

# Zebrafish Locomotor Responses Reveal Irritant Effects of Fine Particulate Matter Extracts and a Role for TRPA1

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## ABSTRACT

Exposure to fine particulate matter (PM) air pollution causes adverse cardiopulmonary outcomes. Yet, the limited capacity to readily identify contributing PM sources and associated PM constituents in any given ambient air shed impedes risk assessment efforts. The health effects of PM have been attributed in part to its capacity to elicit irritant responses. A variety of chemicals trigger irritant behavior responses in zebrafish that can be easily measured. The purposes of this study were to examine the utility of zebrafish locomotor responses in the toxicity assessment of fine PM and its chemical fractions and uncover mechanisms of action. Locomotor responses were recorded in 6-day-old zebrafish exposed for 60 min in the dark at 26 °C to the extractable organic matter of a compressor-generated diesel exhaust PM (C-DEP) and 4 of its fractions (F1–F4) containing varying chemical classes of increasing polarity. The role of the transient receptor potential (TRP) cation channel TRPA1, a chemical sensor in mammals and zebrafish, in locomotor responses to C-DEP, was also examined. Acrolein, an environmental irritant and known activator of TRPA1, and all extracts induced concentration-dependent locomotor responses whose potencies ranked as follows: polar F3 > weakly polar F2 > C-DEP > highly polar F4 > nonpolar F1, indicating that polar and weakly polar fractions that included nitro- and oxy-polyaromatic hydrocarbons (PAHs), drove C-DEP responses. Irritant potencies in fish positively correlated with mutagenic potencies of the same extracts in strains of *Salmonella* sensitive to nitro- and oxy-PAHs, further implicating these chemical classes in the zebrafish responses to C-DEP. Pharmacologic inhibition of TRPA1 blocked locomotor responses to acrolein and the extracts. Taken together, these data indicate that the zebrafish locomotor assay may help expedite toxicity screening of fine PM sources, identify causal chemical classes, and uncover plausible biological mechanisms.

**Key words:** air pollution; particulate matter; zebrafish; locomotor; irritant responses; TRPA1.

In a recent global survey of chronic disease, exposure to particulate matter (PM) air pollution was the ninth leading risk factor, driven largely by cardiovascular and respiratory disease outcomes (Lim *et al.*, 2012). Although ambient PM mass and size are major determinants of PM-related adverse health effects,

leachable components of PM, including metals, and polyaromatic hydrocarbons (PAHs), are the key initiators of PM-induced biological responses (Brook *et al.*, 2010; Thomson *et al.*, 2016; Xia *et al.*, 2004), and are associated with measurable clinical effects apart from those attributable to PM mass (Delfino *et al.*, 2010;

Kraus *et al.*, 2011; Ostro *et al.*, 2007; Ruiz-Vera *et al.*, 2015; Wittkopp *et al.*, 2016; Wu *et al.*, 2016; Zhang *et al.*, 2016). PM composition, however, is exceedingly complex, varying considerably across air sheds, and heavily influenced by weather, time of day, and proximity to emission sources (Brook *et al.*, 2010). Thus, information on the relative contribution of various PM sources and associated PM constituents in the toxicity of ambient PM is limited or unavailable, impeding risk assessment.

The most common experimental approaches used to assess the *in vivo* toxicity of PM sources are rodent inhalation and instillation studies, which are constrained by requirements for large amounts of PM, often in limited supply, and are low throughput. Higher throughput *in vitro* methods using mammalian cells fail to recapitulate the complexity of an intact organism, including the crosstalk of organ systems and the resulting multisystem physiological responses elicited by PM exposure. The zebrafish (*Danio rerio*), a fully intact vertebrate *in vivo* model, is a higher throughput alternative to mammalian models, with a high degree of functional conservation with mammals (Peterson and MacRae, 2012), and offers technical simplicity on par with *in vitro* cell assays. Moreover, zebrafish skin is sensitive to many chemicals known to irritate mammalian lung epithelium, including the aldehyde acrolein (Prober *et al.*, 2008), and expresses the transient receptor potential (TRP) cation channel TRPA1, which like its mammalian ortholog, mediates chemosensation to exogenous and endogenous agents (Prober *et al.*, 2008), as well as chemicals that cause oxidative stress (Takahashi and Mori, 2011). Importantly, PM is rich in chemical triggers of oxidative stress including PAHs (Baulig *et al.*, 2003), and TRPA1 has been described as a major sensor of oxidative stress in rodent pulmonary airways (Bessac *et al.*, 2008).

The purpose of this study was to examine the utility of zebrafish locomotor behavior responses in the toxicity assessment of fine PM. Locomotor responses in zebrafish, wherein the distance moved reflects the potency of the irritant (Padilla *et al.*, 2011; Prober *et al.*, 2008), have been previously measured in response to a variety of chemicals, including bisphenol A (Wang *et al.*, 2015), organophosphate flame retardants (Jarema *et al.*, 2015; Oliveri *et al.*, 2015; Sun *et al.*, 2016), benzo[a]pyrene (Gao *et al.*, 2015), atrazine (Liu *et al.*, 2016), and crude oil (Perrichon *et al.*, 2016). In this study, zebrafish larvae were exposed to a whole organic extract or 1 of 4 chemical fractions of 1 PM sample, ie, compressor-generated diesel exhaust PM (C-DEP), all of which were previously analyzed for organic chemical content (Mutlu *et al.*, 2013). Diesel exhaust (DE) PM is a major contributor to traffic-derived ambient PM and is mostly composed of inorganic carbon and a variety of organic compounds such as PAHs and nitroarenes (nitro-PAHs) (IARC, 2014). C-DEP has been previously analyzed for mutagenicity (Mutlu *et al.*, 2013), electrophilic and redox properties (Shinyashiki *et al.*, 2009), and induction of matrix metalloproteinases in human bronchial epithelial cells (Li *et al.*, 2009). By evaluating chemical fractions of C-DEP, we inferred which chemical classes were most responsible for the irritant responses. We also examined the role of TRPA1 in irritant responses to C-DEP in zebrafish, given that TRPA1-mediated adverse cardiac responses to DE in rodents (Hazari *et al.*, 2011) and TRPA1 in mammalian cells *in vitro* is activated by various DE PM chemicals (Deering-Rice *et al.*, 2011).

## MATERIALS AND METHODS

**Chemicals and particulate extracts.** Dimethyl sulfoxide (DMSO) was obtained from Sigma Aldrich (St Louis, Missouri). Acrolein was

obtained from O2SI Smart Solutions (Charleston, South Carolina). The TRPA1 antagonist HC030031 was obtained from Tocris Bioscience (Ellisville, Missouri). C-DEP was generated, extracted, and fractionated as previously described (Mutlu *et al.*, 2013); the same extracts prepared by Mutlu *et al.* (2013), stored frozen in DMSO at 4°C, were used for this study. Previous studies have shown that diesel exhaust extracts are highly stable after many years of storage, eg, National Institute of Standards and Technology standard SRM1650 had almost identical mutagenic potencies in strains TA98 and TA100 of *Salmonella* after 5 years of storage (Claxton *et al.*, 1992; Hughes *et al.*, 1997). Briefly, C-DEP was produced by combusting low-sulfur petroleum diesel in a 30-kW (40 hp) 4-cylinder Deutz BF4M1008 diesel engine; particles were collected in a bag house and stored in sealed glass containers under nitrogen at 4°C until extracted by Mutlu *et al.* (2013). The particles were extracted by sonication with dichloromethane (DCM), and the DCM whole extract was fractionated sequentially on a silica-gel column with hexane (nonpolar fraction 1 [F1]), 75% hexane: 25% DCM (weakly polar fraction 2 [F2]), DCM (polar fraction 3 [F3]), and methanol (highly polar fraction 4 [F4]; Table 1). The percent extractable organic matter (EOM) was determined gravimetrically, and the extracts were solvent-exchanged into DMSO at 10 mg EOM/ml for the whole extract and 250 µg EOM/ml for the 4 fractions. Note: gravimetric analysis showed that the % EOM of the organic (DCM) extract of the C-DEP particle was 23%, which means that 23% of the weight of the particles was extractable by DCM, and it was this organic material (ie, whole C-DEP extract) that was fractionated.

**Experimental animals.** Wild type adult zebrafish (*D. rerio*; undefined, outbred stock originally obtained from Aquatic Research Organisms, Hampton, New Hampshire 03842 and EkkWill Waterlife Resources Ruskin, Florida 33575) were housed in an Association for Assessment and Accreditation of Laboratory Animal Care-approved animal facility with a 14:10 h light:dark cycle (lights on at 08:30 h). We have previously determined (unpublished data) that behavioral responses in our outbred line of fish are very similar to those in the AB line of fish. Adult fish (males and females housed together, approximately 8 per liter) were kept in 1 of several 9-liter (l) colony tanks (Aquaneering Inc, San Diego, California) with a water temperature of 28°C. For group spawning, all the adults from 2 different home tanks (approximately 80–100 total fish) were placed in 15 l static tanks set up approximately 3 PM, with embryos collected the next morning between 8:30 and 9 AM. All embryos were gathered from the breeder tank and placed in a 26°C water bath and washed twice with 0.06% bleach (vol/vol) in 10% Hanks' Balanced Salt Solution (HBSS; consisted of a mix of NaCl, KCl, Na<sub>2</sub>HPO<sub>4</sub>, CaCl<sub>2</sub>, MgSO<sub>4</sub>, and NaHCO<sub>3</sub>, all chemicals from Sigma Aldrich) for 5 min.

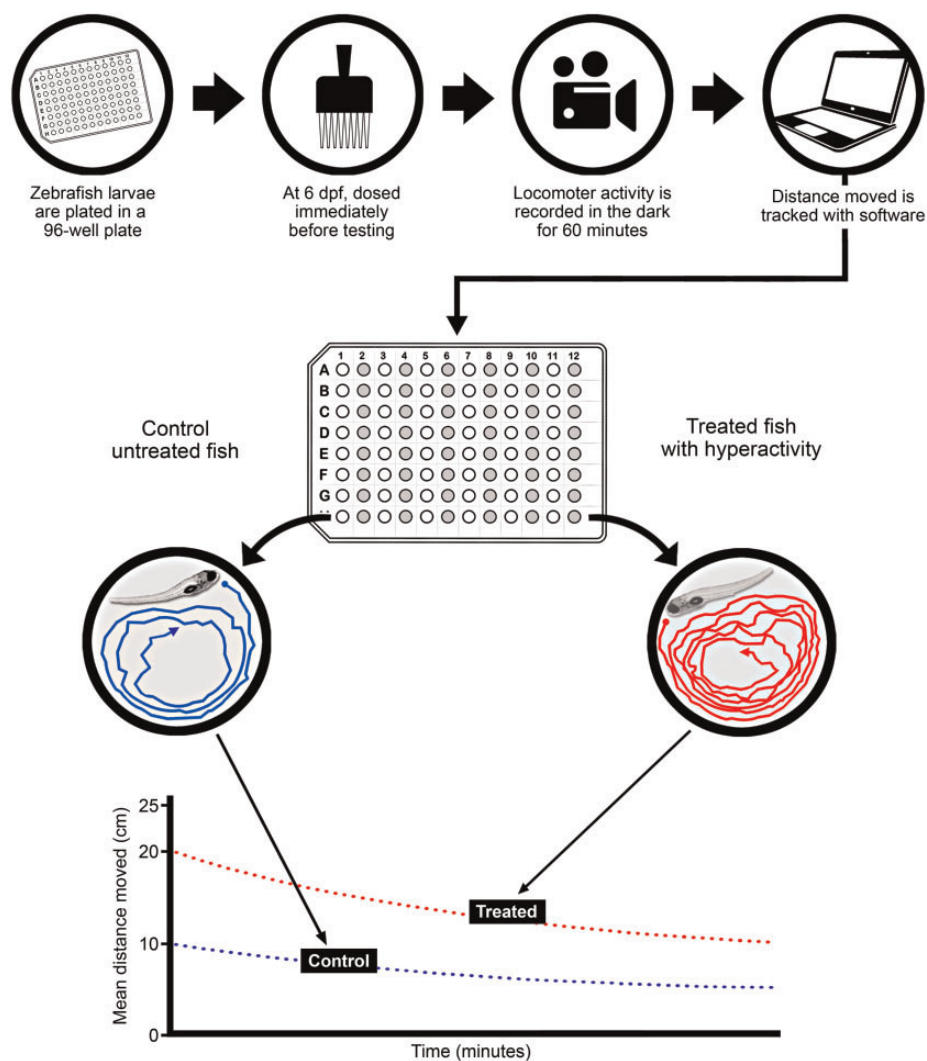
All studies were carried out in accordance with the guidelines of, and approved by, the Institutional Animal Care and Use Committee at the U.S. Environmental Protection Agency's National Health and Environmental Effects Research Laboratory.

**Group size determinations and controlling for plate variability.** We performed sample size analysis using R Studio software (R Studio, Inc) with the "pwr" package (<https://cran.r-project.org/web/packages/pwr/pwr.pdf>) and "pwr.anova.test" command and group sizes were calculated based on effect size indexes obtained from pilot studies. We performed all experiments with duplicate plates, and we combined data from both plates from

Table 1. Chemical Characteristics of C-DEP Fractions<sup>a</sup>

	Fraction 1	Fraction 2	Fraction 3	Fraction 4
Polarity	Nonpolar	Weakly polar	Moderately polar	Highly polar
Example compounds	Alkanes/alkenes	PAHs and nitro-PAHs	Nitro- and oxy-PAHs	Oxy-PAHs
% EOM of C-DEP	Approximately 52	Approximately 5	Approximately 15	Approximately 28

<sup>a</sup>Adapted from Mutlu et al. (2013).



**Figure 1.** Experimental protocol for treatment and behavior assessment of zebrafish larvae. Zebrafish were plated at 6–8 h postfertilization, and then dosed at 6-days postfertilization (dpf), immediately before assessment. Locomotor activity was recorded in the dark for 60 min, and analysis software was used to derive distance moved (cm)/unit time (min). Control (unshaded wells) and treated (shaded wells) fish were dosed in alternating columns of each 96-well plate.

each experiment only after ensuring that there were no plate-to-plate differences. A series of control plates were also run to compare the impact of the various vehicles (1) HBSS versus 0.4% DMSO and (2) 0.4% DMSO versus 0.8% DMSO.

*Concentration-response studies with acrolein, whole C-DEP, and C-DEP fractions.* For the dose-response studies, embryos were plated into single wells of a 96-well plate (1 fish/well) on day 0 and then at 6-days postfertilization (dpf;  $n = 15\text{--}32$ /experimental condition) acutely exposed to vehicle (0.4% DMSO, final nominal concentration), a positive control acrolein (1–30  $\mu\text{M}$ ), whole C-DEP extract (0.125–40  $\mu\text{g}/\text{ml}$ , final nominal concentrations in

wells), or fractions 1 thru 4 (0.125–40  $\mu\text{g}/\text{ml}$ , final nominal concentrations in wells) and activity was assessed in the dark at 26°C for 60 min using Noldus (Leesburg, Virginia) Ethovision video tracking software (Figure 1). Working stock solutions of C-DEP and F1 thru F4 in DMSO and of acrolein in high-performance liquid chromatography-grade water were prepared prior to dosing. Dosing was achieved by adding 1  $\mu\text{l}$  of test solution to 250  $\mu\text{l}$  of HBSS media in each well of a 96-well plate.

For the whole extract and 4 fractions, we performed linear regressions over the linear portion of the dose-response curves to determine slopes, which were used to indicate the irritant potency of each extract and were expressed as mean distance

moved (cm)/ $\mu\text{g}$  EOM/ml. This mirrors a similar approach used previously to quantify the mutagenic potencies of the same PM extracts (Mutlu et al., 2013). Slopes were selected based on the following criteria: (1) correlations that provided the highest goodness of fit (ie,  $r^2$  value) and (2) the highest concentration before the concentration curve plateaued.

**TRPA1 antagonism studies.** Locomotor activity was also measured in acrolein, C-DEP, F2, or F3-exposed zebrafish pretreated with a TRPA1 antagonist (HC030031). One microliter of HC030031 in 0.4% DMSO (final well-concentration of HC030031 = 50  $\mu\text{M}$ ) or 0.4% DMSO alone were added to each well 15 min before dosing with acrolein, C-DEP, F2, or F3. The TRPA1 antagonist pretreatment regimen and treatment concentration were adapted from Curtright et al. (2015). For the antagonism studies, final DMSO concentration was 0.8% for all wells because DMSO was used as the vehicle for both the TRPA1 antagonist and the PM extracts, which were administered separately.

**Behavioral testing.** All testing was performed on 6 dpf larvae in 96-well plates (Figure 1). Video recording of fish behavior was essentially as described previously (MacPhail et al., 2009).

On the morning of testing, after the rearing solution was changed, plates were moved to the dark in the behavioral testing room for 1 h. Temperature in the testing room was kept at 26°C. For testing, the plate was transferred to the light box for 25 min, with dosing taking place during the last 5 min of the light program. Because we were interested in irritant effects of exposure and because organisms often adapt to irritant responses over time, the dark program was started immediately after dosing using a behavior-recording system (Noldus Information Technology, Leesburg, Virginia) to record immediate irritant-induced movement of each fish for 60 min in response to treatment.

**Analysis of fish movement.** Fish movement (locomotion) was tracked from videos using Ethovision XT (Noldus Information Technology) software Version 8.5. Tracking rate was 5 samples/s (ie, an image was captured every 200 ms). A dynamic subtraction method was used to detect objects that were darker than background, with a minimum object size of 10 pixels. Tracks were analyzed for total distance moved (cm). The software provides a distance output per 2 min period. An input filter of 0.135 cm (minimum distance moved) was used to remove system noise. All locomotion data are expressed as distance moved (cm) per unit time (Figure 1). We calculated slopes from linear regressions of these data to determine the irritant potency (mean distance moved [cm]/ $\mu\text{g}$  EOM) of each extract.

**Statistics.** Data were analyzed as follows: (1) data (2 min locomotor activity over time) presented in Supplementary Figure 1 were assessed using a repeated ANOVA with time and dose as the independent variables and locomotor activity (distance moved/time) as the dependent variable using GraphPad Prism Software Version 6.0 (La Jolla, California). Significance was set at  $P \leq .05$ . (2) Data (10 min locomotor activity over time) presented in Figures 2–4 were first assessed using a repeated ANOVA with time and dose as the independent variables and locomotor activity (distance moved/time) as the dependent variable using Statview (SAS Institute, Inc, Cary, North Carolina; version 5.0.1). Significance was set at  $P \leq .05$ . In the case of a significant time  $\times$  dose interaction, step-down ANOVAs were performed to assess lower order effects. This involved first assessing if there was a significant effect of dose at each 10 min behavioral

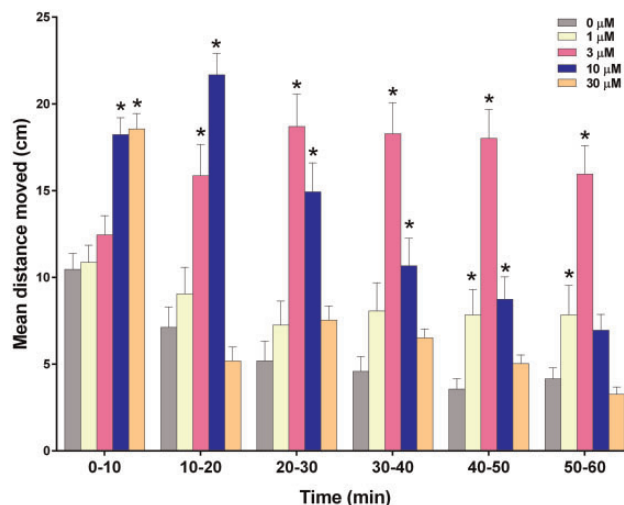


Figure 2. Effect of acrolein on zebrafish larval locomotor activity measured in the dark. Zebrafish were treated with 1, 3, 10, or 30  $\mu\text{M}$  acrolein in 0.4% dimethyl sulfoxide (DMSO) or 0.4% DMSO alone approximately 5 min before monitoring locomotor behavior. Data are presented as mean  $\pm$  SEM for the entire 60 min of recording. Data from 2 replicate plates were combined for a total of 31–32 zebrafish per treatment group. \*Significantly greater than vehicle alone at corresponding time ( $P < .05$ ).

interval, and if so, Fisher's Protected Least-Squares Difference comparisons were conducted to compare among dosage groups. (3) Linear regressions of dose-response data for whole C-DEP and F1–F4 (Table 2) and of plots of irritant potency in zebrafish versus mutagenic potency data (Table 3) were performed using GraphPad Prism Software Version 6.0. For a comparison of the potency (ie, slope) data in Table 2, the slopes were compared via 1-way ANOVA followed by Tukey's post hoc test (note: for this analysis, the slopes derived from the linear regression of the concentration-response curve for each extract were entered as the mean, the standard error for the slope also derived from the linear regression data was entered as the SEM, and the degree of freedom [df] value was entered as "n" or group size—the df for all groups in this analysis was 1). (4) Data (60-min locomotor activity) presented in Figure 5 was first assessed using an ANOVA with extract and antagonist treatment as the independent variables and locomotor activity (mean distance moved [cm] in 60 min) as the dependent variable using Statview (SAS Institute, Inc; version 5.0.1). If there was a significant interaction between extract and antagonist treatment, step-down ANOVAs were done to assess which treatment groups were different from one another. Significance was set at  $P \leq .05$ . All data are presented as mean  $\pm$  SEM. The number of independent observations are provided in the figure legends.

## RESULTS

### Locomotor Responses to Negative and Positive Controls

The locomotor responses to 0.4% DMSO were not different from the responses to HBSS (Supplementary Figure 1A). Likewise, responses were not different among exposures to 0.4% or 0.8% DMSO (Supplementary Figure 1B), indicating that various vehicles do not perturb activity on their own.

Administration of acrolein caused an immediate statistically significant dose-dependent hyperactive locomotor response in zebrafish during the first 10 min of treatment ( $P < .05$ ; Figure 2) with no effect of the lower concentrations (ie, 1 and 3  $\mu\text{M}$ ) during

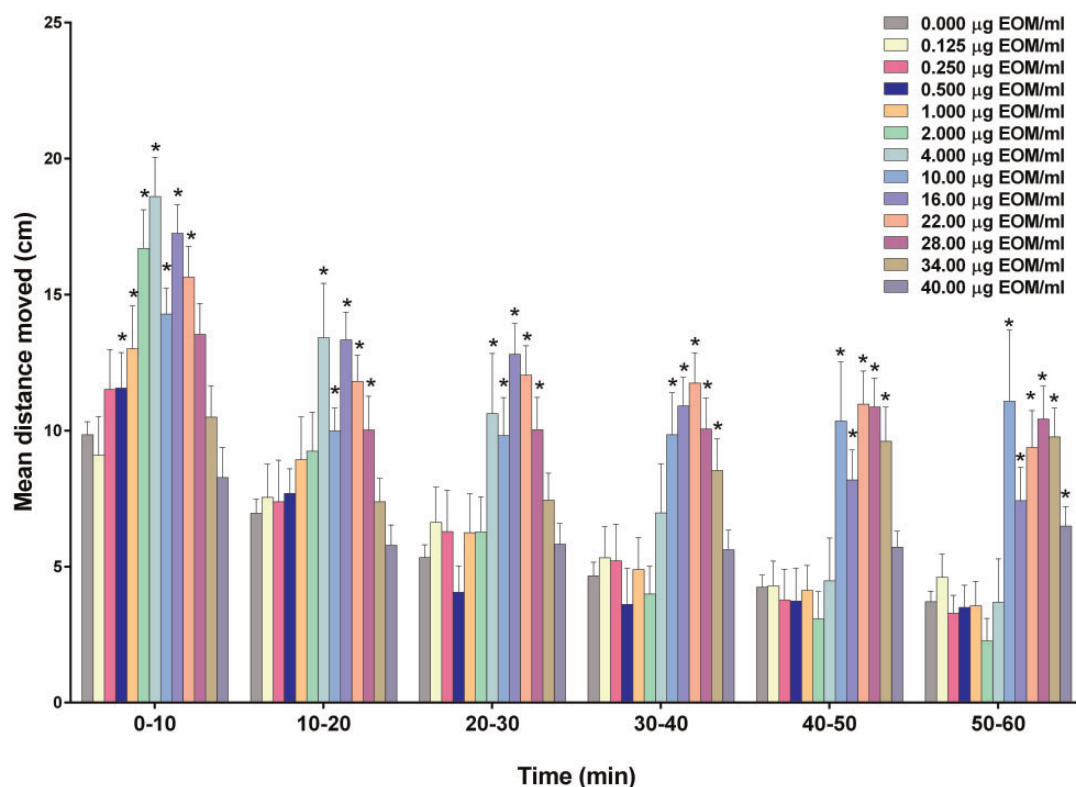


Figure 3. Effect of compressor-generated diesel exhaust particulate matter (C-DEP) on zebrafish larval locomotor activity measured in the dark. Zebrafish were exposed to 0.125–40 µg/ml C-DEP in 0.4% DMSO or 0.4% DMSO alone approximately 5 min before monitoring locomotor behavior. Data are presented as mean ± SEM for the entire 60 min of recording. The wide concentration-range necessitated the splitting of groups over 2 plates (0.0125–4 µg/ml on 1 plate and 10–40 µg/ml on the second plate); each plate was done in duplicate. Data from 2 replicate plates were combined for a total of 15–40 zebrafish per exposure group. \*Significantly greater than vehicle alone at corresponding time ( $P < .05$ ).

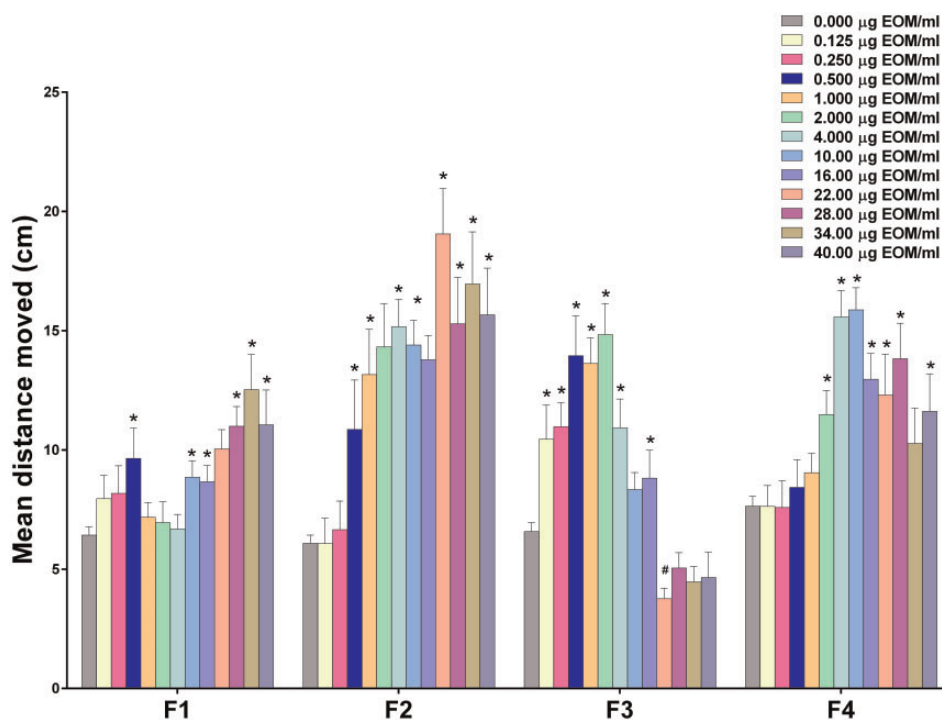


Figure 4. Effects of C-DEP fractions 1 (F1), 2 (F2), 3 (F3), and 4 (F4) on zebrafish larval locomotor activity measured in the dark during the first 10 min of recording. Zebrafish were exposed to 0.125–40 µg/ml of fractions 1–4 in 0.4% DMSO or 0.4% DMSO alone approximately 5 min before monitoring locomotor behavior. Data are presented as mean ± SEM. The wide concentration-range necessitated splitting the groups between 2 plates (0.0125–4 µg/ml on 1 plate and 10–40 µg/ml on the second plate), which were each done in duplicate. Data from 2 replicate plates were combined for a total of 15–32 zebrafish per treatment group. \*Significantly greater than vehicle alone ( $P < .05$ ). #Significantly less than vehicle alone ( $P < .05$ ). Note: each fraction was run on separate plates with respective vehicle controls.

this period. However, during the remaining 50 min of treatment, locomotor activity in the zebrafish treated with 3  $\mu\text{M}$  became significantly greater than control ( $P < .05$ ), while zebrafish activity in the group treated with 30  $\mu\text{M}$  was no longer significantly greater than control. In addition, zebrafish treated with 10  $\mu\text{M}$  acrolein had significantly greater activity than control throughout the first 50 min after treatment ( $P < .05$ ) but were not significantly different from control during the final 10 min of treatment. Interestingly, the locomotor activity in the zebrafish treated with 1  $\mu\text{M}$  acrolein became significantly greater than control only during the last 20 min of monitoring ( $P < .05$ ), ie, 40 min after the beginning of treatment.

#### Locomotor Responses to Whole C-DEP

Treatment of zebrafish with C-DEP induced a concentration-dependent response that peaked soon after treatment (Figure 3), with significant increases in the 0.5–22  $\mu\text{g}/\text{ml}$  treatment groups during the first 10 min of exposure ( $P < .05$ ). Interestingly, there were no significant effects on locomotor activity with the 2 highest concentrations (34 and 40  $\mu\text{g}/\text{ml}$ ) during the first 30 min of exposure. During the final 30 min of exposure ( $P < .05$ ), however, zebrafish treated with C-DEP at 34  $\mu\text{g}/\text{ml}$  had significantly greater locomotor activity, whereas fish treated with 40  $\mu\text{g}/\text{ml}$  had significantly greater locomotor activity only during the final 10 min of exposure ( $P < .05$ ). C-DEP at 10, 16, 22, and 28  $\mu\text{g}/\text{ml}$  induced hyperactivity during the entire 60 min of exposure. Locomotor behavior in response to acrolein and select

C-DEP concentrations are shown in Supplementary Figures 1C and 1D to illustrate changes by 2-min increments.

#### Locomotor Responses to C-DEP Fractions

Treatment of zebrafish with each of the 4 fractions yielded concentration-dependent hyperactivity responses that varied considerably in activity levels (Figure 4). Zebrafish treated with F1, containing nonpolar constituents, yielded only mild responses that were largely restricted to the higher concentrations (10  $\mu\text{g}/\text{ml}$  or greater) during the first 10 min of exposure ( $P < .05$ ), with little evidence of hyperactivity after the first 10 min. Zebrafish treated with F2 in contrast, showed greater increases in activity at more doses, with significant increases in locomotor activity at concentrations as low as 0.5  $\mu\text{g}/\text{ml}$  ( $P < .05$ ) and extending through the highest concentration (ie, 40  $\mu\text{g}/\text{ml}$ ). Significantly increased activity was evident during the entire testing period for F2 concentrations ranging from 2 to 40  $\mu\text{g}/\text{ml}$  ( $P < .05$ ). Zebrafish treated with F3 were also very responsive as evidenced by significant increases in activity compared with control ( $P < .05$ ) with the lowest concentration tested, ie, 0.125  $\mu\text{g}/\text{ml}$ . Interestingly, however, zebrafish treated with F3 concentrations of 22  $\mu\text{g}/\text{ml}$  and greater did not show increased activity at any point during treatment. Zebrafish treated with F4 had peak responses with 4 and 10  $\mu\text{g}/\text{ml}$ , with significant effects occurring at concentrations as low as 1  $\mu\text{g}/\text{ml}$  ( $P < .05$ ). Locomotor activity in zebrafish treated with F4 concentrations at or above 4  $\mu\text{g}/\text{ml}$  was significantly greater than the control group during the entire monitoring period (except for the first 10 min of activity in zebrafish treated with 34  $\mu\text{g}/\text{ml}$ ;  $P < .05$ ).

Linear regressions over the linear portion of the 0- to 10-min concentration-response data for C-DEP and all fractions (Table 2) showed that the potencies (ie, slopes; mean distance moved [ $\text{cm}$ ]/ $\mu\text{g}$  EOM/ml) varied over 2 orders of magnitude (103-fold). Note: linear regressions were performed on data from the first 10 min because this was the most active period with all extracts. The linear range of each curve used to perform the regressions ranged from 0  $\mu\text{g}$  EOM/ml to the highest concentration just before the curve plateaued for each dataset and one that yielded a regression with the highest goodness of fit (ie,  $r^2$  value). The linear ranges were as follows: C-DEP: 0–4  $\mu\text{g}$  EOM/ml, F1: 0–22  $\mu\text{g}$  EOM/ml, F2: 0–2  $\mu\text{g}$  EOM/ml, F3: 0–0.5  $\mu\text{g}$  EOM/ml, and F4: 0–4  $\mu\text{g}$  EOM/ml. The extracts ranked as follows according to their irritant potencies (mean cm moved/ $\mu\text{g}$  EOM/ml): F3 (11.35) > F2 (4.96) > C-DEP (2.43) > F4 (1.79) > F1 (0.11). Thus, the

**Table 2.** Potency of C-DEP and Fractions in the Zebrafish Locomotor Assay

Sample	Potency/Slopes <sup>a</sup>	$r^2$	P-Value*
C-DEP	2.43	0.224	< .0001
F1	0.11	0.036	.0024
F2	4.96 <sup>b</sup>	0.175	< .0001
F3	11.35 <sup>c</sup>	0.169	< .0001
F4	1.79	0.237	< .0001

<sup>a</sup>Data are slopes of linear regressions of dose-response data.

<sup>b</sup>The difference in potency between F2 and F1 approached statistical significance ( $P = .18$ ).

<sup>c</sup>Potency of F3 statistically greater than potency of C-DEP ( $P = .02$ ), F1 ( $P = .01$ ), and F4 ( $P = .02$ ) and nearly F2 ( $P = .07$ ).

\*Indicate P-values of linear regressions.

**Table 3.** Correlations Between Irritant Potencies in Zebrafish and Mutagenic Potencies in Salmonella<sup>a</sup> in Response to the Same Extracts

	C-DEP <sup>b</sup>	F1 <sup>b</sup>	F2 <sup>b</sup>	F3 <sup>b</sup>	F4 <sup>b</sup>	$r^2$ for Fish Potency Versus Mutagenic Potency/Strain	P-Value
Potency in zebrafish ( $\text{cm}/\mu\text{g}$ EOM/ml)	2.43	0.11	4.96	11.35	1.79	—	—
Mutagenic potency in TA100-S9 <sup>c</sup> (Rev./ $\mu\text{g}$ EOM)	13.10	0.00	41.30	60.40	8.00	0.93	.0085
Mutagenic potency in TA98-S9 <sup>d</sup> (Rev./ $\mu\text{g}$ EOM)	7.50	0.00	18.00	23.20	4.30	0.88	.019
Mutagenic potency in TA98NR-S9 <sup>e</sup> (Rev./ $\mu\text{g}$ EOM)	4.70	0.00	12.80	17.50	2.40	0.90	.013
Mutagenic potency in YG1041-S9 <sup>f</sup> (Rev./ $\mu\text{g}$ EOM)	67.90	0.00	192.50	236.80	28.80	0.85	.025

<sup>a</sup>Mutagenicity potency values were obtained from Mutlu et al. (2013).

<sup>b</sup>Data are slopes of linear regressions of dose-response data.

<sup>c</sup>Sensitive to non-PAH base-substitution mutagens.

<sup>d</sup>Sensitive to nitroarenes.

<sup>e</sup>Sensitive to frameshift mutagens other than nitroarenes.

<sup>f</sup>Sensitive to nitroarenes.

S9 is a supernatant fraction of a rat liver homogenate containing metabolizing enzymes; it is added to mimic mammalian metabolism of test substances since some mutagens require metabolism.

Rev., revertants.

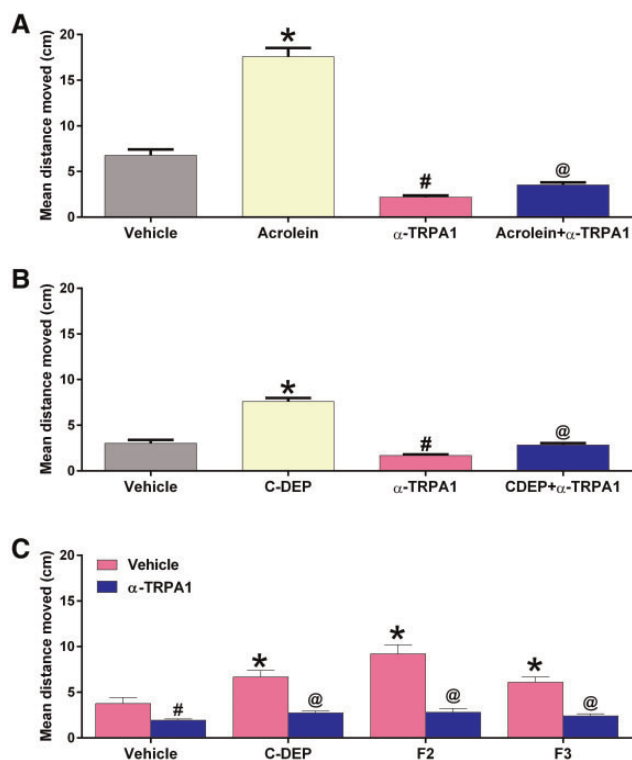


Figure 5. Effect of anti-TRPA1 pretreatment on zebrafish larval locomotor responses to acrolein (10  $\mu$ M) (A), C-DEP (22  $\mu$ g/ml) (B), and fractions F2 (10  $\mu$ g/ml) and F3 (1  $\mu$ g/ml) (C). Zebrafish were treated with 50  $\mu$ M of the TRPA1 antagonist HC030031 beginning 15 min before treatment with each diesel extract. Data are presented as mean  $\pm$  SEM for the entire 60-min period of recording. Data from 2 replicate plates were combined for a total of 24–48 zebrafish per exposure. vehicle = 0.8% DMSO. \*Significantly greater than vehicle alone ( $P < .05$ ). #Significantly less than vehicle alone. @Significantly less than corresponding chemical/extract-treated group without antagonist ( $P < .05$ ).

fractions containing compounds that were weakly polar and polar (F2 and F3, respectively) elicited the most activity, whereas fractions at the extreme of polarity caused the least activity (F1 is nonpolar and F4 is highly polar).

#### Effects of TRPA1 Antagonism on Locomotor Responses to Acrolein and Extracts

We assessed the ability of a TRPA1 antagonist to block irritant responses to single concentrations of acrolein, C-DEP, and the 2 most potent fractions, F2 and F3 (we selected concentrations of acrolein and diesel extracts that elicited robust responses during the entirety of the monitoring periods in the concentration-response studies for the antagonism studies). Although the antagonist caused a significant reduction in locomotor activity relative to the DMSO control, it also caused a significant reduction in acrolein-induced locomotor responses relative to the DMSO control during the 60-min monitoring period ( $P < .05$ ; Figure 5A). The antagonist also significantly reduced the locomotor responses induced by C-DEP and F2 and F3 ( $P < .05$ ; Figs. 5B and 5C, respectively). Figure 5 shows the average values for the entire 60-min monitoring period.

## DISCUSSION

Zebrafish treated with the positive control acrolein had both conventional and unconventional concentration-response patterns. Within the first 10 min of treatment, acrolein elicited a

prototypical concentration-dependent increase in locomotor activity, consistent with previous studies (Prober *et al.*, 2008). This immediate increase in activity was likely due to the irritant effects of acrolein at the superficial skin surface. Acrolein is a potent environmental pollutant with well-documented irritant effects in the respiratory tract (Bascom, 1991; Moretto *et al.*, 2012; Morris *et al.*, 1999). With time after treatment, however, the effects of higher doses diminished while that of lower doses increased. The nature of the mechanisms driving this divergence is unclear but may demonstrate the limits of this assay, wherein high doses elicit a lethargy-like state, perhaps due to overt toxicity or receptor downregulation from overstimulation.

The whole organic extract of C-DEP caused concentration-dependent hyperactive locomotor responses in zebrafish larvae that plateaued at intermediate concentrations. Except for effects during the latter part of the assay, C-DEP at higher concentrations, did not elicit hyperactivity, similar to the responses with high concentrations of acrolein and potentially owing to overt toxicity. These findings highlight the sensitivity of the assay and point to potential utility in discerning effects at PM concentrations in the ng range. These responses were immediate and consistent with well-documented PM-induced respiratory irritant responses in humans (Xu *et al.*, 2013) and analogous responses in rodents (Farraj *et al.*, 2011; Filep *et al.*, 2016; Hemmälä *et al.*, 2010; Sussan *et al.*, 2014). Although responses decreased in magnitude by the end of the assessment, C-DEP's effects remained elevated above controls, likely due to the continued presence of PM extract in the wells. The greater activity with high C-DEP concentrations late in the assay may have been due to a diminution of overt effects or a failure to habituate (a primitive form of learning and memory) because zebrafish treated with lower concentrations became hypoactive. Attenuation of habituation is an endpoint that has been noted in adult zebrafish treated with drugs (Wong *et al.*, 2010). Given the similarity in responses to PM across species, and homologous sensory and inflammatory effects, zebrafish larval skin epithelial responses may be predictive of mammalian lung epithelial responses as previously proposed (McLeish *et al.*, 2010).

C-DEP hyperactivity responses were likely driven by the presence of weakly polar and polar components as indicated by a comparison of responses to the fractions and whole C-DEP on an equal mass basis. F3, which was previously determined to have the highest concentration of polar nitro-PAHs and oxy-PAHs (Mutlu *et al.*, 2013) of any of the fractions, caused the most activity in zebrafish. F2, which contained the highest concentration of weakly polar nitro-PAHs, was next most active. Similar locomotor responses were reported with exposure to PAHs found in oil-contaminated aquatic ecosystems (Vignet *et al.*, 2014a,b). Oxy-PAHs have also been linked to developmental toxicity in zebrafish (Wincent *et al.*, 2015). Furthermore, fine PM-induced increases in metabolizing enzymes (Olivares *et al.*, 2013) and developmental effects (Mesquita *et al.*, 2016) in zebrafish strongly correlated with PAH content. By contrast, the relative inactivity of F1 was likely due to a predominance of inert nonpolar unburned fuel components (ie, alkanes and alkenes). The reasons for reduced activity with F4, which contained highly polar compounds and some oxy-PAHs, are unclear and may relate to surface reactivity of such compounds and/or their capacity for penetration through hydrophobic zebrafish skin. Overall, these findings are striking in that weakly polar and polar components that account for only approximately 20% of the total EOM of C-DEP (approximately 52% and 28% of the mass eluted in F1 and F4, respectively) were responsible for most of the observed responses with C-DEP. Although PAH constituents

are known drivers of PM-induced cytotoxicity and oxidative stress (Baulig et al., 2003; Li et al., 2002), non-PAH constituents (eg, quinones) may have contributed to the observed irritant responses given that PAHs were only a small fraction (0.3%) of the total EOM of a similar diesel particulate (Stevens et al., 2009).

Because the diesel-particle extracts used in this study have been previously evaluated for mutagenicity in strains of *Salmonella* that are sensitive to different chemical classes (Mutlu et al., 2013), we examined if the potencies of the extracts correlated between the *Salmonella* and zebrafish assays to provide further insight on the potential offending chemical classes that drove irritant responses in zebrafish. Among the 16 *Salmonella* strain/S9 combinations for which mutagenic potencies were determined (Mutlu et al., 2013), we found highly significant correlations between mutagenic potencies and fish irritant potencies for 4 strain/S9 combinations (Table 3). Two of these (TA98-S9 and YG1041-S9) respond to nitro-PAHs (nitroarenes), supporting the fact that F2 and F3, which are enriched in nitro-PAHs, were the most potent fractions in both assays. In addition, the mutagenic potencies of the extracts in strain TA100-S9 and TA98NR-S9, which detect base-substitution mutagens and frameshift mutagens that are not nitroarenes, respectively, also correlated with the irritant potencies of the extracts in the fish (Table 3). The precise identity of the chemicals that drove responses in zebrafish, however, require further study.

C-DEP hyperactivity responses were inhibited with an antagonist of TRPA1, a cation channel found on pain-sensing C-fiber neurons that is activated by chemical modification of N-terminal cysteinyl sulfhydryl groups (Takahashi and Mori, 2011). TRPA1 senses exogenous (eg, cinnamaldehyde) and endogenous (eg, inflammatory mediator 4-hydroxynonenal) chemicals (Takahashi and Mori, 2011). Importantly, DEP is pro-oxidant, and the same C-DEP used in this study activated proinflammatory signaling in human epithelial cells in an oxidative stress-dependent manner (Cao et al., 2007), providing a plausible mechanism for activation of TRPA1 in zebrafish given that zebrafish larvae express TRPA1 as early as 30-h postfertilization (Prober et al., 2008). Furthermore, whole diesel exhaust inhalation in rodents caused cardiac responses in part through TRPA1 activation (Hazari et al., 2011) and diesel particulates and diesel particulate PAHs activated TRPA1 in HEK293 cells transfected with human TRPA1 and in dorsal root ganglion neurons (Deering-Rice et al., 2011). Although other receptors and mechanisms (eg, inflammation) may play a role, the immediacy of the responses and the antagonism of effects suggest that C-DEP locomotor responses in zebrafish larvae were mediated by TRPA1.

Although behavior responses elicited after PM treatment were likely mediated by sensory pathways triggered in zebrafish skin, this study did not examine the role internalization of PM components played, particularly in the elicitation of responses recorded in the latter parts of the assay. Although there was no evidence of lethality, many organic components of PM, particularly hydrophobic chemicals such as PAHs, are absorbed by zebrafish (Kühnert et al., 2013), and once internalized are likely to trigger inflammatory responses (Bai and van Eeden, 2013). Future studies should include assessments of the amount of absorbed PM components in zebrafish and measures of inflammation. Similarly, although there were no differences in locomotor responses among various DMSO concentrations, the higher concentration of DMSO used in the antagonism studies may have influenced chemical uptake and should be examined in future studies.

Although this study examined the irritant potential of a diesel exhaust particulate, this zebrafish behavior assay is likely suitable for the assessment of other organic-rich PM sources and those dominated by water-soluble fractions containing metals. High sensitivity and reproducibility, correlation of results with other established assays, and assay simplicity make this model ideal for screening approaches requiring greater throughput. Taken together, zebrafish behavior assessments may help expedite toxicity determinations of PM sources, uncover mechanisms of action, and when coupled with chemical fractionation, identify causal chemical classes.

## SUPPLEMENTARY DATA

Supplementary data are available at *Toxicological Sciences* online.

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