



## Research

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**Author for correspondence:**

Hannah B. Blair

e-mail: [hannah.blair@stonybrook.edu](mailto:hannah.blair@stonybrook.edu)

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# Evidence for ship noise impacts on humpback whale foraging behaviour

Hannah B. Blair<sup>1</sup>, Nathan D. Merchant<sup>2</sup>, Ari S. Friedlaender<sup>3</sup>, David N. Wiley<sup>4</sup> and Susan E. Parks<sup>1</sup>

<sup>1</sup>Syracuse University, Syracuse, NY, USA

<sup>2</sup>Centre for Environment Fisheries and Aquaculture Science, Lowestoft, Suffolk, UK

<sup>3</sup>Marine Mammal Institute, Oregon State University, Newport, OR, USA

<sup>4</sup>Stellwagen Bank National Marine Sanctuary, National Oceanic and Atmospheric Administration, Scituate, MA, USA

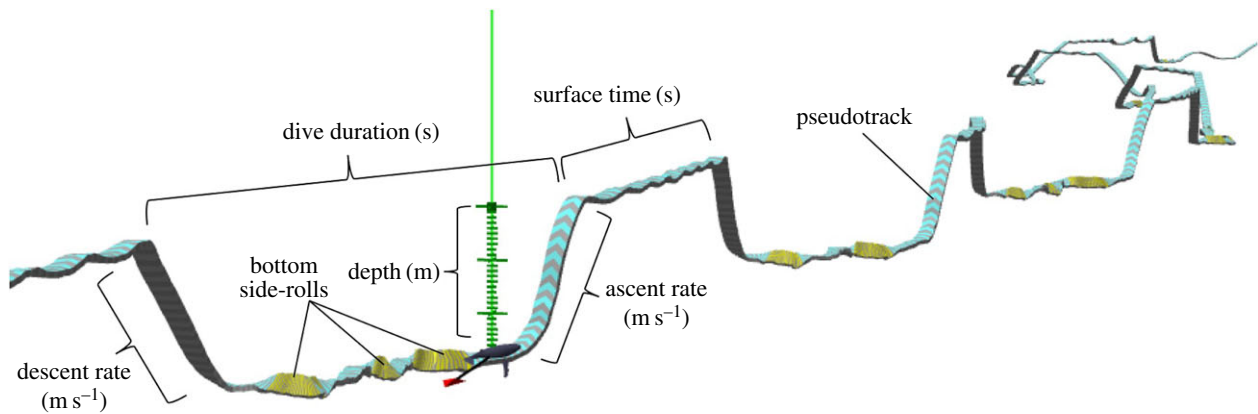
HBB, 0000-0003-2895-1951

Noise from shipping activity in North Atlantic coastal waters has been steadily increasing and is an area of growing conservation concern, as it has the potential to disrupt the behaviour of marine organisms. This study examines the impacts of ship noise on bottom foraging humpback whales (*Megaptera novaeangliae*) in the western North Atlantic. Data were collected from 10 foraging whales using non-invasive archival tags that simultaneously recorded underwater movements and the acoustic environment at the whale. Using mixed models, we assess the effects of ship noise on seven parameters of their feeding behaviours. Independent variables included the presence or absence of ship noise and the received level of ship noise at the whale. We found significant effects on foraging, including slower descent rates and fewer side-roll feeding events per dive with increasing ship noise. During 5 of 18 ship passages, dives without side-rolls were observed. These findings indicate that humpback whales on Stellwagen Bank, an area with chronically elevated levels of shipping traffic, significantly change foraging activity when exposed to high levels of ship noise. This measurable reduction in within-dive foraging effort of individual whales could potentially lead to population-level impacts of shipping noise on baleen whale foraging success.

## 1. Introduction

Increased levels of anthropogenic noise have become a chronic condition in both terrestrial and marine environments [1,2]. Noise pollution has been shown to alter acoustic communication [3], distribution patterns [4] and stress responses [5,6] in a wide range of taxonomic groups. Noise has also been shown to impact foraging behaviours by masking sound produced by prey movement [7], by eliciting an avoidance response or a cessation of foraging [8], or by altering prey behaviour [9]. These wide ranging effects are raising concerns about the impacts of anthropogenic noise on species survival [1,2].

The impact of noise has been a major focus of cetacean research over the past two decades, as whales and dolphins are highly dependent on sound for critical life functions including communication and foraging [10]. Cetaceans are exposed to a variety of anthropogenic noise sources [10] and have been shown to respond in several ways, including physiological and context-dependent changes in behaviour [5,8,10]. Some evidence suggests that odontocetes (toothed whales and dolphins) may alter foraging behaviour in response to noise exposure [11,12]. However, relatively few studies have investigated the effects of ship noise on foraging behaviour in mysticetes (baleen whales). Many mysticetes are found in coastal areas with high levels of ship traffic, resulting in frequent mortalities from collisions [13]. Investigations of foraging



**Figure 1.** TrackPlot still image demonstrating ribbon track and dive measurements for one bottom-feeding dive.

blue (*Balaenoptera musculus*), finback (*Balaenoptera physalus*) and humpback whales using surface behavioural observations have found no detectable responses to loud low-frequency noise [14,15]. The advent of multi-sensor tags has allowed for exploration of subsurface behaviours of baleen whales [16]. In response to mid-frequency sonar playbacks, blue whales show termination of feeding events at depth while humpbacks demonstrate avoidance behaviours [17,18]. Preliminary evidence suggests that close ship passage might result in decreased foraging time in blue whales [19].

Humpback whales are generalist predators with diverse diets and foraging behaviours intended to aggregate and engulf small numerous prey [20]. One well-described foraging method used by humpbacks in the Gulf of Maine is bottom side-roll feeding on sand lance (*Ammodytes* spp.) near the sea floor, particularly during night-time hours [21]. Humpbacks have demonstrated flexible acoustic behaviour in the presence of noise [3,22,23], but few studies have investigated noise effects on their subsurface foraging behaviours. Here, we use subsurface tag data to assess the impacts of shipping noise on this foraging behaviour.

## 2. Methods

### (a) Data collection and analysis

Field data were collected in the southern Gulf of Maine on and around the Stellwagen Bank National Marine Sanctuary (Massachusetts, USA) in June–July from 2006 to 2009 and April 2009. Data were collected using archival digital acoustic recording tags, Dtags [16], to simultaneously record the whale's three-dimensional behaviours and the acoustic environment. Details on field data collection methods are documented in Friedlaender *et al.* [24].

Ten tag deployments were included in the analyses (electronic supplementary material, table S1). These tag records contained data between 1 h after sunset to 1 h before sunrise and included the passage of at least one large ship on the acoustic record. Two of the deployments came from the same individual in different years. A total of 218 dives were analysed: 83 occurred in ship noise exposure periods while 135 occurred with no ship noise. Subsurface behaviours were quantified using the software application TrackPlot (figure 1) [25]. Seven behavioural measurements were extracted from each dive for use as dependent variables in the models: duration, rate of descent and ascent, maximum depth, number of bottom side-roll feeding events, time between dives and surface time immediately following each dive (figure 1).

The ship presence was determined through visual and aural detection of ship noise in Dtag audio recordings (figure 2a). To minimize effects of flow noise, we measured the received level (RL) of ship noise within the 2–3 kHz frequency band for a 1 min period during the bottom time of each dive (RAVENPRO v. 1.5) (electronic supplementary material).

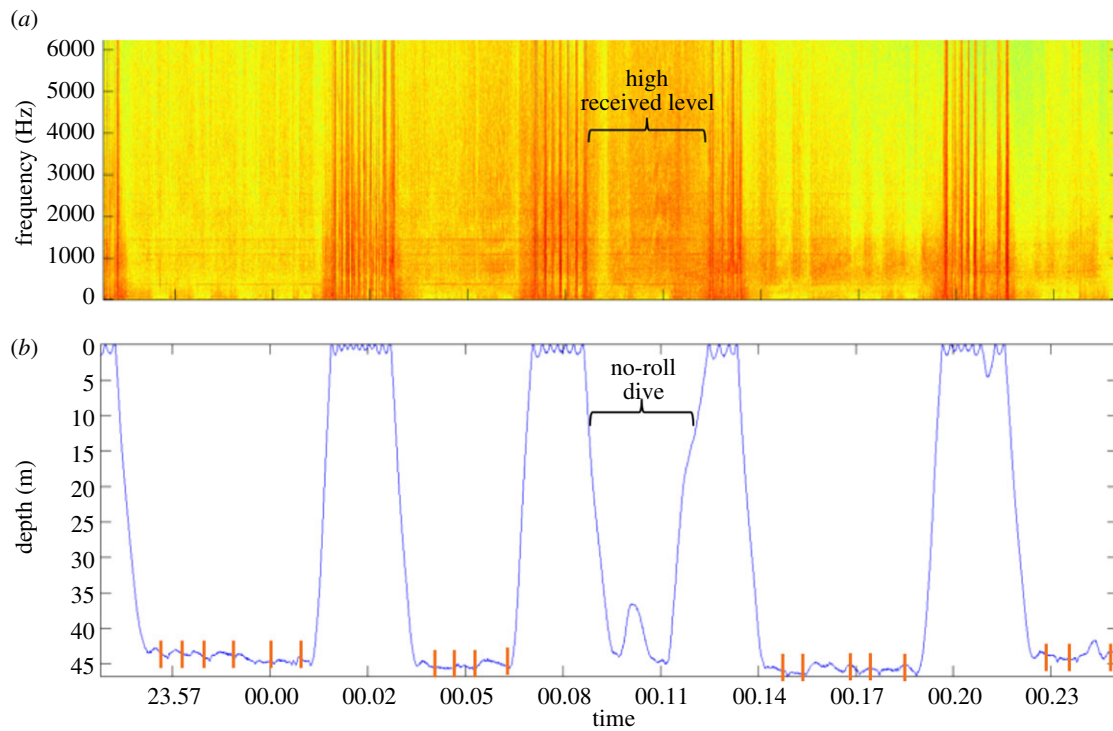
### (b) Statistics

To test whether the ship noise altered bottom foraging behaviours, we performed linear mixed-effects models using the seven dive metrics as dependent variables. Dependent variable data were square root transformed to approximate normality. The two fixed effects included the RL and the presence/absence of ship noise in each dive (SN), with tag deployment as a random effect. Best model fit was evaluated using Akaike's Information Criterion corrected for small sample sizes (AICc) [26]. Variable importance values were calculated by summing Akaike weights ( $w_i$ ) of all models including a particular variable (electronic supplementary material) [26]. All statistical analyses were performed in R (v. 2.15.3).

## 3. Results

The presence of ship noise significantly affected three of the seven dependent variable metrics tested (table 1; electronic supplementary material, table S2). In the best-fit models, as RL increased, the number of rolls decreased by 29% ( $t = -2.12$ , d.f. = 207,  $p$ -value = 0.04) (electronic supplementary material, figure S1a), and the descent rate decreased by 14.5% ( $t = -4.17$ , d.f. = 207,  $p$ -value < 0.01) (electronic supplementary material, figure S1b). Ascent rate also decreased by 12.8% as RL increased ( $t = -2.40$ , d.f. = 206,  $p$ -value = 0.02); however, ascent rate was faster by  $0.002 \text{ m s}^{-1}$  during SN exposure periods than in periods of no ship noise ( $t = 2.07$ , d.f. = 206,  $p$ -value = 0.04). The interaction of RL and SN was positive but non-significant ( $t = 0.28$ , d.f. = 205,  $p$ -value = 0.78). RL was the most important variable influencing all three response variables (electronic supplementary material, table S3).

In five out of nine individuals, one or more dives without bottom side-rolls occurred in the presence of ship noise. All of these responding individuals were adult females: two with their dependent calf, one pregnant and two who were neither pregnant nor lactating. The individual with two deployments showed this response in 2009, but not in 2006. These dives lacked bottom-feeding side-rolls despite a maximum depth near that of usual feeding dives immediately before and after, though some no-roll dives were shallower than the



**Figure 2.** (a) Spectrogram and (b) depth profile of whale mn08\_182a demonstrating atypical dive in the presence of ship noise. Orange tick marks indicate bottom side-rolls.

**Table 1.** Model coefficients, standard errors, *t*-values and *p*-values for best-fit models for descent rate, ascent rate and number of rolls.

response	model	variables	estimate	s.e.	<i>t</i> -value	<i>p</i> -value
descent rate	RL	intercept	1.558	0.127	12.271	<0.01
		RL	−0.006	0.001	−4.171	<0.01
	RL, SN	intercept	1.624	0.148	10.968	<0.01
		RL	−0.007	0.002	−3.395	<0.01
		SN	0.013	0.019	0.858	0.392
ascent rate	RL, SN	intercept	1.474	0.186	7.910	<0.01
		RL	−0.005	0.002	−2.400	0.017
		SN	0.040	0.019	2.069	0.040
	RL × SN	intercept	1.528	0.267	5.723	<0.01
		RL	−0.006	0.003	−1.847	0.066
		SN	−0.053	0.328	−0.162	0.872
		RL × SN	0.001	0.004	0.283	0.778
number of rolls	RL	intercept	3.413	0.752	4.541	<0.01
		RL	−0.018	0.009	−2.119	0.035
	RL, SN	intercept	3.984	0.879	4.531	<0.01
		RL	0.025	0.010	−2.457	0.015
		SN	0.114	0.094	1.208	0.228

surrounding dives (figure 2). Whales did not demonstrate bottom side-rolls in 11 out of 218 dives. Of the 11 no-roll dives, seven occurred in ship noise exposure periods (7/83) and only four without ship noise (4/135). A McNemar's test with continuity correction indicates that the percentage of dives with no rolls significantly differed in periods of ship noise exposure versus periods of no ship noise (McNemar's  $\chi^2(1, N = 218) = 109.63$ ,  $p$ -value < 0.01). Five of 18 ship passages resulted in a no-roll

dive. Whales did not compensate by increasing side-rolls following ship passage (Wilcoxon test,  $p$ -value > 0.05).

## 4. Discussion

While numerous studies have demonstrated modifications of acoustic communication in cetaceans exposed to noise, few

have assessed changes in baleen whale foraging behaviour. We show that humpback whales decrease the number of bottom-feeding events per dive and reduce feeding dive descent rate as the intensity of ship noise increases, indicating that ship noise can impact foraging rates and efficiency. Ship passages were also correlated with dives without bottom side-rolls, which implies either a cessation of feeding or a switch from bottom side-roll feeding to another method. Our results provide some of the first evidence to show statistically significant alterations in baleen whale foraging behaviour from ship noise exposure.

There are several potential explanations for the observed results. Whales may modify their diving behaviour in response to a perceived threat from ship noise, given that they require surface access to breathe. Ship noise could also affect prey behaviour; in the Gulf of Maine, sand lance seek refuge into sandy seafloor substrate in response to disturbance [24]. If sand lance retreat into substrate in increased ship noise, this could affect the prey availability for foraging whales. Further, if whales coordinate bottom feeding using paired burst vocalizations, which are within the same low-frequency band as ship noise [27], ship noise could cause masking that further reduces foraging efficiency.

Given the adaptability of humpback whales [23,24], we expect the Gulf of Maine population to potentially show habituation to human disturbance from ship noise, as they have been regularly exposed to commercial and whale watching vessels for decades [28]. Therefore, it is especially interesting that alterations to foraging behaviours were detectable in this study, as it suggests that humpbacks are unable to completely adjust to this disturbance. Short and potentially chronic cessations of feeding can result in biologically relevant decreases in balaenopterid foraging efficiency [17], which could manifest to decrease fitness. These behavioural changes were also observed at night when there are fewer ship interactions compared with the daytime hours

based on acoustic records; therefore, our results likely reflect the lower limit of disturbance of ship noise on foraging behaviour over the course of 24 h. Humpback whales forage during both day and night, albeit with different strategies dependent on prey behaviour, and each time period is likely critical to help satisfy their large energetic demands [17,24]. Further research on the impacts of noise on daytime foraging activities and variation in the sensitivity to different age and sex classes is needed, as mother–calf pairs are often more sensitive to disturbance [3,10]. Yet, these results are among the first support that ship noise can impact humpback whales' foraging, making this source of disturbance a management concern. Chronic impacts of even small reductions in foraging efficiency could affect individual fitness and translate to population-level effects on humpback whales exposed to ship noise in critical foraging areas.

**Ethics.** All animal use was conducted in accordance with institutional guidelines under IACUC approval at Duke University, the Pennsylvania State University and Syracuse University. Permission was obtained through federal research no. 775-185 (Northeast Fisheries Science Center) and 605-1904 (Whale Center of New England) issued by the United States National Marine Fisheries Service.

**Data accessibility.** We have deposited the data analysed in this study with the external repository Dryad: <http://dx.doi.org/10.5061/dryad.18637> [29].

**Authors' contributions.** H.B.B. and S.E.P. designed the study; A.S.F., D.N.W. and S.E.P. collected the data; and all authors participated in the data analysis and manuscript preparation. All authors approved the final version of the manuscript and agree to be held accountable for the content therein.

**Competing interests.** We have no competing interests.

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## References

- Slabbekoorn H, Bouton N, van Opzeeland I, Coers A, ten Cate C, Popper AN. 2010 A noisy spring: the impact of globally rising underwater sound levels on fish. *Trends Ecol. Evol.* **25**, 419–427. (doi:10.1016/j.tree.2010.04.005)
- Francis CD, Barber JR. 2013 A framework for understanding noise impacts on wildlife: an urgent conservation priority. *Front. Ecol. Environ.* **11**, 305–313. (doi:10.1890/120183)
- Dunlop RA, Cato DH, Noad MJ. 2010 Your attention please: increasing ambient noise levels elicits a change in communication behaviour in humpback whales (*Megaptera novaeangliae*). *Proc. R. Soc. B* **277**, 2521–2529. (doi:10.1098/rspb.2009.2319)
- McClure CJW, Ware HE, Carlisle J, Kaltenecker G, Barber JR. 2013 An experimental investigation into the effects of traffic noise on distributions of birds: avoiding the phantom road. *Proc. R. Soc. B* **280**, 20132290. (doi:10.1098/rspb.2013.2290)
- Rolland RM, Parks SE, Hunt KE, Castellote M, Corkeron PJ, Nowacek DP, Wasser SK, Kraus SD. 2012 Evidence that ship noise increases stress in right whales. *Proc. R. Soc. B* **279**, 2363–2368. (doi:10.1098/rspb.2011.2429)
- Tennessen JB, Parks SE, Langkilde T. 2014 Traffic noise causes physiological stress and impairs breeding migration behaviour in frogs. *Conserv. Physiol.* **2**, 1–8. (doi:10.1093/conphys/cou032)
- Siemers BM, Schaub A. 2011 Hunting at the highway: traffic noise reduces foraging efficiency in acoustic predators. *Proc. R. Soc. B* **278**, 1646–1652. (doi:10.1098/rspb.2010.2262)
- Isojunno S, Curé C, Kvasdheim PH, Lam FPA, Tyack PL, Wensveen PJ, Miller PJOM. 2016 Sperm whales reduce foraging effort during exposure to 1–2 kHz sonar and killer whale sounds. *Ecol. Appl.* **26**, 77–93. (doi:10.1890/15-0040)
- Chan AAY-H, Giraldo-Perez P, Smith S, Blumstein DT. 2010 Anthropogenic noise affects risk assessment and attention: the distracted prey hypothesis. *Biol. Lett.* **6**, 458–461. (doi:10.1098/rsbl.2009.1081)
- Nowacek DP, Thorne LH, Johnston DW, Tyack PL. 2007 Responses of cetaceans to anthropogenic noise. *Mamm. Rev.* **37**, 81–115. (doi:10.1111/j.1365-2907.2007.00104.x)
- Miller PJO, Johnson MP, Madsen PT, Biassoni N, Quero M, Tyack PL. 2009 Using at-sea experiments to study the effects of airguns on the foraging behavior of sperm whales in the Gulf of Mexico. *Deep. Res. Part I Oceanogr. Res. Pap.* **56**, 1168–1181. (doi:10.1016/j.dsr.2009.02.008)
- Williams R, Lusseau D, Hammond PS. 2006 Estimating relative energetic costs of human disturbance to killer whales (*Orcinus orca*). *Biol. Conserv.* **133**, 301–311. (doi:10.1016/j.biocon.2006.06.010)
- Laist DW, Knowlton AR, Mead JG, Collet AS, Podestà M. 2001 Collisions between ships and whales. *Mar. Mammal Sci.* **17**, 35–75. (doi:10.1111/j.1748-7692.2001.tb00980.x)
- Croll DA, Clark CW, Calambokidis J, Ellison WT, Tershy BR. 2001 Effect of anthropogenic low-frequency noise on the foraging ecology of *Balaenoptera* whales. *Anim. Conserv.* **4**, 13–27. (doi:10.1017/S1367943001001020)
- Todd S, Stevick P, Lien J, Marques F, Ketten D. 1996 Behavioural effects of exposure to underwater

- explosions in humpback whales (*Megaptera novaeangliae*). *Can. J. Zool.* **74**, 1661–1672.
16. Johnson MP, Tyack PL. 2003 A digital acoustic recording tag for measuring the response of wild marine mammals to sound. *IEEE J. Ocean. Eng.* **28**, 3–12. (doi:10.1109/JOE.2002.808212)
  17. Goldbogen JA *et al.* 2013 Blue whales respond to simulated mid-frequency military sonar. *Proc. R. Soc. B* **280**, 20130657. (doi:10.1098/rspb.2013.0657)
  18. Sivle LD *et al.* 2015 Severity of expert-identified behavioural responses of humpback whale, minke whale, and northern bottlenose whale to naval sonar. *Aquat. Mamm.* **41**, 469–502. (doi:10.1578/AM.41.4.2015.469)
  19. McKenna MF. 2011 Blue whale response to underwater noise from commercial ships. UC San Diego: Oceanography, b7068914. See <https://escholarship.org/uc/item/4rv0q1mv>.
  20. Jurasz CM, Jurasz VP. 1979 Feeding modes of the humpback whale, *Megaptera novaeangliae*, in southeast Alaska. *Sci. Rep. Whales Res. Inst.* **31**, 69–83.
  21. Ware C *et al.* 2014 Bottom side-roll feeding by humpback whales (*Megaptera novaeangliae*) in the southern Gulf of Maine, U.S.A. *Mar. Mammal Sci.* **30**, 494–511. (doi:10.1111/mms.12053)
  22. Miller PJ, Biassoni N, Samuels A, Tyack PL. 2000 Whale songs lengthen in response to sonar. *Nature* **405**, 903. (doi:10.1038/35016148)
  23. Dunlop RA. 2016 The effect of vessel noise on humpback whale, *Megaptera novaeangliae*, communication behaviour. *Anim. Behav.* **111**, 13–21. (doi:10.1016/j.anbehav.2015.10.002)
  24. Friedlaender AS, Hazen EL, Nowacek DP, Halpin PN, Ware C, Weinrich MT, Hurst T, Wiley D. 2009 Diel changes in humpback whale *Megaptera novaeangliae* feeding behavior in response to sand lance *Ammodytes* spp. behavior and distribution. *Mar. Ecol. Prog. Ser.* **395**, 91–100. (doi:10.3354/meps08003)
  25. Ware C, Arsenault R, Plumlee M, Wiley D. 2006 Visualizing the underwater behavior of humpback whales. *IEEE Comput. Graph. Appl.* **26**, 14–18. (doi:10.1109/MCG.2006.93)
  26. Burnham KP, Anderson DR. 2002 *Model selection and multimodel inference: a practical information-theoretic approach*. New York, NY: Springer-Verlag.
  27. Parks SE, Cusano DA, Stimpert AK, Weinrich MT, Friedlaender AS, Wiley DN. 2014 Evidence for acoustic communication among bottom foraging humpback whales. *Sci. Rep.* **4**, 7508. (doi:10.1038/srep07508)
  28. Hatch L, Clark C, Merrick R, Van Parijs S, Ponirakis D, Schwehr K, Thompson M, Wiley D. 2008 Characterizing the relative contributions of large vessels to total ocean noise fields: a case study using the Gerry E. Studds Stellwagen Bank National Marine Sanctuary. *Environ. Manage.* **42**, 735–752. (doi:10.1007/s00267-008-9169-4)
  29. Blair H, Merchant N, Friedlaender A, Wiley DN, Parks SE. 2016 Data from: Evidence for ship noise impacts on humpback whale foraging behaviour. Dryad Digital Repository: <http://dx.doi.org/10.5061/dryad.18637>.