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Transposable elements in response to environmental stressors&

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Abstract

Transposable elements (TEs) comprise a group of repetitive sequences that bring positive, negative, as well as neutral effects to the host organism. Earlier considered as "junk DNA," TEs are now well-accepted driving forces of evolution and critical regulators the of expression of genetic information. Their activity is regulated by epigenetic mechanisms, including methylation of DNA and histone modifications. The loss of epigenetic control over TEs, exhibited as loss of DNA methylation and decondensation of the chromatin structure, may result in TEs reactivation, initiation of their insertional mutagenesis (retrotransposition) and has been reported in numerous human diseases, including cancer. Accumulating evidence suggests that these alterations are not the simple consequences of the disease, but often may drive the pathogenesis, as they can be detected early during disease development. Knowledge derived from the *in vitro, in vivo*, and epidemiological studies, clearly demonstrates that exposure to ubiquitous environmental stressors, many of which are carcinogens or suspected carcinogens, are capable of causing alterations in methylation and expression of TEs and initiate retrotransposition events. Evidence summarized in this review suggests that TEs are the sensitive endpoints for detection of effects caused by such environmental stressors, as ionizing radiation (terrestrial, space, and UV-radiation), air pollution (including particulate matter [PM]-derived and gaseous), persistent organic pollutants, and metals. Furthermore, the significance of these effects is characterized by their early appearance, persistence and presence in both, target organs and peripheral blood. Altogether, these findings suggest that TEs may potentially be introduced into safety and risk assessment and serve as biomarkers of exposure to environmental stressors. Furthermore, TEs also show significant potential to become invaluable surrogate biomarkers in clinic and possible targets for therapeutic modalities for disease treatment and prevention.

Conflict of interests

 $\&$ The views expressed in this manuscript do not necessarily represent those of the US Food and Drug Administration * correspondence to: Igor Koturbash, phone: 1-501-526-6638; ikoturbash@uams.edu.

The authors have no conflicts of interest to declare.

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Keywords

air pollution; DNA methylation; environmental stressors; ionizing radiation; metals; retrotransposition

1. Introduction

1.1. Classification

Repetitive sequences account for about 40-45% of mammalian genomes, with some estimates suggesting 66-69% of the human genomes to be repetitive or repetitive elementsderived [1]. The majority of these repetitive elements are transposable elements (TEs) – repeated and mobile DNA sequences, capable of moving and invading genomes [2]. TEs are represented as retrotransposons and transposons, while the rest of repetitive elements are represented with satellite DNA and tandem repeats, which are immobile (Table 1).

TEs comprise a group of repetitive sequences that bring positive, negative, as well as neutral effects to the host organism. Previously considered as "junk DNA," it has become increasingly evident that TEs carry a set of important gene regulatory functions, including serving as recombinogenic structures for rapid genome remodeling, maintaining centromere and telomere integrity, providing alternative promoters, silencing by transcriptional or RNA interference (RNAi), and creating cryptic splice sites and polyadenylation signals [3-7]. TEs are also considered evolutionary precursors of many genes in mammalian genomes [8]. Of particular concern are the deleterious effects of TEs exerted by their transposition that may result in potential insertions and deletions within the coding sequences that may disrupt gene expression, as well as damaging recombination events [9,10].

All transposable elements can be classified into retrotransposons (or those that relocate via an RNA intermediate in a "copy-and-paste" mechanism, Class I) and transposons (or those that use "cut-andpaste" mechanism, Class II) (Fig. 1). Retrotransposons are clustered into Long Terminal Repeat (LTR) and non-LTR elements, depending on their structure, and can be further subdivided into autonomous (capable of propagating themselves throughout the genome) and non-autonomous (those that use machinery of other TEs). In this review, we will concentrate solely on retrotransposons, the most abundant and only active in mammalian genomes group of repetitive elements.

1.1.1. Long Interspersed Nuclear Elements (LINE)—Autonomous Long Interspersed Nuclear Elements (LINE) represent the most abundant group, not only among non-LTR retrotransposons, but of all mammalian TEs, reaching 17 to 23% of their genomes [6,7]. They include low-copy archaic inactive elements such as LINE-2 and LINE-3 and currently active and abundant LINE-1 (L1) elements.

Around 516,000 copies of L1 elements are present in the human genome; however, only about 100 of them are functional full-length (6 Kb long) sequences. The majority of L1s are 5'-truncated (0.9 Kb in length on average), incapable of retrotransposition, elements. Full length L1 contains four functional units: a ~900 bp 5'-untranslated region (UTR), a bicistronic open reading frame that encodes two proteins - ORF1p (a 40 kDa trimeric protein

with RNA binding and nucleic acid chaperone activity) and ORF2p (a 150 kDa responsible for retrotransposition protein that encodes an endonuclease, reverse transcriptase, and zinc finger-like protein) and a $3'-UTR$ with a poly(A) tail $[11,12]$ (Fig. 1). The $5'-UTR$ of L1 is an important regulatory unit that contains sense and antisense promoters and transcription factors binding sites, including those for p53, YY1, RUNX, SRY, and SOCS1 [13-15]. While the sense promoter regulates the expression of L1, the function of the antisense promoter is still largely unknown. However, some studies report the role of L1 antisense promoter in gene transcription regulation [16,17]. Accumulating evidence also indicates that L1 antisense promoter is involved in suppression of transcription from its own sense promoter, as well as in control over L1 retrotransposition [18,19]. A 3'-UTR is represented as a conserved poly-G tract with unknown functions.

The important feature of mammalian retrotransposons and L1, particularly, is their functional retrotransposition activity by which they are initially transcribed from the genome and then are reverse transcribed into a new location using a transposon-encoded reverse transcriptase [20]. L1 retrotransposition is initiated by transcription of its full-length mRNA from the internal promoter located in the 5'UTR between nucleotide positions +390 and +526 [21], and this process is mediated by the RNA polymerase II (Fig. 2-1). The newly transcribed mRNA is transported to the cytoplasm, where it is translated to L1-encoded proteins ORF1p and ORF2p by the ribosomal machinery (Fig. 2-2). ORF1p and ORF2p then anneal with an mRNA molecule, creating a ribonucleoprotein particle (Fig. 2-3). The ribonucleoprotein particle is formed in a *cis* preference, meaning that although ORF1p and ORF2p have the ability to bind any cellular mRNA molecule, there is, in fact, a strong bias observed towards annealing with L1 mRNAs specifically [22]. The ORF1p/ORF2p/mRNA ribonucleoprotein particle then enters the nucleus, where reintegration into genomic DNA occurs [12]. At this stage, due to the ORF2p endonuclease activity, a single-stranded nick is produced in genomic DNA. The exposed free 3'-hydroxyl residue serves as a primer, and the associated L1 mRNA is reverse-transcribed into cDNA. This process is referred to as "target-primed reverse transcription" (Fig. 2-4). The end product is a new L1 insertion into genomic DNA (Fig. 2-5). The site of insertion is a function of the endonuclease moiety of ORF2p, with minor grove width and T_nA_n content of the genomic DNA sequence being major factors [12]. For more detailed mechanisms of retrotransposition, we refer the readers to excellent reviews [20,23,24].

L1 RNA and associated proteins have been detected primarily in germ cells and embryos, while their presence in differentiated cells under normal conditions is rare [25,26]. L1 retrotransposition occurs mainly in embryogenesis and has been associated with somatic mosaicism, but can be also detected in the germline [26]. The estimated rate of retrotransposition in humans is between 1 in 95 and 270 live births [26]. Importantly, L1 retrotransposition usually results in 5'-truncated elements incapable of future retrotransposition [3,27].

1.1.2. Long Terminal Repeats (LTR)—LTRs are named for their long terminal repeats flanking the internal proviral sequence on both sides of the element. They comprise 8% of the human genome and about 10% of the mouse genome [28]. Structurally, LTRs are related to exogenous retroviruses; however, they lack the ability to move between cells and,

therefore, are also called endogenous retroviruses (ERVs). LTRs encode *gag, pol, pro*, and *env* genes or may use the retroviral genes encoded by other ERVs (Fig. 1). While the activity of LTRs in humans remains controversial, at least two families are reported to be currently active in mice [29,30].

1.1.3. Short Interspersed Nucleotide Elements (SINE)—The cluster of nonautonomous retrotransposons includes Short Interspersed Nuclear Elements (SINE), which in humans are represented as *Alu* elements, the only active family of primate-specific SINE, and SINE-R, VNTR, and *Alu* (SVA) [31,32]. These elements lack their own retrotransposition machinery and, thus, utilize L1-encoded proteins for their own mobilization due to homologies between the 5' ends of the two elements (Fig. 1).

Alu elements comprise up to 13.7% of human genomes (~1.1 million copies) with their *de novo* insertion rates exceeding those of L1 elements, reaching 1 in 20 live births [33-35]. They are 300 bp in length and contain two almost identical monomer sequences separated by a centrally-located A-rich region (A_5TACA_6) . *Alu* elements derive from the 7SL RNA gene, are transcribed by RNA Polymerase III, and require L1 ORF1p for their mobilization [35,36]. In rodents they correspond to SINE elements – SINE B1 and SINE B2.

Accumulating evidence clearly demonstrates that SINE/*Alu* are also important regulators of genetic information, as they may possess an insulator/boundary activity, repress transcription by disrupting contacts between RNA polymerase II and promoter DNA, and cause epigenetic reprogramming of adjacent gene promoters [37-40].

1.2. Regulation of expression of transposable elements

The expression of TEs is regulated by both TE- and host-mediated mechanisms. TEmediated control is usually linked to the ability to produce truncated TE suppressor copies for transposase-mediated autoregulation or utilization of host regulatory factors for activation [5,41]. Host-mediated mechanisms of control over the TEs are primarily associated with epigenetic mechanisms and include methylation of DNA, histone modifications, and regulation by small RNAs.

DNA methylation is important for normal development and maintenance of cellular homeostasis. It is involved in the regulation of the proper expression of genetic information in a sex-, tissue-, and cell type-dependent manner and serves as a key mechanism in silencing of TEs. DNA methylation is primarily associated with cytosine residues of cytosine and guanine base sequences (so called CpG sites). Approximately 70–90% of CpG sites in the mammalian genome are methylated, where TEs represent the most highly methylated sequences [42,43]. In turn, L1s are among the most highly methylated TEs in mammalian genomes; however, the methylation status of L1 may differ between the families and evolutionary age of the elements [44,45]. On the other hand, tandem repeats are relatively CpG-poor elements, but the extent of their methylation is relatively high given their length.

Earlier studies considered hypermethylation of TEs as a primary mechanism for their silencing. For instance, methylation of the CpG island located within the 5'-UTR of

mammalian L1 has been shown to repress the expression of L1 [46,47]. Some studies, however, report reactivation of retrotransposons without alterations in DNA methylation, suggesting that other epigenetic mechanisms are also involved. These findings have led to the development of the hypothesis that methylation may serve rather as a locking mechanism over the other established epigenetic silencing marks [43]. Further studies have also indicated that while DNA methylation may be crucial for maintaining inactive transcription status of some TEs (i.e., LTR/ERVs), other TEs such as L1 may be less dependent on it [5,48]. It has also been demonstrated that different epigenetic controls are necessary for the regulation of TEs in somatic and embryonic stem cells [48,49].

Histone modifications are among the most important epigenetic mechanisms and have recently been recognized to play a key role in regulation of TEs. Covalent histone modifications shape the heterochromatin ubiquitously associated with TEs. Transcriptionally silent TEs are often associated with repressive histone lysine methylation marks (lysines 9 and 27 on histone 3 and lysine 20 on histone 4) and histone H2A.Z; however, different marks are specifically enriched in different TEs and cell types [50-52].

Non-coding RNAs provide another conserved mechanism of suppression of TEs and include control by PIWI-interacting RNA (piRNA), specific for germline and spermatogenesis as well as siRNA-, and miRNA-regulated silencing [42,53,54]. Additionally, some elements, such as L1, can use their antisense promoter as a mechanism of self-regulation [19].

1.3. Transposable elements in human disease

Failure of epigenetic control over TEs is associated with numerous pathological states. Loss of methylation of L1 and *Alu*/SINE has been reported in virtually all human and rodent experimental cancers (reviewed in [35,42,45]). Global genomic hypomethylation, to which TEs are the main contributors, is now a well-accepted hallmark of cancer. In turn, hypomethylation of 5'-UTR of L1 is frequently associated with its reactivation and increased expression of L1 has been detected in human cancer tissues [55].

Although only small subset of L1 are the full length and capable of retrotransposition elements, their insertions accumulated during the millions years of evolution may have dramatic effects on genomic loci in which retrotransposition occur. In general, L1 are associated with AT-rich and gene-poor regions of the genome [56]; however, many genes acquired L1 insertions either within them or in their proximal vicinity, and these genes are usually lowly expressed [57,58]. Alterations in the methylation status of such truncated and retrotransposition-inactive L1 elements may result in reactivation of these genes, if the insertion occured within the transcription start sites [59]. For instance, hypomethylation of the L1 element located within the *MET* oncogene results in activation of an alternate transcript of the latter in blade tumors and across the entire uroepithelium in tumor-bearing bladders [60]. Similar findings were reported in patients with colorectal cancer, where hypomethylation of the L1 sequence within the *MET* gene inversely correlated with induction of *MET* expression in metastatic tissue [61].

Retrotransposition of a given TE can also result in disruption of the ORF with subsequent inactivation of a functional gene, development of genomic instability and genome

amplification, and thus serve as a causative factor in life-threatening human disorders, including cancer. For instance, Miki and colleagues reported insertion of the 3' portion of L1 into the last exon of the APC gene, leading to the disruption of its function in colorectal cancer [62]. L1 retrotransposition has been reported in a number of other human cancers, including lung, prostate, and ovarian cancers [63,64]. However, it still remains unknown whether retrotransposition serves as a driving force of tumorigenesis or merely a consequence, occurring after tumor initiation. Taking into account the existing evidence, it is plausible to hypothesize that both scenarios may take place. For instance, a retrotransposition event that occurs within a critical gene with its further inactivation, like in the case of *APC* in colorectal cancer [62], can be considered as a driving mechanism. On the other hand, Solyom and colleagues reported that in about 60% of cases, L1 insertion identified in the first section of the colorectal tumor was not identified in the second section of the same tumor [65], suggesting that this particular case of retrotransposition may play role in tumor heterogeneity, but, most likely, was not a tumor-initiating mechanisms.

L1 retrotransposition is not associated with cancers only. Independent mutagenic L1 insertion into exon 14 of the *Factor VIII* gene disrupts synthesis of functional coagulation factor and results in hemophilia A [66]. In addition, human diseases, such as diabetes, chronic granulomatous disease, and β-thalassemia are associated with L1 retrotransposition [67,68].

1.4. Transposable elements in response to environmental stressors: detection methodologies

1.4.1. Analysis of TEs methylation—Because of their potential for transposition in human genomes, L1 and Alu/SINE received the most attention in the studies dedicated to the effects of environmental stressors on TEs. The majority of studies have concentrated on the studies of TEs' methylation, as this epigenetic mechanism is critical for transcriptional silencing of TEs, and because the loss of global DNA methylation is a generally-accepted hallmark of cancer, can be detected in carcinogenesis and even shortly after exposure to carcinogens.

Combined bisulfite restriction analysis (COBRA) is the methodology based on a standard bisulfite modification of genomic DNA, followed by polymerase chain reaction (PCR) amplification and digestion of PCR product with restriction endonculeases. This methodology was extensively utilized in the L1 methylation status analysis. Sodium bisulfite treatment with subsequent PCR amplification results in methylation-dependent creation of new restriction endonucleases sites (Rsa I) or methylation-dependent retention of preexisting sites (BstUI) within the rat L1 elements. The digested products are then separated on agarose gel and the band intensity serves as a qualitative measurement of L1 methylation. The limitations of COBRA include low sensitivity and limited number of CpG sites that can be evaluated. *Methylation-sensitive quantitative PCR (MS qPCR)* is also based on the sensitivity of restriction endonucleases to differentially methylated DNA. The digested DNA is amplified with a set of primers specific for the sequence of interest. Utilization of realtime PCR allows for more precise and quantitative assessment of DNA methylation compared to COBRA; however, evaluation will be still limited to CpG sites to which

restriction endonucleases are selected. *Pyrosequencing* is a recently developed sequencebased technology that allows for accurate quantification of locus-specific DNA methylation. It relies on the detection of pyrophosphate release on nucleotide incorporation. Advantages of pyrosequencing is its very high sensitivity and analysis of the methylation status of all CpG sites within the investigated locus, but the methodology is limited to assessment of only short DNA sequences and relatively high costs.

1.4.2. Analysis of retrotransposition—Increases in TEs expression that can be assessed by designing specific assays for the TE of interest and real-time PCR may lead to retrotransposition events. Investigating the mobilization of endogenous TEs is associated with certain difficulties; therefore, an exogenous cell culture retrotransposition assay was developed and adapted for studies on L1 mobilization, first - in cells lines and later – in animal models. The history of development and success of this assay has been described in detail in an outstanding review by Rangwala and Kazazian [69]. Recent technological advances, including next-generation sequencing, now allow for comprehensive analysis of somatic retrotranspositions [70,71]. A number of tools have been developed in the last several years to analyze L1 insertions in both normal tissues and tumors, including *ME-Scan* [72], *droplet digital PCR* [73], and *Transposeq* [74]. Similarly, tools such as *RepEnrich* have been developed to study genome-wide transcriptional regulation of TEs [75]. Finally, the current knowledge on L1 insertions in humans has been summarized in *euL1db* – the European database of L1 retrotransposon insertions that contains over 140,000 sample-wise insertions and almost 9,000 distinct merged insertions [76].

Research in the last twenty years has convincingly demonstrated that alterations in TEs may not be only the consequences of the disease, but are among those that are involved in the pathogenesis. Further knowledge that environmental stressors can affect TEs has allowed a connection to be drawn between the environmental exposures, TEs, and disease development. Below, we review the current knowledge on the effects of the most ubiquitous environmental stressors exerted on mammalian TEs. The summary of the reviewed data can be found in Table 2

2. Transposable elements and ionizing radiation

Exposure to ionizing radiation (IR), both man-made and natural, constitutes a serious hazard for human health. Of major concern is medical radiation, utilized as a diagnostic and treatment modality because of the ever-growing number of patients routinely being exposed. Recent reports suggest that a significant risk of radiation-induced health effects, including cancer, exists even after medical procedures associated with exposure to low doses of IR [77,78]. While it is generally accepted that the benefits of medical radiation outweigh the risks, there are, however, considerable concerns, since in the US alone tens of millions of unnecessary computed tomography scans are performed annually, and approximately half of cancer patients receive some form of radiotherapy [79-81]. Additionally, large cohorts of people can be exposed to radiation because of accidents, such as Chornobyl and Fukushima-Daiichi.

2.1. Effects of terrestrial radiation on methylation and expression of TEs

Terrestrial radiation includes primarily low- linear energy transfer (LET) photon radiation, such as X- and γ-rays and is characterized by relatively diffuse energy deposition patterns. It is a potent genotoxic agent capable of inducing DNA single- and double-strand breaks, oxidative damage, and mutation induction.

IR is also capable of targeting the cellular epigenome and DNA methylation, in particular. Earlier studies described hypermethylation of tumor-suppressor genes in lung cancers of occupationally exposed workers in Russia [82]. Aside from the gene-specific hypermethylation associated with exposure to IR, a wealth of data from *in vivo* studies also indicate loss of global DNA methylation [83-90]. Further investigations have revealed that alterations in global DNA methylation are primarily associated with the alterations in the methylation status of TEs in a tissue-, dose-, and radiation quality-dependent manner [91-94].

Less consistency has been observed in DNA methylation response to IR *in vitro* [95-97]. For instance, Goetz et *al* reported loss of L1 methylation in RKO human colorectal carcinoma cells after exposure to 1 Gy of X-rays (250 kV peak, 13 mA; half-value layer, 1.65 mm copper; 2 keV/µm; dose rate of 2.4 Gy min⁻¹); however, the same dose of radiation led to hypermethylation of L1 in AG01522D primary human diploid skin fibroblasts [96]. Importantly, these changes were detected at \sim 20 population doublings after irradiation, suggesting a persistent nature of cell-dependent radiation-induced alterations in L1 methylation. At the same time, no significant changes in methylation of *Alu* elements were detected in either cell line.

A lack of low dose-associated alterations in methylation of TEs was observed in *in vivo* studies. Exposure to 10 mGy of X-ray (using an attenuated 140 kVp X-ray beam (8 mm Al half value layer) from a Gulmay D3150 superficial X-ray unit; dose-rate of 13.9 mGy/min) did not result in any persistent alterations in the methylation of L1, SINE B1, and IAP TEs in peripheral blood, spleen, or liver of C57BL/6 mice [98]. Exposure to a higher dose of 1 Gy of X-irradiation led to strain-, tissue-, sex-, and time-dependent alterations in methylation of the TEs [99]. Interestingly, hypermethylation of L1 was identified in relatively radio-resistant C57BL/6 mice at day 1 after irradiation, followed by hypomethylation at day 14. At the same time, radiosensitive BALB/c mice were characterized by hypermethylation of L1 in spleen tissue at 1 and 14 days after exposure. Methylation of SINE B1 and IAP was affected to a lesser extent [99]. A limitation of this study was that the mice of different strains analyzed at different time-points were irradiated at somewhat different conditions. Specifically, C57BL/6 mice, analyzed at day 1 after irradiation, were exposed with 100 kVp Philips RT100 SXR X-ray unit, while the mice analyzed at day 14 after irradiation were irradiated with a 140 kVp Gulmay D3150 X-ray unit. BALB/c mice were exposed using a 6 MeV X-ray beam from a Varian 600 CD Linear Accelerator. Exposures to doses higher than 1 Gy are usually associated with the loss of global and repetitive elements-associated DNA methylation [84,85,87-89]. Persistent global and TEs-associated hypomethylation after exposure to high doses of IR has been observed even in bystander tissues, as evident in a study reporting a 5 to 10% decrease in methylation

and 60% increase in expression of L1 element in the spleen tissue 7 months after localized hippocampal irradiation (20Gy (5cGy/s) of X-rays, 90 kV, 5 mA) [91].

2.2. Effects of space radiation on methylation and expression of TEs

The recent introduction of proton radiation into clinical practice and interest in space exploration has inspired the investigation of biological effects and molecular mechanisms of response to space radiation exposure. It is now clear that exposure to proton and heavy ion irradiation, the sources of space radiation, is associated with more complex clustered and often irreparable DNA damage, leading to greater relative biological effectiveness in cell death (reviewed in [100]).

Studies in cell culture have confirmed that space radiation epigenetic effects are also distinct from the effects elicited by terrestrial radiation. Exposure to low mean absorbed doses of protons (150 MeV/n, LET 0.55 keV/µm; dose rate 2 cGy min⁻¹) or ⁵⁶Fe ions (⁵⁶Fe⁺²⁶, 1) GeV/n, LET 150 keV/μm; 10 cGy min−1) lead to hypomethylation of L1 and *Alu* elements in AG01522D and RKO cells at 16-20 population doublings after irradiation [96]. Similar effects were observed in the human-hamster hybrid cell line GM10115, where exposure to 1 Gy of ${}^{56}Fe$ (${}^{56}Fe+{}^{26}$, 1GeV/amu, 150keV/µm, dose rate 2 cGy min⁻¹) led to hypomethylation in both L1 and *Alu* [95]. At the same time, exposure to 0.1 Gy of ⁵⁶Fe did not affect the methylation status of these retrotransposons.

Exposure to mean absorbed doses of ⁵⁶Fe ions within10-100 cGy; ⁵⁶Fe⁺²⁶, LET 175 keV/mm did not affect the methylation status of LINE-1 in the livers of C3H/HeN mice for up to 120 days after exposure [92]. This is a particularly interesting finding because exposure to 56Fe within this range of doses is associated with the development of liver tumors in this mouse strain [101,102], and the global and L1-associated DNA hypomethylation is a well-recognized hallmark of cancer and hepatocellular carcinoma in particular [103]. It is possible that L1 hypomethylation is a signature of later time-points, or high-LET hepatocarcinogenesis may not be necessarily associated with L1 hypomethylation at all. Another explanation is that assays specific for two L1 families – TF and A – were used in this study, while it has been suggested that L1s from different families may differ in their response to external stressors [44].

The same study has reported weak hypomethylation of both TF and A L1s shortly after exposure to $56Fe$ in the C3H/HeN mice lungs (day 1- day 7) and subtle hypermethylation of TF at day 120 [92]. L1 hypermethylation (26% increase) was also observed in the lungs of C57BL/6 mice 5 months after exposure to 0.4 Gy of ${}^{56}Fe$ (${}^{56}Fe+{}^{26}$, 600 MeV/nucleon), but not after lower doses [94]. Similar effects were observed in other TEs as well, including SINE B1, Charlie, and Mariner (15-25% increase), where the latter two present TEs that are currently transpositionally-inactive in mammals. This hypermethylation was clearly associated with further transcriptional silencing of these TEs. Interestingly, a recent study reported induction of lung tumors in mice exposed only to ${}^{56}Fe$ doses below 0.4 Gy [104], suggesting that hypermethylation and silencing of TEs may have a protective effect against heavy ion-induced carcinogenesis.

Of particular interest are the effects of high-LET radiation on TEs in the hematopoietic system because exposure to $56Fe$ and $28Si$ has been associated with increased rates of acute myeloid leukemia (AML) in mice [101,102], and studies indicate alterations in L1 and *Alu* methylation in patients with AML [105,106]. Exposure to 56 Fe ion radiation selectively targeted transposable elements in the less differentiated hematopoietic stem and progenitor cells, while the effects in mononuclear cells that are primarily comprised of terminally differentiated cells were miniscule [93]. Interestingly, while global, L1-, and SINE B1 associated dose-dependent hypermethylation was observed in hematopoietic stem and progenitor cells 30 days after exposure (similar to abovementioned findings in the lung by Lima et al. [92] at the same time-point), global hypomethylation (37% decrease) was observed 22 weeks after exposure to 0.4 Gy of ${}^{56}Fe$ (${}^{56}Fe$ + 26 , 600 MeV/nucleon) – a dose associated with increased rates of AML [101]. Importantly, this hypomethylation was associated with a dramatic reactivation of both L1 and SINE B1, suggesting a possible link between the epigenetic alterations and development of AML.

2.3. Exposure to ionizing radiation and L1 retrotransposition

Previous studies have indicated that hypomethylation of L1 elements after exposure to radiation is associated with its transcriptional reactivation and may potentially result in its *de novo* retrotransposition [91]. Indeed, accumulating evidence clearly demonstrates that both low- and high-LET irradiation may induce L1 retrotransposition in *vitro*; however, radiation quality-dependent differences were noted. Both X-rays (0, 2.5, 5 or 10 Gy; dose rate of 0.86 Gy/min; 200 kV, 20 mA) and carbon-ion beams (0, 1, 2 or 4 Gy; 18.3 MeV/nucleon, 108 keV/μ m) similarly increased frequencies of transcription in L1/reporter knock-in human glioma cell line NP2; at the same time, X-ray induced the *de novo* insertions of 5'-truncated L1s, while carbon-ion beams promoted insertions of full-length or long-sized insertions [107]. Interestingly, radiation-induced retrotransposition has been shown to regulate gene expression in the human hybrid endothelial cell line EA.hy926 with the human L1 [108]. Although it is expected that radiation-induced L1 hypomethylation may serve as one of the key mechanisms for L1 retrotransposition, no studies, to the best of our knowledge, have been performed to analyze simultaneously the methylation, transcription, and retrotransposition of L1 in response to irradiation.

2.4. Effects of UV-radiation on TEs

UV-radiation, derived primarily from sun, is known to target cellular processes in human skin, including induction of inflammation, immune-suppression, cell death, and premature aging and can serve as both tumor initiator and promoter in skin carcinogenesis [109-111]. Chronic exposure to sub-apoptotic doses of UV results in transformation of the normal human keratinocyte cell line, HaCaT, and upregulation of L1 ORF2 [112]. Importantly, among all the exposed cells, increased mRNA transcripts of L1 ORF2 were detected only in transformed cells, suggesting L1's role in UV-induced tumorigenesis. Another study reported L1 activation in human cervical carcinoma cells (HeLa) upon exposure to UV [113]. Although DNA methylation was not analyzed in any of these studies, the results from the recent report suggest the possibility of hypomethylation-induced reactivation of L1 in response to UV exposure since a decrease in peripheral blood lymphocytes L1 methylation was associated with increased solar UV exposure [114].

Of particular interest are several investigations that report involvement of human endogenous retroviruses (HERV) in melanoma, the most malignant type of skin cancer. Increased activity of the HERV-K(HML-2), an HERV family member, has been suggested to play a critical role in the etiology of melanoma, most probably due to increased promoter activity [115,116]. Additionally, several melanoma cell lines have been characterized by the production of retrovirus-like particles with reverse transcriptase activity [117]. A recent study by Schmitt et al. identified an HML-2 locus transcribed in melanomaderived samples only [118]. At the same time, it has been demonstrated that exposure to UV can activate HERVs, and HML-2 transcripts, in particular [118-120].

2.5. Conclusions

Overall, the exposure to terrestrial radiation at doses above 1 Gy is characterized by the loss of global DNA methylation, which stems from the TEs, and L1, particularly. At the same time, global DNA hypermethylation is characteristic for exposure to space radiation – proton and heavy ions. Limited numbers of studies suggest that this hypermethylation originates primarily from TEs, rather than from specific genes, but more studies are needed to confirm this notion. The principal differences in methylation of TEs in response to either terrestrial or space radiation may underlie the differences in biological responses to these types of radiation. Both terrestrial and space radiation cause long-term alterations in TEs-associated DNA methylation and expression, as has been shown both *in vitro* and *in vivo*, and lead to increased rates of L1 retrotransposition, suggesting their involvement in the development of delayed radiation-induced health outcomes. At the same time, it has to be taken into consideration that the data on methylation and expression of TEs is derived from *in vitro* and *in vivo* experiments, and the successful studies on retrotransposition were performed *in vitro* only. Thus, the translational relevance of these studies should be interpreted with caution, since no data, to our knowledge, exist on the effects of exposure to radiation on TEs methylation, expression, or retrotransposition in normal tissues in humans. Another considerable limitation of the studies that used rodent models is utilization of mainly male animals, while sex differences in response radiation exposure, including effects of DNA methylation are well-documented [89]. It also has to be acknowledged that the effects of low doses of radiation that are characterized by non-linear responses remain to be a challenge in assessment of the effects exerted on TEs in experimental systems.

3. Transposable elements and environmental contaminants

3.1. Particulate matter (PM)

The inhalation of atmospheric pollutants has been associated with eye, nose, and throat irritation, wheezing and breathing difficulties, allergies, worsening of respiratory diseases, such as asthma, and chronic obstructive pulmonary disease, lung cancer, increased cardiovascular morbidity, and hospitalization and emergency room visits for heart attack and stroke [121]. In addition, long-term exposure can cause impaired function of the respiratory tract, and significant damage to immune, neurological, and reproductive systems [122,123]. *A priori*, poor air quality affects everybody; however, specific groups of people (newborn and children, older adults, and those with compromised health) may be more susceptible to air pollution [124].

Ambient PM is a spatiotemporally variant and heterogeneous suspension of tiny solid particles and liquid droplets. The size, shape, and chemical content of PM depend on emissions from various sources and physicochemical transformations during atmospheric transport [125]. Exposures to different particle types and sizes [10-2.5 μ m: PM₁₀-_{2.5} (thoracic/coarse), 2.5-0.1 μ m: PM_{2.5} (fine), and; >0.1 μ m: PM_{0.1} (ultrafine)] are associated with premature mortality, increased hospitalizations, and the development/progression of sub-clinical and clinical disease in the respiratory and circulatory tracts as well as in unexpected organs, such as liver and brain [126-129]. The International Agency for Research on Cancer (IARC) has recently classified outdoor air pollution as carcinogenic to humans (Group 1 carcinogen) [130]. PM originating from traffic depletes antioxidants, with transition and heavy metals and oxidized organic compounds being the primary determinants of the observed responses [131-133]. PM from biomass burning appears to proceed through the same mechanism of action [134], although the chemical content and size distribution is different than that observed for traffic, dust, and biogenic particles [135,136]. Secondary inorganic aerosols (sulfate and nitrate formed from the oxidation of SO2 and NOx, respectively) have shown less toxicity *in vitro*; however, statistically significant associations of sulfates and nitrates with various health outcomes have been reported [137].

There have been numerous epidemiological studies linking ambient air pollution to an altered methylation status of L1 and *Alu* in the blood of exposed humans. For instance, occupational exposure at one of the largest steel, oil refinery, and petrochemical complexes at the Ma Ta Phut industrial estate in Southeastern Asia is associated with lower levels of L1 methylation and 3-(2-deoxy-β-D-erythropentafuranosyl) pyrimido[1,2-α]purin10(3H)-one deoxyguanosine adducts [138]. This was evident by lower levels of L1 in the blood leucocytes of 67 Ma Ta Phut industrial estate workers in comparison with 65 Ma Ta Phut residents (74.8% vs 78%; *p*<0.001). Similar effects were observed in 63 workers in an electric furnace steel plant in Brescia, Northern Italy, where both L1 (β = -0.19 %5methylcytosine; $p=0.04$) and $\Delta l \mu$ (β = -0.34 %5-methylcytosine; $p=0.04$) were found to be hypomethylated in blood leukocytes after 3 days of work in comparison with those obtained on the first day of a work week [139]. Although the authors suggest that these effects were primarily caused by PM_{10} exposure, they could not exclude that exposures other than PM (heat, polycyclic aromatic hydrocarbons, carbon monoxide, and non-ionizing radiation) may potentially contribute to the observed effects. Short-term (130 minutes) exposure to concentrated ambient particles in healthy adult volunteers led to decreased methylation in *Alu* (β =0.74, adjusted-*P*=0.03), but not L1 elements, and was associated with increased diastolic blood pressure [140]. However, the low number (15) and wide age range (18-60) between the healthy participants do not allow for generalization to different population strats, as suggested by the authors [140].

Several studies in mice and cell culture have confirmed that exposure to ambient PM can trigger an epigenetic response, including alterations within the TEs; however, this response was not as uniform as in epidemiological studies. For instance, in contrast to a general tendency towards DNA hypomethylation observed in the majority of human studies, persistent global DNA hypermethylation was detected in the germline of C57BL/CBA mice

exposed *in situ* for 3 to 10 weeks to ambient air near two integrated steel mills and a major highway [141].

A recent study investigated the short-term (24 hours) *in vitro* exposure of murine RAW264.7 macrophages (the cells that comprise the first line of defense against inhaled particles) to various concentrations of the aqueous extract of ambient PM_{10} (10-200 μ g/ml). Although exposure did not result in changes of global DNA methylation, it did lead to redistribution of methylation patterns between L1, SINE B1, and SINE B2 [142]. This study also reported a dose-dependent loss of expression of all three DNA methyltransferases – *Dnmt1, Dnmt3a*, and *Dnmt3b* – the enzymes involved in the maintenance and *de novo* DNA methylation, suggesting possible global and TEs-associated hypomethylation at later timepoints. Therefore, a study that would investigate the effects of the prolonged exposure to PM would be of particular interest. At the same time, independent of methylation status, exposure to PM resulted in reactivation of SINE B2 in a dose-dependent manner, suggesting the possible role of histone modifications and non-coding RNAs in transcriptional regulation of TEs in the short-term response [142].

3.2. Traffic exhausts including diesel emissions

Traffic is one of the most important sources of urban air pollution contributing carbon mono- and dioxide, PM, and volatile organics and hydrocarbons [133]. Epidemiological studies clearly demonstrate an association between the ambient level of black carbon particles, a tracer for traffic pollution, and human diseases [143,144]. Accumulating evidence indicates that exposure to traffic-related black carbon affects the methylation of TEs. Exposure to black carbon led to decreased methylation of L1 (β = -0.11; 95%) confidence interval, -0.18 to -0.04; *p*=0.002 [145] and *Alu* (β= -0.31; 95% confidence interval, 0.12- 0.50%) in 1,097 blood samples from 718 elderly participants in the Boston area Normative Aging Study [146]. Another study reports that truck drivers from Beijing, China (20 participants) exposed to trafficderived elemental carbon (exposure to elemental carbon $16.6 \,\mathrm{\mu g/m^3}$) exhibited hypermethylation of one of the most youngest *Alu* elements – Yb8 (mean difference=0.4%, *p*=0.039), compared to indoor office workers (20 participants with elemental carbon exposure $16.6 \,\mu\text{g/m}^3$) suggesting that evolutionary age of RE subfamilies may determine differential susceptibility of DNA methylation to airborne pollutants [44].

Diesel exhaust, one of the most toxic traffic-derived pollutants, is the major contributor to fine PM in urban environments and has been classified by the United States Environmental Protection Agency (EPA) as a likely carcinogen [147,148]. Indeed, a number of cohort and case-control epidemiologic studies link diesel exhaust with lung cancer (reviewed in [148] and [149]). The mechanisms of the lung carcinogenesis are largely unknown, mainly due to the complex composition of diesel exhaust: its gaseous phase, which contains 1,3-butadiene, polycyclic aromatic hydrocarbons (PAHs), formaldehyde, benzene and acetaldehyde, and diesel exhaust particles – the largest source of emitted airborne particulate matter. Recent studies indicate that exposure to diesel also leads to altered methylation status of TEs, suggesting the epigenetic alterations may be also involved in lung carcinogenesis. Jiang and colleagues, in their controlled human crossover study, reported that short-term diesel

exhaust inhalation was associated with changes in DNA methylation of circulating mononuclear cells in 16 asthmatics 6 and 30 hours postexposure [150]. Similar to the findings in RAW264.7 macrophages [142], the response to diesel exhaust exposure within L1 and *Alu* elements was not uniform, with sites of both hypo- and hypermethylation being identified [150]. Specifically, out of the 31 L1differentially methylated positions in the genome, 13 increased, while 18 decreased in DNA methylation after exposure. Out of the 25 Alu differentially methylated in response to short-term diesel exposure, 12 were characterized by increase and 13 by decrease in DNA methylation.

3.3. Volatile organic compounds

3.3.1. Benzene—Benzene is a known human carcinogen (classified as Group I carcinogen by IARC), causing acute myeloid and lymphoblastic leukemia and chronic myeloid leukemia [151-153]. Exposure to benzene has been also associated with other hematological disease, including myelodysplastic syndrome and aplastic anemia, bone marrow abnormalities, and neural tube defects [135,137]. Occupational exposure to lowbenzene levels in 78 gasoline filling workers and 77 urban traffic officers was associated with significant hypomethylation of L1 (-2.33%, *p*=0.009) and *Alu* (-1.00%, *p*=0.027) in comparsion with 58 unexposed referents in Milan, Italy [154]. Further work has reported an association of urine benzene biomarkers and hypomethylation of TEs. A study by Fustinoni et al. [155] reported a negative association between the benzene metabolite t,t-muconic acid and L1 and *Alu* methylation in occupationally exposed workers from the abovementioned study [154]. Weak, but significant L1 hypomethylation (-0.15%, *p*<0.01) was also reported with another urinary biomarker of benzene exposure, S-phenylmercapturic acid, in 158 Bulgarian petrochemical workers compared to 50 unexposed office workers [156]. Subsequent studies have demonstrated that loss of L1 methylation is associated primarily with its evolutionary older subfamilies [44].

3.3.2. 1,3-butadiene—1,3-Butadiene is a high-volume industrial chemical used in the production of synthetic rubber, resins, and plastics and is also a component of traffic-related air pollution and cigarette smoke [133,157]. It is a ubiquitous environmental pollutant and, according to IARC, known to be carcinogenic in humans [158]. The genotoxicity of 1,3 butadiene, attributed to its highly reactive metabolites, has been considered a critical event in the initiation of tumorigenesis [159]. Solid evidence of 1,3-butadiene-induced epigenetic alterations also suggests its epigenotoxicity. The latter was tightly associated with alterations in DNA methylation, primarily linked to the loss of L1 ORF1, SINE B1, and SINE B2 methylation in livers of male C57BL/6 mice after 2 weeks inhalational exposure to 1,3 butadiene [160]. Importantly, this study demonstrated the dose-dependent response, where exposure to higher concentrations of 1.3-butadiene was associated with the more pronounced loss of methylation within the TEs. Subsequent studies have shown that these effects were strain- [161] and tissue- [162] specific. Furthermore, 1,3-butadiene-induced hypomethylation of TEs was associated with their subsequent reactivation [160].

3.4. Tobacco smoke

Tobacco smoke is a potent carcinogen responsible for the vast majority of lung cancers in both men and women [163,164]. A number of known carcinogens are present in tobacco

smoke, including aldehydes, polycyclic aromatic hydrocarbons (PAH), N-nitrosamines and aromatic amines, as well as 1,3- butadiene and benzene [149,165].

It is well accepted that tobacco smoke induces both genetic and epigenetic changes (reviewed in [166,167]). Among the latter, alterations in DNA methylation, associated with hypermethylation and subsequent silencing of tumor-suppressor genes, have received the most attention [168,169]. Tobacco smoke has been shown to target DNA methylation within the TEs as well. For instance, exposure to cigarette smoke condensate for up to 9 months under potentially relevant exposure conditions has led to loss of DNA methylation in L1 in human bronchial epithelial cells (HBEC) as well as in fully transformed A549 lung cancer cells [170]. This was associated with the decreased function of DNMT1 DNA methyltransferase, suggesting a possible mechanism for L1 hypomethylation.

Alterations in TEs associated with tobacco smoke are not only limited to the respiratory system. The levels of L1 in the normal esophagus mucosa from 105 patients with esophageal squamous cell carcinoma with history of cigarette smoking was inversely correlated with the Brinkman index (daily number of cigarettes \times years) compared to non-smokers ($p=0.037$) [171]. Additionally, tobacco smoke during pregnancy has been shown to affect placental TEs. As reported by Wilhelm-Benartzi and colleagues [172], the methylation status of one of the youngest *Alu* elements, Yb8, was positively associated with tobacco smoke exposure (*p*<0.01) in 380 placental samples from full-term deliveries at the Women and Infants Hospital in Providence, Rhode Island (USA). At the same time, placental methylation of L1 was not affected by tobacco, as reported in the same studies.

3.5. Conclusions

Overall, it is without a doubt that exposure to different components of atmospheric pollution, including PM, may cause alterations in the methylation and expression of TEs, although we lack the knowledge in regard to the retrotransposition potential of such exposures. However, ambient PM is composed of different types of particles, such as soil and road dust, traffic exhausts, and biomass burning, to name a few. It is unclear, currently, whether different types of PM may cause differential responses in the methylation and expression of TEs and whether these potential differences may serve as biomarkers of exposure for certain types of ambient PM.

As indicated by epidemiological studies, exposures to atmospheric contaminants are mainly associated with hypomethylation of L1 in peripheral blood leukocytes, while some studies also indicate hypermethylation of *Alu* elements. Further *in vitro* investigations showed that the response to the same stressor is not uniform, with both hypo- and hypermethylation detected between the different TEs in the same experimental system [142]. The response to the same contaminant may vary even between the same TEs, depending on their genomic location, as shown by Jiang et al in their study on the effects of the short-term diesel exhaust inhalation in asthmatics [150]. The latter two studies provided clear evidence that alterations in the methylation status of TEs are not unidirectional, as thought before; however, the biological relevance of the TE-associated differentially methylated loci will have to be elucidated. As mentioned above, peripheral blood cells served as a material for the majority of these epidemiological studies. The authors of these studies, however, acknowledge that

the observed differences in TEs methylation, although potentially representing markers of exposure or biological effects, are limited to leukocytes and, therefore, may not necessarily correlate with aberrant DNA methylation in the target tissues. Also, it has to be taken into consideration that the DNA used for the analyses of TEs comprises a pool of DNA from different types of circulating blood cells, primarily – leukocytes. Therefore, the alterations in DNA methylation may be a consequence of the shift in blood cell populations, resulting from the exposure.

3.6. Other environmental pollutants

3.6.1. Persistent organic pollutants (POPs)—Today, almost the entire general population is exposed to various POPs, such as organophosphates, carbamates, and pyrethroids through diet, inhalation, and dermal exposures. Accumulating evidence suggests that exposure to POPs, many of which are known endocrine disruptors, may be associated with increased risks of pancreatic and prostate cancer and non-Hodgkin lymphoma [173-175]. Epidemiological studies have also linked exposure to POPs and obesity [176] and have shown statistically significant correlations between pesticides exposures, preterm labor/fetal death [177], impaired semen quality [178], cryptorchidism [179], and thyroid hormones in women [180]. The Agricultural Health Study ("a prospective study of cancer and other health outcomes in a cohort of licensed pesticide applicators and their spouses from Iowa and North Carolina") provides evidence of an association of pesticides exposures to prostate cancer in farmers [181,182] and childhood cancers [183].

Although the mechanisms of POPs-induced disease remain largely unknown, it is becoming increasingly evident that many of them may be underlined by an altered cellular epigenome [184]. A study performed on the blood samples collected from 71 Greenlandic Inuit, a population highly exposed to POPs, showed an inverse correlation between the blood percent methylcytosine (global DNA methylation) and the concentrations of a number of measured POPs [185]. Furthermore, the same study reported that the global DNA hypomethylation was associated with hypomethylation of *Alu* and that this was affected by increased concentrations of *p*,*p*'-DDT (β= -0.26, p=0.01), *p*,*p*'-DDE (β= -0.38, *p*=0.01), βhexachlorocyclohexane (β= -0.48, *p*=0.01), oxychlordane (β= -0.32, *p*=0.01), α-chlordane (β = -0.75, *p*=0.05), and mirex (β = -0.27, *p*=0.01) in the blood, measured individually, and for the sum of all POPs (β = -0.48, *p*=0.01). The effects observed in L1 were not statistically significant [185].

A subsequent study in a cohort of 86 healthy adults (≥ 40 years of age) Koreans, randomly selected from 1,007 participants as controls in case-control studies on associations of POPs with diabetic or metabolic syndrome, confirmed these findings by reporting *Alu* hypomethylation, associated with the increased blood levels of oxychlordane (β= -0.28, *p*<0.05), trans-nonachlor (β= -0.28, *p*<0.05), and *p*,*p'*- DDE (β= -0.29, *p*<0.01) [186]. At the same time, exposure to POPs was not associated with the methylation status of L1. Importantly, most POPs were lower in the subjects by several orders of magnitude when compared with Greenland Inuit [185]; however, they were still able to cause significant alterations in *Alu* elements methylation. The most recent study in a birth cohort of Mexican-American children (n=358) reports that higher prenatal exposure to *o*,*p'*-DDT and *p*,*p'*-DDE

was associated with lower *Alu* methylation in blood at birth, especially after adjusting for cell type composition $(p=0.02)$ [187]. Weak associations of POPs with L1 methylation were only identified after examining the coexposure to DDT and DDE with polybrominated diphenyl ethers.

3.6.2. Benzo[a]pyrene (BaP)—BaP, a prototypical PAH, is a ubiquitous environmental pollutant and IARC group I carcinogen [164,188]. It is present in tobacco smoke and charcoal-grilled meat and is also a byproduct of fossil fuel combustion. Metabolized to reactive epoxides, BaP covalently binds to DNA causing DNA adducts and is capable of inhibiting DNA repair [189]. Accumulating evidence demonstrates that it is also a potent epigenotoxic carcinogen, and its epigenotoxicity, at least in part, is mediated by effects elicited on TEs.

Earlier studies have shown that exposure to BaP can result in reactivation and initiation of L1 retrotransposition in murine vascular smooth muscle cells vSMCs [174] and in human HeLa cells [190]. Later studies have demonstrated that these effects are tightly associated with the ability of BaP to cause L1 hypomethylation in HeLa cells by inducing early enrichment of the transcriptionally-active chromatin markers histone H3K4me3 and H3K9 acetylation and reducing *DNMT1* association with the L1 promoter [191]. The effects of BaP are not limited to L1 only. Hypomethylation of SINEs and LTR elements have been reported in MCF-7 and HCC1806 breast cancer cell lines after exposure to BaP [192]. Of particular interest in this study is the identification of SINE elements hypomethylated at 16p13.3, the region of the chromosome 16 that includes the *TSC2* tumor suppressor gene, in both investigated cell lines. The expression of the gene, however, was not addressed in this study, and the qualitative nature of the analysis does not allow for evaluation of the extent of the BaP-induced SINE hypomethylation.

3.6.3. Conclusions—In the limited studies, exposure to POPs was associated with significant loss of *Alu* methylation in the peripheral blood cells, independently of the levels of POPs, geographical location of the study, and racial and age differences. At the same time, none of the studies could identify alterations in methylation of L1 except for the study in newborns where weak L1 hypomethylation was identified after examining the coexposure to DDT and DDE with polybrominated diphenyl ethers [187]. These findings suggest that *Alu*, but not L1 elements, are primarily targeted by POPs; therefore, the methylation status of *Alu* may serve as a surrogate biomarker for exposure to POPs.

Exposure to BaP is associated with the loss of L1, SINE, and LTR elements, suggesting global DNA hypomethylation effects; however, another study performed on C3H/10T1/2 mouse embryonic fibroblast cells, reports global DNA hypermethylation in response to BaP treatment [193]. One possible explanation of this discrepancy is that in the latter study, cells were exposed to BaP for 4 weeks, while the former studies aimed to evaluate relatively short-term effects. It has to be noted that all the investigations on BaP effects on TEs were performed *in vitro*, and *in vivo* experiments using the rodent models are clearly needed to confirm these findings.

3.7. Metals

Massive utilization of metals in industry has led to increased occupational and environmental exposure to them, many of which are considered human carcinogens [123]. Other, non-cancerous health effects of exposure to metals include pulmonary and cardiovascular disease and various neurodevelopmental deficits. Carcinogenic metals are typically weak mutagens and do not form DNA adducts; however, they are capable of causing oxidative damage to DNA. Accumulating evidence clearly indicates that exposure to metals, independent of their genotoxicity and mutagenicity, is associated with epigenetic alterations. The vast majority of investigated metals is capable of altering DNA methylation patterns, both *in vitro* and *in vivo*, and this feature is also documented in epidemiological studies [194-197].

3.7.1. Arsenic—The US EPA has listed arsenic as the number 1 chemical in its Comprehensive Environmental Response, Compensation and Liability Act Priority List of Hazardous Substances [198]. Arsenic is a ubiquitous environmental contaminant present in soil, rocks, aquatic environments, and even as airborne particles. Arsenic has been also used in insecticide production and treatment of acute promyelocytic leukemia. The most common exposure of humans to arsenic occurs via consumption of contaminated underground water and food in the form of either arsenite (As^{III}) or arsenate (As^{V}) [199,200]. Health effects associated with arsenic exposure are various and include cardiovascular, respiratory, reproductive, and gastrointestinal disease and neurologic defects [201-203]. Exposure to inorganic arsenic is also associated with skin, liver, urinary, and lung cancers; and it has been, therefore, classified as a Group 1 human carcinogen by IARC [204,205].

Arsenic can affect a variety of cellular processes, including signal transduction pathways and redox status; it can also induce deletion mutations and chromosomal aberrations, but not point mutations (reviewed in [194]). Accumulating evidence also suggests that arsenic may promote cellular transformation via epigenetic mechanisms and DNA methylation particularly. Indeed, arsenic, in order to be detoxified and excreted from the cell, needs to be reduced and methylated. Studies indicate that chronic exposure to arsenic results in global DNA hypomethylation, decreased levels of S-adenosyl methionine (the main donor of methyl groups for DNA methylation), and reduced DNA methyltransferase activity [206,207]. *In vitro* and epidemiological studies suggest this global DNA hypomethylation originates from TEs, such as L1, whose methylation was negatively associated with the increasing arsenic exposure [208-211]. For instance, levels of arsenic in the 90th percentile were associated with reduced L1 methylation ($p=0.04$) in the blood of 459 adult participants of a population-based incident case control study in New Hampshire (USA) [208]. *In utero* exposure to arsenic, however, despite the significantly higher levels of arsenic in cold blood, fingernails, toenails, and hair of newborns, was not characterized by altered methylation status of L1 in their cord blood lymphocytes as reported by Intarasunanont and colleagues [209]. On the other hand, the same study reports that *in vitro* exposure of human lymphoblasts resulted in loss of L1 methylation after short-term (4-8 hrs) and long-term (8 weeks) exposure with arsenite [209]. Toenail levels of arsenic showed a trend of negative association with L1methylation (β = - 0.05 per one interquartile range [0.06 μg/g] increase in toenail arsenic; 95% confidence interval $= -0.11$ to 0.02) in 581 participants from the

Normative Aging Study in Boston [210]. Interestingly, hypermethylation of *Alu* elements was reported in the same study (β = 0.08 per interquartile range [0.06 μg/g] increase in toenail arsenic; 95% confidence interval $= 0.03$ to 0.13) [210], suggesting that alterations in DNA methylation patterns upon arsenic exposure are also not unidirectional.

Arsenic can also trigger retrotransposition events. It has been shown that exposure to low levels of arsenic trioxide resulted in L1 retrotransposition in HepG2 cells [212]. Whether or not this was caused by L1 hypomethylation remains unknown since L1 methylation status was not addressed in this study. Another study reported arsenic-induced retrotransposition of viral-like 30 (VL30) LTR in mouse NIH3T3 cells that was inhibited upon treatment with antioxidant N-acetyl-cysteine [213].

3.7.2. Cadmium—Cadmium is a common environmental and occupational toxicant and is one of the most widespread environmental pollutants among metals. Battery manufacturing, tobacco usage, and consumption of certain foods are the main source of exposure to cadmium. The exposure has been associated with adverse health effects, including bone and kidney damage and developmental impairment [214-216]. Cadmium has been also classified as human carcinogen [217]. It is capable of causing damage to DNA only at high concentrations and does not form stable DNA adducts. *In vitro* studies clearly indicated that cadmium is a potent epigenotoxic stressor with abilities to affect both global and genespecific methylation [218-221]. An epidemiological study in 202 non-smoking Argentine women reported that DNA hypomethylation in peripheral blood was associated with dietary exposure to cadmium and indicated that this was associated with hypomethylation of L1 (β= –0.50, $p = 0.0070$) [197].

3.7.3. Lead—Lead is a heavy metal still considered as one of the most harmful environmental toxicants for young children, pregnant women, and other adults, despite great legislative efforts to reduce the use of lead and to control exposure pathways [222]. Sources for lead exposure include deteriorated lead based paints in older homes, drinking water, burning of fossil fuels, trades, hobbies (e.g., making stained-glass windows), ammunition, batteries involved in construction and mining, consumer products, and cultural home health remedies (e.g., azarcon, ayurvedic) [223,224].

The absorption of lead can lead to irreversible neurological damage, especially in children. Even at low exposures, over time, loss of intelligence and problems with hearing, memory, and balance have been documented [225,226]. Pregnant and lactating women are also at greater risk to low levels of lead exposure, where the weight-of-evidence is convincing that exposures can lead to lasting adverse neurotoxic outcomes on the fetus and on infant development because lead crosses the placenta and is found in breast milk [227,228].

Health outcomes of lead exposure have been related to oxidative stress and effects on the immune system based on genetic determinants, even at low levels [229,230]. Exposure to lead has also been associated with alterations in DNA methylation in peripheral blood. Specifically, L1 (β = –0.25; *p*< 0.01), but not *Alu* (β = –0.03; *p*= 0.4), hypomethylation was associated with patella lead levels in 2280 male participants of the Normative Aging Study [231]. In the Early Life Exposures in Mexico to Environmental Toxicants (ELEMENT)

study, an inverse dose-response relationship was found in which quartiles of patella lead correlated with umbilical cord blood L1 methylation (*p* for trend 0.01) 103 analyzed samples [232]. At the same time, tibia lead correlated with *Alu* methylation (*p* for trend 0.05) [232]. Furthermore, maternal tibia lead was negatively associated with the methylation status of *Alu* elements in umbilical cord blood (β = –0.027; *p*= 0.01). A recent case-control study that involved 53 workers from a battery plant reported a significant decrease in blood L1 methylation (-10,3%, *p*<0.001), as compared to 57 healthy volunteers with matching age and gender [233]. The same study also reports a significant loss of L1 methylation in human embryonic kidney HEK293 cells 24 and 48 hrs after *in vitro* exposure to lead acetate [233].

3.7.4. Mercury—Mercury is among the World Health Organization (WHO)'s top 10 chemicals of major public health concern [234]. Exposure to mercury, a heavy metal pollutant, is a concern in metal and electric industries and can also occur with consumption of seafood and exposure from burning fossils. Additionally, it has been estimated that over 6,000 tons of mercury are released into the environment annually worldwide, suggesting mercury evaporation from contaminated sites as another source of exposure [235].

Biological effects elicited by mercury may be mediated via modulation of epigenetic mechanisms. Exposure to methylmercury has led to global DNA hypomethylation in neural stem cells [236]. Limited data exist on the effects of mercury on TEs. Occupational exposure to mercury among 131 dental professionals was not characterized by alterations in L1 methylation in DNA isolated from buccal mucosa; however, hypomethylation of *SEPP1* promoter, the gene known to bind mercury, was detected among male participants [237]. Interestingly, a recent study reported activation of L1 in BE (2)-M17 human neuroblastoma cells, exhibited as increase in mRNA, protein expression, and genomic retrotransposition upon exposure to mercury [238]. Mercury was not able to induce similar events in three other non-neuronal cell lines, underlining the neurotoxic effects of mercury exposure.

Other environmentally ubiquitous heavy metals targeting the cellular epigenome include nickel and chromium; however, the knowledge on how they impact TEs is limited. For instance, chromium has been shown to cause global DNA hypomethylation in A549 lung cells and B lymphoblastoid cells, as well as in red blood cells of industrial chromate workers, but the methylation status of TEs was not addressed in these studies [239,240]. One study indicates that exposure to nickel chloride increased the L1 retrotransposition rates about 2.5 times in human HeLa cells [241].

3.7.5. Conclusions—The existing data on the effects of metals on TEs is derived primarily from epidemiological studies. Analysis of data obtained from human studies holds a number of advantages over the analysis of data obtained from *in vitro* and *in vivo* models. On the other hand, analysis of L1 and *Alu* methylation was performed mainly in peripheral blood cells, suggesting this data can be further utilized in development of biomarkers of exposure; however, it provides limited knowledge in regards to the molecular effects of metal exposure in their correspondent target organs.

Overall, exposure to metals is associated with the loss of blood L1 methylation, while the methylation status of *Alu* elements varies between the exposures and cohorts. Similar to the

effects observed in response to PM exposure, different TEs may be differentially methylated in response to exposure to the same metal within the same experimental system. The results are also dependent on the tissue analyzed. For instance, lead patella levels were associated with L1 hypomethylation but not *Alu*, while tibia lead levels were associated with hypomethylation of *Alu*, but not L1 elements [232]. What determines such differences is currently unknown.

Accumulating evidence has also indicated that metals may induce retrotransposition events. Arsenic has been shown to be capable of causing retrotransposition of L1 and LTRs in human HepG2 and mouse NIH 3T3 cells, respectively, while mercury initiated L1 retrotransposition in human BE (2)-M17 cells, but not in three other non-neuronal cell lines. Both cell lines used for assessing of arsenic- and mercury-induced L1 retrotransposition are genomically unstable tumor cell lines; therefore, studies that will utilize the non-tumorous cell lines will be needed to confirm the L1 mobilization as an effect of metal exposure.

4. Concluding remarks

Evidence summarized in this review suggests that TEs are the sensitive endpoints for detection of effects caused by environmental stressors. Hypomethylation of L1 elements is the most frequently reported consequence of exposure observed in the studies. It was detected in numerous *in vitro* and *in vivo*, as well as in epidemiological studies; in target tissue and in peripheral or cord blood; in the tissue of exposed subjects in comparison to the tissue of control subjects and was inversely correlated with the levels of an environmental contaminant detected in the same subject. Loss of DNA methylation within the TEs, and L1 particularly, contributes to global DNA hypomethylation [91,93,160], given their abundance in mammalian genomes. Hypermethylation of TEs is a much less frequently reported effect of exposures, and the biological consequences of accumulation of methyl groups within the TEs still need to be elucidated. Limited studies that report hypermethylation of TEs link it to their transcriptional silencing [94], possibly suggesting formation of more condensed heterochromatic structures.

Accumulating evidence also demonstrates that loss of epigenetic control over TEs caused by environmental stressors may result in reactivation of TEs and initiation of retrotransposition. Numerous environmental stressors, including terrestrial, space, and UV-radiation, B[a]P, and metals were shown to cause L1 and ERV/LTR retrotransposition in experimental systems, suggesting the possible link between the exposure and development of disease.

4.1. Current challenges

Despite the significant progress in the field and rapid accumulation of data in the last ten years, some challenges exist in the comprehensive analysis and understanding of the effects exerted by the environment on TEs.

While in cancer both *Alu* and L1 are usually found to be hypomethylated, the early response to environmental toxicants and carcinogens, as evident from Table 2, often results in hypomethylation of L1 and hypermethylation of *Alu*. Furthermore, in the wealth of studies, only some classes of TEs are reported to be affected, while no significant alterations have

been identified in the others. For instance, hypermethylation of *Alu* was frequently reported in response to exposure to POPs, while the methylation status of L1, analyzed within the same studies, remained unchanged. What underlies such differences remains unknown, but one of the possible explanations is the fact that different TEs belong to different regions of chromosomes: *Alu* are preferentially enriched in hypomethylated euchromatic regions, while L1 are found in hypermethylated heterochromatic and transcriptionally silent regions.

Differences in response to environmental cues may also be predetermined by differences in the tissue-specific methylation status of TEs. For instance, L1 in placenta have been reported hypomethylated in comparison with other tissues [42]. This might explain the lack of alterations in placental L1 methylation in response to *in utero* tobacco smoke exposure, discussed in this review earlier [172].

Furthermore, it has to be taken into consideration that certain differences in TEs methylation and expression are described between the different cell lines and even within the same tissue in different mouse strains [42,167]. Similarly, a comparison of results obtained from epidemiological studies is often challenging due to the high heterogeneity in human populations involved in these studies. Additionally, human studies are usually limited to biological fluids only, thus, often not allowing for investigations within the target organs.

Other challenges include differences in techniques and approaches for evaluation of the same endpoint that may yield incongruent results; complex structure and different evolutionary age of TEs; high variability of TEs in their retrotransposition activity; exhibited polymorphism of TEs; and extrapolation of results obtained from *in vitro* and *in vivo* studies to humans [45,49,242-244].

4.2. Future directions

4.2.1. TEs in risk assessment and as biomarkers of exposure—Epigenetic parameters hold a number of advantages over the genetic end-points and, therefore, have been proposed to be incorporated into current platforms for risk and safety assessment and as biomarkers of exposure to environmental stressors [167,245,246]. One of the particular concerns that has to be addressed before incorporation of epigenetics into regulatory practice is the large number of epigenetic alterations caused by exposures and observed in diseases. Therefore, the need for identification of specific parameters exists. Based on the growing evidence of numerous environmental stressors ability to affect the methylation of TEs and the persistent nature of these effects, TEs may serve as a unique platform for risk and safety assessment, as well as serve as biomarkers of exposure to environmental stressors with potential epi-mutagenic effects. However, it has to be determined which specific TEs and their families are the most sensitive targets for environmental stressors, as it has been shown on the example of L1 element that exposure to PM affects the methylation status of specific L1 families [44]. Furthermore, there is a need to determine a generally available methodology for TEs methylation assessment that will yield reproducible and comparable results across different laboratories.

4.2.2. TEs: role in cancer and biomarkers in cancer diagnostics—Global genomic hypomethylation that is usually originated from TEs is associated with

decondensation of the chromatin structure, may affect the expression of genes that acquired TE insertions during evolution and lead to genomic instability. Furthermore, hypomethylation itself may predispose to tumor development, as evident from the mouse models of erythroleukemia and lymphoma [247,248].

Alterations in the methylation status of TEs observed after exposure may also result in the loss of proper control over the TEs expression and may subsequently lead to new somatic retrotransposition events. Up to date, a number of L1 retrotranspositions have been reported in human cancers [64,65,249]. Although retrotransposition rates in cancer are relatively low, insertional mutagenesis that results in inactivation of tumor-suppressor gene may have a detrimental effect on cancer initiation and promotion, as has been shown in the case of *APC* gene and colorectal cancer [62]. A recent study indicating that the TE-associated insertional mutagenesis in cancerous tissues preferentially occurs at genes commonly mutated in cancer and substantially disrupts their expression [64] further supports this hypothesis.

As has been shown above, environmental stressors, many of which are carcinogens or suspected carcinogens, are capable of causing persistent alterations in methylation of TEs and initiate retrotransposition. Furthermore, accumulating evidence demonstrates that L1 becomes a useful tool in cancer diagnostics, including utilization of L1 as a surrogate biomarker in identifying the cancer stage, tumor response to chemo- and radiotherapy and determination of tumor metastatic potential (reviewed in [45]). Similarly, L1 and *Alu*/SINEs present a promising platform for the development of surrogate biomarkers of exposures to environmental stressors with carcinogenic potential.

In conclusion, exposures to environmental stressors contribute to over 13 million deaths each year due to cancer and non-cancerous diseases [249]. Although the etiology and pathogenesis of these diseases vary, they are all characterized by alterations in TEs, including their hypo- and hypermethylation, reactivation, and retrotransposition. Accumulating evidence clearly demonstrates that these alterations are not the simple consequence of the disease, but often may drive the pathogenesis, as they can be detected early during disease development and even shortly after exposure to a given environmental stressor. These findings suggest that TEs may potentially be introduced into safety and risk assessment and serve as biomarkers of exposure to environmental stressors, as well as surrogate biomarkers in clinic and possible targets for therapeutic modalities for disease treatment and prevention.

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Abbreviations

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Figure 1. The structure of the main transposable elements in mammalian genomes L1 – LINE-1, Long Interspersed Nucleotide Element 1; **SINE** – Short Interspersed Nucleotide Element; **ERV (LTR)** – Endogenous Retrovirus (Long Terminal Repeat); **UTR** – Untranslated Region; **ORF** – Open Reading Frame; *gag*, *pol*, *env* – encoded genes; **IR** – Inverted Repeats.

Miousse et al. Page 40

Figure 2. The mechanisms of LINE-1 (L1) and *Alu* **retrotransposition**

L1 – LINE-1, Long Interspersed Nucleotide Element 1, autonomous retrotransposon; *Alu* – human SINE element, nonautonomous retrotransposon that utilizes L1 machinery for its own retrotransposition; **UTR** – Untranslated Region; **ORF** – Open Reading Frame; **AAAAA** – polyA tail.

Table 1

Transposable elements in mammalian genomes.

Table 1

Effects of exposure to environmental stressors on transposable elements.

Effects of exposure to environmental stressors on transposable elements.

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Miousse et al. Page 44

smoking asthmatic subjects exposed to filtered air and diesel

CpG sites overlapping with *Alu* and L1 elements

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Miousse et al. Page 46

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Miousse et al. Page 47

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