

Impact of air quality on lung health: myth or reality?

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Abstract: The respiratory system is a primary target of the harmful effects of key air pollutants of health concern. Several air pollutants have been implicated including particulate matter (PM), ozone (O₃), nitrogen dioxide (NO₂) polycyclic aromatic hydrocarbons (PAHs) and volatile organic compounds (VOCs). It is well known that episodes of exposure to high concentrations of outdoor air pollutants can cause acute respiratory exacerbations. However, there is now increasing evidence suggesting that significant exposure to outdoor air pollutants may be also associated with development of lung cancer and with incident cases of chronic obstructive pulmonary disease (COPD) and respiratory allergies. Here we provide a critical appraisal of the impact of air pollution on respiratory diseases and discuss strategies for preventing excessive exposure to harmful air pollutants. However, the evidence that significant exposure to air pollutants is causing COPD, lung cancer or respiratory allergies is not conclusive and therefore regulators must be aware that execution of clean air policies may not be that cost-effective and may lead to unintended consequences. Addressing the lung health effects of air pollution must be considered work in progress.

Keywords: air pollution, asthma, COPD, lung cancer, respiratory diseases

Introduction

Air pollution may be an important environmental risk factor with global, public health implications. The World Health Organization (WHO) estimates that outdoor air pollution may have caused 3.7 million premature deaths worldwide in 2012 [WHO, 2014].

Given that an adult inhales on average 10–15 m³ of air per day, it is obvious that inhalation represents the fundamental exposure mechanism to airborne pollutants in man [Phalen, 2008]. Consequently, it is not surprising that the respiratory system becomes a primary target of the harmful effects of key air pollutants of health concern including particulate matter (PM), ozone (O₃) and nitrogen dioxide (NO₂) [Spirić *et al.* 2012], as well as new and emerging pollutants such as polycyclic aromatic hydrocarbons (PAHs) and volatile organic compounds (VOCs) [Ferrante *et al.*, 2012a]. Most importantly, there is now emerging evidence that co-exposures have the potential to synergistically augment the individual effects of several air pollutants [Sava and Carlsten, 2012].

The overall impact of these exposures on lung health has been the subject of intense research. Epidemiological studies suggest that significant exposure to air pollutants appears to be associated with respiratory exacerbations. Less convincing evidence is available for the causation of chronic obstructive pulmonary disease (COPD), lung cancer and respiratory allergies (Table 1).

Particulate matter with an aerodynamic diameter less than 2.5 µm (PM_{2.5}) and O₃ are amongst the most studied air pollutants of health concern [Huang, 2014]. It seems as if the size of particles is directly linked to their potential for causing health problems. Particles less than 10 µm in diameter can get deeper into the lungs and some may even get into the bloodstream [Nemmar *et al.* 2002]. However, based on 23 estimates for all causes mortality, a 10 µg/m³ increment in PM_{2.5} was associated with only a 1.04% increase in the risk of death [Atkinson *et al.* 2014]. Last but not least, it should be noted that specific changes in climate appears to have a negative impact on overall air quality with substantial

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Table 1. Effects of common air pollutants on lung function and respiratory diseases.

Outdoor pollutants	Lung health outcomes				
	Lung function decrements	Lung function growth	Asthma and COPD exacerbation	Acute respiratory symptoms	Lung cancer
PM10	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation
PM2.5	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation
UFP	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation	Positive Correlation
Ozone	Positive Correlation		Positive Correlation	Positive Correlation	Positive Correlation
Nitrogen oxides (NO _x)	Positive Correlation		Positive Correlation	Positive Correlation	
SO ₂			Positive Correlation	Positive Correlation	
PAHs					Positive Correlation
Benzene*			Positive Correlation	Positive Correlation	
1,3-Butadiene*	Positive Correlation		Positive Correlation	Positive Correlation	Positive Correlation
Aldehydes	Positive Correlation		Positive Correlation	Positive Correlation	Positive Correlation
Inorganic arsenic					Positive Correlation
Chromium	Positive Correlation		Positive Correlation	Positive Correlation	
Nickel			Positive Correlation	Positive Correlation	
Cadmium	Positive Correlation		Positive Correlation	Positive Correlation	
Phthalates			Positive Correlation	Positive Correlation	

*Volatile organic compound (VOC)
COPD, chronic obstructive pulmonary disease; PAH, polycyclic aromatic hydrocarbon; PM, particulate matter; UFP, ultrafine particle.

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increase in respiratory morbidity and mortality of patients with chronic lung conditions [Bernstein and Rice, 2013].

There is indication that the implementation of strategies to reduce air pollution may lead to significant health benefits. For example, the US Environmental Protection Agency (US EPA) estimated that the implementation of measures to reduce emissions from diesel engines might result in 12,000 fewer mortalities and 8900 less hospital admissions in the United States each year [US EPA, 2004]. However, while legislation to control air pollution has vastly improved air quality in many regions of the world, there still remain many countries with heavily polluted cities and increasing vehicle exhaust emissions. In China, ambient air pollution is associated with more than 300,000 deaths and 20 million cases of respiratory illnesses annually [World Bank, 2007]. Outdoor air pollution resulted in 1.2 million premature deaths in China in 2010, or nearly 40% of premature deaths worldwide due to pollution in the world [Lim *et al.* 2012]. Alarming levels of air pollution in China have been reported in as many as 31 Chinese cities; the median of daily PM_{2.5} concentration exceeds up to 5 times the air quality levels recommended by the World Health Organization (WHO) [Wang *et al.* 2014].

Here we provide a critical appraisal of the impact of outdoor air pollution on respiratory diseases and discuss strategies for preventing excessive exposure to harmful air pollutants.

Air quality and COPD

COPD is a progressive and debilitating disease characterized by a persistent inflammatory response that cannot be reversed and generally leads to progressive decline in lung function, respiratory failure, cor pulmonale, and death [Morjaria *et al.* 2010].

In order to reduce the significant economic burden of COPD [Mannino and Braman, 2007], it is important to identify each contributing factor. Cigarette smoking is currently considered as the most important cause of COPD. However, cigarette smoking is not the sole cause for COPD. Besides disease exacerbation, some authors have also suggested a role for air pollution as an important factor in the development of COPD [Salvi and Barnes, 2009; Mackay and Hurst, 2013]. However, there is insufficient evidence to prove a

causal relationship because of the influence of powerful confounding factors, including tobacco smoking and climatic changes.

Only a few studies have investigated the relationship between outdoor air pollutants and objectively defined COPD [Schikowski *et al.* 2005; Pujades-Rodriguez *et al.* 2009; Schikowski *et al.* 2010] (Table 2). For example, a 4.5% prevalence of COPD was found in 4757 women living in the Rhine-Ruhr Basin (Germany) in a consecutive cross-sectional study conducted between 1985 and 1994. In this study, COPD and pulmonary function were markedly affected by PM₁₀ and traffic-related exposure. For women living less than 100 m from a busy road, COPD was 1.79 times more likely than for those living farther away [Schikowski *et al.* 2005]. In a subsequent long-term follow up of a subgroup of 402 women, the same researchers were able to show a lower COPD prevalence in association with improved air quality (i.e. decreasing PM₁₀ level) [Schikowski *et al.* 2010]. In contrast, a study from Nottingham (UK) involving a cohort of 2644 adults found no significant cross-sectional associations between living in close proximity to traffic or NO₂ levels and spirometry confirmed COPD [Pujades-Rodriguez *et al.* 2009]. Most recently, a 35-year prospective study of more than 57,000 participants in a Danish cohort reported a significant but small positive association between long-term exposure to traffic-related air pollution [NO₂ or nitrogen oxides (NO_x)] and incident COPD in the period from 1993 to 2006 [Andersen *et al.* 2011]. This study considered extensive control of confounders, but was limited by the lack of objective spirometric measurement for the diagnosis of COPD.

If the evidence for an important association between air pollution and COPD development is not very clear, its role as an important trigger of COPD exacerbations it is widely acknowledged. Most of the published studies in this area of research have focused on associations between air pollution and hospital admissions.

In a very large study of hospital admissions related to heart and lung diseases in 10 US cities, a 2.5% increase in COPD admissions was observed for every 10 µg/m³ increase in PM₁₀ [Zanobetti *et al.* 2000]. Another US study found that a sudden increase in PM_{2.5} was associated with a risk of about 0.9% for COPD hospitalizations [Dominici *et al.* 2006]. The role of O₃ was addressed in a

Table 2. Traffic-related exposure studies on COPD.

Study	Date	Location	Pollutants	Subject/location	Results
Schikowski <i>et al.</i> [2005], Schikowski <i>et al.</i> [2010]	1985–1994	Rhine-Ruhr Basin (Germany)	PM10	4757 women living less than 100 m from a busy road	4.5% prevalence of COPD
Pujades-Rodriguez <i>et al.</i> [2009]	1991–2000	Nottingham (UK)	NO ₂	2644 adults aged 18–70 living in close proximity to traffic	spirometry confirmed COPD
Andersen <i>et al.</i> [2011]	1993–2006	Denmark	NO ₂ /NO _x	57,000 adults	incident COPD
Zanobetti <i>et al.</i> [2000]	1986–1994	10 US cities	PM10	adults aged >65 years living in a metropolitan county	2.5% increase in hospital admissions for AECOPD
Dominici <i>et al.</i> [2006]	1999–2002	204 US urban counties	PM2.5	11.5 million adults aged >65 years	risk of about 0.9% for COPD hospitalization
Medina-Ramon <i>et al.</i> [2006]	1986–1999	36 US cities	PM10	warm season	1.47% increase in hospital admissions for AECOPD
Fusco <i>et al.</i> [2001]	1995–1997	Rome (Italy)	NO ₂ and O ₃	residents of all ages and among children (0–14 years)	4.3% increase in hospital admissions for AECOPD
Tao <i>et al.</i> [2014]	2001–2005	Lanzhou, China	PM10, SO ₂ , NO ₂	females and aged ≥65 years	increases in hospital admissions for AECOPD

AECOPD, acute exacerbation of COPD; COPD, chronic obstructive pulmonary disease; NO_x, nitrogen oxides; PM, particulate matter.

large multicity ($n = 36$) study in the US, in which a 2-day cumulative effect of a 5 parts per billion (ppb) increase in O₃ in the course of warm sunny days was associated with a 0.27% increase in admissions for acute exacerbation of COPD (AECOPD). Likewise, a 10 µg/m³ increase in PM10 during the warm season was associated with 1.47% [95% confidence interval (CI): 0.93–2.01] immediate increase in AECOPD [Medina-Ramon *et al.* 2006].

A study assessing the data on admissions for COPD in 6 European cities showed that the relative risk for a 50 µg/m³ increase in daily mean level of SO₂, black smoke, total suspended particulates, NO₂ and O₃ for AECOPD admissions was 1.02, 1.04, 1.02, 1.02 and 1.04, respectively [Anderson *et al.* 1997]. A study in Rome (Italy) reported that carbon monoxide (CO) and the photochemical pollutants, NO₂ and O₃, were important determinants for acute respiratory conditions with a 4.3% increase in COPD admissions [Fusco *et al.* 2001]. A recent study from a rural county of England, where the pollutant concentration is lower than that in the urban area, found that increases in ambient CO, nitric oxide (NO), NO₂ and NO_x concentrations were associated with increases in hospital admissions for

AECOPD, similar in extent to that in the urban areas [Sauerzapf *et al.* 2009]. Other published European studies have investigated the effect of air pollutants on asthma and COPD admissions grouped together instead of analysing them separately, making it difficult to estimate the accurate effect on COPD admissions [Atkinson *et al.* 2001; Halonen *et al.* 2009; 2010].

A report including a systematic and quantitative assessment of 82 time-series Asiatic studies of daily mortality and hospital admissions for cardiovascular and respiratory disease observed that all-cause mortality was associated with increase in ambient PM10, total suspended particles and SO₂ levels [HEL, 2010]. In addition, respiratory admissions were associated with NO₂ and SO₂ levels. However, COPD admissions or mortality were not separately addressed in this study. A single city study in Hong Kong focused specifically on the effect of air pollutants on hospital admissions due to AECOPD from 2000 to 2004 and included 119,225 admissions for AECOPD [Ko *et al.* 2007]. The study observed that the relative risk of hospital admissions for every 10 µg/m³ increase in SO₂, NO₂, O₃, PM10 and PM2.5 was 1.007, 1.026, 1.034, 1.024 and 1.031, respectively. In Lanzhou, one of the most air-polluted

Table 3. Traffic-related exposure studies on lung cancer.

Study	Date	Location	Pollutants	Subject/location	Results
Vineis <i>et al.</i> [2006]	1993–1998	10 European countries	NO ₂	adults aged 35–74 residing near heavy traffic roads	46% increase in lung cancer
Chiu <i>et al.</i> [2006]	1994–2003	Taiwan	PM ₁₀ , SO ₂ , NO ₂	females	28% increased risk of lung cancer
Edwards <i>et al.</i> [2006]	2000–2004	Teesside, northeast England	PM ₁₀ , SO ₂ , NO ₂	women aged <80 years living for >25 years close to highly industrialized area	83% increased risk of lung cancer
Loomis <i>et al.</i> [2014]	2003–2012	31 provincial capital cities in China	PM _{2.5}	71,000 adults	increased risk of lung cancer
Raaschou-Nielsen <i>et al.</i> [2013]	2008–2011	17 separate European cohorts	PM ₁₀	312,944 adults	increased risk of lung cancer

PM, particulate matter.

city in China, stronger effects of air pollutants on respiratory hospital admissions were observed, particularly in females aged ≥ 65 years [Tao *et al.* 2014].

In conclusion, time series studies appear to support a relationship between AECOPD and increasing ambient air pollutant levels, but the role of additional confounding factors has to be considered.

Air quality and lung cancer

Although the most common cause of lung cancer is long-term exposure to tobacco smoke, an estimated 10–25% of cases worldwide occur in never smokers [Couraud *et al.* 2012] and there is now increasing evidence that also exposure to air pollution may contribute to lung cancer in at-risk individuals (Table 3).

NO₂ is a good marker of traffic-related pollution [Vineis *et al.* 2007; NSW Government Advisory Committee on Tunnel Air Quality, 2014]. Near-roadway (within about 50 m) concentrations of NO₂ have been measured to be approximately 30–100% higher than concentrations away from roadways [US EPA, 2014]. The Health Effects Institute (HEI) identified that an exposure zone extending up to 300–500 m from a major road was the most highly affected by traffic emissions [HEI Traffic Panel, 2010]. In a European nested case-control study of non- and ex-smokers, a 46% increase in lung cancer was reported in those residing near heavy traffic roads [Vineis *et al.* 2006]. When individual pollutants were

examined, exposure to each increment of 10 ppb NO₂ produced a 14% increase in lung cancer. Exposure to concentrations greater than 30 ppb resulted in a 30% increase. These findings did not change after controlling for occupational factors and cotinine (a short-term marker of tobacco exposure). In another case-control study examining the risk of outdoor air pollution, women living in the group of Taiwan municipalities with the highest levels of air pollution had a 28% increased risk of lung cancer [Chiu *et al.* 2006]. Likewise, lung cancer risk among women with prolonged (>25 years) residence in a highly industrialized area of northeast England was increased by 83% [Edwards *et al.* 2006].

In 2013, the International Agency for Research on Cancer (IARC) classified outdoor air pollution and related PM as a class I human carcinogen [IARC, 2013]. The IARC evaluation showed an increasing risk of lung cancer with rising environmental levels of PM [Costa *et al.* 2014]. Data from epidemiologic studies in Asia, Europe and North America consistently show positive association between lung cancer and PM exposure and other indicators of air pollution, which persist after adjustment for important lung cancer risk factors such as tobacco smoking [Loomis *et al.* 2014].

Another emerging air pollutant is diesel exhaust particles (DEPs), a complex mixture of thousands of chemicals. Most DEPs have aerodynamic diameters falling within a range of 0.1 to 0.25 μm and are classified as class I human carcinogen by the IARC based on epidemiological evidence for lung cancer [Leem and Jang, 2014].

Table 4. Traffic-related exposure studies on respiratory allergies.

Study	Date	Location	Pollutants	Subject/location	Results
Gehring <i>et al.</i> [2010], McConnell <i>et al.</i> [2010]	1996–1997	Netherlands	PM2.5	3863 yearly from birth until age 8 years	significant increase in incidence and prevalence of asthma
Penard-Morand <i>et al.</i> [2010]	1998–2000	6 French cities	benzene, SO ₂ , PM10, NO _x , and CO	6683 children (9–11 years)	increased risk of asthma and allergic rhinitis
Yamazaki <i>et al.</i> [2014]		Japan	elemental carbon	10,069 school children 6–9-year old	increased risk of asthma incidence
Kunzli <i>et al.</i> [2009]	1991–2002	Switzerland	PM10	adult aged 18–60 years	increased risk of asthma incidence
Young <i>et al.</i> [2014]		U.S.	PM2.5	50,884 women	increased risk of asthma incidence

PM, particulate matter.

In urban settings worldwide, diesel exhaust emissions are known to contribute substantially to the PM quota of air pollution and the possible association of exposures to diesel exhaust and an increased incidence of lung cancer has been raised. However, current epidemiological evidence from population-based case-control studies [Olsson *et al.* 2011; Villeneuve *et al.* 2011] and cohort studies of bus and truck drivers [Birdsey *et al.* 2010; Merlo *et al.* 2010; Petersen *et al.* 2010] fails to develop any confident quantitative estimate of cancer risk due to several methodological problems. Specifically, the population-based case-control studies suffer from inherent defects in job groupings and exposure estimations, insufficient latency periods, inconsistent *a posteriori* subanalyses, nonsignificant exposure–response trends after adjustment for potential confounders, and failures to adjust for the rates of dieselization or for the evolution of diesel engines and fuels (and thus exposure levels) over time. In the cohort studies of bus and truck drivers, incorrect adjustments for confounding factors such as current smoking produced spuriously elevated odds ratios (ORs) that were incorrectly attributed to diesel exhaust exposure.

The European Study of Cohorts for Air Pollution Effects (ESCAPE) allowed collection and analysis of clinical and exposure data from 312,944 members enrolled in 17 separate European cohorts [Raaschou-Nielsen *et al.* 2013]. The association between long-term exposure to ambient air pollution and lung cancer was investigated in the 2095 incident lung cancer cases diagnosed over a follow-up period averaging

12.8 years. The meta-analyses showed a statistically significant association between risk for lung cancer and PM10 [hazard ratio (HR) 1.22 (95% CI 1.03–1.45) per 10 µg/m³]. For PM2.5, a non-significant HR of 1.18 (0.96–1.46) per 5 µg/m³ was reported. Of note, proxies for traffic intensities failed to show a significant association with lung cancer.

New and emerging pollutants should be also considered. VOCs, PAHs that are known to react with other common air pollutants such as O₃, NO_x and SO₂ yielding diones, nitro- and dinitro-PAHs, sulfonic acids, and heavy metals bound to inorganic and organic compounds and/or adsorbed on PM may have mutagenic and genotoxic potential, and may induce DNA adduct formation *in vitro* and *in vivo*, thus qualifying for being probable (group 2A) or possible (group 2B) human carcinogens [Ferrante *et al.* 2012]. However, these have been little studied in epidemiological population studies. Due to known role of nanoparticles in inducing an intense lung inflammation, studies to clarify the role of nanoparticles on the respiratory tract are significant [Bakand *et al.* 2012; Ferrante *et al.*, 2012b].

Air quality and respiratory allergies

It is widely accepted that exposure to ambient concentrations of air pollutants can cause short-term exacerbations in those who already have respiratory allergies (i.e. asthma and rhinitis). In particular, asthma exacerbations – measured as visits to emergency departments – have been frequently reported on days with higher levels of O₃ and other pollutants [Weisel *et al.* 1995].

However, whether air pollutants play a role in the initiation of new cases of asthma in those previously free from the condition is less clear. Main attention has been concentrated on gaseous materials such as O₃ and NO₂, as well as PM, generated by urban traffic and industry (Table 4). The fast expansion of the global vehicular fleet during the past 50 years exposes billions of people to unhealthy and dangerous levels of motor vehicle generated air pollutants. Among the various pollutants emitted from motor vehicle exhausts, airborne PM has been long suggested to be an important contributing factor for the increased prevalence of respiratory allergies in recent years [Polosa *et al.* 2002; Diaz-Sanchez *et al.* 2003; Oliveri *et al.* 2011]. The largest single source of airborne PM is that derived diesel engine emissions; diesel vehicles emit up to 100 times more PM than catalyst-equipped petrol cars of corresponding performance. In cities such as Los Angeles (USA), 40% of the 10 µm particles (PM₁₀) in the atmosphere are derived from diesel vehicle engines [Diaz-Sanchez, 2000]. DEPs have the ability to bind proteins and may serve as a potential carrier of allergens, penetrating deep into respiratory tract [Anderson *et al.* 2013]. In allergic subjects challenged nasally with pollen ±0.3 mg DEP, allergic antibody production is up to 50 times greater if DEP present [Diaz-Sanchez, 2000].

Although initial studies of traffic-related air pollution and asthma were conflicting due to a number of methodological flaws including misclassification of pollution exposure to the individual cases [Jenerowicz *et al.* 2012], subsequent studies using more refined methodologies such as land-use regression (LUR) or dispersion modelling produced some significant associations in both children and adults with chronic pollution exposure.

Two recent studies using these sophisticated techniques to investigate the association between childhood asthma and exposure to traffic-related air pollution at home, school or both have shown increased risk of childhood asthma incidence even after controlling for relevant confounders [Gehring *et al.* 2010; McConnell *et al.* 2010].

The prospective birth cohort study of 3863 children has investigated the association between traffic-related air pollution and the development of objectively diagnosed asthma and allergies during the first 8 years of life. PM_{2.5} levels were associated with a significant increase in incidence

of asthma (OR, 1.28; 95% CI, 1.10–1.49), prevalence of asthma (OR, 1.26; 95% CI, 1.04–1.51), and prevalence of asthma symptoms (OR, 1.15; 95% CI, 1.02–1.28). Similar findings were observed for NO₂. Differences in air pollution effects between different age groups were small with the exception of asthma incidence, for which associations became somewhat stronger at 6–8 years of age [Gehring *et al.* 2010].

In a smaller study, Carlsten and colleagues recruited infants at high familial risk for asthma and examined birth year home exposures to NO, NO₂, black carbon and PM_{2.5} with follow up at 7 years of age. Birth year PM_{2.5} (interquartile range, 4.1 µg/m³) was associated with a markedly increased risk of asthma with an OR of 3.1 (95% CI, 1.3–7.4). NO and NO₂ demonstrated similar associations, but black carbon did not [Carlsten *et al.* 2011].

A study of self-reported allergic disease, using the International Study of Asthma and Allergies in Childhood (ISAAC) questionnaire, and home traffic density based on distance to major roadways found approximately 1.5–3 fold prevalence ratios for heavy traffic density for wheeze, asthma, rhinitis and rhinoconjunctivitis, with no associations for children who slept in air-conditioned homes [Zuraimi *et al.* 2011].

Using a cross-sectional design and an enhanced ISAAC protocol for outcomes, 6683 children in the French Six Cities Study were studied with exposures based on a 3-year dispersion model for each school address to assign individual school exposures [Penard-Morand *et al.* 2010]. Asthma was significantly associated with benzene, SO₂, PM₁₀, NO_x and CO levels; allergic rhinitis was only associated with PM₁₀ levels. Sensitization to pollens was associated with benzene and PM₁₀ levels. This is in accordance with the increase in the proportion of positive skin prick test and total immunoglobulin E (IgE) levels reported in traffic warden of the city of Catania (Sicily) with a well-characterized occupational history of road traffic fumes exposure [Proietti *et al.* 2003].

A cohort study in Japan examined the association between traffic-related air pollution and the development of asthma in school children [Yamazaki *et al.* 2014]. Subjects were 10,069 school children in their first through third years of compulsory education (6–9 years old). As surrogates of traffic-related air pollution, the estimation target was the

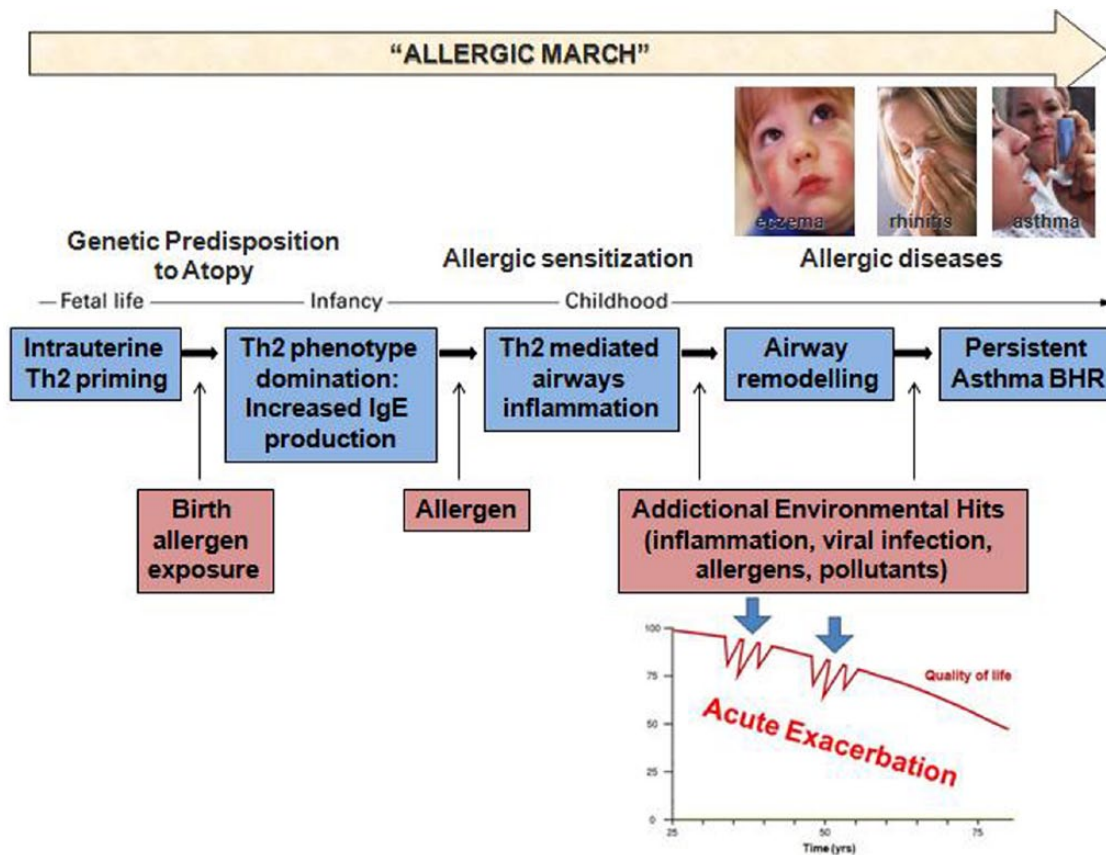


Figure 1. Hypothetical network of interaction between air pollution and asthma. Subjects with genetic predisposition to atopy tend to develop a prevalent Th2 immune phenotype, as a result of birth allergen exposure. Increased IgE production shapes a state of allergic sensitization, with Th2 mediated airways inflammation. Continuous environmental exposure to pollutants or/with other factors (viral infection, allergens, inflammation, etc.), inducing persistent inflammation and airway remodeling, contributes to acute exacerbations on sensitized subjects.

BHR, bronchial hyperresponsiveness; IgE, immunoglobulin E; Th2, type 2 helper T cells.

annual average individual exposure of automobile exhaust-originating NO_x and elemental carbon (EC). The OR (95% CI) for asthma incidence was significant at 1.07 (1.01–1.14) for each 0.1 µg/m³ EC, but not significant for NO_x (OR = 1.01 (0.99–1.03) for each 1 ppb NO_x).

The Swiss Cohort Study on Air Pollution and Lung Diseases in Adults was a population-based cohort of adult lung disease-free nonsmokers initiated in 1991 with 11-year follow up in 2002 [Kunzli *et al.* 2009]. Using a dispersion model that included hourly meteorological and emissions data on industrial, construction, heating, agricultural and forestry, and traffic emissions, the latter separated by type of vehicle (truck *versus* car), each participant was assigned an exposure to PM₁₀. An HR for doctor-diagnosed asthma of

1.30 (95% CI, 1.05–1.61) new cases was found for a given (1 µg/m³ as PM₁₀) change in traffic pollution over the 10 years; this was more frequent in those with baseline atopy or bronchial hyperreactivity.

Trupin and colleagues looked at the simultaneous effect on forced expiratory volume in 1 second (FEV₁) percent predicted and an asthma severity score of diverse social and physical environmental exposures on adult asthma in 176 subjects. Their final model had an R² value of 0.30 for FEV₁ percent predicted and 0.16 for the severity of asthma score and distance to the nearest road was a significant predictor of FEV₁ [Trupin *et al.* 2010].

A recent study estimated the association between traffic-related air pollution exposure (PM_{2.5}) and

incident adult asthma in a nationwide cohort of US women ($n = 50,884$) [Young *et al.* 2014]. For an interquartile range (IQR) difference ($3.6 \mu\text{g}/\text{m}^3$) in estimated PM_{2.5} exposure, the adjusted odds ratio (aOR) was 1.20 (95% CI, 0.99–1.46; $p = 0.063$) for incident asthma. Results suggest that PM_{2.5} exposure may increase the risk of developing asthma in adult women.

The above mentioned studies support an association between various aspects of traffic-related air pollution and new-onset asthma or asthma exacerbation in children and adults. This relationship, at least for the acute exposure domain, has been further strengthened by the findings of recent pollution intervention studies. The Beijing Olympics of 2008 were a great opportunity for environmental researchers to study the effects of greater than usual degrees of changes in air pollution on human health [Cai and Xie, 2011]. One study examined visits for outpatient treatment of asthma at a Beijing Hospital [Li *et al.* 2010]. Although somewhat sparse in clinical detail, they reported a reduction from 12.5 visits per day to 7.3 visits per day, a 41.6% reduction during the Olympic Games.

A mechanistic hypothesis to explain the interaction between air pollution and asthma has been proposed [Gowers *et al.* 2012] (Figure 1). Subjects with a genetic predisposition to atopy are known to develop a strong Th2 inflammatory phenotype with increased IgE production. In presence of continuous environmental exposure, these individuals might end up developing a full-blown allergic inflammatory process and to develop respiratory disorders such as allergic rhinitis and asthma [Bates and Maksym, 2011; Holgate, 2011]. When exposed to allergen, air pollutants or viruses [Willart and Lambrecht, 2009], these individuals are more likely to develop acute inflammatory responses and related respiratory symptoms (i.e. acute exacerbations).

Prevention and public health measures

Air pollution may affect the health of millions of people. Current understanding of the role of air pollutants, both in isolation and in association with climate changes, on human health is unclear [Prüss-Üstün and Corvalán, 2006]. Nonetheless, effective and proportionate policies to reduce environmental air pollution may still be contemplated. The case for action to reduce air pollution can take many forms including urban planning,

technological developments (e.g. the design of new vehicles that produce less pollution), introduction of warning systems and new policies. This strategy must work both at the population level, with preventive policies and at the individual level, by informing and educating the population at risk about reducing exposure to air pollutants. Individuals with heart or lung disease (such as coronary artery disease, congestive heart failure, and asthma or COPD), older adults and children should be warned about higher exposure risk when exercising outdoors. Also pregnant women are at risk of exposing their vulnerable unborn child to the harmful effects of air pollutants. It has been recently shown that exposure to road traffic-related air pollution (i.e. higher levels of benzene and NO₂) during early stages of pregnancy may be associated with lung function defects in children later in life [Morales *et al.* 2014].

At population level, some health benefits can be obtained by implementing specific policies as for the case of congestion charges in large cities such as London and Stockholm. Likewise, Aphekom (Improving Knowledge and Communication for Decision Making on Air Pollution and Health in Europe), a multicountry EU project investigating the effects of EU legislation to reduce inner cities air pollution in several EU capitals [Aphekom Project, 2011], shows that substantial reduction in ambient SO₂ levels is estimated to prevent some 2200 premature deaths valued at €192 million. These findings underscore the combined health and monetary benefits deriving from implementing effective EU policies on air pollution and ensuring compliance with them over time. Since the 1973, the EU has also developed a series of Environment Action Programmes (EAPs) to prevent and reduce the burden of air pollution-related diseases. The most recent ones, 5th EAP (1993–2000) and 6th EAP (2002–2012), are most relevant and set the scene for developing specific policies and directives to control air pollution and improve air quality in the past two decades, with a specific focus on reducing emissions from both the road traffic and industrial sector by 2020 [EEA, 2012]. However, it is important to understand the true impact to human health of specific exposure thresholds to key air pollutants before investing huge amount of taxpayers money in meaningless prevention programs. This is particularly relevant if we consider that there are more common sources of pollution that are not accounted for. For example, living with a smoker exposes an individual to levels of toxicants that

are comparable with those of a heavily polluted city [Semple *et al.* 2014].

Conclusion

If in principle it makes sense to ensure that the air we breathe should be clean as possible, there is however no conclusive evidence that significant exposure to outdoor air pollutants is causing COPD, lung cancer or respiratory allergies. The current available literature has revealed unforeseen complexities and methodological limitations thus increasing, rather than reducing, contradictions and doubts. The extraordinary difficulty of demonstrating the benefits of salt reduction, mammography or various diets, for example, ought to serve as cautionary lessons.

Given the uncertainties about the exposure levels known to be dangerous to lung health and about mortality/morbidity estimations from air pollution, it does not seem right to spend years characterizing and debating ancillary risks of air pollutants that are almost certainly much less serious than the known risks of smoking. Nonetheless, for areas where knowledge is lacking may still be sensible to invoke the precautionary principle. But regulators must be aware that execution of clean air policies may not be that cost-effective and may lead to unintended consequences.

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