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Maternal cadmium, placental PCDHAC1, and fetal development

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Abstract

Cadmium (Cd) is a ubiquitous environmental contaminant implicated as a developmental toxicant, yet the underlying mechanisms that confer this toxicity are unknown. Mother-infant pairs from a Rhode Island birth cohort were investigated for the potential effects of maternal Cd exposure on fetal growth, and the possible role of the *PCDHAC1* gene on this association. Mothers with higher toenail Cd concentrations were at increased odds of giving birth to an infant that was small for gestational age or with a decreased head circumference. These associations were strongest amongst those with low levels of DNA methylation in the promoter region of placental *PCDHAC1* expression to be inversely associated with maternal Cd, and *PCDHAC1* expression positively associated with fetal growth. Our findings suggest that maternal Cd affects fetal growth even at very low concentrations, and some of these effects may be due to the differential expression of *PCDHAC1*.

Keywords

maternal cadmium; fetal growth; small for gestational age; placenta; methylation; expression; *PCDH*; *PCDHAC1*

1. Introduction

Cadmium (Cd) is a ubiquitous environmental pollutant and well recognized public health hazard [1]. Cigarette smoking and certain occupations can result in high Cd exposure [2], though most people are exposed chronically to low levels of Cd via diet, primarily from

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cereal grains, vegetables, potatoes, and products made from these foods [1,3]. Cd has been long established as a toxicant to the kidney[4] and extremely high Cd intake was the cause of Itai-itai disease, a severe condition characterized by bone and renal injuries [5]. Currently Cd has a dietary intake recommendation of no more than 25 μ g/kg body weight per month (corresponding to 5.8 μ g/kg body weight per week), which protects against these extreme health outcomes; though some recommend that the weekly intake be reduced further to 2.5 μ g/kg [3]. It is also classified as a group I human carcinogen by the International Agency for Research on Cancer [6] and more recently has been implicated as a developmental toxicant and possible teratogen [2,7].

Animal studies have shown that Cd exposure during pregnancy decreases neonatal birthweight [8], decreases trophoblast proliferation in the placenta [9], and induces preeclamptic conditions in the placenta [10]. These models have suggested that some of the developmentally toxic effects of cadmium may be related to increased glucocorticoid concentrations in placentae and plasma [8], repression of placental lactogens [9], oxidative damage to placental tissue [10], or inhibition of essential metals transfer from mother to fetus [11,12]. Multiple epidemiologic studies have also related maternal Cd exposures to restricted fetal development. To date, human maternal Cd has been associated with infants born small for gestational age (SGA) [13] or a lower birth weight [14–16], decreased head circumference [14,17], and overall growth in the first 3 years of life [17]. There are also indications that maternal Cd, in combination with other toxic metals, may result in mental [18] and possibly psychomotor impairment [19]. Although evidence of Cd as a developmental toxicant is mounting, the mechanisms through which it may exert these effect are still being uncovered. Thus, we hypothesized that Cd exposure may be working through the alteration of placental function to impact fetal growth.

Recent work has shown that low level Cd exposures induce or repress the expression of a wide array of genes [20,21], while others have suggested that Cd affects DNA methylation (DNA-M) [22–24], an epigenetic mechanism that regulates gene transcription, at the global and gene-specific levels [25]. Human, animal, and in vitro studies have illustrated that Cd may affect the functions of multiple epigenetic mechanisms [25]. Also, Cd may be responsible for the mis-expression of multiple cellular adhesion molecules (such as cadherins and catenins), whose activities are integral to normal embryo implantation as well as in post-implantation cellular signaling, remodeling and migration [2]. Cadherins are key regulators of trophoblast behavior and organization; altered expression and/or function of some cadherins has been suggested to play a role in fetal growth restriction and preeclampsia [26]. An *in vitro* study of human bronchial epithelial cells found that Cd altered the expression of many cellular adhesion and junction molecules, likely disrupting tight junction integrity [27]. Similarly, murine studies have shown Cd to affect the abundance of, and interactions between cadherin molecules [28,29]. Protocadherins (PCDH) comprise the largest subfamily of the cadherin cell-adhesion molecules, some of which are encoded very close in genomic proximity forming PCDH gene-clusters [30]. These PCDH clusters can produce multiple unique transcripts that could impart various functionalities and distinct cellular identities [31]. The PCDH genes also appear to be susceptible to epigenetic modifications in response to toxic heavy metal exposures, including Arsenic (As) and Cd [32], and the transcription of *PCDH-a* isoforms are regulated by promoter and 1^{st} exon

DNA-M [31,33]. The *PCDH-a* genes are predominantly expressed in the central nervous system (CNS) and likely play important roles in the innervation of serotonergic projections [34].

Given the importance of cellular-adhesion molecules in fetal development, as well as the apparent responsiveness of *PCDH*-genes to heavy metal exposures, altered *PCDH-a* gene activities could play a role in maternal Cd-restricted fetal development. We hypothesized that maternal Cd could influence the expression of placental *PCDHAC1*, which in turn may be associated with restricted fetal growth. Therefore, this study aimed to characterize the relationship between maternal toenail Cd concentrations and placental *PCDHAC1* expression, and to investigate whether the expression of *PCDHAC1* was related to various markers of fetal development.

2. Methods

2.1 The Rhode Island Child Health Study

The Rhode Island Child Health Study (RICHS) is a birth cohort that included nonpathologic term pregnancies at born at 37 weeks gestation at the Women and Infants' Hospital in Providence, RI, USA with enrollment for this analysis from September 2010 through February 2013. Infants classified as large for gestational age (LGA) (90th BW percentile) and SGA (10th BW percentile) were oversampled for inclusion and adequate for gestational age (AGA) (between the 10th and 90th BW percentiles) infants that matched on sex, gestational age (± 3 days), and maternal age (± 2 years) were enrolled coincidentally. Mothers younger than 18 years of age, with life threatening conditions, or infants with congenital/chromosomal abnormalities were excluded. All protocols were approved by the institutional review boards at the Women and Infants Hospital and Dartmouth College and all participants provided written informed consent. An interviewer administered questionnaire was used to collect self-reported sociodemographic, lifestyle, and medical history data, and a structured medical records review was employed to collect anthropometric and clinical data. Birthweight (BW) percentiles were calculated while accounting for gestational age, infant sex, birthweight, head circumference, and length [35]. The samples used in this study included those mother-infant pairs for which maternal and newborn toenails had been collected for metals analyses. From this metals-sample (n=242) a sub-sample of those with the highest and lowest maternal Cd concentrations were selected for gene-expression assays (n=95).

2.2 Cadmium Measurements

Toenail clippings from all toes from both mothers and newborns were requested following discharge, and mailed back to the study office. Average time from birth to collection was 2.8 months and ranged from 0.3 to 7.1 months. From these samples, μ g of Cd per gram of toenail were measured at the Dartmouth Trace Element Analysis Core following procedures, involving HNO₃:HCl acid digestion of the nails followed by ICP-MS (Agilent 7700×, Santa Clara, CA) analysis of the resulting digests. The ICP-MS method and quality control (QC) followed that outlined in EPA 6020A and QC involved initial and continuing calibration verification and blanks, digestion blanks, fortified blanks and hair powder certified reference

material (NIES # 13 and GBW07601, certified at 0.23 and 0.11 µg/g Cd respectively), analysis duplicates and spikes. Average Cd percent recovery for the reference materials across the digestion batches was 102 ± 11 % (n= 22). Detection limits were determined for each digestion batch by either the instrument detection limit or the method detection limit (average of digestion blanks + 3σ), whichever was the higher value. Samples that were below the limit of detection (LOD) were assigned a value equal to half the lowest LOD observed for that batch.

2.3 Placenta Sampling

Full-thickness sections of placenta were taken from the fetal side of the placenta, 2 cm from the umbilical cord insertion site and free of maternal decidua, from each of the four quadrants around the cord insertion within two hours of birth. These samples were immediately placed in RNAlater[™] (Applied Biosystems, Inc., AM7020). Following 72 hours at 4°C, samples were blotted dry, snap-frozen in liquid nitrogen. The four biopsied samples were then pulverized to homogenize across the samples then stored at −80°C until analysis.

2.4 Nanostring Gene Expression Assay

The nCounter Panel-Plus custom gene expression panel (NanoString Technologies, Seattle, WA) was used to quantitate gene expression and was performed according to manufacturer's instructions. Total RNA was isolated from placenta with 500 nanograms of RNA used per sample in this assay. Briefly, two sets of oligonucleotide probes were designed to target a specific gene sequence, along with a reporter probe and a capture probe this platform multiplexes the detection of 24 genes (3 of which were placenta housekeeping genes) per sample in a 96-sample format; only one of these 21 candidate genes (*PCDHAC1*) was the involved in this study. A table showing variation in relative expression by maternal Cd for all 21 genes is included in the supplemental materials (Supplemental Table 1).

Following an overnight hybridization (67°C), samples were vertically pooled, and placed into a cartridge for further processing. The cartridge was placed on the automated nCounter sample prep station where excess reporter and capture probes were washed away and probe/ target complexes were aligned and immobilized. The cartridge was then transferred to the nCounter Digital Analyzer for data collection. This yielded raw transcript counts for the candidate genes, housekeeping genes, and QC probes. The lane-specific mean of the negative control probes was utilized to estimate the background signal (Supplemental Table 2); only one sample was assigned a value for *PCDHAC1* expression that was indistinguishable from background.

One sample was a technical replicate, used to evaluate the reliability of the counts (Supplemental Figure 1), which was excluded from subsequent analyses and another sample was excluded due to extremely low transcript counts across all genes, yielding 94 samples for analyses. We then applied positive-control normalization to each lane based on the sum of its pooled positives relative to the sum of the pooled positives in all the other lanes. Then the data were normalized on the geometric mean of the three housekeeping genes (*SDHA*, *TBP* and *YWHAZ*), which have been shown to be stable housekeeping genes in placental

tissue [36]. These housekeeping also genes exhibited no differential expression by maternal Cd concentrations in our sample (T-test p-values > 0.05; Supplemental Table 3).

2.5 Placenta Illumina 450 DNA Methylation Array

Genome-wide DNA-M was measured at the University of Minnesota Genomics Center via Illumina Infinium Human Methylation450K BeadArray (Illumina) and the full QA/QC and analytical procedures described elsewhere [37]. Briefly, samples were randomized across multiple batches, stratified by birthweight group and gender. Arrays were processed and normalized using standard methods; poorly-detected probes, probes measuring DNA-M at X- and Y-linked loci, and SNP-associated loci were excluded [38], and data were standardized across batches to remove technical variations [39]. DNA-M data were analyzed as β -values, which can be interpreted as the proportion of methylated alleles for that individual CpG site. DNA-M array data for the RICHS placenta are available via the NCBI Gene Expression Omnibus (GEO) accession number GSE75248; we provide a supplemental file herein which includes log-Cd, log-*PCDHAC1*, and a variable for cross-linking these data (Supplemental Data).

2.5 Statistical Analyses

Due to the skewed distributions of both Cd and PCDHAC1 expression, both were logtransformed for all continuous analyses. Also, because the study design for the sub-sample selected mother-infant pairs with the highest and lowest measures of toenail Cd concentrations, we ran all regression models with continuous measures of log-Cd and with dichotomous Cd (High vs. Low based on median split) for analyses of the sub-sample. We compared the distributions of all independent and dependent variables, as well as potential confounders between the sub-sample (n=94) and the metals-sample (n=242) using Student's t-test for continuous measures and Chi-squared tests for categorical measures. Fisher's Exact tests were used to compare categorical measures with low frequencies. Kruskal-Wallis rank sum tests were used to compare continuous measures across categorical groups for data that did not approximate a normal distribution. For all exposure-outcome regression analyses, models were fit using robust linear models for continuous outcomes and robust generalized linear models for categorical outcomes via the robustbase package in R [40], to limit the effects of outliers. Structural equation modeling was used to evaluate the direct, indirect and total effects via mediation analyses using the lavaan package in R [41]. For all analyses, determinations of statistical significance were set at a level of 0.05. In our adjusted models we included maternal BMI and maternal age, which have been shown to be associated with placental Cd concentrations [42], as well as infant sex and gestational age which are strong determinants of birth weight and size.

3. Results

3.1 Comparing the sub-sample (n=94) to the metals-sample (n=242)

The expression sub-sample (n=94) drawn for this study was representative of the metalssample (n=242) in terms of maternal smoking during pregnancy, maternal BMI, maternal age, size for gestational age, infant sex, gestational time, and infant birthweight percentiles (Table 1). In the metals-sample (n=242), Cd concentrations ranged from 0.00062 μ g/g to

 $0.0846 \ \mu g/g$, with a right skewed distribution. Within the sub-sample, Cd concentrations among mothers with 'low' Cd (n=47) ranged from $0.00062 \ \mu g/g$ to $0.0029 \ \mu g/g$, with a median of $0.00104 \ \mu g/g$. Those with 'high' Cd (n=47) had concentrations ranging from $0.0088 \ \mu g/g$ to $0.0846 \ \mu g/g$, with a median of $0.0159 \ \mu g/g$. Maternal toenail Cd concentrations were modestly correlated with offspring toenail Cd concentrations (Kendall's tau = 0.30, p-value < 0.001).

3.2 Associations between maternal cadmium and fetal growth

All results herein with p-values < 0.05 were determined to be statistically significant. We tested whether maternal Cd was associated with indicators of fetal growth in the metals-sample (n=242) using robust linear regressions. Newborns that were smaller in length (p = 0.31), had smaller head circumference (p = 0.08), lower BW (p = 0.13), and lower BW percentiles (p = 0.28) tended to have mothers with higher log-Cd, though these associations were not statistically significant. However, maternal log-Cd was significantly associated with size for gestational age (p-value = 0.002), with higher log-Cd among SGA newborns than AGA (p-value = 0.001) or LGA (p-value = 0.003) newborns. In contrast, there was no difference in log-Cd between AGA and LGA (p-value = 0.96) newborns (Figure 1), so we grouped AGA and LGA together and used logistic regression to estimate the odds of SGA while adjusting for confounders. We found that increasing maternal log-Cd significantly increased the odds of SGA (OR = 2.44, 95% CI = 1.53 - 3.89), adjusted for infant sex, maternal BMI, maternal age, and gestational age. Of note, in this multivariable model, maternal log-Cd and maternal age (p-value = 0.037) were the only significant predictors of SGA.

We then tested whether we could reproduce the associations with SGA in the sub-sample (n=94) for which we had obtained *PCDHAC1* expression. The associations between maternal Cd and SGA were robust (Figure 1), with increasing maternal log-Cd associated with increased odds of SGA (OR = 2.12, 95% CI = 1.24 - 3.64), adjusted for infant sex, maternal BMI, maternal age, and gestational age.

3.3 PCDHAC1 Expression analyses

PCDHAC1 expression was below the detection limit for one sample, ranged from 5.8 to 114.1 transcript counts, exhibited a right-skewed distribution with a median of 15.3, and appeared to be more highly expressed among placenta of mothers with lower Cd exposure (Supplemental Figure 2). We found that log-*PCDHAC1* expression was inversely associated with log-Cd ($\beta_{log-Cd} = -0.064$; p-values = 0.0087) and positively associated with maternal BMI (β MatBMI = 0.015; p-values = 0.0032) in a multivariable linear regression model including both terms. *PCDHAC1* expression was not associated with any of the other covariates (infant sex, gestational age, or maternal age) in this study.

We then investigated whether *PCDHAC1* expression was associated with indicators of fetal development: infant head circumference, length, BW, BW percentiles, and odds of SGA. Log-*PCDHAC1* expression was not a significant predictor of SGA (p-value=0.164); however, when grouped into tertiles, those within the highest tertile of *PCDHAC1* expression had lower odds of SGA, compared to the lowest tertile (OR = 0.13, 95% CI =

0.018 - 0.89), adjusted for infant sex, maternal BMI, maternal age, and gestational time. Log-*PCDHAC1* and tertiles of *PCDHAC1* had significant linear associations with increased

By $(\beta_{log-PCDHAC1} = 250.01, p-value = 0.049; \beta_{HvL} = 358.58, p-value = 0.038)$ and BW percentiles $(\beta_{log-PCDHAC1} = 15.70, p-value = 0.026; \beta_{HvL} = 22.23, p-value = 0.026)$, while adjusting for infant sex, gestational time, maternal age, and maternal BMI. We also found that log-*PCDHAC1* and tertiles of *PCDHAC1* were associated with increased head circumference ($\beta_{log-PCDHAC1} = 0.623$; p-values = 0.016; $\beta_{MvL} = 0.540$, p-value = 0.0084; $\beta_{HvL} = 0.682$, p-value = 0.0017), while adjusting for BW, infant sex, gestational time, maternal age, and maternal BMI. Neither log-*PCDHAC1* nor tertiles of *PCDHAC1* were associated with infant length.

3.4 Summarizing DNA-M across PCDHAC1

We had hypothesized that the association between Cd and *PCDHAC1* expression may depend on DNA-M levels within *PCDHAC1* regulatory regions, for which data were available at 37 CpG sites. We calculated the average DNA-M across all CpGs within 1500 bps of the transcription start site (TSS1500; 3 sites), 200 bps of the transcription start site (TSS200; 6 sites), the 1st exon (5 sites), the gene-body (22 sites), and the 3'-untranslated region (3'-UTR; 1 site). We then evaluated the representativeness of these regional averages with a pairwise correlation matrix of all 37 sites, as well as the correlations between individual sites and the average of the regulatory regions in which they reside (Figure 2). The CpGs within the TSS200 were highly correlated with each other, while those sites within the TSS1500 and TSS200 were highly correlated their regional averages: correlation coefficients ranged from 0.69 to 0.79 and 0.81 to 0.91, respectively, all with p-values < 0.001. Whereas CpGs within the 1st exon and gene body did not have strong correlation patterns and were not as well represented by their regional averages.

3.5 DNA-M dependent associations

To test whether the effect of log-Cd on log-*PCDHAC1* expression was dependent on regional DNA-M, we produced five linear regression models while including an interaction term between log-Cd and DNA-M at the TSS1500, the TSS200, the 1st exon, the gene body, and the 3'UTR, and adjusting for BMI (Table 2). The association with log-Cd appeared to be dependent on TSS1500 DNA-M (interaction p-value = 0.053) and TSS200 DNA-M (interaction p-value = 0.149). Thus we ran regression models, stratified by those with 'high' and 'low' DNA-M (median split), within the TSS1500 and the TSS200. In the stratified models, log-Cd was inversely associated with log-*PCDHAC1* expression, but only among those with lower DNA-M within the TSS1500 and TSS200 (Table 3). We observed positive associations between maternal BMI and *PCDHAC1* expression, but this association was independent of DNA-M strata.

Then we tested whether the associations of maternal Cd with odds of SGA and decreased head circumference were dependent on *PCDHAC1* DNA-M strata. We found that higher log-Cd was significantly associated with increased odds of SGA and decreased head circumference among the low DNA-M strata, but not in the high DNA-M strata (Table 4). Interestingly, the associations of maternal Cd with SGA became stronger after adjustment for potential confounders, and exemplified the potential dependence on DNA-M. Among

those with lower DNA-M within the TSS200 or TSS1500, a one unit increase in log-Cd was associated with 5.64 or 3.00 times increased odds of SGA, respectively. In the high DNA-M strata these odds ratios were much smaller, 1.28 and 1.87 for the TSS200 and TSS1500 respectively, and not statistically significant. Similarly, among those with lower DNA-M within the TSS200 or TSS1500, a one unit increase in log-Cd was associated with a 0.143cm or 0.191cm decrease in head circumference. These associations were substantially attenuated, decreases of 0.052cm or 0.038cm, among those with higher DNA-M at TSS200 and TSS1500, and were not statistically significant.

3.6 Dichotomous Cd as the Exposure Variable

Because the sub-sample was selected based on mothers with the highest or lowest toenail Cd concentrations, we investigated whether the above associations were reproducible using dichotomous maternal Cd (median split) as the primary exposure in place of continuous maternal log-Cd. Mothers with the highest Cd levels were more likely to give birth to SGA offspring than mothers with the lowest Cd levels (OR = 4.70, 95% CI = 1.23 - 18.05). Similarly, high maternal Cd was associated with lower expression of *PCDHAC1* (β = -0.16; p-values = 0.036). We were also able to reproduce the *PCDHAC1* DNA-M dependent effects for almost all adjusted regression models (Supplemental Materials Table 4), only the association between dichotomized Cd and head circumference was no longer statistically significant within the low DNA-M strata of TSS200 ($\beta_1 = -0.353$, p-value = 0.093).

3.7 Structural Equation Modeling (SEM)

If these associations represent causal relationships, which we cannot verify given the observational nature of these data, we wanted to explore whether some of the effect of maternal Cd on increased odds of SGA was mediated through the repression of *PCDHAC1* expression using SEM. For these models we used dichotomized Cd (High vs. Low) as the primary exposure, lowest vs mid-and-highest *PCDHAC1* tertiles as the mediator, and SGA vs. non-SGA as the outcome (Figure 3). The regression for the effects of Cd on *PCDAHC1* was adjusted for maternal BMI, while the regressions for effects on SGA were also adjusted for maternal age, gestational time and infant sex. Only 10% of the effect of Cd on SGA appeared to work through the repression of *PCDHAC1* as a mediator, and this indirect effect was not statistically significant. Whereas the total effect p-value = 0.020; Direct Effect p-value = 0.032) in the SEM. We did attempt to fit the above SEMs stratified by DNA-M in the promoter of *PCDHAC1*; although we again observed stronger total effects of maternal Cd on SGA among those with lower DNA-M, none of the stratified direct and indirect effects were statistically significant (all p-values > 0.05).

4. Discussion

This study adds to the increasing body of evidence that maternal Cd exposures can influence fetal development and suggests a possible role for *PCDH* genes in fetal growth. Currently, *PCDH* genes are primarily known for their functions in neurons and have been implicated in neurological disorders [43]. Five SNPs within the *PCDH-a* cluster have been associated with autism [44] and microdeletions at chromosomal region 5q31.3, where the *PCDH* gene

clusters are located, cause developmental delay and other cerebral abnormalities [45]. Because PCDH-a gene activity has not been well-characterized in placental tissue, we explored two publicly available placenta gene-expression datasets from the Gene Expression Omnibus (GEO): GSE9984, which evaluated placental gene-expression at different time points throughout pregnancy [46] and GSE7434, which compared placental gene-expression between mothers that did and did not smoke [47]. Both of publically available datasets contain transformed counts produced via the Affymatrix Human Genome U133 Plus 2.0 Array. Many of the PCDH-a genes were expressed by the placenta at relatively consistent levels throughout pregnancy and at birth (Supplemental Table 5). Specifically, PCDHAC1 exhibited consistently mid-level expression in trimester 1, trimester 2, and at term, where mid-level was defined as being within the middle tertile of average expression across the entire array. Furthermore, in our sample the placental expression of *PCDHAC1* did not differ by gestational age (Kruskall-Wallis rank sum p-value = 0.36) at term. Placental expression of the PCDH-a genes also did not differ between mothers who smoked and mothers that did not smoke (Supplemental Table 6). We add further evidence that *PCDHAC1* is expressed by placenta collected at term, and that it tends to be expressed in relatively low levels. Due to differences in placental sampling and expression assay technologies, we could not directly compare and determine whether expression levels in our samples were similar to those in the publicly available datasets.

We are the first to suggest that maternal Cd may be inversely associated with *PCDHAC1* expression in the placenta, and that DNA-M patterns within *PCDHAC1* may impart an epigenetic susceptibility for the associations between Cd and fetal development. We observed that those with lower DNA-M in the promoter region of *PCDHAC1* may be more susceptible to Cd-associated repression of *PCDHAC1*, increased odds of SGA, and decreased head circumference. We are also the first to suggest that placental expression of *PCDHAC1* may be positively associated with higher birth weight percentiles and larger newborn head circumferences. These findings were robust after adjustment for the potential confounding effects of infant sex, gestational age, maternal age, maternal BMI, and other measures of fetal growth when appropriate.

Despite being the first to associate maternal Cd concentrations *in toenail* to SGA and smaller head circumference, our findings are consistent with other research showing restricted fetal growth with increasing levels of Cd in maternal urine [14], maternal blood [13,16,17], and placental tissue [15]. Toenails provide a unique source for biomonitoring, as they are simple and non-invasive to collect, and may be more representative of long-term chronic exposures to trace elements compared to other tissues [48]. In our sample, maternal and child toenail Cd concentrations were modestly but significantly correlated with each other. We considered the possibility that maternal toenail Cd may merely have acted as a marker of direct fetal exposure to Cd, and that the direct fetal exposure may instead be causing the observed effects on restricted growth. To address this, we ran a sensitivity analysis in which we remodeled our previous regressions for odds of SGA, BW percentiles, and head circumference but used newborn toenail log-Cd levels instead of maternal toenail log-Cd levels, as the primary exposure. None of these outcomes were significantly associated with newborn Cd in crude models or after adjustment for the same covariates that were used for the maternal-Cd models. Thus, the associations we observed were not driven by fetal toenail Cd, a potential

biomarker of fetal exposure to Cd *in utero*, but were consistently related to maternal exposure.

Other investigators have proposed that the effect of maternal or placental Cd on fetal growth could be due to impaired zinc-transfer to the fetus [49], or through endocrine disruption via the repression of 11 β -hydroxysteroid dehydrogenase type 2 gene (*HSD11B2*) in human trophoblasts [50], which is important for its role in protecting against high levels of maternal glucocorticoids which can inhibit fetal growth [51]. We present another possible mechanism, epigenetic susceptibility within and repression of *PCDHAC1*, through which maternal Cd may alter placental functions and restrict fetal growth. Due to the observational nature of this study, we could not experimentally verify whether Cd does in fact repress *PCDHAC1* and whether that repression is dependent on promoter DNA-M patterns. Thus these findings should be validated in an independent cohort or studied further with an experimental design.

These findings should be interpreted within the limitations of the study. The majority of analyses were conducted with a relatively small sample (n=94) in which only those with the highest and lowest maternal Cd concentrations were included. Thus there is a possibility that selection bias could have contributed to our findings. However, we did show that our subsample was representative of the metals-sample across all variables used in the analyses of the sub-sample. Also, one of the most interesting findings was that DNA-M patterns in the promoter region of placental PCDHAC1 may confer an epigenetic susceptibility to Cdrelated SGA and decreased head circumference. Although we only had PCDHAC1 expression data within our sub-sample, we did have a larger number of mother-infant pairs (n=176) for which placental DNA-M was available. Thus, as a sensitivity analysis, we reproduced the DNA-M dependent effects of log-Cd on head circumference and SGA using these data. In this larger sample, again, we observed stronger and statistically significant effects of Cd on odds of SGA and on head circumference in the low DNA-M strata compared to the high DNA-M strata (Supplemental Table 7). Thus it is unlikely that selection of our sub-sample biased our findings, and PCDHAC1 may indeed play a role in Cd-related effects on fetal growth.

Maternal smoking was considered a possible confounder of our observed associations between maternal-Cd and restricted growth, but due to very low frequency of smokers, we could not include this as an adjustment covariate in our robust regression analyses; only four (4%) and eighteen (8%) mothers smoked cigarettes during pregnancy in the sub-sample and metals-sample, respectively. Thus we re-ran our adjusted models for maternal log-Cd on the odds of SGA, with and without maternal smokers, and found that the parameter estimates and p-values were barely perturbed by excluding smokers. We therefore concluded that our observed effects of Cd on restricted fetal growth were independent of the effects of smoking.

Also, we did not observe statistically significant indirect effects of maternal log-Cd on SGA, mediated through repression of *PCDHAC1* using SEM. However, our analyses involved a small sample size, which was likely underpowered to identify a mediated effect, particularly within separate DNA-M strata. The consistent results across the individual regression models combined with the successful sensitivity analyses for the DNA-M dependent effects of Cd restricted fetal growth, suggest that *PCDHAC1* may still play some role in Cd-related

restricted fetal growth and warrant follow-up in other studies with larger samples. Also, since Cd is known to cause a number of effects on the placenta, including inhibition of trophoblast invasion, decreased hormone production, and altered nutrient metal transport [2], we caution that our results may not represent a direct effect, but a sign of this wider response of placental trophoblasts to Cd exposure. Further studies would be needed, particularly those with controlled timing of exposures and molecular endpoint, to dissect the true causal pathway.

5. Conclusions

Higher levels of maternal Cd exposure are associated with increased odds of giving birth to an infant that is SGA or that has decreased head circumference, and these associations appear to be strongest among those with lower levels of DNA-M in the promoter of placental *PCDHAC1*. Also, maternal Cd was inversely associated with placental *PCDHAC1*, which in turn may be associated with birth weight percentiles, odds of SGA, and head circumference of newborns. Further epidemiologic research is necessary to determine whether similar associations can be observed in different populations and experimental studies are necessary to infer causality.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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Abbreviations

AGA	Adequate for gestational age
bps	basepairs
Cd	Cadmium
CpG	cytosine-phosphate-guanine
DNA-M	DNA methylation
LGA	Large for gestational age
PCDH	protocadherin
RICHS	Rhode Island Child Health Study
SGA	Small for gestational age

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	Highlights
•	Maternal cadmium is associated with restricted fetal growth
•	Placental <i>PCDHAC1</i> expression is inversely associated with maternal cadmium
•	Placental <i>PCDHAC1</i> expression is positively associated with fetal growth
•	Associations were strongest for those with lower methylation in the <i>PCDHAC1</i> promoter



Figure 1.

Distribution of maternal cadmium exposure by infant size for gestational age. Cadmium (μ g/g) distribution by size for gestational age (Small = 10th BW percentile, Adequate = between the 10th and 90th BW percentiles, Large: 90th BW percentile); Ttests were used for pairwise comparisons across groups.



Figure 2.

Correlation matrix of DNA-M levels for the PCDHAC1 gene.

Pairwise Pearson correlations between DNA-M levels at CpG sites within the regulatory regions of the *PCDHAC1* gene; dark purple and dark green represent strong positive and negative correlations, whereas white represents no correlation.



Total Effect $\beta = 0.185$; p-value = 0.020

Figure 3.

Assessment of whether expression of PCDHAC1 mediated the effects of maternal cadmium on infant size for gestational age.

Structural equation model assessing the potential mediation of maternal Cd on size for gestational age, through low placental *PCDHAC1* expression, adjusted for confounders.

Table 1

Comparison of this study's sub-sample to the metals-sample across important maternal and infant characteristics.

Maternal and Infant Characteristics	n trapression (n:	=94)	-11)	=242)	
	Z	%	Z	%	χ^2 p-value
Mother Didn't Smoke	87	95.60	218	92.37	0.46
Smoked During Pregnancy *	4	4.40	18	7.63	
Female infant	53	56.38	119	49.17	0.29
Male infant	41	43.62	123	50.83	
SGA	18	19.15	39	16.12	0.56
AGA	48	51.06	139	57.44	
LGA	28	29.79	64	26.45	
	Mean	Std. Dev.	Mean	Std. Dev.	T-test p-value
BW percentiles	55.3	35.07	57.61	33.02	0.58
BW (g)	3523.55	683.67	3558.95	648.41	0.67
Maternal BMI	26.48	6.72	26.72	6.52	0.77
Maternal Age	31.68	4.07	31.70	4.40	0.97
Gestational time (weeks)	39.02	0.84	39.04	0.94	0.88
Maternal Cadmium (µg/g) **	0.01	0.02	0.01	0.01	0.86

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** Wilcox Rank-Sum Test used to compare across samples due to skewed distribution.

Interactions between DNA-M and log-Cd on log-PCDHAC1 expression.

	log	g-Cd	ā	N-M	Interact	ion Term
PCDHACI Regional DNA-M	B1	p-value	₿₂	p-value	₿1*2	p-value
TSS1500	-0.16	0.003	2.99	0.043	0.49	0.053
TSS200	-0.20	0.041	4.63	0.085	0.70	0.149
1 st Exon	-0.33	0.240	2.56	0.259	0.50	0.341
Gene Body	-0.30	0.140	2.91	0.268	0.59	0.246
3′UTR	-0.06	0.300	0.87	0.550	0.0034	0.990

DNA-M = DNA methylation; TSS1500 = within 1500 base pairs of the transcription start site; TSS200 = within 200 base pairs of the transcription start site, UTR = untranslated region.

Table 3

Linear associations between Cd and log-*PCDHACI* expression, stratified by High and Low DNA-M at the TSS200 and the TSS1500; β_1 may be interpreted as the change in log-*PCDHAC1* expression associated with a one unit increase in exposure, log(Cd µg/g) or maternal BMI (kg/m²).

		SSL	200	SSL	1500
Predictors	<i>PCDHAC1</i> DNA-M Strata	р,	p-value	Ъ	p-value
Crude Models					
log-Cd	Low DNA-M High DNA-M	-0.0901 -0.0283	0.0028 0.5245	- 0.1029 -0.0243	0.0042 0.5133
Multivariable N	Model				
log-Cd	Low DNA-M High DNA-M	-0.0980 -0.0188	0.0005 0.6507	-0.1060 -0.0248	0.0026 0.4599
Maternal BMI	Low DNA-M High DNA-M	0.0136 0.0168	0.0146 0.0399	0.0139 0.0155	0.0231 0.0633

TSS200 = within 200 base pairs of the transcription start site; TSS1500 = within 1500 base pairs of the transcription start site; Cd = Cadmium.

Table 4

Logistic and linear regressions for the effects of log-Cd on SGA and head circumference, respectively, stratified by high and low DNA-M levels within exposure, $\log(Cd \mu g/g)$; exponentiation of β_1 for logistic models can be interpreted as the odds of SGA associated with a one unit increase in exposure, the promoter region of *PCDHACI*; β_1 for linear models may be interpreted as the change in head circumference associated with a one unit increase in log(Cd µg/g).

Predictor	Outcome	<i>PCDHAC1</i> DNA-M strata	β	p-value	βī	p-value
Crude Mode	ls					
	SGA	Low DNA-M	0.756	0.022	0.695	0.034
Č		High DNA-M	0.584	0.118	0.807	0.068
log-Cd	Head Circumference	Low DNA-M	-0.346	0.059	-0.464	0.031
		High DNA-M	-0.070	0.686	-0.030	0.848
Adjusted Mo	dels					
	SGA *	Low DNA-M	1.732	0.017	1.110	0.024
Ċ		High DNA-M	0.253	0.501	0.626	0.107
log-Cd	Head Circumference **	Low DNA-M	-0.143	0.026	-0.191	0.003
		High DNA-M	0.052	0.664	0.038	0.772

transcription start site; Cd = Cadmium; SGA = Small for Gestational Age.

 $^{**}_{Adjusted}$ for birth weight, infant sex, gestational time (weeks), maternal BMI and maternal age.