite rates varied significantly among species, but also varied among sites (Table 3), whereby the feeding rates for all but *C. trifascialis* were higher at Stephen's Hole than at North Reef or Potholes. For *C. lunulatus*, bite rates recorded at Stephen's Hole ( $27.00$  bites per 3-min  $\pm 7.22$  SE) were twice those recorded at Potholes ( $12.56$  bites per 3-min  $\pm 1.66$  SE). For *C. melannotus*, bite rates recorded at Stephen's Hole ( $5.36$  bites per 3-min  $\pm 2.51$  SE) were three times higher than recorded at Potholes ( $1.72$  bites per 3-min  $\pm 0.71$  SE) or North Bay ( $1.78$  bites per 3-min  $\pm 0.67$  SE). For *C. trifascialis*, bite rates were consistently high across all sites, but were highest at North Bay ( $19.9$  bites per 3-min  $\pm 1.02$  SE). Even after accounting for those individuals that did not feed at all, the mean number of bites taken by *C. melannotus* ( $5.78$  bites per 3-min  $\pm 1.69$  SE) and *C. tricinctus* ( $11.98$  bites per





**Figure 3.** Principal coordinates analysis (PCO) of spatial variation in the abundance and composition of the (A) *Chaetodon* assemblage, and (B) coral reef habitat for 36 transects spread across three sites (North Bay = filled triangles, Potholes = open circles, Stephen's Hole = open squares) at Lord Howe Island. Vectors are variables (*Chaetodon* species and substratum categories, respectively) most correlated (Pearson's correlation coefficient,  $r > 0.4$ ) with the PCO axes.

3-min  $\pm 0.84$  SE) were much lower than for the other three species (Table 2).

For *C. tricolor*, feeding rates differed significantly among fishes in different size classes (ANOVA,  $F_{2,138} = 1434.25$ ,  $P < 0.001$ ), being highest for the smallest fishes (14.60 bites per 3-min  $\pm 2.67$  SE) and declining with increasing TL (Fig. 5). All individuals  $< 5$  cm TL remained in close proximity to the benthos feeding

almost continually on scleractinian corals throughout feeding observations. Among *C. tricolor* of 6–10 cm TL, 20 individuals (of 86 in total) did not feed; larger individuals that did feed under observation exhibited sustained feeding on scleractinian corals, taking a mean of 11.12 bites per 3-min ( $\pm 1.05$  SE). For individuals  $> 10$  cm, only 2 (of 25) individuals were seen to feed on benthic substrata and these fishes took only 1 and 2 bites, respectively, throughout a 3-min observation. For the most part, all individuals  $> 10$  cm TL remained in schools in mid-water and rarely approached or searched the substratum during our diurnal observations. While it is possible that they were opportunistically feeding on passing plankton, as they did occasionally open and close their mouths, they tended to move very slowly rather than making any darting movements to actively seek out planktonic prey.

Four (of five) dominant *Chaetodon* butterflyfishes (*C. lunulatus*, *C. plebeius*, *C. tricolor*, and *C. trifascialis*) at Lord Howe Island were classified as obligate corallivores (following Cole et al. 2008) due to them taking almost 100% of recorded bites from the surface of live corals (Table 2). The exception was *C. melannottus*, which took only 6.28% of bites from the surface of scleractinian corals, with most of their bites taken on soft corals. All of the obligate corallivore species fed predominantly on *Acropora* (Table 2), which was prevalent across all sites. However, all four species of butterflyfishes clearly distinguished between different types of *Acropora*, consuming tabular *Acropora* disproportionately more than expected based on availability across the three sites, while they consumed arborescent *Acropora* in lower or equal proportions to availability (Table 2).

All butterflyfishes exhibited significant levels of dietary selectivity (Table 2), consuming some corals disproportionately to their availability. *Chaetodon melannottus* avoided all scleractinian corals in preference for soft corals (Table 2), but still consumed an average of 1.90 different coral types per 3-min observation (Fig. 6B). *Chaetodon lunulatus* was the least selective of the four obligate corallivores, consuming an average of 2.25 different coral genera per 3-min observation (Fig. 6B). While most bites were taken from arborescent *Acropora*, *C. lunulatus* preferentially consumed tabular *Acropora*, *Porites*, and *Pocillopora* (Table 2). *Chaetodon plebeius* exhibited intermediate levels of dietary selectivity, consuming an average of 2.19 different coral genera per 3-min observation (Fig. 6B) and preferentially consumed preferentially consumed tabular *Acropora*, *Isopora*, and *Porites*. *Chaetodon tricolor* and *C. trifascialis* were the most specialized coral feeders (Table 2), generally consuming only 1–2 different coral genera during feeding observations. Both species took most bites from arborescent *Acropora*, but preferred tabular *Acropora* to the exclusion of most other

**Table 1.** Summary of (A) marginal tests and (B) distance-based linear model (DISTLM) selection, based upon Akaike Information Criteria for finite samples (AICc) to select “best” model combinations of habitat variables (i.e., best solutions for models with 1, 2, 3, or 4 variables) to explain spatial variation in *Chaetodon* assemblages at Lord Howe Island. Marginal tests are for higher groupings of substratum variables to explore overall trends in multivariate variation (following Anderson et al. 2008). Abiotic categories (sand/rubble and pavement) were excluded from DISTLM selection due to strong (negative) correlations with biotic categories (following Anderson et al. 2008).

(A) Marginal tests						
Group	SS	Residual df	Regression df	% variation	Pseudo-F	P
Scleractiniancoral	7.035	23	13	46.0	1.64	0.008
Sand/rubble/pavement	3.123	33	3	20.4	4.24	0.001
Soft coral	0.903	34	2	0.06	2.14	0.063
Macroalgae	0.649	34	2	0.04	1.51	0.145

(B) Best DISTLM solutions			
Habitat variables	AICc	SS (resid.)	r <sup>2</sup>
<i>Porites</i>	29.16	14.05	0.08
<i>Porites</i> + <i>Acropora</i> (arborescent)	31.05	12.60	0.18
<i>Porites</i> + <i>Acropora</i> (arborescent) + <i>Pocillopora</i>	30.93	11.78	0.23
<i>Porites</i> + <i>Acropora</i> (arborescent) + <i>Pocillopora</i> + <i>Cyphastrea</i>	30.24	11.14	0.27

coral prey (Table 2), while most strongly avoiding *Isopora* and *Porites*.

### Correlates of species abundance

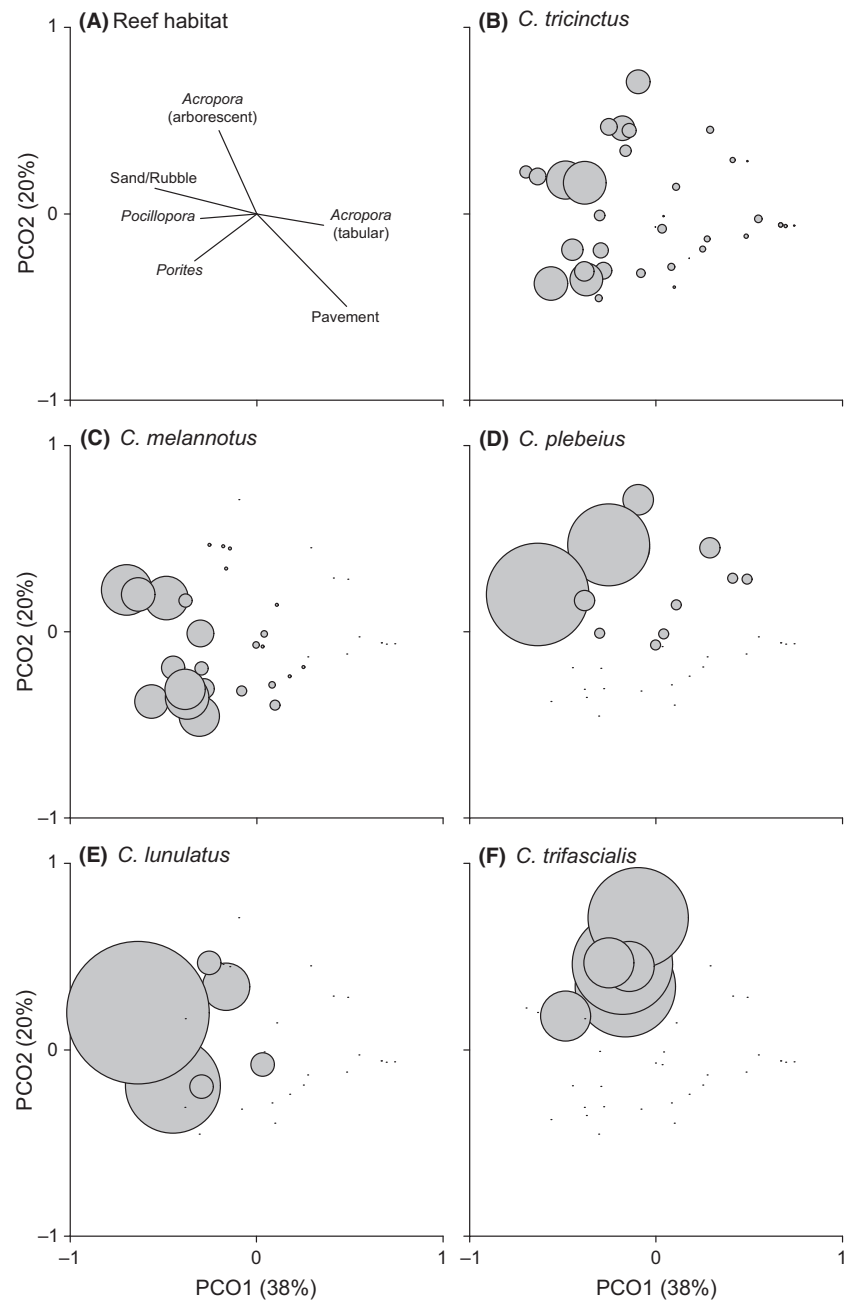
*Chaetodon tricinctus* was the dominant butterflyfish at all study sites, and while their abundance varied, it tended to be >3 times more abundant than any other butterflyfish species present. Found only at Lord Howe Island, nearby Elizabeth and Middleton Reefs and Norfolk Island, *C. tricinctus* geographic range is <5% of the next smallest range species, *C. uentheri*. The most widespread species recorded at Lord Howe Island, *Chaetodon auriga* and *C. trifascialis*, have widespread geographic ranges that extend across the entire Indo-Pacific and are >200 times larger than that of *C. tricinctus*, but both these widespread species are rare at Lord Howe Island (especially compared to *C. tricinctus*). Mean abundance of coral reef butterflyfishes at Lord Howe Island (averaged across the three sites) was weakly negatively correlated ( $r = -0.40$ ,  $n = 13$ ,  $P = 0.18$ ) with geographic range (Fig. 6A). This relationship appeared to be driven by the high abundance and limited geographic range of *C. tricinctus*. Indeed excluding *C. tricinctus* from the analysis resulted in no relationship between abundance and geographic range ( $r = 0.09$ ,  $n = 12$ ,  $P = 0.78$ ).

Regardless of the metric, *C. trifascialis* and *C. tricinctus* have the most specialized diets at Lord Howe Island. Notwithstanding the apparent lack of feeding among larger individuals, *C. tricinctus* used available coral prey in very similar proportions to *C. trifascialis*, feeding predominantly on arborescent *Acropora*, but selectively targeting tabular *Acropora* (Table 2). The main difference was that

*C. trifascialis* avoided eating *Pocillopora* corals, whereas *C. tricinctus* consumed *Pocillopora* in approximate accordance with its' availability. Despite similarities in their selectivity and dietary composition, *C. tricinctus* was >50 times more abundant than *C. trifascialis*, being the most and least abundant (respectively) of the five species for which dietary composition was analyzed. Other coral-feeding butterflyfishes (*C. melannotus*, *C. plebeius*, and *C. lunulatus*) were less selective and less abundant compared to *C. tricinctus*, suggesting that if there was any relationship between mean abundance and diet breadth it would be negative (Fig. 6B). However, the actual relationship based on these five species was nonsignificant ( $r = -0.34$ ,  $n = 5$ ,  $P = 0.58$ ).

### Discussion

The extent to which patterns of local abundance in coral reef fishes can be related to ecological specialization and/or geographical range size is uncertain, given the wide variety of relationships detected among taxonomic groups and locations (e.g., Hawkins et al. 2000; Bean et al. 2002; Hobbs et al. 2010; Berkström et al. 2012). Here, we reveal that marked interspecific variations in the local abundance of coral reef butterflyfishes at Lord Howe Island are weakly correlated to the geographic range size of species, but unrelated to levels of feeding specialization. Much of this range–abundance relationship hinges upon the most abundant species, *C. tricinctus*, which is a regional endemic with >3 times higher abundance than any other butterflyfish species at Lord Howe Island, and is the dominant species across all of our study sites. While high local abundances are often thought to be linked to high



**Figure 4.** Optimized principal coordinate analysis (PCO) of spatial variation in *Chaetodon* abundance and composition across 36 transects at Lord Howe Island. (A) Reef habitat variables most correlated (Pearson's correlation coefficient,  $r > 0.4$ ) with the PCO axes. Bubble sizes indicate proportional abundance of (B) *Chaetodon tricinctus*, (C) *Chaetodon melannotus*, (D) *Chaetodon plebeius*, (E) *Chaetodon lunulatus*, and (F) *Chaetodon trifascialis* in areas characterized by tabular *Acropora* and/or pavement (indicated on panel A as transects toward bottom right quadrants of each panel), arborescent *Acropora* (top left quadrants) or *Porites*, *Pocillopora* and sand/rubble (bottom left quadrants).

levels of preferred resource availability (Brown 1984; Brown et al. 1995; Gregory and Gaston 2000), in *C. tricinctus* we find unusual foraging behavior that is, unlike any other butterfly-fish classed as an obligate corallivore (Cole et al. 2008).

Despite their vulnerability to coral loss (e.g., Pratchett et al. 2006), butterflyfish assemblages are often dominated by obligate coral-feeding species (reviewed by Pratchett 2014). At Lord Howe Island, obligate coral-feeding species (including *C. tricinctus*) accounted for 77.43% of all but-

terflyfishes (580 of 749), and three of four of the most abundant species were all obligate coral-feeding species. Obligate corallivores also dominate butterflyfish assemblages at many other locations throughout the Indo-Pacific (Emslie et al. 2010; Pratchett et al. 2013a; Cole and Pratchett 2014), but it is less clear to what extent specialist versus generalist corallivores dominate butterflyfish assemblages.

Highly specialized species are expected to be much less abundant than generalist counterparts because they are



**Table 2.** Bite rates, coral use, and feeding selectivity of five *Chaetodon* butterflyfishes at Lord Howe Island, ordered according to increasing selectivity. Significance of prey selection was assessed using forage selection ratios and Bonferroni corrected 95% confidence intervals (“=” indicates prey that were used in proportion to availability, “+” indicates prey used significantly more than expected, “-” indicates prey used less than expected, and “0” indicates prey that were not used at all). Overall significance of feeding selectivity was tested using total forage ratios, comparing relative use of different prey categories to their availability across the three study sites (Manly et al. 2002).

Species	n	Bite rate	Hard corals (%)	Arborescent <i>Acropora</i>	Tabular <i>Acropora</i>	<i>Isopora</i>	<i>Pocillopora</i>	<i>Porites</i>	Soft corals	Total Forage Ratio	Sig.
<i>Chaetodon melannotus</i>	67	2.85	6.28	3.14% (-)	1.05% (-)	0.52% (-)	1.05% (-)	0.52% (-)	45.03% (+)	1797.53	<0.001
<i>Chaetodon lunulatus</i>	51	16.69	99.76	30.55% (-)	6.46% (+)	3.06% (=)	29.38% (+)	20.92% (+)	0.00% (0)	2849.47	<0.001
<i>Chaetodon plebeius</i>	65	15.85	99.90	23.20% (-)	16.21% (+)	26.70% (+)	19.42% (=)	9.81% (+)	0.00% (0)	3499.17	<0.001
<i>Chaetodon tricinctus</i>	141	8.33	100	51.57% (=)	22.38% (+)	1.96% (-)	19.23% (=)	0.68% (-)	0.00% (0)	3940.70	<0.001
<i>Chaetodon trifascialis</i>	73	17.63	100	55.71% (=)	38.54% (+)	0.39% (-)	3.89% (-)	0.23% (-)	0.00% (0)	4552.50	<0.001

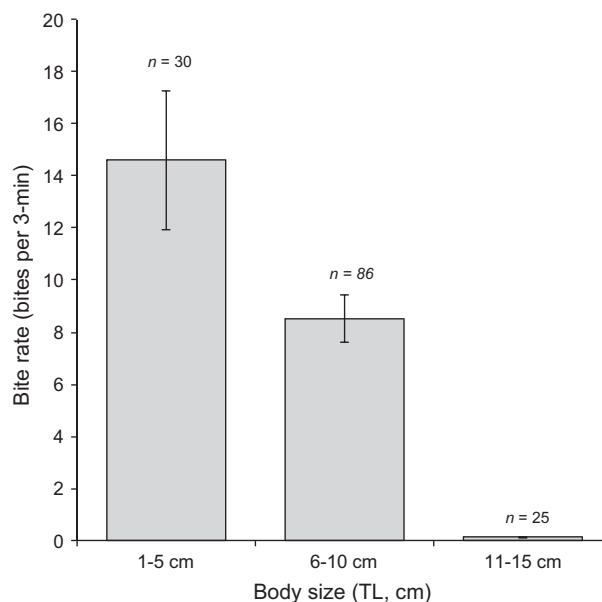
**Table 3.** Two-way factorial ANOVAs testing for differences in (A) bite rates and (B) the range of prey types consumed among species (see Table 2 for details) and among the three distinct study locations (North Bay, Stephen’s Hole and Potholes). Given that both the total number of bites and the number of distinct prey types consumed within a 3-min period is highly constrained, data were square-root transformed prior to analyses.

A) Bite rate					
Source	SS	df	MS	F	Sig.
Species	518.74	4	129.68	58.95	<0.001
Sites	35.37	2	17.68	8.04	<0.001
Species × sites	23.20	8	2.90	1.32	0.23
Error	840.40	382	2.20		
Total	4534.00	396			

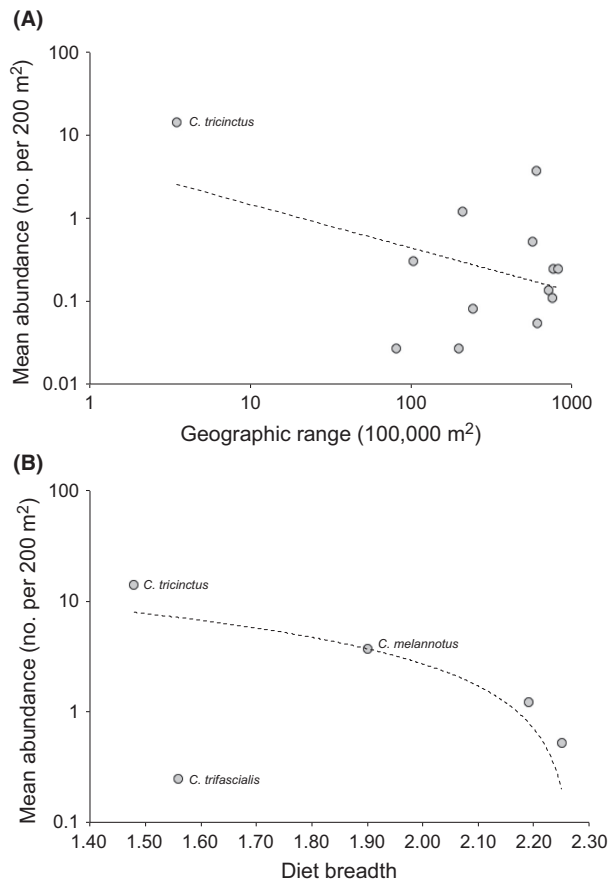
B) Range of prey types					
Source	SS	df	MS	F	Sig.
Species	6.58	4	1.64	16.56	<0.001
Sites	0.29	2	0.15	1.49	0.23
Species × sites	0.63	8	0.08	0.79	0.61
Error	30.20	382	0.10		
Total	545.00	396			

assumed to be more constrained by a narrower range of possible resources (Brown 1984; Gaston et al. 1997). While such trends have been recorded in some coral reef fishes (Hawkins et al. 2000; Bean et al. 2002), the relative abundance of generalist versus specialists species within a specific location will depend upon the availability of different resources (Munday 2004); consequently, specialist species may be more abundant where their preferred resources are also abundant (Brown 1984; Emslie et al. 2010; Pratchett et al. 2013a). At Lord Howe Island, four species of obligate coral-feeding butterflyfishes (*C. lunula-*



**Figure 5.** Size-based variation in mean (±SE) bites rates of *Chaetodon tricinctus*. Total length (TL) was visually estimated (to the nearest cm) for all fishes for which 3-min feeding observations were conducted. Data were pooled across site, and the number of fishes within each size class (n) is shown.

*tus*, *C. plebeius*, *C. tricinctus*, and *C. trifascialis*) all consumed tabular *Acropora* disproportionately to its availability, as shown elsewhere (Berumen and Pratchett 2006; Cole et al. 2012; Pratchett et al. 2013a). Given that proportional consumption of tabular *Acropora* was highest for the two most specialized species, *C. tricinctus* and *C. trifascialis* (Table 2), it may be that a predominance of *Acropora* corals at Lord Howe Island (which accounted



**Figure 6.** Correlations of (A) geographic range and (B) dietary breadth versus mean abundance (averaged across all sites) for *Chaetodon* butterflyfishes at Lord Howe Island. Abundance and geographic range are shown on a log-scale. Dietary breadth was estimated only for the five most abundant butterflyfishes at Lord Howe Island.

for up to 94% of coral recorded on individual transects) confounds the expected negative relationship between dietary specialization and abundance. While it is clear that specialist butterflyfishes are numerically dominant in some locations (e.g., Pratchett et al. 2013a), this is not necessarily the case at Lord Howe Island. The dominant species, *C. tricinctus*, does feed on a relatively restricted range of different corals, but it is not altogether clear how this species derives sufficient energy, especially as adults.

While it has long been assumed that *C. tricinctus* consumes mainly scleractinian corals (Kuitert 1996), which is consistent with its abundance in coral-rich habitats (Lieske and Myers 2001; Hobbs et al. 2009; Hoey et al. 2014), this is the first detailed study of their foraging behavior. Based on phylogenetically conserved patterns of feeding (e.g., Bellwood et al. 2010) one would assume *C. tricinctus* is an obligate corallivore. Bellwood et al.

(2010) showed that *C. tricinctus* is within a clade containing all obligate hard-coral-feeding butterflyfishes. Clearly, when *C. tricinctus* feeds on corals (e.g., as juveniles) it is very selective, and preferentially targets *Acropora* and *Pocillopora*. Bite rates of small (<5 cm TL) *C. tricinctus* (14.60 bites per 3-min  $\pm$  2.67 SE) are also consistent with bite rates recorded for other obligate coral-feeding butterflyfishes (Gregson et al. 2008). However, the adult foraging behavior is very different to other obligate coral-feeding butterflyfishes. Obligate coral-feeding butterflyfishes typically exhibit sustained high levels of diurnal feeding upon hard corals (Gregson et al. 2008), which is attributed to physical constraints on the amount of coral tissue that can be effectively removed with each bite (Tricas 1989). It is possible that cooler water temperatures at this high-latitude coral reef may be reducing metabolic rates and altering the energetic budgets of these tropical fishes (Beamish 1981; Harmelin-Vivien 2002; Pörtner 2002), which may manifest as different types of foraging behaviors among these butterflyfish species (Clarke 2003). Size-based declines in feeding rates have been recorded among other functional groups of fishes (e.g., van Rooij et al. 1996; Bonaldo et al. 2006), and may reflect declines in energetic requirements among large and mature individuals, whereas juveniles invest substantially into growth and development (Harmelin-Vivien 2002). It is also possible that adult *C. tricinctus* feed mainly at night, as has been suggested for some other coral-feeding butterflyfishes (Zekeria et al. 2002). Alternatively, *C. tricinctus* may fundamentally alter its foraging behavior with ontogeny, as shown for some coral-feeding wrasses (Cole 2010).

The schooling behavior of *C. tricinctus* is also very unique, especially among corallivorous butterflyfishes. Aside from Lord Howe Island, we know that *C. tricinctus* is also very abundant and often forms large schools at Elizabeth and Middleton Reefs (Hobbs et al. 2009; Hoey et al. 2014), but is generally rare and occurs singly or in pairs at Norfolk Island (van der Meer et al. 2013). In reviewing the social organization of butterflyfishes, Hourigan (1989) reported that schooling is restricted to planktivorous butterflyfishes, whereas obligate corallivores tend to form pairs that aggressively maintain distinct feeding territories (Hourigan 1989; Roberts and Ormond 1992). Schooling behavior among coral reef fishes is generally considered to be a strategy to decrease search times for patchily distributed resources, provide increased protection from predators, and/or save on the energetic costs of locomotion (Ward et al. 2002; Liao 2007; Pereira and Ferreira 2013). Without further evidence (e.g., observations of nocturnal behavior) it is difficult to conclude whether this behavior plays a role in driving the extreme abundance of *C. tricinctus* at Lord Howe Island (especially, compared to other butterflyfishes).

Aside from resource use and availability, interspecific differences in abundance of coral reef fishes may be explained by contrasting population dynamics and key demographic rates. In particular, the relative abundance of different fishes is fundamentally dependent upon species-specific rates of recruitment (e.g., Schroeder 1987; Doherty and Williams 1988; Doherty 1991; Caselle and Warner 1996) and this is likely to be even more important at relatively isolated locations, such as Lord Howe Island. Small and isolated coral reefs, like islands, often contain a high proportion of endemic species (Jones *et al.* 2002; Allen 2008). Moreover, endemic marine fishes are often more (not less) abundant than their widespread counterparts (e.g., Hourigan and Reese 1987; Randall 1998; Jones *et al.* 2002; DeMartini 2004; DeMartini and Friedlander 2004; Hobbs *et al.* 2010, 2011). One obvious explanation for this pattern is that restricted range species have reproductive strategies that minimize dispersal and advection of larvae away from their natal reefs, thereby limiting the capacity for range expansion, but also ensuring effective self-recruitment (e.g., DeMartini 2004; DeMartini and Friedlander 2004; Eble *et al.* 2009; Hobbs *et al.* 2011). Consistent with this hypothesis, we recorded few (if any) very small (<5 cm TL) individuals, assumed to represent new recruits, for any species, except *C. tricinctus*. Moreover, van der Meer *et al.* (2013) showed that there are very high rates of self-recruitment at each of the reefs (Lord Howe Island, Elizabeth and Middleton Reefs) where *C. tricinctus* is the predominant butterflyfish species. However, interspecific comparisons of recruitment rates will require systematic surveys over multiple recruitment seasons, as well as detailed demographic studies to account for possible interspecific differences in growth rates.

There is increasing evidence that terrestrial macroecological relationships between abundance and range size do not necessarily apply to coral reef fishes (e.g., Hobbs *et al.* 2010, 2011, 2012; Berkström *et al.* 2012). Contrary to expectations, the most abundant species of butterflyfish at Lord Howe Island, *C. tricinctus*, is a restricted range endemic and also appears to be among the most specialized of butterflyfishes recorded at this location. Endemic species may predominate at isolated locations because they are uniquely adapted to the local conditions (Blackburn *et al.* 1997; Thiollay 1997; Reif *et al.* 2006). Similarly, highly specialized species may be particularly abundant at locations with very high availability of their preferred habitat and/or food resources. *Chaetodon tricinctus*, however, remains an enigmatic species that contradicts much of the established understanding of coral-feeding butterflyfishes. Future research needs to consider whether the energetic demands (metabolic rates) of *C. tricinctus* are fundamentally different from that of other coral-feeding

butterflyfishes, or how adult fishes derive necessary energy despite infrequent bouts of benthic feeding. This research is necessary to clearly establish the vulnerability of *C. tricinctus* to increasing degradation of coral reef environments. Specialist coral-feeding butterflyfishes are extremely vulnerable to sustained and ongoing coral loss (Pratchett *et al.* 2008) that is, occurring on reefs throughout the world (Hughes *et al.* 2003), but flexible foraging (Noble *et al.* 2014) and highly resilient population dynamics may help to buffer against species extinctions (Lawton *et al.* 2011).

## Acknowledgments

We thank I. Kerr and S.-A. Gudge from the New South Wales Marine Park Authority, and B. Busted and the crew from Howea Divers, for advice and logistical support during this project. Funding was provided by ARC Centre of Excellence for Coral Reef Studies.

## Conflict of Interest

None declared.

## References

- Allen, G. R. 2008. Conservation hotspots of biodiversity and endemism for Indo-Pacific coral reef fishes. *Aquat. Conserv.* 18:541–556.
- Anderson, M. J., K. E. Ellingsen, and B. H. McCordle. 2006. Multivariate dispersion as a measure of beta diversity. *Ecol. Lett.* 9:683–693.
- Anderson, M. J., R. N. Gorley, and K. R. Clarke. 2008. PERMANOVA+ for PRIMER: guide to software and statistical methods. PRIMER-E, Plymouth.
- Beamish, F. W. H. 1981. Swimming performance and metabolic rate of three tropical fishes in relation to temperature. *Hydrobiologia* 83:245–254.
- Bean, K., G. Jones, and M. Caley. 2002. Relationships among distribution, abundance and microhabitat specialization in a guild of coral reef triggerfish (family Balistidae). *Mar. Ecol. Prog. Ser.* 233:263–272.
- Bellwood, D. R., S. Klanten, P. F. Cowman, M. S. Pratchett, N. Konow, and L. van Herwerden. 2010. Evolutionary history of the butterflyfishes (f: Chaetodontidae) and the rise of coral feeding fishes. *J. Evol. Biol.* 23:335–349.
- Berkström, C., G. Jones, M. McCormick, and M. Srinivasan. 2012. Ecological versatility and its importance for the distribution and abundance of coral reef wrasses. *Mar. Ecol. Prog. Ser.* 461:151–163. doi: 10.3354/meps09788.
- Berumen, M. L., and M. S. Pratchett. 2006. Effects of resource availability on the competitive behaviour of butterflyfishes (Chaetodontidae). *Proc 10th Int Coral Reef Symp 2004*, Okinawa, Japan.

- Berumen, M. L., M. S. Pratchett, and M. I. McCormick. 2005. Within-reef differences in diet and body condition of coral-feeding butterflyfishes (Chaetodontidae). *Mar. Ecol. Prog. Ser.* 287:217–227.
- Blackburn, T. M., K. J. Gaston, R. M. Quinn, H. Arnold, and R. D. Gregory. 1997. Of mice and wrens: the relation between abundance and geographic range size in British mammals and birds. *Philos. Trans. R. Soc. Lond., B, Biol. Sci.* 352:419–427.
- Blowes, S. A., M. S. Pratchett, and S. R. Connolly. 2013. Heterospecific aggression and dominance in a guild of coral-feeding fishes: the roles of dietary ecology and phylogeny. *Am. Nat.* 182:157–168.
- Bonaldo, R. M., J. P. Krajewski, C. Sazima, and I. Sazima. 2006. Foraging activity and resource use by three parrotfish species at Fernando de Noronha Archipelago, tropical West Atlantic. *Mar. Biol.* 149:423–433.
- Brook, B. W., N. S. Sodhi, and C. J. Bradshaw. 2008. Synergies among extinction drivers under global change. *Trends Ecol. Evol.* 23:453–460.
- Brown, J. H. 1984. On the relationship between abundance and distribution of species. *Am. Nat.* 124:255–279.
- Brown, J. H., D. W. Mehlman, and G. C. Stevens. 1995. Spatial variation in abundance. *Ecology* 76:2028–2043.
- Caselle, J. E., and R. R. Warner. 1996. Variability in recruitment of coral reef fishes: the importance of habitat at two spatial scales. *Ecology* 77:2488–2504.
- Clarke, A. 2003. Costs and consequences of evolutionary temperature adaptation. *Trends Ecol. Evol.* 18:573–581.
- Cole, A. J. 2010. Cleaning to corallivory: ontogenetic shifts in feeding ecology of tubelip wrasse. *Coral Reefs* 29:125–129.
- Cole, A. J., and M. S. Pratchett. 2014. Diversity in diet and feeding behaviour of butterflyfishes; reliance on reef corals versus reef habitats. Pp. 107–139 *in* M. S. Pratchett, M. L. Berumen and B. Kapoor, eds. *Biology of butterflyfishes*. CRC Press, Boca Raton, FL.
- Cole, A. J., M. S. Pratchett, and G. P. Jones. 2008. Diversity and functional importance of coral-feeding fishes on tropical coral reefs. *Fish Fish.* 9:286–307. doi: 10.1111/j.1467-2979.2008.00290.x.
- Cole, A. J., R. J. Lawton, S. K. Wilson, and M. S. Pratchett. 2012. Consumption of tabular acroporid corals by reef fishes: a comparison with plant–herbivore interactions. *Funct. Ecol.* 26:307–316.
- Davies, K. F., C. R. Margules, and J. F. Lawrence. 2004. Asynergetic effect puts rare, specialized species at greater risk of extinction. *Ecology* 85:265–271.
- DeMartini, E. E. 2004. Habitat and endemism of recruits to shallow reef fish populations: selection criteria for no-take MPAs in the NWHI Coral Reef Ecosystem Reserve. *Bull. Mar. Sci.* 14:185–205.
- DeMartini, E. E., and A. M. Friedlander. 2004. Spatial patterns of endemism in shallow-water reef fish populations of the Northwestern Hawaiian Islands. *Mar. Ecol. Prog. Ser.* 271:281–296.
- Devictor, V., J. Clavel, R. Julliard, S. Lavergne, D. Mouillot, W. Thuiller, et al. 2010. Defining and measuring ecological specialization. *J. Appl. Ecol.* 47:15–25.
- Doherty, P. J. 1991. Spatial and temporal patterns in recruitment. Pp. 261–293 *in* P. F. Sale, ed. *The ecology of fishes in coral reefs*. Academic Press, San Diego, CA.
- Doherty, P. J., and D. M. Williams. 1988. The replenishment of coral reef fish populations. *Oceanogr. Mar. Biol. Ann. Rev.* 26:447–551.
- Dulvy, N. K., J. R. Ellis, N. B. Goodwin, A. Grant, J. D. Reynolds, and S. Jennings. 2004. Methods of assessing extinction risk in marine fishes. *Fish Fish.* 5:255–276.
- Eble, J. A., R. J. Toonen, and B. W. Bowen. 2009. Endemism and dispersal: comparative phylogeography of three surgeonfishes across the Hawaiian Archipelago. *Mar. Biol.* 156:689–698.
- Emslie, M. J., M. S. Pratchett, A. J. Cheal, and K. Osborne. 2010. Great Barrier Reef butterflyfish community structure: the role of shelf position and benthic community type. *Coral Reefs* 29:705–715.
- Futuyma, D. J., and G. Moreno. 1988. The evolution of ecological specialization. *Annu. Rev. Ecol. Syst.* 19:207–233.
- Gaston, K. J. 1994. *Rarity*. Chapman and Hall, London, UK.
- Gaston, K. J. 1996. Species-range-size distributions: patterns, mechanisms and implications. *Trends Ecol. Evol.* 11:197–201.
- Gaston, K. J. 1998. Species-range size distributions: products of speciation, extinction and transformation. *Philos. Trans. R. Soc. Lond., B, Biol. Sci.* 353:219–230.
- Gaston, K. J., T. M. Blackburn, and J. H. Lawton. 1997. Interspecific abundance-range size relationships: an appraisal of mechanisms. *J. Anim. Ecol.* 66:579–601.
- Gregory, R. D., and K. J. Gaston. 2000. Explanations of commonness and rarity in British breeding birds: separating resource use and resource availability. *Oikos* 88:515–526.
- Gregson, M., M. Pratchett, M. Berumen, and B. Goodman. 2008. Relationships between butterflyfish (Chaetodontidae) feeding rates and coral consumption on the Great Barrier Reef. *Coral Reefs* 27:583–591.
- Harmelin-Vivien, M. L. 2002. Energetics and fish diversity on coral reefs. Pp. 265–274 *in* P. F. Sale, ed. *Coral reef fishes: dynamics and diversity in a complex ecosystem*. Academic Press, San Diego, CA.
- Hobbs, J. P., J. Neilson, and J. J. Gilligan. 2009. Distribution, abundance, habitat association and extinction risk of marine fishes endemic to the Lord Howe Island region. Report to Lord Howe Island Marine Park.
- Hobbs, J. P.-A., G. P. Jones, and P. L. Munday. 2010. Rarity and extinction risk in coral reef angelfishes on isolated islands: interrelationships among abundance, geographic range size and specialisation. *Coral Reefs* 29:1–11.

- Hobbs, J.-P. A., G. P. Jones, and P. L. Munday. 2011. Extinction risk in endemic marine fishes. *Conserv. Biol.* 25:1053–1055.
- Hobbs, J.-P. A., G. P. Jones, P. L. Munday, S. R. Connolly, and M. Srinivasan. 2012. Biogeography and the structure of coral reef fish communities on isolated islands. *J. Biogeogr.* 39:130–139.
- Hoey, A. S., M. S. Pratchett, and C. Cvitanovic. 2011. High macroalgal cover and low coral recruitment undermines the potential resilience of the world's southernmost coral reef assemblages. *PLoS ONE* 6:e25824.
- Hoey, A. S., M. S. Pratchett, J. Johansen, and J. Hoey. 2014. 2014 marine ecological survey of Elizabeth and Middleton reefs, Lord Howe Commonwealth marine reserve. Report to the Department of the Environment, Canberra, ACT, Australia.
- Hourigan, T. F. 1989. Environmental determinants of butterflyfish social systems. *Environ. Biol. Fishes* 25:61–78.
- Hourigan, T. F., and E. S. Reese. 1987. Mid-ocean isolation and the evolution of Hawaiian reef fishes. *Trends Ecol. Evol.* 2:187–191.
- Hughes, T. P., D. R. Bellwood, and S. R. Connolly. 2002. Biodiversity hotspots, centres of endemism, and the conservation of coral reefs. *Ecol. Lett.* 5:775–784.
- Hughes, T. P., A. H. Baird, D. R. Bellwood, M. Card, S. R. Connolly, C. Folke, et al. 2003. Climate change, human impacts and the resilience of coral reefs. *Science* 301:929–933.
- Johnson, C. N. 1998. Species extinction and the relationship between density and distribution. *Nature* 394:272–274.
- Jones, G. P., M. J. Caley, and P. L. Munday. 2002. Rarity in coral reef fish communities. Pp. 81–101 in P. F. Sale, ed. *Coral reef fishes dynamics and diversity in a complex ecosystem*. Academic Press, San Diego, CA.
- Koh, L. P., N. S. Sodhi, and B. W. Brook. 2004. Ecological correlates of extinction proneness in tropical butterflies. *Conserv. Biol.* 18:1571–1578.
- Kuiter, R. H. 1996. *Guide to sea fishes of Australia*. New Holland, Sydney, NSW.
- Lawton, J. H. 1993. Range, population abundance and conservation. *Trends Ecol. Evol.* 8:409–413.
- Lawton, R. J., V. Messmer, L. K. Bay, and M. S. Pratchett. 2011. High gene flow across large geographic scales reduces extinction risk for highly specialised coral feeding butterflyfishes. *Mol. Ecol.* 17:3584–3598.
- Lawton, R. J., A. J. Cole, M. L. Berumen, and M. S. Pratchett. 2012. Geographic variation in resource use by specialist versus generalist butterflyfishes. *Ecography* 35:566–576.
- Liao, J. C. 2007. A review of fish swimming mechanics and behaviour in altered flows. *Philos. Trans. R. Soc. Lond., B, Biol. Sci.* 362:1973–1993. doi: 10.1098/rstb.2007.2082.
- Lieske, E., and R. Myers. 2001. *Collins pocket guide: coral reef fishes*. Collins, London.
- Malcolm, J. R., C. Liu, R. P. Neilson, L. Hansen, and L. Hannah. 2006. Global warming and extinctions of endemic species from biodiversity hotspots. *Conserv. Biol.* 20:538–548. doi: 10.1111/j.1523-1739.2006.00364.x.
- Manly, B. F., L. McDonald, D. L. Thomas, T. L. McDonald, and W. P. Erickson. 2002. *Resource selection by animals: statistical design and analysis for field studies*. Springer, Dordrecht, the Netherlands.
- Manne, L. L., and S. L. Pimm. 2001. Beyond eight forms of rarity: which species are threatened and which will be next? *Anim. Conserv.* 4:221–229. doi: 10.1017/S1367943001001263.
- McKinney, M. L. 1997. How do rare species avoid extinction? A paleontological view. Pp. 110–129 in W. E. Kunin and K. J. Gaston, eds. *The biology of rarity*. Chapman and Hall, London.
- van der Meer, M. H., J. B. Horne, M. G. Gardner, J.-P. A. Hobbs, M. S. Pratchett, and L. van Herwerden. 2013. Limited contemporary gene flow and high self-replenishment drives peripheral isolation in an endemic coral reef fish. *Ecol. Evol.* 3:1653–1666.
- Munday, P. L. 2004. Habitat loss, resource specialization, and extinction on coral reefs. *Glob. Chang. Biol.* 10:1642–1647.
- Noble, M. M., M. S. Pratchett, D. J. Coker, C. Cvitanovic, and C. J. Fulton. 2014. Foraging in corallivorous butterflyfish varies with wave exposure. *Coral Reefs* 33:351–361. doi: 10.1007/s00338-014-1140-7.
- Olden, J. D., N. LeRoyPoff, and K. R. Bestgen. 2008. Trait synergisms and the rarity, extirpation, and extinction risk of desert fishes. *Ecology* 89:847–856.
- Päivinen, J., A. Grapputo, V. Kaitala, A. Komonen, J. S. Kotiaho, K. Saarinen, et al. 2005. Negative density–distribution relationship in butterflies. *BMC Biol.* 3:1–13.
- Pereira, P. H. C., and B. P. Ferreira. 2013. Effects of life phase and schooling patterns on the foraging behaviour of coral-reef fishes from the genus *Haemulon*. *J. Fish Biol.* 82:1226–1238.
- Pimm, S. L., C. N. Jenkins, R. Abell, T. M. Brooks, J. L. Gittleman, L. N. Joppa, et al. 2014. The biodiversity of species and their rates of extinction, distribution, and protection. *Science* 344:1246752–1–1246752–10. doi: 10.1126/science.1246752.
- Pörtner, H. O. 2002. Climate variations and the physiological basis of temperature dependent biogeography: systemic to molecular hierarchy of thermal tolerance in animals. *Comp. Biochem. Physiol. A. Mol. Integr. Physiol.* 132:739–761.
- Pratchett, M. S. 2005. Dietary overlap among coral-feeding butterflyfishes (Chaetodontidae) at Lizard Island, northern Great Barrier Reef. *Mar. Biol.* 148:373–382.
- Pratchett, M. S. 2014. Feeding preferences and dietary specialization among obligate coral-feeding butterflyfishes. Pp. 140–179 in M. S. Pratchett, M. L. Berumen and B. Kapoor, eds. *Biology of butterflyfishes*. CRC Press, Boca Raton, FL.
- Pratchett, M. S., A. S. Hoey, S. K. Wilson, V. Messmer, and N. A. J. Graham. 2011. Changes in the biodiversity and



- functioning of reef fish assemblages following coral bleaching and coral loss. *Diversity* 3:424–452.
- Pratchett, M. S., S. K. Wilson, M. L. Berumen, and M. I. McCormick. 2004. Sublethal effects of coral bleaching on an obligate coral feeding butterflyfish. *Coral Reefs* 23:352–356.
- Pratchett, M. S., S. K. Wilson, and A. H. Baird. 2006. Declines in the abundance of *Chaetodon* butterflyfishes following extensive coral depletion. *J. Fish Biol.* 69:1269–1280.
- Pratchett, M. S., P. L. Munday, S. K. Wilson, N. A. J. Graham, J. E. Cinner, D. R. Bellwood, et al. 2008. Effects of climate-induced coral bleaching on coral-reef fishes: ecological and economic consequences. *Oceanogr. Mar. Biol. Ann. Rev.* 46:251–296.
- Pratchett, M. S., N. A. J. Graham, and A. J. Cole. 2013a. Specialist corallivores dominate butterflyfish assemblages in coral-dominated reef habitats. *J. Fish Biol.* 82:1177–1191.
- Pratchett, M. S., A. S. Hoey, D. A. Feary, A. G. Bauman, J. A. Burt, and B. Riegl. 2013b. Functional composition of *Chaetodon* butterflyfishes at a peripheral and extreme coral reef location, the Persian Gulf. *Mar. Pollut. Bull.* 72:333–341.
- Randall, J. E. 1976. The endemic shore fishes of the Hawaiian Islands, Lord Howe Island and Easter Island. *Colloque Commerson 1973 O.R.S.T.O.M. Travaux et Documents No.* 47:49–73.
- Randall, J. E. 1998. Zoogeography of shore fishes of the Indo-Pacific region. *Zool. Stud.* 37:227–268.
- Reif, J., D. Horak, O. Sedlacek, J. Riegert, M. Pesata, K. Hrazsky, et al. 2006. Unusual abundance-range size relationship in an Afrotropical bird community: the effect of geographical isolation? *J. Biogeogr.* 33:1959–1968.
- Roberts, C. M., and R. F. G. Ormond. 1992. Butterflyfish social behaviour, with special reference to the incidence of territoriality: a review. *Environ. Biol. Fishes* 34:79–93.
- Roberts, C. M., C. J. McClean, J. E. N. Veron, J. P. Hawkins, G. R. Allen, D. E. McAllister, et al. 2002. Marine biodiversity hotspots and conservation priorities for tropical reefs. *Science* 295:1280–1284.
- van Rooij, J. M., E. de Jong, F. Vaandrager, and J. J. Videler. 1996. Resource and habitat sharing by the stoplight parrotfish, *Sparisomaviride*, a Caribbean reef herbivore. *Environ. Biol. Fishes* 47:81–91.
- Schroeder, R. E. 1987. Effects of patch reef size and isolation on coral reef fish recruitment. *Bull. Mar. Sci.* 41:441–451.
- Thiollay, J. M. 1997. Distribution and abundance patterns of bird community and raptor populations in the Andaman archipelago. *Ecography* 20:67–82.
- Tricas, T. C. 1989. Prey selection by coral-feeding butterfly-fishes: strategies to maximize the profit. *Environ. Biol. Fishes* 25:171–185.
- Ward, J. F., R. M. Austin, and D. W. Macdonald. 2002. A simulation model of foraging behaviour and the effect of predation risk. *J. Anim. Ecol.* 69:16–30.
- Williams, Y. M., S. E. Williams, R. A. Alford, M. Waycott, and C. N. Johnson. 2006. Niche breadth and geographical range: ecological compensation for geographical rarity in rainforest frogs. *Biol. Lett.* 2:532–535.
- Wilson, S. K., N. A. J. Graham, M. S. Pratchett, G. P. Jones, and N. V. C. Polunin. 2006. Multiple disturbances and the global degradation of coral reefs: are reef fishes at risk or resilient? *Glob. Change Biol.* 12:2220–2234.
- Wilson, S. K., N. A. J. Graham, and M. S. Pratchett. 2014. Susceptibility of butterflyfish to habitat disturbance: do ‘chaets’ ever prosper? Pp. 226–245 in M. S. Pratchett, M. L. Berumen and B. Kapoor, eds. *Biology of butterflyfishes*. CRC Press, Boca Raton, FL
- Zann, L. P. 2000. The eastern Australian region: a dynamic tropical/temperate biotone. *Mar. Pollut. Bull.* 41:188–203.
- Zakeria, Z. A., Y. Dawit, S. Ghebremedhin, M. Naser, and J. J. Z. Videler. 2002. Resource partitioning among four butterflyfish species in the Red Sea. *Mar. Freshw. Res.* 53:63–168.