

Received: 25 August 2017 Accepted: 15 December 2017 Published online: 10 January 2018

# **OPEN** Rare Earth Elements Removal from **Water Using Natural Polymers**

Adina Negrea<sup>1</sup>, Andreea Gabor<sup>1</sup>, Corneliu Mircea Davidescu<sup>1</sup>, Mihaela Ciopec<sup>1</sup>, Petru Negrea<sup>1</sup>, Narcis Duteanu 10 & Alina Barbulescu2

Adsorption of rare earth metals, Eu (III) and Nd (III) was investigated on a new environmental friendly material, thiourea functionalized cellulose. Before usage, the synthesized material was characterized by Fourrier Transform Infrared spectroscopy and energy dispersive X-ray analysis. The influence of adsorption parameters (adsorbent dosage, time, temperature and initial metal concentration) on adsorption capacity was investigated. Experimental data were fitted by using the pseudo-first-order and pseudo-second-order kinetic models. Simultaneously thermodynamic and equilibrium studies have been carried out using Langmuir, Freundlich and Sips isotherm. Maximum adsorption capacities were reached in 30 minutes at 298 K having the value of 27 mg/g for Eu (III) and 73 mg/g for Nd (III).

Rapid industrial development during last decades involves development and extensive usage of special materials in different commercial products<sup>1-3</sup>. The usage of heavy metals in industry introduces a large amount of toxic metals into atmosphere as well as into the aquatic and terrestrial environment4. Physical and chemical properties of the rare earth elements (REEs), correlated with their electronic, optical, and metallurgical properties make them perfect candidates for specific applications in cutting-edge technologies<sup>5-7</sup>. Wide spread of rare earth elements is due to their applications in many fields as special alloys, magnets, catalysts, and to their usage into equipments as batteries, sensors, energy efficient lighting, electric vehicles<sup>7–9</sup>, nuclear technologies, telecommunications, security systems<sup>1-3</sup>, medical application, and as active components of some fertilizers<sup>10</sup>. Industrial development is leading at real increase of consumption that can lead at some technical issues of waste disposal, especially due to the heavy metals content. Inadequate waste disposal leads at higher increase of public exposure at toxic compounds having negative effects on environment and human health11. It was proved the REE's bioaccumulation through the food chain can cause ailments due to the exposure of humans at low concentration of REEs<sup>6,10</sup>.

A viable alternative is the proper treatment of electrical and electronic compounds which have reached their lifetime, and that can become an alternative resource for many metals including REEs8. Applications from high technology field require high purity of REEs, with the content of non-rare earth impurity strictly limited, due to the negative influence of the impurities on the properties of desired rare earth material<sup>5</sup>. Nowadays, the extensive usage of REEs leads to an increase of environmental pollution.

Last decades, several methods (chemical precipitation, ion exchange extraction, coagulation, flocculation, liquid - liquid extraction, solid - phase extraction, biosorption, classical adsorption on different adsorbents) have been developed and used for metal removal from wastewaters<sup>7,12</sup>. The most effective and attractive one is represented by the adsorption, due to the advantages associated with the inexistence of chemical sludge and its highly removal efficiency<sup>12</sup>

Taking into account the environmental protection, new classes of economic and environmental friendly adsorbents must be developed and produced. Biopolymers represent a possible new class of adsorbents. Polysaccharides are widely available, biological polymers presenting a remarkable structural diversity. Based on their properties, they became an indispensable material for medicine, pharmaceutical industry, food, textile industry and also for environmental protection <sup>13,14</sup>. One of the simplest (biopolymer) polysaccharide is cellulose, which is a linear polysaccharide presenting no branches or substituents 15 and is often considered as the most abundant polymer produced in nature<sup>13,15</sup>. Being a natural product, represents a renewable biopolymer and can be considered as a promising environmental friendly adsorbent 16. Compared with other bi-products, cellulose

<sup>1</sup>Politehnica University Timisoara, Faculty of Industrial Chemistry and Environmental Engineering, Victoriei Square Nr. 2, 300006, Timişoara, Romania. <sup>2</sup>Ovidius University of Costanta, Romania, 124, Mamaia Blvd., 900527, Constanta, Romania. Correspondence and requests for materials should be addressed to C.M.D. (email: corneliu.davidescu@ upt.ro) or N.D. (email: narcis.duteanu@upt.ro)

Eu				Nd				
Ci = 50  mg/L			Ci = 50 mg/L					
ratio cellulose:thiourea	1:0.05	1:0.1	1:0.2	ratio cellulose:thiourea	0.05:1	0.1:1	0.2:1	
Q	10.4	10.3	10.6	q	11.6	11.6	11.7	

**Table 1.** Preliminary attempts for determination of optimum ratio cellulose: thiourea. (Ci – initial concentration, q – adsorption capacaity).

presents a bigger number of hydroxyl groups which can be easily modified, improving this way the synthesis yield and increasing his commercial value by making it eligible for new technological applications<sup>16,17</sup>.

In this study, cellulose was functionalized by impregnation with thiourea in order to get a new environmental friendly adsorbent material and was used for the adsorption of Nd (III) and Eu (III) ions from aqueous solutions. By performing kinetic, equilibrium and thermodynamic studies, we proved that beside a proper solid support, usage of thiourea as extractant has improved adsorption properties of cellulose.

#### Methods

**Chemicals and characterization.** All used chemicals are analytical grade and were purchased for Sigma-Aldrich. Functionalized adsorbent material was characterized by using FTIR and also EDX, in order to prove that the used technique leads at functionalized cellulose. After all adsorption experiment concentration of RREs into the effluent was measured using ICP- MS.

ICP-MS represent the most powerful analytical technique for detecting REEs, having the capability of multi-element detection over a wide concentration range, achieving extremely low detection limits and low mass interference<sup>6</sup>.

**Functionalization of the solid support.** In preliminary attempt was studied the influence of cellulose thiourea ratio on to the maximum adsorption capacity, so same amount of cellulose (1 gram) was mixed with three different solutions containing different amount of adsorbent (5 mL solution containing 0.05, 0.1 and respectively 0.2 g of thiourea dissolved in pure ethanol). Cellulose samples were maintained in contact with adsorbent solution for 24 hours, and after that was filtered, washed, and dried for minimum 24 hours at a suitable temperature. Obtained modified cellulose was used in adsorption experiment in order to find how the ratio cellulose: thiourea is influencing the adsorption of europium and neodymium ions. So in this stage synthesized material was kept in contact for 60 minutes with solutions containing 50 mg  $\rm L^{-1}$  Eu and Nd ions, after that the solution was filtered and were determined the residual concentrations of Eu and Nd ions after adsorption, obtaining in this way the maximum adsorption capacity (results presented in Table 1). Based on depicted data can observe that the maximum adsorption capacity have a constant value, so we choose for further experiment the ratio cellulose: thiourea 1:0.1.

Taking in account the results obtained during preliminary attempts cellulose was functionalized by mixing 5 g of cellulose with 25 mL solution containing 0.5 g of thiourea dissolved in 25 mL ethanol. Cellulose was maintained in contact with the adsorbent for 24 hours, and then the modified cellulose was filtered, washed, and dried for 24 hours at 323 K. The new prepared material was characterized using two different methods: Fourrier Transform Infrared Spectroscopy (FTIR) and Energy Dispersive X-Ray Analysis (EDX). FTIR spectra was recorded on Shimadzu Presige-21 FTIR spectrophotometer in the range 4000–400 cm<sup>-1</sup> by using KBr pellets technique, and the EDX spectra was recorded using Quanta FEG 250 scanning electron microscopy.

**Adsorption experiments.** Adsorption experiments were carried out to define the influence of important parameters: adsorbent dosage, contact time, temperature and initial concentration of metal ions onto the adsorption efficiency.

All Eu (III) and Nd (III) solutions used during adsorption experiments were prepared from stock solutions with  $1 \, \mathrm{g} \, \mathrm{L}^{-1}$  concentration, through proper dilution.

Adsorbent dosage influence was studied by using different amounts of functionalized cellulose (0.05, 0.1, 0.2, 0.3, 0.4, 0.5 g) which were mixed one hour with 25 mL solution containing  $50 \,\mathrm{mg} \,\mathrm{L}^{-1}$  of metal ions.

Contact time influence was studied using four samples 0.1 g of functionalized material mixed with 25 mL solutions containing 50 mg metallic ions per liter, and kept in contact for 15, 30, 45 and 60 minutes.

In order to investigate the temperature influence, the same amount of metallic solution, with the same concentration was mixed with 0.1 g of functionalized adsorbent and kept at 298, 308, and 318 K for 30 minutes.

The influence of the initial metal concentration was followed up by mixing for 30 minutes at 298 K, samples of 0.1 g functionalized material with 25 mL solution containing 10, 50, 100, 150, 200, 250, 300 mg of ionic metals per liter, for both studied metallic ions.

All samples were mixed in a Julabo SW23 mechanical shaker bath at 200 rotation min<sup>-1</sup>, filtered after and the metallic ions residual concentration was analyzed by inductively coupled plasma mass spectrometer – ICPMS Bruker Aurora M90.

The stages of our research are the following:

- (1) Characterization of the functional material, by using the Fourrier transformed infrared spectroscopy (FTIR) and the electron dispersive X-Ray analysis (EDX)
- (2) Study of the influence of adsorbent dosage on the adsorption of Eu (III) and Nd (III)
- (3) The kinetic studies and activation energy

In order to determine adsorption process kinetics for the studied material, experimental data were modeled using two different kinetic models: Lagergren pseudo-first -order model and Ho & McKay pseudo-second -order model.

The pseudo-first -order model $^{18}$  is described by the equation (1):

$$\ln(q_e - q_t) = \ln q_e - k_1 t \tag{1}$$

where t is the contact time (min).  $q_e$  is the adsorption capacity at equilibrium (mg/g),  $q_t$  is the adsorption capacity at the time t, and  $k_1$  is the adsorption constant rate (1/min).

The pseudo-second -order model<sup>19</sup> can be described by equation (2)

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \tag{2}$$

where t is the contact time (min),  $q_e$  is the adsorption capacity at equilibrium (mg/g),  $q_t$  is the adsorption capacity at time t and  $k_2$  the adsorption rate constant (g/mg·min).

The most important role in adsorption processes is played by the activation energy, which shows if the studied adsorption is a chemical or a physical process<sup>20–27</sup>. This energy can be evaluated by using the Arrhenius equation (equation (3)) using the adsorption rate constant, computed from the pseudo-second-order kinetic model:

$$ln k_2 = ln A - \frac{E}{RT}$$
(3)

where  $k_2$  is the pseudo-second-order rate constant (g/min·mg), A is the Arrhenius constant (min·g/mg), E is the activation energy (kJ/mol), E the ideal gas constant (8.314 J/mol·K), and E represent absolute temperature (K).

#### (4) Thermodynamic studies

Thermodynamic studies were performed in order to establish if the studied adsorption processes are spontaneous or not. Therefore, it was necessary to determine the value of the free Gibbs energy for studied adsorption processes.

Gibbs free energy can be computed from the Gibbs-Helmholtz relation (equation 4):

$$\Delta G^{\circ} = \Delta H^{\circ} - T \Delta S^{\circ} \tag{4}$$

where  $\Delta S^{\circ}$  is the standard entropy change and  $\Delta H^{\circ}$  is the standard enthalpy change. Standard entropy and enthalpy changes can be evaluated from van't Hoff equation (equation 5):

$$\ln K_{\rm d} = \frac{\Delta S^{\rm o}}{R} - \frac{\Delta H^{\rm o}}{RT},\tag{5}$$

where  $K_d$  is the equilibrium constant, T is the absolute temperature (K), R the ideal gas constant (8.314 J/mol·K).

The equilibrium constant  $K_d$  is given by the equation (6)

$$K_{\rm d} = \frac{q_{\rm e}}{C_{\rm e}},\tag{6}$$

 $q_e$  being the equilibrium adsorption capacity (mg/g), and  $C_e$  the equilibrium concentration (mg/L).

### (5) Equilibrium studies

Firstly, the adsorption isotherms have been determined in order to elucidate the adsorption mechanism<sup>19,28</sup>. Secondly, the experimental data were fitted by using three different non-linear isotherm models: Langmuir, Freundlich and Sips.

Langmuir's model<sup>29</sup> assumes that in all cases adsorption is accomplished by coating homogenous adsorbent surface with a monolayer of adsorbate, because in this case all active sites over the adsorbent surface are identical, and in same time all adsorbed molecules have the same activation energy. The non-linear Langmuir isotherm is described by equation (7):

$$q_{\rm e} = \frac{q_{\rm L}K_{\rm L}C_{\rm e}}{1 + K_{\rm L}C_{\rm e}}.\tag{7}$$

where:  $q_e$  is the equilibrium adsorption capacity (mg/g),  $C_e$  is the equilibrium concentration of adsorbent in the solution (mg/L),  $q_L$  (mg/g) represents the maximum adsorption capacity, and  $K_L$  is the Langmuir constant connected to the free energy of adsorption.

Freundlich isotherm model<sup>30</sup> assumes that the adsorption can take place also on heterogeneous surfaces as a multilayer adsorption. The non-linear Freundlich adsorption model is described by equation (8):

$$q_{\rm e} = K_{\rm F} C_{\rm e}^{1/n_{\rm F}}.$$
(8)

where:  $q_e$  is the equilibrium adsorption capacity (mg/g),  $C_e$  is the equilibrium concentration of the adsorbent in solution (mg/L),  $K_F$  and  $n_F$  are specific constants that are connected to the relative adsorption capacity of the adsorbent material and the adsorption intensity.

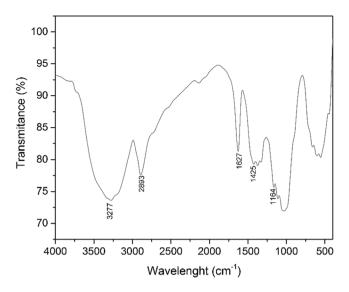


Figure 1. FTIR spectra of cellulose functionalized with thiourea.

Sips' adsorption model<sup>31</sup> represents a combination of the Langmuir and Freundlich models: at low concentrations, the adsorption processes have Freundlich properties, and at higher concentrations have Langmuir properties. Sips' adsorption model is described by equation (9):

$$q_e = \frac{q_s K_s C_e^{1/n_s}}{1 + K_s C_e^{1/n_s}}. (9)$$

where:  $q_s$  is the maximum adsorption capacity (mg/g),  $K_s$  is a constant related to the adsorption capacity of the material (mg/g) and  $n_s$  is the heterogeneity factor.

Adsorption isotherms are used in order to better understand the adsorption process and in order to elucidate the adsorption mechanism.

(6) Mathematical modeling of the evolution of q – adsorption capacity function of the REEs concentration at 298 K and of q function of time and temperature was performed. The first model is of Weibull type and the second one, linear multiple model.

# **Results and Discussions**

**Characterization of the functionalized material.** *FTIR characterization.* In order to identify if was obtained the functionalized cellulose was compared the FTIR spectra of functionalized cellulose with FTIR spectra of pure cellulose and pure thiourea (depicted in Fig. 1). From spectra of pure compounds were identified characteristic bands of cellulose and thiourea. Into the functionalized cellulose spectra is expected that the thiourea specific bands have less intensity in comparison with the pure compound spectra.

In wavelength interval between 3500 and  $3000\,\mathrm{cm^{-1}}$  can observe the large band located at  $3277\,\mathrm{cm^{-1}}$  presenting a small plateau at  $3160\,\mathrm{cm^{-1}}$  corresponding to the vibration of OH bonds. This large cellulose band include the asymmetric vibration of NH $_2$  thiourea group located at  $3395\,\mathrm{cm^{-1}}$  and the symmetric vibration of NH thiourea group located at  $3179\,\mathrm{cm^{-1}}$ . Band located at  $2893\,\mathrm{cm^{-1}}$  is associated with the vibrations of cellulose glucopyranose cycle C-H bonds.

Intense band located at 1627 cm<sup>-1</sup> into the FTIR spectra of functionalized cellulose is due to the elongation vibration of NH<sub>2</sub> group, and is not a band associated with cellulose because in case of pure cellulose is observed a weak band at 1427.3 cm<sup>-1</sup>. Intense bands observed at 1464 cm<sup>-1</sup> (vibration of CN bond + elongation of NH bond) and at 1395 cm<sup>-1</sup> (vibration of C=S bond) in spectra of pure thiourea, appears also into the spectra of functionalized cellulose as bands with medium intensity at 1425, 1380 and 1290 cm<sup>-1</sup>. Bands observed at 1164 and 1100 cm<sup>-1</sup> into the spectra of functionalized cellulose appears also into the spectra of pure cellulose at 1158 and 1104 cm<sup>-1</sup> associated with vibrations of C-O and O-C-O cellulosic bonds.

Based on presented data can conclude that analyzed material is represented by the thiourea functionalized cellulose.

*EDX analysis.* According to the EDX spectrum of cellulose functionalized with thiourea presented in Fig. 2, specific peaks belonging to elements from thiourea are visible (S peaks and N peak). These show a successfully functionalization of the natural polymer with thiourea extractant.

BET analysis. Cellulose and modified cellulose specific surface area were determined using a Micromeritics Brunauer-Emmett-Teller (BET) ASAP 2020 system. From BET adsorption isotherm (picture not showed) recorded for cellulose and functionalized cellulose were evaluated the specific surfaces areas. For pure cellulose

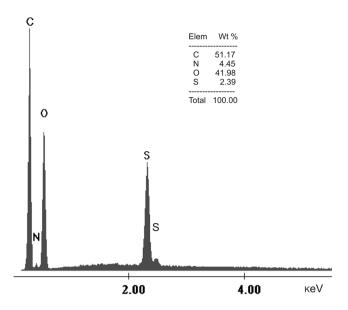


Figure 2. EDX spectra of cellulose functionalized with thiourea.

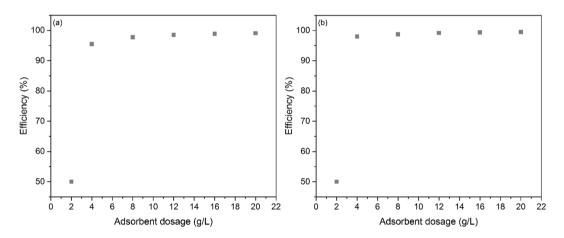


Figure 3. Adsorbent dosage influence on the adsorption of (a) Eu (III) and (b) Nd (III) (normalized at litter).

was obtained a specific surface of  $0.858\,m^2\,g^{-1}$  and in case of functionalized cellulose was obtained a specific surface of  $0.744\,m^2\,g^{-1}$ . From these data can observe a small decrease into the specific surface of the functionalized cellulose due to the thiourea functionalization.

**Eu (III) and Nd (III) adsorption onto cellulose functionalized with thiourea.** In preliminary attempts were effectuated comparative tests in order to see how the functionalization is affecting the maximum adsorption capacity of cellulose. So, by using Eu and Nd solutions with initial concentrations of  $50 \, \text{mg} \, \text{L}^{-1}$  were obtained the maximum adsorption capacities of  $1.69 \, \text{mg}$  Eu ions per g of cellulose and  $2.39 \, \text{mg}$  of Nd per g of cellulose in comparison with the maximum adsorption capacities of  $10.3 \, \text{and} \, 11.6 \, \text{obtained}$  when the functionalized cellulose was used as adsorbent material.

**Influence of the adsorbent dosage.** The influence of the adsorbent dosage on the adsorption of Eu (III) and Nd (III) on thiourea functionalized cellulose is depicted in Fig. 3.

It can be observed that when the used adsorbent was less than  $4\,\mathrm{g}\,\mathrm{L}^{-1}$ , the adsorption efficiency for both metals was around 50%. By increasing the adsorbent quantity, the removal efficiency increases and remains approximately constant at 95%. Based on presented data, it can be concluded that the optimum amount of the adsorbent for maximum adsorption efficiency is  $4\,\mathrm{g}\,\mathrm{L}^{-1}$ , amount used in all next experiments.

**Kinetic studies and activation energy.** As expected, the contact time between the adsorbent and the contaminated solution have an important influence on the adsorption efficiency for Eu (III) and Nd (III) ions. The temperature influences the maximum adsorption capacity of the modified cellulose as well. Based on the data presented in Fig. 4, it can be remarked the increase of the adsorption capacity for both ions when the temperature

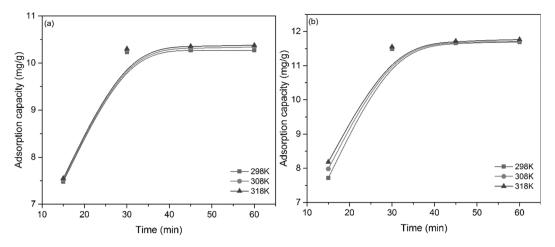


Figure 4. Effect of the contact time on the adsorption of (a) Eu (III) and (b) Nd (III) on the functionalized cellulose

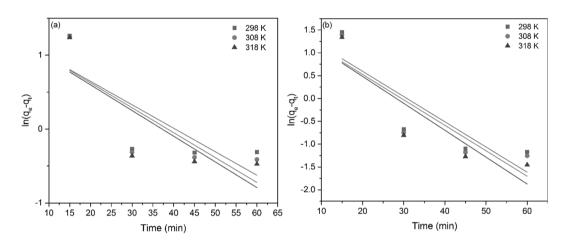


Figure 5. Pseudo-first-order kinetic plots of the adsorption of (a) Eu (III) and (b) Nd (III).

rises from 298 to 318 K. The Eu and Nd ions adsorption efficiencies are reaching the maximum values (10.3 mg Eu (III) for each gram of adsorbent and 11.6 mg Nd (III) for each gram of adsorbent used) after 30 minutes at 298 K. Further increase of the contact time does not lead to the further augmentation of the adsorption efficiencies.

Figure 5 depict the plot of  $\ln(q_e - q_t)$  against t obtained by modeling the experimental data using Lagergren pseudo-first-order model. Adsorption rate constant  $k_1$  and correlation coefficient are calculated from lines slopes and from the intercept points.

In the next stage, the experimental data obtained for Eu (III) and Nd (III) adsorption onto functionalized cellulose were modeled using pseudo-second-order model. Linear plots of  $t/q_t$  against t are depicted in Fig. 6.

Table 2 contains the parameters obtained from Lagergren and Ho and McKay kinetic models used to describe the adsorption processes of Eu and Nd ions from aqueous solutions on the functionalized cellulose. By comparing the results obtained from the two kinetic models, can conclude that the adsorption processes of Eu and Nd ions on thiourea functionalized cellulose are better described by pseudo-second-order kinetic model.

This conclusion is drawn based on the values of the correlation coefficient (that is closed to 1 for the pseudo-second-order equation) and the small differences between the experimental and the computed adsorption capacities ( $q_{e,exp}$  and  $q_{e,calc}$ ) of the used adsorbent.

In Fig. 7 are depicted the Arrhenius plot obtained for the adsorption of Eu and Nd ions onto the functionalized cellulose. The activation energies associated with the studied adsorptions have been determined from the slopes of the Arrhenius linear plots. Based on the fact the Eu ions adsorption needs an activation energy of 2.99  $10^{-3}$  kJ mol<sup>-1</sup> and Nd ions adsorption needs an activation energy of 5.4  $10^{-4}$  kJ mol<sup>-1</sup>, can conclude that the adsorption processes of Eu and Nd ions are physicals ones.

**Thermodynamic studies.** Figure 8 presents the linear depedence of  $\ln K_{\rm d}$  against 1/T associated with the adsorption processes of Eu and Nd ions onto the functionalized cellulose. The standard entropy change ( $\Delta S^{\circ}$ ) and the standard enthalpy change ( $\Delta H^{\circ}$ ) associated with studied adsorption processes are given by the slope of the line, respectively by the intersection of the lines with the Y axis. By using the computed entropy and enthalpy

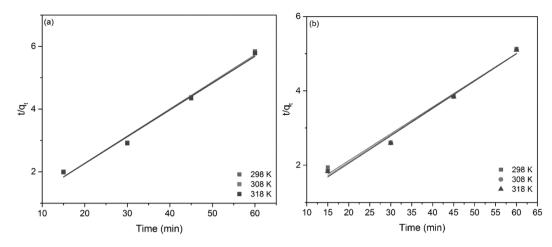


Figure 6. Pseudo-second-order kinetic plots of the adsorption of (a) Eu (III) and (b) Nd (III).

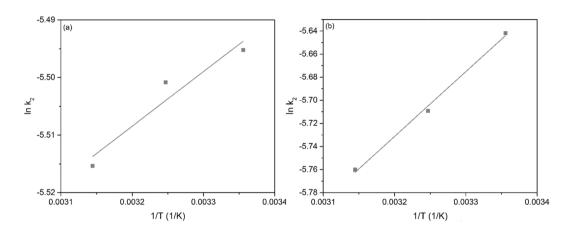


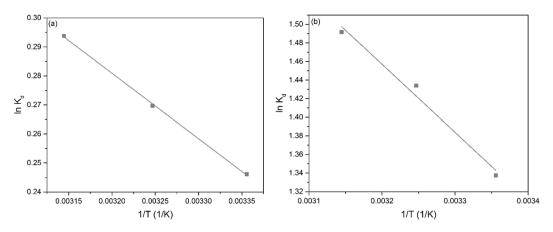
Figure 7. Arrhenius plot of the adsorption of (a) Eu (III) and (b) Nd (III) on the functionalized cellulose.

	Pseudo - first -	order kineti	С	Pseudo - second - order kinetic				
Temperature (K)	q <sub>e,exp</sub> (mg/g)	k <sub>1</sub> (1/min)	q <sub>e,calc</sub> (mg/g)	R <sup>2</sup>	k <sub>2</sub> (g/mg·min)	q <sub>e,calc</sub> (mg/g)	R <sup>2</sup>	
Eu (III)								
298	10.26	0.0317	3.60	0.4342	$4.11 \cdot 10^{-3}$	11.57	0.9849	
308	10.33	0.0336	3.66	0.4722	$4.08 \cdot 10^{-3}$	11.60	0.9856	
318	10.37	0.0347	3.64	0.4786	$4.02 \cdot 10^{-3}$	11.70	0.9858	
Nd (III)								
298	11.68	0.0552	5.48	0.6238	3.55-10-3	13.86	0.9729	
308	11.71	0.0557	5.17	0.6273	$3.31 \cdot 10^{-3}$	13.66	0.9784	
318	11.74	0.0589	5.27	0.6794	$3.15 \cdot 10^{-3}$	13.58	0.9824	

**Table 2.** Kinetic parameters of the adsorption at different temperatures.

changes, the free Gibbs energy values for Eu and Nd adsorptions are obtained. Calculated values of the thermodynamic parameters are summarized in Table 3.

Analyzing these parameters, one can conclude that the adsorptions on Eu and Nd ions on the modified cellulose are spontaneous processes as the free Gibbs energy has negative values in both cases. Since the standard enthalpy change has positive values, the adsorptions of Eu and Nd are endothermic processes. Positive values of standard entropy change suggest an increased disorder of the system. Since  $\Delta H^{\rm o}$  had values under 80 kJ mol $^{-1}$ , can conclude that studied adsorption processes onto modified cellulose are physical adsorptions.



**Figure 8.** Plot of  $\ln K_d$  against 1/T of the adsorption of (a) Eu (III) and (b) Nd (III) on the obtained material.

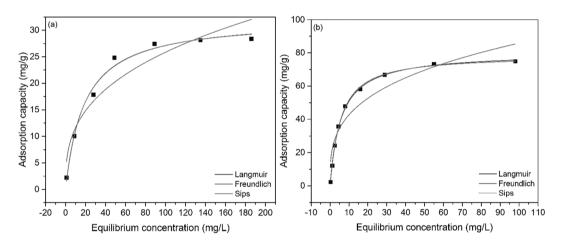


Figure 9. Adsorption isotherm of (a) Eu (III) and (b) Nd (III) on the functionalized cellulose with thiourea.

			ΔG° (kJ/mol)			
	ΔH° (kJ/mol)	ΔS° (J/mol·K)	298.15 K	308.15 K	318.15 K	$R^2$
Eu (III)	1.87	8.33	-0.608	-0.692	-0.775	0.9986
Nd (III)	6.08	31.52	-3.32	-3.64	-3.95	0.9682

**Table 3.** The thermodynamic parameters.

**Equilibrium studies.** Adsorption isotherm of Eu and Nd ions onto thiourea functionalized cellulose are depicted in Fig. 9; based on presented data can observe that the increase of initial concentrations of the metallic ions leads to the augmentation of the adsorption capacity until a maximum is reached.

For Eu ions, the equilibrium stage was reached when the initial concentration of the metallic ions was  $90\,\mathrm{mg}\,L^{-1}$ . Further increasing of the initial concentration leads to a slow augmentation of the adsorption capacity until a maximum value of  $27\,\mathrm{mg}\,g^{-1}$  was reached. For Nd ions, the equilibrium is reached when the initial concentration has a value of  $60\,\mathrm{mg}\,L^{-1}$ . Further increasing of the initial concentration leads to a slow augmentation of the adsorption capacity until reaching a maximum value of  $73\,\mathrm{mg}\,g^{-1}$ .

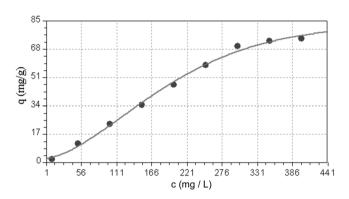
Comparing these values, we can notice a higher equilibrium concentration and a lower adsorption capacity of Eu ions and on the other hand, a lower equilibrium concentration and a higher adsorption capacity of Nd ions. This means that Nd ions are better adsorbed by thiourea functionalized cellulose.

Table 4 contains the parameters obtained by fitting the experimental data with the used isotherm models. The Freundlich isotherm model has the smaller correlation factor for both studied adsorptions, meaning that the other models have better described the adsorption of Eu and Nd on thiourea functionalized cellulose.

The value of the parameter  $1/n_F$  less than 1 indicates that the adsorption process is a favorable one, described by a convex isotherm. In this case, the adsorption sites characterized by highest binding energy are occupied first, followed by the sites with lower binding energy. A value of the parameter  $1/n_s$  approximately 1 suggests that the

	Langmuir isotherm						
	q <sub>m,exp</sub> (mg/g)	$K_{\rm L}$ (L/mg) $q_{\rm L}$ (mg/g		) R <sup>2</sup>			
Eu (III)	32	0.052 32.27			0.9914		
Nd (III)	73	0.17	80.07		0.9981		
	Freundlich isotherm						
	K <sub>F</sub> (mg/g)	$1/n_{ m F}$	$R^2$				
Eu (III)	5.24	0.34	0.9190				
Nd (III)	21.17	0.30	0.8825				
	Sips isotherm						
	K <sub>s</sub>	$q_{\rm s}$ (mg/g)	$1/n_{\rm s}$	$R^2$			
Eu (III)	0.05	32.09	1.01	0.9897			
Nd (III)	0.16	78.30	1.07	0.9985			

Table 4. Parameters of the Langmuir, Freundlich and Sips isotherm model.



**Figure 10.** Dependence of adsorption capacity function of initial concentration.

functionalized material have a low heterogeneity; as consequence, the mathematic equation associated to Sips' isotherm will be the same as that of the Langmuir isotherm. As consequence, the correlation coefficients of the Langmuir and Sips isotherms are almost equals. In both cases, the computed adsorption capacities of Eu and Nd ions are close to the experimental values.

**Mathematical models.** The first model determined was that of the evolution of adsorption capacity (q) function of metallic ions concentration (c) at 298 K. This model is of Weibull type (Fig. 10), with the equation:

$$q = 82.719 - 79.814 e^{-0.000189 \times c^{1.588}}$$
 (10)

The standard error of the model is 2.135 and the correlation coefficient 0.998.

It can be seen that adsorption capacity q is stabilizing at an initial metallic ions concentration of about 300 mg. The second model is that of the dependence of the adsorption capacity function of the adsorption time (t) and temperature (T) (Fig. 10):

$$q = 0.1714 t + 0.1177 T. (11)$$

To test the significance of the model's coefficients and of the model in its whole, the t-test and the F-test have been performed at the significance level of 0.05.

The values of the t-statistics corresponding to the coefficients are respectively 4.932 and 3.874, and the corresponding p-values 0.000 and 0.0043. Since the p-values are less than 0.05, the coefficients are significant. The value of the F-statistics is 174.79 and the p-value = 0.000, proving that the model is significant.

The determination coefficient is  $R^2 = 0.972$ , showing that 97.2% of the variation of q is explained by the variation of T and t.

To validate the model, the residual analysis has been performed as well, at the same confidence level. We found that there is no autocorrelation of the residual, they are Gaussian (the value of the Anderson – Darling statistics is 0.293 and the p-value = 0.542), and homoscedastic (the Levene statistics is 0.02 and the p-value = 0.880). Therefore, the multiple linear model is correct from the viewpoint of statistics.

#### **Conclusions**

Present investigation demonstrate that the new material obtained by functionalization of the cellulose with thiourea represent an effective adsorbent material for Eu (III) and Nd (III) cations removal from aqueous solutions. FTIR and EDX spectra show that the cellulose was successfully impregnated with thiourea. It was proven that for an efficient adsorption of Eu and Nd ions, the optimum adsorption parameters are: adsorbent dosage – 0.1 g of

functionalized material, contact time – 30 minutes, and temperature – 298 K. The equilibrium study proved that the adsorption of rare earth elements is well described by Langmuir isotherm, the maximum adsorption capacities being of  $27\,\mathrm{mg\,g^{-1}}$  for Eu (III) and  $73\,\mathrm{mg\,g^{-1}}$  for Nd (III). Kinetic behavior of Eu and Nd adsorption, are better described by the pseudo–second–order models rather than pseudo–first–order model, confirming that the studied adsorptions are physical one.

The mathematical models can be used for the interpolation of the experimental values of adsorption capacity function of metallic cations initial concentrations and also function time and temperature. Mathematical model proved that the maximum adsorption capacity is reached when the initial concentration was around 300 mg/L.

# References

- Ashour, R. M., Abdel-Magied, A. F., Abdel-khalek, A. A., Helaly, O. S. & Ali, M. M. Preparation and characterization of magnetic iron oxide nanoparticles functionalized by l-cysteine: Adsorption and desorption behavior for rare earth metal ions. *Journal of Environmental Chemical Engineering* 4, 3114–3121, https://doi.org/10.1016/j.jece.2016.06.022 (2016).
- Ashour, R. M. et al. Selective separation of rare earth ions from aqueous solution using functionalized magnetite nanoparticles: kinetic and thermodynamic studies. Chemical Engineering Journal 327, 286–296, https://doi.org/10.1016/j.cej.2017.06.101 (2017).
- 3. Gok, C. Neodymium and samarium recovery by magnetic nano-hydroxyapatite. *Journal of Radioanalytical and Nuclear Chemistry* **301**, 641–651, https://doi.org/10.1007/s10967-014-3193-z (2014).
- Saravanan, R. & Ravikumar, L. The Use of New Chemically Modified Cellulose for Heavy Metal Ion Adsorption and Antimicrobial Activities. *Journal of water resource and protection* 7, https://doi.org/10.4236/jwarp.2015.76042 (2015).
- 5. An, F. et al. Selectively removal of Al (III) from Pr (III) and Nd (III) rare earth solution using surface imprinted polymer. Reactive and Functional Polymers 73, 60–65, https://doi.org/10.1016/j.reactfunctpolym.2012.08.022 (2013).
- Celik, I., Kara, D., Karadas, C., Fisher, A. & Hill, S. J. A novel ligandless-dispersive liquid liquid microextraction method for matrix elimination and the preconcentration of rare earth elements from natural waters. *Talanta* 134, 476–481, https://doi.org/10.1016/j. talanta.2014.11.063 (2015).
- 7. Unal Yesiller, S., Eroglu, A. E. & Shahwan, T. Removal of aqueous rare earth elements (REEs) using nano-iron based materials. *Journal of Industrial and Engineering Chemistry* 19, 898–907, https://doi.org/10.1016/j.jiec.2012.11.005 (2013).
- 8. Yoon, H.-S. *et al.* Solvent extraction, separation and recovery of dysprosium (Dy) and neodymium (Nd) from aqueous solutions: Waste recycling strategies for permanent magnet processing. *Hydrometallurgy* **165**(Part 1), 27–43, https://doi.org/10.1016/j. hydromet.2016.01.028 (2016).
- 9. Parhi, P. K., Park, K. H., Nam, C. W. & Park, J. T. Liquid-liquid extraction and separation of total rare earth (RE) metals from polymetallic manganese nodule leaching solution. *Journal of Rare Earths* 33, 207-213, https://doi.org/10.1016/S1002-0721(14)60404-X (2015).
- Liang, P., Liu, Y. & Guo, L. Determination of trace rare earth elements by inductively coupled plasma atomic emission spectrometry after preconcentration with multiwalled carbon nanotubes. Spectrochimica Acta Part B: Atomic Spectroscopy 60, 125–129, https:// doi.org/10.1016/j.sab.2004.11.010 (2005).
- 11. Yantasee, W. et al. Selective removal of lanthanides from natural waters, acidic streams and dialysate. *Journal of Hazardous materials* **168**, 1233–1238, https://doi.org/10.1016/j.jhazmat.2009.03.004 (2009).
- 12. Li, L., Liu, F., Jing, X., Ling, P. & Li, A. Displacement mechanism of binary competitive adsorption for aqueous divalent metal ions onto a novel IDA-chelating resin: Isotherm and kinetic modeling. *Water Research* 45, 1177–1188, https://doi.org/10.1016/j. watres.2010.11.009 (2011).
- 13. Zhang, Q., Lin, D. & Yao, S. Review on biomedical and bioengineering applications of cellulose sulfate. *Carbohydrate Polymers* 132, 311–322, https://doi.org/10.1016/j.carbpol.2015.06.041 (2015).
- 14. Habibi, Y. & Lucia, L. A. (Wiley, Hoboken, New Jersey, 2012).
- Fox, S. C., Li, B., Xu, D. & Edgar, K. J. Regioselective esterification and etherification of cellulose: A review. Biomacromolecules 12, 1956–1972, https://doi.org/10.1021/bm200260d (2011).
- Gurgel, L. V. A., P de Melo, J. C., de Lena, J. C. & Gil, L. F. Adsorption of chromium (VI) ion from aqueous solution by succinylated mercerized cellulose functionalized with quaternary ammonium groups. *Bioresource Technology* 100, 3214–3220, https://doi. org/10.1016/j.biortech.2009.01.068 (2009).
- 17. Gurgel, L. V. A., Junior, O. K., Gil, R. Pd. F. & Gil, L. F. Adsorption of Cu(II), Cd(II), and Pb(II) from aqueous single metal solutions by cellulose and mercerized cellulose chemically modified with succinic anhydride. *Bioresource Technology* 99, 3077–3083, https://doi.org/10.1016/j.biortech.2007.05.072 (2008).
- 18. Lagergren, S. About the theory of so-called adsorption of soluble substabces. *Kungliga Svenska Vetenskapsakademiens Handlingar* **1898**(24), 1–39 (1898).
- Ho, Y. S. & McKay, G. Pseudo-second order model for sorption processes. Process Biochemistry 34, 451–465, https://doi.org/10.1016/ S0032-9592(98)00112-5 (1999).
- 20. Doğan, M., Özdemir, Y. & Alkan, M. Adsorption kinetics and mechanism of cationic methyl violet and methylene blue dyes onto sepiolite. *Dyes and Pigments* 75, 701–713, https://doi.org/10.1016/j.dyepig.2006.07.023 (2007).
- 21. Hobson, J. P. Physical adsorption isotherms extending from ultrahigh vacuum to vapor pressure. *The Journal of Physical Chemistry* 73, 2720–2727, https://doi.org/10.1021/j100842a045 (1969).
- 22. Kara, A. & Demirbel, E. Kinetic, Isotherm and Thermodynamic Analysis on Adsorption of Cr(VI) Ions from Aqueous Solutions by Synthesis and Characterization of Magnetic-Poly(divinylbenzene-vinylimidazole) Microbeads. *Water, Air, & Soil Pollution* 223, 2387–2403, https://doi.org/10.1007/s11270-011-1032-1 (2012).
- Krishna, B. S., Murty, D. S. R. & Jai Prakash, B. S. Thermodynamics of Chromium(VI) Anionic Species Sorption onto Surfactant-Modified Montmorillonite Clay. *Journal of Colloid and Interface Science* 229, 230–236, https://doi.org/10.1006/jcis.2000.7015 (2000)
- Özcan, A., Özcan, A. S., Tunali, S., Akar, T. & Kiran, I. Determination of the equilibrium, kinetic and thermodynamic parameters of adsorption of copper(II) ions onto seeds of Capsicum annuum. *Journal of Hazardous Materials* 124, 200–208, https://doi. org/10.1016/j.jhazmat.2005.05.007 (2005).
- Tassist, A., Lounici, H., Abdi, N. & Mameri, N. Equilibrium, kinetic and thermodynamic studies on aluminum biosorption by a mycelial biomass (Streptomyces rimosus). *Journal of Hazardous Materials* 183, 35–43, https://doi.org/10.1016/j.jhazmat.2010.06.078 (2010)
- 26. Ünlü, N. & Ersoz, M. Adsorption characteristics of heavy metal ions onto a low cost biopolymeric sorbent from aqueous solutions. *Journal of Hazardous Materials* 136, 272–280, https://doi.org/10.1016/j.jhazmat.2005.12.013 (2006).
- Wang, C.-C. et al. Effects of exchanged surfactant cations on the pore structure and adsorption characteristics of montmorillonite. *Journal of Colloid and Interface Science* 280, 27–35, https://doi.org/10.1016/j.jcis.2004.07.009 (2004).
- 28. Ho, Y.-S. Review of second-order models for adsorption systems. *Journal of Hazardous Materials* 136, 681–689, https://doi.org/10.1016/j.jhazmat.2005.12.043 (2006).

- 29. Langmuir, I. Theadsorption Of Gases On Plane Surfaces Of Glass, Mica And Platinum. *Journal of the American Chemical Society* 40, 1361–1403, https://doi.org/10.1021/ja02242a004 (1918).
- 30. Freundlich, H. M. F. Over the adsorption in solution. J. Phys. Chem. 57, 385-470 (1906).
- 31. Sips, R. On the Structure of a Catalyst Surface. The Journal of Chemical Physics 16, 490-495, https://doi.org/10.1063/1.1746922 (1948).

# Acknowledgements

This research did not receive any specific grant from funding agencies in the public, commercial, or not-for-profit sectors.

# **Author Contributions**

Adina Negrea, Corneliu Mircea Davidescu, Petru Negrea designed the research, Andreea Gabor performed research, collected data, Mihaela Ciopec, Narcis Duteanu carried out analyses, Alina Barbulescu prepared the mathematical models. Adina Negrea and Narcis Duteanu wrote the first draft of the manuscript. All authors contributed to interpretation of experimental data, discussion and the preparation of final manuscript.

# **Additional Information**

**Competing Interests:** The authors declare that they have no competing interests.

**Publisher's note:** Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit <a href="https://creativecommons.org/licenses/by/4.0/">https://creativecommons.org/licenses/by/4.0/</a>.

© The Author(s) 2017