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## Hydroxylated and Sulfated Metabolites of Commonly Observed Airborne Polychlorinated Biphenyls Display Selective Uptake and Toxicity in N27, SH-SY5Y, and HepG2 Cells

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### Abstract

Although neurotoxicity and hepatotoxicity have long been associated with exposure to polychlorinated biphenyls (PCBs), less is known about the selective toxicity of those hydroxylated PCBs (OH-PCBs) and PCB sulfates that are metabolites derived from exposure to PCBs found in indoor air. We have examined the toxicity of OH-PCBs and PCB sulfates derived from PCBs 3, 8, 11, and 52 in two neural cell lines (N27 and SH-SY5Y) and an hepatic cell line (HepG2). With the exception of a similar toxicity seen for N27 cells exposed to either OH-PCB 52 or PCB 52 sulfate, these OH-PCBs were more toxic to all three cell-types than their corresponding PCB or PCB sulfate congeners. Differences in the distribution of individual OH-PCB and PCB sulfate congeners between the cells and media, and the ability of cells to interconvert PCB sulfates and OH-PCBs, were important components of cellular sensitivity to these toxicants.

### Keywords

PCB; hydroxy PCB; OH-PCB; PCB sulfate; neurotoxicity; hepatotoxicity

## 1. Introduction

The man-made environmental contaminants known as polychlorinated biphenyls (PCBs) continue to persist in the environment and cause or contribute to various harmful health effects including neurotoxicity (Schantz et al. 2003; Tilson and Kodavanti 1997, 1998) and

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#### Conflict of Interest

The authors state that there are no conflicts of interest.

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hepatotoxicity (Cave et al. 2010). Although global levels of PCBs have decreased since their production and use have been limited worldwide (Nyberg et al. 2014), the lower-chlorinated PCBs (i.e., 4 chlorine atoms per congener) are found in environmental samples (Basu et al. 2009; Hu et al. 2011; Rodenburg et al. 2010) including indoor and outdoor air from both urban and rural areas (Egsmose et al. 2016; Pedersen et al. 2016) and as unintended byproducts from current production of consumer products such as pigments and dyes (Hu and Hornbuckle 2010; Shang et al. 2014; Vorkamp 2016). Exposure to these semi-volatile compounds has been proposed to occur through multiple pathways that include inhalation (Hu et al. 2014) as well as the consumption of contaminated food or water (Ampleman et al. 2015; Chen et al. 2017).

Lower-chlorinated PCBs are highly susceptible to metabolic transformation and have often been considered transient species in the body. This metabolic vulnerability, however, also carries with it the potential for production of reactive and toxic compounds (Grimm et al. 2015b; Hansen 2001; Sethi et al. 2017). Hydroxylated PCB metabolites (OH-PCBs) have been detected in human blood samples and in biological samples from various species (Gutleb et al. 2010; Koh et al. 2016; Marek et al. 2014; Schafer et al. 2009). The oxidation of PCBs to hydroxylated metabolites allows for further metabolism, of which sulfation represents a potentially important route (Grimm et al., 2015b). Although sulfation of a phenolic compound is traditionally considered a mode for its removal from the body due to increased polarity, water solubility, and excretion of the sulfated product, the potential for biological retention may also exist. A study in Sprague-Dawley rats has indicated that hydroxylation followed by sulfation accounts for more than half of the metabolic fate after exposure to the monochlorinated PCB 3 (Dhakal et al. 2012). Additional studies in rats following administration of 4-PCB 11 sulfate indicated that some PCB sulfates, however, may be retained *in vivo* (Grimm et al. 2015a). Furthermore, the presence of 4-PCB 11 sulfate in human serum samples has recently been reported (Grimm et al. 2017). *In vitro* studies have shown that, while OH-PCBs can inhibit the sulfation of endogenous molecules including dehydroepiandrosterone (DHEA) and estradiol, many OH-PCBs also serve as substrates for sulfate conjugation (Ekuase et al. 2011; Kester et al. 2002; Liu et al. 2006; Parker et al. 2018). The resulting PCB sulfates bind to the thyroid hormone carrying protein transthyretin, where, in some cases, they bind with similar affinity to that observed with thyroxine (Grimm et al. 2013). Moreover, PCB sulfates have been shown to bind with high affinity to the major drug binding sites of human serum albumin (HSA), the most abundant protein in human plasma (Rodriguez et al. 2016). In the case of both transthyretin and HSA, protein binding was influenced by the degree of chlorination and the substitution patterns of the PCB congeners, and PCB sulfates generally bound with a higher or equal affinity than the corresponding PCB or OH-PCB, thereby potentially increasing their retention and distribution in the body. These studies suggest that, contrary to the general assumption that sulfation of a xenobiotic is simply a mode for its excretion, the sulfates derived from lower-chlorinated OH-PCBs may be retained, transported, and have distinct biological and/or toxicological activities.

While little is known about the *in vivo* toxic effects of PCB metabolites, the neurotoxic and hepatotoxic effects of various PCB mixtures and individual congeners have been well documented in the scientific literature. Exposure to PCBs has been associated with non-

alcoholic fatty liver (Cave et al. 2010), and PCBs have been identified as promoting and initiating agents in hepatic carcinogenesis (Ludewig et al., 2008). Epidemiological studies on the neurotoxic effects of PCB exposure indicate correlations with neurodevelopmental dysfunction and with incidences of neurodegenerative diseases (Hatcher-Martin et al. 2012; Steenland et al. 2006). Environmental PCB exposure-related effects on mood, depression, social and reproductive behaviors, cognition and motor function have also been reported (Berghuis et al. 2015; Berghuis et al. 2013; Jurewicz et al. 2013; Polanska et al. 2013). *In vitro* studies using cultured neuronal cells have often focused on the cytotoxic effects of higher-chlorinated PCB congeners and Aroclor mixtures (Tilson and Kodavanti 1997; Tilson et al. 1998). Lower-chlorinated PCBs are, however, of increasing interest, as seen in the recent study of the effect of PCB 11 and its hydroxylated and sulfated metabolites on axonal and dendritic growth in cultured primary rat neuronal cells (Sethi et al. 2017).

We hypothesized that OH-PCB and corresponding PCB sulfate metabolites of lower-chlorinated PCBs exhibit toxicity in cultured cells that is influenced by PCB congener, metabolite, and target cell type. The cell lines used in this study were of neural (rat midbrain N27 and human neuroblastoma SH-SY5Y) and hepatic (human hepatic HepG2) origins. Cellular toxicity was measured using two orthogonal cell viability assays (i.e., the reduction of MTT and the release of lactate dehydrogenase (LDH)). The PCBs and PCB metabolites included in this study (Figure 1) were chosen to represent some of the most frequently detected PCB congeners in air samples and encompass varying degrees of chlorination and substitution patterns (Grimm et al. 2015b). Moreover, since the presence of *ortho*-substituted chlorines among PCB congeners can dictate their three dimensional structure by influencing the dihedral angle between the phenyl rings, with significant influences on their biological effects (Van den Berg et al. 2006), we have examined both *ortho*-substituted and non *ortho*-substituted congeners. The fate of these molecules *in vitro* was monitored by HPLC to determine their distribution between cells and extracellular medium. Finally, to determine the effects of albumin-binding on cytotoxicity, studies were performed with HSA supplementation in the incubation media.

The studies presented here probe the roles that metabolism of lower-chlorinated PCBs, particularly hydroxylation and subsequent sulfation, may play in the toxic effects of certain PCB congeners. These changes impart complex differences regarding toxicity profiles, distribution of the metabolites into cells from different tissues, as well as their potential for further metabolic reactions within those cells that influence toxic outcomes.

## 2. Materials and Methods

Cell culture media and media components that were obtained from Gibco (Life Technologies, Madison, WI, USA) included: Roswell Park Memorial Institute (RPMI) medium, Dulbecco's Modified Eagle's Medium (DMEM), Opti-MEM, Dulbecco's phosphate buffered saline (DPBS), Trypsin -EDTA (0.25%), penicillin/streptomycin, sodium pyruvate (100mM), fetal bovine serum (FBS), horse serum (HS), and MEM non-essential amino acids (MEM NEAA). Corning Falcon tissue culture 100 mm<sup>2</sup> petri dishes, Corning Costar 96-well plates, and dimethyl sulfoxide (DMSO, 99.9%) were purchased from Fisher Scientific (Radnor, PA, USA). Collagen Type I, rat tail, was purchased from BD Biosciences

(San Jose, CA, USA). Human serum albumin (HSA, fatty acid and globulin free, 99%), 3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide (MTT, 98%), and NADH (99% by HPLC) were purchased from Sigma Aldrich (St. Louis, MO, USA). Authenticity of the human cell lines was confirmed by analysis of genomic DNA conducted by the University of Arizona Genetics Core (Arizona Research Laboratories, Tucson, AZ). Cell culture and incubations were performed under the standard conditions of 37°C in a 5% CO<sub>2</sub> atmosphere. All PCBs and their corresponding hydroxylated and sulfated metabolites were prepared and characterized as previously described (Lehmler et al. 2013; Lehmler and Robertson 2001; Li et al. 2010; Rodriguez et al. 2016). Spectroscopic analyses were performed using a Spectramax M5 fluorimeter (Molecular Devices, Sunnyvale, CA, USA), and statistical analyses and sigmoidal dose response cytotoxicity analyses were obtained using SigmaPlot v. 11.0, Systat Software (Chicago, IL, USA).

### **N27 cells**

Rat midbrain-derived immortalized N27 cells were a generous gift from Dr. Jau-Shyong Hong, Neuropharmacology Group, National Institute of Environmental Health Sciences. N27 cells were seeded at a density of  $1 \times 10^6$  cells in collagen-coated 100 mm<sup>2</sup> tissue culture dishes and maintained in RPMI 1640 medium supplemented with 10% heat-inactivated HS, penicillin (100 I.U./mL) and streptomycin (100 µg/mL), Medium was changed every other day until the cells were near confluence (approximately four days). The N27 cells used in this study were between passages 17 through 30.

### **SH-SY5Y cells**

The human neuroblastoma-derived SH-SY5Y cells were grown in collagen-coated 100 mm<sup>2</sup> tissue culture dishes and maintained in Opti-Mem medium supplemented with 10% heat-inactivated FBS, non-essential amino acids, sodium pyruvate (500 µM), penicillin (100 I.U./mL), and streptomycin (100 µg/mL). Medium was changed every other day until the cells were near confluence (approximately seven days). The SH-SY5Y cells used in this study were between passages 15 through 30.

### **HepG2 cells**

The immortalized human liver-derived HepG2 cells were provided by Ms. Susanne Flor of the University of Iowa Superfund Research Center. HepG2 cells were grown in 100 mm<sup>2</sup> tissue culture dishes and maintained in DMEM supplemented with 10% heat-inactivated FBS, penicillin (100 I.U./mL), and streptomycin (100 µg/mL). Medium was changed every other day until the cells were near confluence (approximately 4 days). The HepG2 cells used in this study were between passages 21 through 32.

**2.1 Effects of PCBs and PCB Metabolites on Cell Viability**—Following culture in 100 mm<sup>2</sup> tissue culture dishes, cells were seeded at  $25 \times 10^3$  cells/well in 96-well plates (plates were collagen-coated for the N27 and SH-SY5Y cells) and allowed to grow for 48 h with a change to fresh medium after 24 h. The cells were then washed twice with 100 µL DPBS, and treated with 100 µL phenol red-free, serum-free medium (prepared respective to cell line-specific medium composition), containing the desired PCB, OH-PCB, or PCB sulfate concentration. The experiments in this study were performed under serum-free

conditions because serum proteins bind with high affinity to PCBs, OH-PCBs, and PCB sulfates (Grimm et al. 2013; Rodriguez et al. 2016), which might alter their effective concentration *in situ*. The solutions of PCBs, OH-PCBs, and PCB sulfates were obtained using serial dilutions of a stock solution (i.e., 100  $\mu\text{M}$  in serum-free medium). The concentrated stock solutions of PCBs and metabolites were made in DMSO at a concentration of 200 mM, and the resulting concentration of DMSO in the media stock solutions following dilution was 0.1% (v/v). The untreated controls in these experiments contained the respective DMSO amounts (i.e., 0.1% v/v), although no change in cell viability was seen at these DMSO concentrations. Each plate contained four sets of three control wells. The cells were incubated for 24 h in the presence of compound. Toxicity was assessed by measuring both the cellular reduction of MTT and the release of LDH.

For the HSA supplementation and toxicity studies, these same conditions were used, however, stock solutions of 50  $\mu\text{M}$  compound were made in serum free media, and the HSA stock solutions were made in the same medium at 100  $\mu\text{M}$  based on a molecular weight of 66.5 kDa. Untreated serum free medium was used to make the final dilution to the desired concentration.

Cellular release of LDH was measured according to a previously described procedure (Vassault 1974). After 24 h of exposure to each concentration of PCB, OH-PCB, or PCB sulfate, 10  $\mu\text{L}$  of a Triton-X 100 solution (20% v/v in serum free medium) was added to the high-control wells (2% v/v final concentration of Triton-X 100 in each well) in the plate, and the plate was incubated for 15 min. Lysis of untreated cells and subsequent complete release of cellular LDH provided the high control for this assay. A 50  $\mu\text{L}$  volume of medium from each well was then transferred to a new 96-well plate. This step was performed carefully so as to not disrupt the cells that were to be assessed using the MTT-reduction assay (see method below). To each 50  $\mu\text{L}$ -containing well in the new 96-well plate was added 50  $\mu\text{L}$  of a DPBS solution containing 200  $\mu\text{M}$  NADH (100  $\mu\text{M}$  final concentration) and 3.2 mM sodium pyruvate (1.6 mM final concentration). The oxidation of NADH was determined by monitoring absorbance at 340 nm, and results were reported as a percent of control catalytic activity of NADH oxidation (i.e., pyruvate reduction catalyzed by total LDH from lysed cells).

The reduction of MTT as a measure of cell viability was determined according to the procedure described by (van Meerloo et al. 2011). After the 24 h exposure of cells to PCBs or PCB metabolites, 50  $\mu\text{L}$  of the medium was carefully removed and placed in a separate 96-well plate for analysis of LDH release (see method above). To the remaining cells in each well, 50  $\mu\text{L}$  of a 1 mg/mL MTT solution in DPBS was added, and they were incubated for 1 h. The plate was removed, 100  $\mu\text{L}$  of acetonitrile was added to each well, and the water-insoluble formazan crystals were dissolved with a multi-pipettor. Visible absorption of each well was then determined at 570 nm and 650 nm. The absorbance at 650 nm was subtracted from that at 570 nm, and the resulting absorbance values were then plotted as a percent of control versus the ligand concentration.

**2.2 Graphical and statistical analyses**—The graphs of the MTT-reduction and LDH-release experiments are shown in the Supplementary Material in Figures S1 through S3.

Results were plotted as the percentage of vehicle-treated control values (for either MTT reduction or LDH release) vs. the log of the concentration of PCB or PCB metabolite. Plotting the data in a logarithmic form allowed for analysis by fitting to a sigmoidal dose response curve from which an effective concentration that afforded half of the total effect ( $EC_{50}$ ) was calculated. For those cases where a sigmoidal dose response model was appropriate, curve-fitting and statistical analyses were performed using SigmaPlot v. 11.0 (Systat Software, Chicago, IL). Statistical significance of differences in  $EC_{50}$  values was determined by one way ANOVA with Bonferoni post hoc analysis.

### 2.3 Analysis of OH-PCBs and PCB sulfates in cellular and extracellular compartments

Cells were grown, seeded, and treated with solutions containing 25  $\mu\text{M}$  of the desired test compound as described in the previous sections. After 24 h, the 96-well plate was centrifuged at  $2300 \times g$  for 5 min. The media from two identically treated wells were carefully removed and pooled in microcentrifuge tubes (0.5-mL high-clarity polypropylene microcentrifuge tubes). The cells were gently washed with 100  $\mu\text{L}$  DPBS, centrifuged at  $2300 \times g$  for 5 min, and the supernatant medium was removed. This washing procedure was repeated once. HPLC analyses indicated that these washes showed no detectable amounts of OH-PCBs or PCB sulfates. One hundred  $\mu\text{L}$  of trypsin solution was added to the cells remaining in each well and the plate was incubated for 30 min at room temperature. The plate was subjected to water bath-sonication for 10 min, cells were dislodged from each well using a multi-pipettor, and quantitative removal of the cells was verified using a light microscope. The contents of two identically treated wells were pooled in a microcentrifuge tube and subjected to sonication (Fisher Scientific Model 100 hand-held sonicator). To each of the extracellular and cellular pooled extracts was added 200  $\mu\text{L}$  acetonitrile and 15 mg NaCl. Each mixture was then thoroughly vortexed and centrifuged at  $4400 \times g$  for 5 minutes. The organic acetonitrile (MeCN) layer was collected and analyzed by HPLC using a Shimadzu Model LC-20-AT liquid chromatograph equipped with an SPD-20-AT UV/VIS detector and a C18 AQ 5 $\mu\text{m}$  (4.6  $\times$  250 mm) column (Grace, Deerfield, IL). Elution of the HPLC column was accomplished with a mobile phase containing 0.04% (v/v) triethylammonium acetate (pH 7.4) and the indicated concentration of MeCN: from 0–1 min, 15% MeCN; from 1–10 min, a linear gradient from 15% to 100% MeCN; from 10–14 min, 100% MeCN; and from 14–15 min, a linear gradient from 100% to 15% MeCN. Analysis of the eluate was carried out by absorbance at 254 nm. Concentrations were determined by relating the peak area to standard curves, ranging from 0.8  $\mu\text{M}$  to 50  $\mu\text{M}$ , obtained using synthetic standards. Standard curves exhibited acceptable linearity with  $r^2$  values ranging between 0.97 and 0.99. The limit of detection (LOD) and limit of quantitation (LOQ) for the compounds were calculated using the relationships 3:1 and 10:1 signal to noise (S/N) ratio, respectively. The S/N ratio was calculated from the standard curve by the relationship:  $\frac{S}{N} = S_y/K$ , where  $S_y$  is the y-intercept standard deviation and K is the slope of the best fit line. The LOD and LOQ for the compounds in this study were: 4'-OH-PCB 3, 4-PCB 52 sulfate (0.02 nmol/100  $\mu\text{L}$  and 0.07 nmol/100  $\mu\text{L}$ ); 4'-PCB 3 sulfate, 4-PCB 8 sulfate, 4-PCB 11 sulfate (0.03 nmol/100  $\mu\text{L}$  and 0.10 nmol/100  $\mu\text{L}$ ); 4-OH-PCB 8 (0.05 nmol/100  $\mu\text{L}$  and 0.17 nmol/100  $\mu\text{L}$ ); 4-OH-PCB 11 (0.06 nmol/100  $\mu\text{L}$  and 0.20 nmol/100  $\mu\text{L}$ ) and 4-OH-PCB 52 (0.01 nmol/100  $\mu\text{L}$  and 0.04 nmol/100  $\mu\text{L}$ ), respectively.

### 3. Results

#### 3.1 Cytotoxicity as assessed by MTT reduction

As shown in Table 1 and Supplemental Material Figures S1 through S3 (panels A), the parent PCBs in this study exhibited little toxicity, within the range of concentrations tested, to the three cell lines as assessed by the reduction of MTT. The di-*ortho*-substituted congener PCB 52, however, showed moderate toxicity to the neural N27 cell line. In fact, in the N27 cells, all PCBs indicated some toxicity within the range of concentrations (i.e., exposures of up to 100  $\mu$ M PCB), however to a lesser extent than PCB 52.

Exposure to OH-PCBs elicited greater toxic responses than the corresponding PCBs in all three cell lines, however to a lesser extent in the hepatic cells. The mono-chlorinated 4'-OH-PCB 3 showed the least potent toxicity throughout, and the *ortho*-chlorinated congeners (4-OH-PCB 8 and 4-OH-PCB 52) presented the highest. The dichlorinated non-*ortho*-substituted 4-OH-PCB 11 exhibited selective toxicity to the non-hepatic cells with significantly higher potency in the N27 cells. Interestingly, the tetra-chlorinated di-*ortho*-substituted 4-OH-PCB 52 showed the highest toxicity in all three cell lines, with the greatest effect in the SH-SY5Y cells. With the exception of the effects seen with 4-OH-PCB 52, the rat N27 cells generally exhibited the highest susceptibility to cytotoxicity following exposure to the OH-PCBs examined in this study.

Similar to the effects observed with PCBs, most of the PCB sulfates showed limited toxicity as determined by the MTT assay. Interestingly, exposure to 4-PCB 52 sulfate elicited a toxic response in all cell lines within the concentration range tested. Furthermore, the N27 cell line exhibited a significantly lower EC<sub>50</sub> value (i.e., higher potency) for 4-PCB 52 sulfate than either the SH-SY5Y or HepG2 cells. Of further note is that, in the N27 cells, 4-PCB 52 sulfate exhibited a toxic response similar to that of some OH-PCBs.

#### 3.2 Cytotoxicity as assessed by LDH release

As seen in Table 2 and Supplementary Material Figures S1 through S3 (panels B), incubation with PCBs resulted in a minimal release of LDH. It was noted, however, that although exposure of N27 cells to PCB 52 at the highest concentration (100  $\mu$ M) indicated a complete loss of cell viability by a lack of mitochondrial reduction of MTT, a complete release of cellular LDH was not observed at this concentration. None of the PCBs exhibited inhibition of LDH catalytic activity (data not shown), indicating that some cell membrane integrity remained, even when the cells were not metabolically active. Therefore, in many cases, minimal changes in LDH release upon exposure to these PCBs precluded the quantitative assessment of toxicity using this method within the range of doses examined.

A comparison of the results of cell viability assays measuring LDH release and MTT reduction following incubation with OH-PCBs yielded similar trends between the two methods. By both methods, 4'-OH-PCB 3 exhibited the lowest toxicity to all cells tested. The assays for LDH release indicated that the di- and tetra-chlorinated OH-PCBs displayed a selective toxicity to the neural cells versus the hepatic cells, with the *ortho*-substituted congeners (i.e., 4-OH PCB 8 and 4-OH PCB 52) exhibiting a greater potency. For example, the di-*ortho*-substituted 4-OH-PCB 52 showed a significantly lower EC<sub>50</sub> value for LDH

release in the N27 cells when compared with either the SH-SY5Y or HepG2 cells. The relative EC<sub>50</sub> values for LDH release following exposure to 4-OH-PCB 52 in the two neural cell lines were reversed from those seen with the MTT assay, however, in both assay methods, selective toxicity was observed for the non-hepatic cells.

With the exception of N27 cells treated with 4-PCB 52 sulfate, cells exposed to PCB sulfates exhibited similar results between the two cell viability methods. For the experiments in which LDH release-EC<sub>50</sub> values could be calculated (i.e., in the SH-SY5Y and HepG2 cells), the data corresponded well with the experimental values observed for MTT reduction.

### 3.3 Supplementation of N27 cells with serum proteins during exposure to PCB metabolites

Since previous studies have shown that OH-PCBs and PCB sulfates bind with high affinity to human serum albumin (HSA), we examined the effect of addition of either HSA or horse serum (HS; a common constituent of growth medium for N27 cells) to the cell culture medium during exposure of N27 cells to OH-PCBs and PCB sulfates. Results of these experiments are shown in Supplementary Material Figures S4 through S7. Incubation of the cells with 25 μM 4-OH-PCB 52 or its corresponding sulfate (Figure S4) resulted in a significant decrease in cell viability that correlated with the results seen in Table 2. This concentration of 25 μM for PCB metabolites was chosen to facilitate comparison across all compounds in the N27 cells (i.e., it includes both toxic and non-toxic responses). The concentrations of HSA were chosen to represent a 1:2, 1:1, 2:1 molar ratio of metabolite:HSA, and the HS supplementation was chosen to be in a range to mimic the media composition of serum (i.e., 2%–10%) in previously published PCB-exposure studies (Angus and Contreras 1996; Costa et al. 2007; De et al. 2010; Lee and Opanashuk 2004; Shimokawa et al. 2006). These results, as well as those for the remaining PCB metabolites included in this study (Figures S5 through S6), indicate that even at a 2:1-metabolite:HSA ratio, the protective effects of albumin-binding were sufficient to decrease toxicity substantially to a level that was statistically indistinguishable from untreated controls.

### 3.4 The disposition of OH-PCB and PCB sulfates between media and cells

In the analysis of our results on the toxicity of OH-PCB and PCB sulfate metabolites, the question arose as to the extent to which these metabolites are taken up into the cells and/or further metabolized within those cells. To address this question, each cell-type was treated for 24 h with 25 μM (2.5 nmol in 100 μL of serum-free medium) of the OH-PCB or PCB sulfate. The cellular and extracellular compartments were analyzed by HPLC, and the distributions of these metabolites between the cells and the extracellular media are presented in Figures 2 through 5, with the treatment compound denoted with an arrow.

The results for 4'-OH-PCB 3 and 4-OH-PCB 11 (i.e., the hydroxylated metabolites without *ortho*-chlorine substitutions) are shown in Figures 2 and 3, respectively. A comparison across cell lines indicates that, after 24 h, there was a distribution of 4'-OH-PCB 3 into all three cell-types, however, the cell lines differed in the amount of OH-PCB found within the cells and the total amount recovered in the two compartments. For N27 and SH-SY5Y cells, 4'-OH-PCB 3 and 4-OH-PCB 8 were found in the cell lysate, although less of these two



OH-PCBs were detected in the SH-SY5Y cell lysates. While exposure of the HepG2 cells to either 4'-OH-PCB 3 or 4-OH-PCB 11 resulted in the sulfation of these hydroxylated congeners and further redistribution of the PCB sulfate into the medium, there was no evidence of sulfation of these two OH-PCBs in either neural cell line. Exposure of N27 and SH-SY5Y cells to either 4'-PCB 3 sulfate or 4-PCB 11 sulfate indicated that these sulfated congeners did not enter the neural cells, given their near quantitative recovery in the media. However, with the HepG2 cells, exposure to 4'-PCB 3 sulfate or 4-PCB 11 sulfate resulted in its cellular uptake and distribution between the cellular and extracellular compartments. Moreover, the corresponding hydroxylated derivatives were also detected in these cells. These OH-PCBs are most likely present due to hydrolysis of the sulfates catalyzed by sulfatases within the HepG2 cells, since control experiments showed no spontaneous sulfate-hydrolysis in the medium under the experimental conditions. In those cases where there was a lack of quantitative recovery of the OH-PCB and PCB sulfate, it is likely that further metabolism may account for this discrepancy. Although binding of the OH-PCB and PCB sulfate to the plastic surfaces of the culture plates was considered, this was unlikely due to the near quantitative recovery of several OH-PCB and PCB sulfate congeners from some cell-types under identical extraction conditions.

Exposure of the three cell-types to the metabolites of PCBs containing *ortho*-chlorine substitutions (4-OH-PCB 8, 4-OH-PCB 52, 4-PCB 8 sulfate, and 4-PCB 52 sulfate) exhibited qualitative and quantitative differences from the results obtained with the metabolites of PCB 3 and PCB 11. As seen in Figures 4 and 5, 4-OH-PCB 8 distributed well into the N27 cells while relatively lower concentrations were found in the lysate of SH-SY5Y cells exposed to this metabolite. Furthermore, overall recovery of the 4-OH-PCB 8 after incubation was lower in the SH-SY5Y cells. As observed for the sulfated metabolites of PCB 3 and PCB 11, 4-PCB 8 sulfate was not found in the lysate of either N27 or SH-SY5Y cells. The 4-PCB 52 sulfate, however, was taken up by the N27 cells and the presence of 4-OH-PCB 52 in both the medium and cell lysates indicated metabolic hydrolysis of the sulfate and redistribution between the extracellular and intracellular compartments. While both 4-OH-PCB 8 and 4-OH-PCB 52 were taken up by HepG2 cells, only 4-OH-PCB 8 was detected as its corresponding sulfate conjugate. Incubation of HepG2 cells with either 4-PCB 8 sulfate or 4-PCB 52 sulfate resulted in substantial amounts of the corresponding 4-OH-PCB metabolites. This indicated that hydrolysis of the PCB sulfate occurred in these cells.

## 4. Discussion

### 4.1 Selectivity in the toxicities of lower-chlorinated PCBs and their metabolites

Our results on cytotoxicity indicated that the lower-chlorinated OH-PCBs examined were more toxic than their corresponding PCBs or PCB sulfates. Moreover, in most cases, rat and human cells of neural origin exhibited increased toxicity when compared with the hepatic HepG2 cells, and the *ortho*-substituted OH-PCBs elicited the most potent toxic responses. Although the present experiments were not designed to distinguish among specific mechanisms of toxicity, it is noteworthy that OH-PCBs have been shown to be potent inducers of reactive oxygen species (ROS) in both neural and hepatic cells (Dreiem et al.

2009; Spencer et al. 2009; Xu et al. 2015), and this may represent a potential starting point for future studies on the mechanisms of these toxic responses.

Most of the PCBs and PCB sulfates examined here did not exhibit a significant toxic response to any cell line, as determined by either viability assay. The exception in both cases was the di-*ortho*-substituted PCB 52 and its sulfate conjugate, where the latter exhibited a toxic response in N27 cells that was comparable to that observed for most of the OH-PCBs in all cell lines. The toxicity of PCBs with *ortho-chlorine* substituents has been reported (Lilienthal et al. 2014; Tan et al. 2004), and toxic responses to the di-*ortho*-substituted PCB 52 have been proposed to proceed through multiple pathways, such as cell membrane dysfunction, disruption of cellular calcium homeostasis, and covalent binding to proteins and DNA (Lin et al. 2000; Pessah et al. 2010; Tan et al. 2003). A comparison of PCB 52- and PCB 52 sulfate-mediated cytotoxicity in N27 cells (Figure S1) indicates that, under the current experimental conditions, mitochondrial dysfunction may have been more substantial than alteration in cell membrane integrity when considering cellular insult. While this interpretation was supported by the lack of quantitative LDH release in the medium at concentrations where the MTT-reduction capability was essentially eliminated, the exact cause of this difference between the two responses remains to be elucidated.

#### 4.2 Serum supplementation and PCB metabolite toxicities

Upon supplementing N27 cells with either human serum albumin (HSA) or horse serum (HS; commonly used in culture of N27 cells) the PCB-mediated toxic responses were mitigated. Even at molar concentrations of protein that were half those of the PCB derivatives, the toxic response was eliminated. Serum proteins such as HSA bind lower-chlorinated PCBs, OH-PCBs, and PCB sulfates with high affinity (Rodriguez et al. 2016), and their presence would likely decrease the concentration of the free form of each toxicant in the cell culture media (vs. protein-bound which is not available for diffusion or transport across the cell membrane). Many published studies on PCB-related assessments of cellular viability, calcium homeostasis, ROS formation, and other endpoints use serum supplemented media (ranging from 2–10% horse or fetal calf serum, or some mix thereof) throughout the duration of PCB exposure. Although it is unknown the extent to which binding to serum proteins might protect either neural or hepatic cells from exposure to these PCB metabolites *in vivo*, previous studies indicate that this binding may serve as a mode for their retention and distribution (Cutler et al. 1967; Kragh-Hansen et al. 2002). Moreover, our results highlight the observation that *in vitro* studies on PCBs and PCB metabolites may be affected by serum proteins in media components. These results also have implications *in vivo* since, variations in the concentrations of albumin may impart organ-specific susceptibility to PCB-mediated toxicities. For example, the steady state concentration of albumin in cerebrospinal fluid is only approximately 0.5% of that found in plasma (Cutler et al., 1967).

#### 4.3 Selectivity in cellular distribution of PCB metabolites

The further metabolism and distribution of the lower-chlorinated PCB metabolites between the cells and the surrounding medium may contribute to the observed selective toxicity in the neural versus the hepatic cell lines. Using a mass balance approach, our experiments gave some insight into the movement of these PCB metabolites into the cells, as well as their

potential for further metabolism, degradation, or other modifications (e.g., covalently bound to protein). It is important to note that our approach did not allow accounting for all possible metabolites in all cells.

The OH-PCBs that showed the highest toxicity to all three cell lines generally exhibited an appreciable distribution into the cells. The extent of distribution, however, could not be correlated with toxic potency. These experiments were not designed to discern between a lack of uptake into and an efficient efflux out of the cells. Likewise, metabolic reactions other than sulfation or de-sulfation could not be distinguished by this method. Previous results from *in vitro* and *in vivo* studies indicate that oxidized metabolites of lower chlorinated PCBs may become covalently adducted to proteins or DNA (Flor and Ludewig 2010; Ludewig et al. 2008; Zhao et al. 2004). Moreover, the formation of such covalent adducts in the livers and brains of rats treated with PCB 52 have been reported (Lin et al., 2000). Protein/DNA-binding, or other sequestration of metabolites, may explain the lack of quantitative recovery for many of the experiments that, nonetheless, resulted in highly potent toxic responses (e.g., 4-OH-PCB 52 in the SH-SY5Y and HepG2 cells). However, this explanation cannot account for all trends in toxicity such as that seen with exposure to 4-OH-PCB 8, where the N27 cells exhibited higher quantitative recovery, better distribution into the cells, and higher potency when compared to the HepG2 cells.

HepG2 cells have been reported to display multiple drug-metabolizing enzymes and transporters (Guo et al. 2011), and our studies indicated substantial metabolic capacity in the sulfation of OH-PCBs and hydrolysis of PCB sulfates. This capability was seen for all compounds tested, and it may contribute to the reduced toxicity of these OH-PCBs to the HepG2 cells (i.e., due to sulfation or other conjugation reactions, thereby decreasing the intracellular concentration of the OH-PCB). The ability of the HepG2 cells to catalyze hydrolysis of the sulfate conjugates, however, is likely responsible for the toxic response to 4-PCB 52 sulfate associated with the intracellular formation of 4-OH-PCB 52. The N27 cells displayed similar metabolic potential with the 4-PCB 52 sulfate, and this may also explain the toxic effects of this metabolite to these cells, where the relatively high potency of 4-PCB 52 sulfate (as judged by the MTT assay) may be due to the presence of 4-OH-PCB 52, as evidenced by the HPLC analysis.

## 5. Conclusions

The number of studies concerning the cytotoxic potential of individual PCB congeners with lower numbers of chlorine atoms is small in comparison to those studies involving the higher-chlorinated PCBs and Aroclor mixtures. This is certainly true for investigations on PCB metabolites, where previous experiments have primarily focused on hydroxylated PCBs. The results of our present studies suggest that metabolism of lower-chlorinated PCBs by hydroxylation and subsequent sulfation may mediate or influence PCB-related health effects. This was particularly evident in the differential effects on neural cells, indicating a need for further investigation of potential effects on neurological function. The use of cell lines derived from different origins (i.e., midbrain vs. neuroblastoma vs. liver) and from different species (i.e., rat and human) in the present experiments indicate that hydroxylated

and sulfated metabolites derived from PCBs exhibit distinct toxicological profiles that differ by congener and cell-type.

The conclusions from our study of the distribution of OH-PCBs and PCB sulfates in these cells are summarized in Scheme 1. Generally, with all three cell lines, OH-PCBs were able to enter the cells and elicit variable toxic responses. However, only in the hepatic cell line was there evidence of subsequent sulfation, presumably through sulfotransferase-mediated metabolism, which may explain the reduced toxicity in these cells when compared to those of neural origin. The PCB sulfates were also able to enter the HepG2 cells, where, in some cases, their hydrolysis resulted in detectable OH-PCBs. The neural cells were not permeant to all PCB sulfates, but exhibited sulfatase activity for those congeners that were able to enter the cells. Sulfatase-catalyzed formation of the OH-PCBs may explain the toxicity detected upon exposure of N27 cells to 4-PCB 52 sulfate. Thus, our results indicate the need for more information on the transport of PCB sulfates and OH-PCBs into and out of different cell types as well as the congener selectivity in relation to intracellular sulfatases and sulfotransferases.

Although this study does not provide details on the exact mechanism(s) of toxicity for these compounds, it does highlight the need to consider that metabolism may play a key role in the toxic effects associated with PCB-exposure to cells of neural origin. Previous studies using rat hippocampal slices and primary cultured hippocampal neurons have indicated no primary metabolism of parent PCBs, however no such studies with lower-chlorinated PCBs were reported (Wu et al. 2013; Yang et al. 2014). Furthermore, while OH-PCBs have not been identified in human brain samples, they have been reported in brain samples collected from East Greenland polar bears (Gebbink et al. 2008). The CNS may not be a site for the initial oxidative metabolism of lower chlorinated PCBs, however, our studies indicate that it may be a site for further metabolism, and the production of potentially toxic species. In addition to the direct toxic insult of these PCB metabolites, the sulfation of hormones and neurotransmitters is a crucial component of cellular maintenance and regulation in the CNS, and disruption of this component of the neuroendocrine system may have damaging and chronic effects (Coughtrie 2002).

An additional consideration is the fact that the  $EC_{50}$  values reported in this study are generally higher than PCB concentrations commonly reported in an exposed population. A chronic exposure to these compounds may, however, allow for their accumulation, resulting in high localized concentrations that might exhibit similar toxic outcomes. A recent study on the elimination kinetics of PCB 28 metabolites in a PCB-exposed human population indicated half-lives on the time-scale of years, with some of the OH-PCB metabolites showing a longer half-life than the parent PCB (Quinete et al. 2017). Thus, our results on the distinct toxicity profiles of OH-PCBs and PCB sulfates, and the growing evidence that these metabolites may play roles in the dysregulation of hormonal, neurotoxicological, and neurodevelopmental events, indicate the need for more extensive studies on the mechanism(s) of these toxic effects and the dependence upon congener structure and metabolic fate.

## Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

## Acknowledgements

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## Abbreviations

<b>DMEM</b>	Dulbecco's modified eagle's medium
<b>DPBS</b>	Dulbecco's phosphate buffered saline
<b>HS</b>	horse serum
<b>HSA</b>	human serum albumin
<b>LDH</b>	lactate dehydrogenase
<b>MTT</b>	3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyltetrazolium bromide
<b>OH-PCB</b>	hydroxylated polychlorinated biphenyl
<b>4'-OH PCB 3</b>	4-chloro-4'-hydroxybiphenyl
<b>4-OH-PCB 8</b>	2,4'-dichloro-4-hydroxybiphenyl
<b>4-OH-PCB 11</b>	3,3'-dichloro-4-hydroxybiphenyl
<b>4-OH PCB 52</b>	2,2',5,5'-tetrachloro-4-hydroxybiphenyl
<b>PCB</b>	polychlorinated biphenyl
<b>PCB 3</b>	4-chlorobiphenyl
<b>PCB 8</b>	2,4'-dichlorobiphenyl
<b>PCB 11</b>	3,3'-dichlorobiphenyl
<b>PCB 28</b>	2,4,4'-trichlorobiphenyl
<b>PCB 52</b>	2,2',5,5'-tetrachlorobiphenyl
<b>4'-PCB 3 sulfate</b>	4-chloro-4'-biphenylsulfate
<b>4-PCB 8 sulfate</b>	2,4'-dichloro-4-biphenylsulfate
<b>4-PCB 11 sulfate</b>	3,3'-dichloro-4-biphenylsulfate

<b>4-PCB 52 sulfate</b>	2,2',5, 5'-tetrachloro-4-biphenylsulfate
<b>RPMI medium</b>	Roswell Park Memorial Institute medium.

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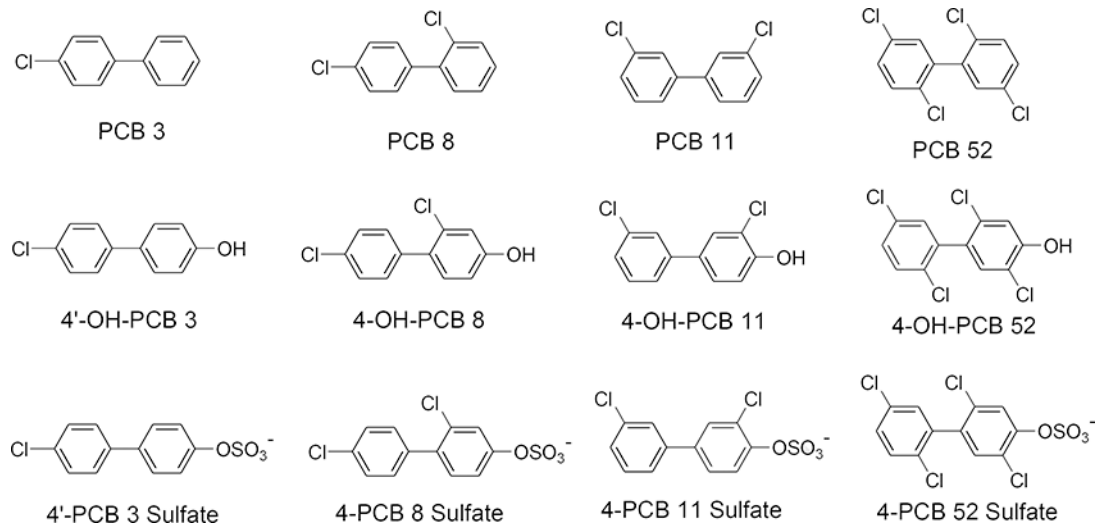
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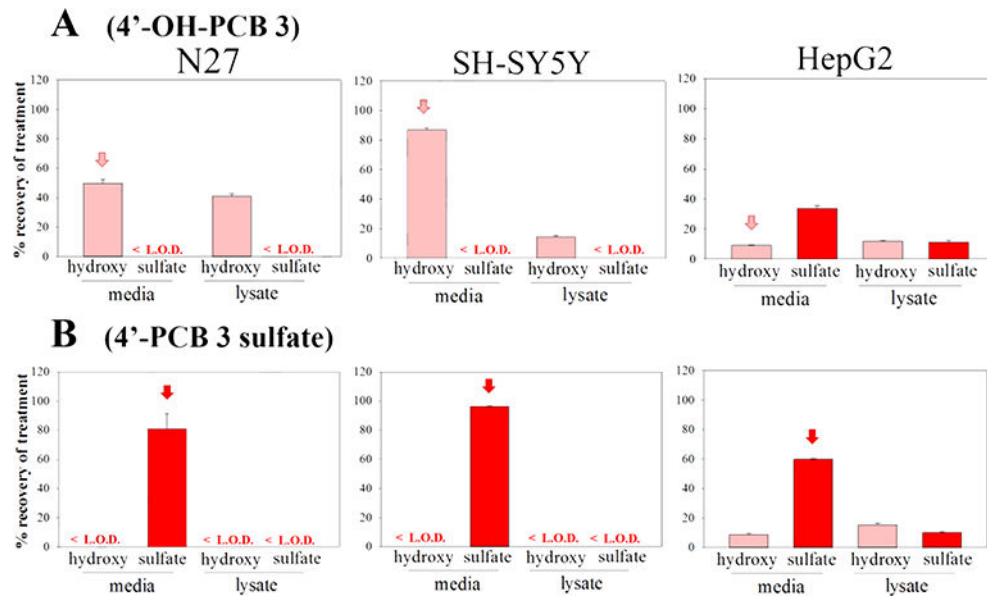
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### Highlights

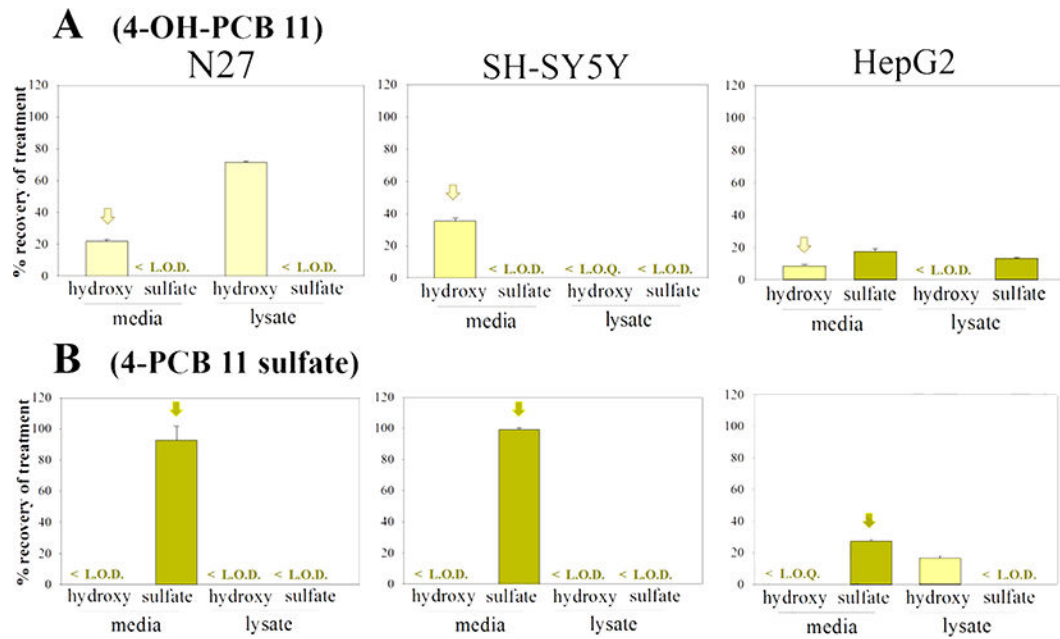
- Phenolic and sulfated metabolites of PCBs 3, 8, 11, and 52 were studied in vitro.
- The OH-PCBs displayed toxicity to N27, SHSY-5Y, and HepG2 cells.
- OH-PCBs were generally more toxic than the corresponding PCBs or PCB sulfates.
- 4-PCB 52 sulfate, however, showed significant toxicity to neural and hepatic cells.
- Mechanisms for 4-PCB 52 sulfate toxicity may require its metabolism to 4-OH-PCB 52.



**Fig. 1.**  
Chemical structures of the PCBs, OH-PCBs, and PCB sulfates used in these studies.

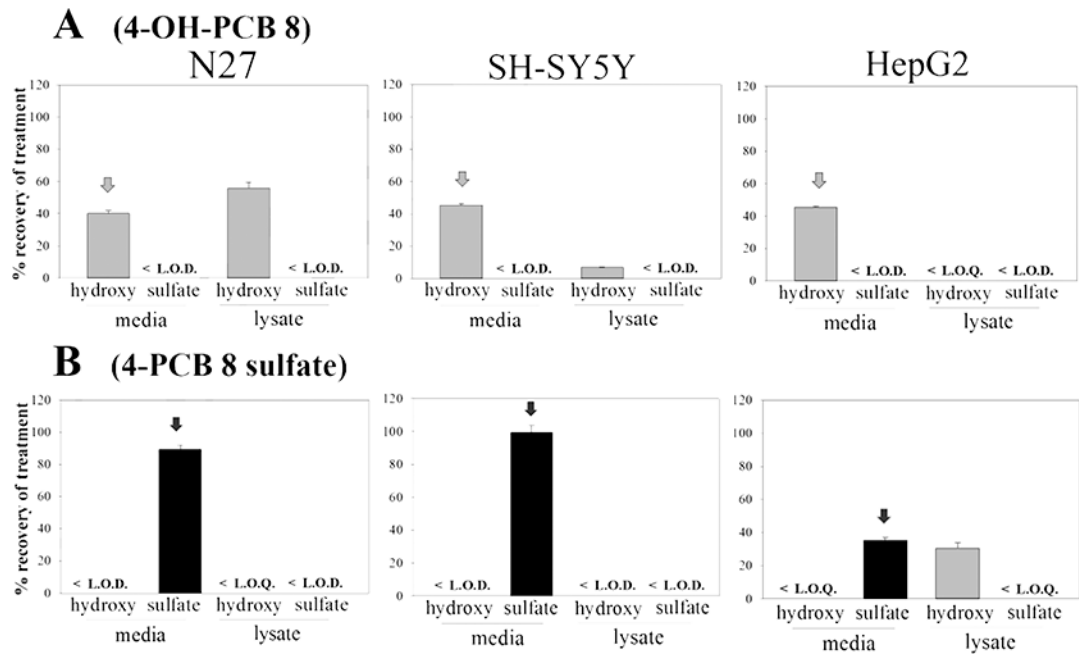
**Fig. 2.**

The distribution of: A) 4'-OH-PCB 3 and B) 4'-PCB 3 sulfate, in N27, SH-SY5Y, and HepG2 cells as determined by HPLC analysis. Cells were treated with 25  $\mu$ M (2.5nmol) compound for 24h, then extracellular media and intracellular contents were collected and analyzed. The treatment compound is annotated with an arrow. The values shown are the means  $\pm$  SE, n=3.

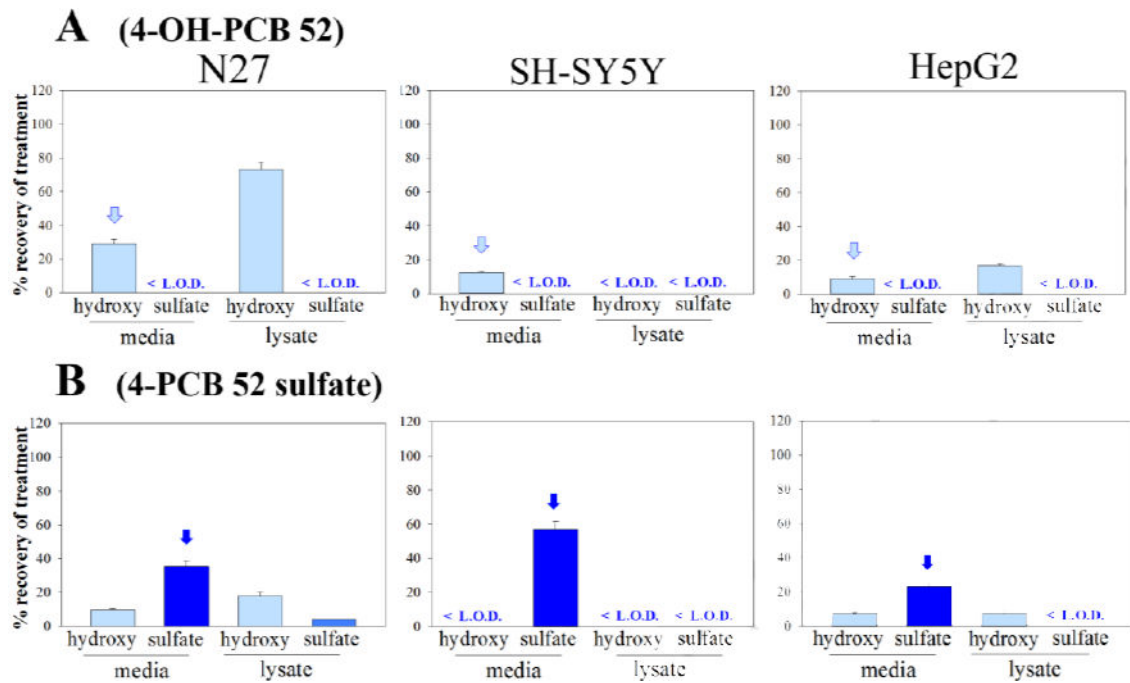


**Fig. 3.**

The distribution of: A) 4-OH-PCB 11 and B) 4-PCB 11 sulfate, in N27, SH-SY5Y, and HepG2 cells as determined by HPLC analysis. Cells were treated with 25 $\mu$ M (2.5nmol) compound for 24h, and subjected to analysis of the extracellular media and intracellular contents. The treatment compound is annotated with an arrow. The values shown are the means  $\pm$  SE, n=3.

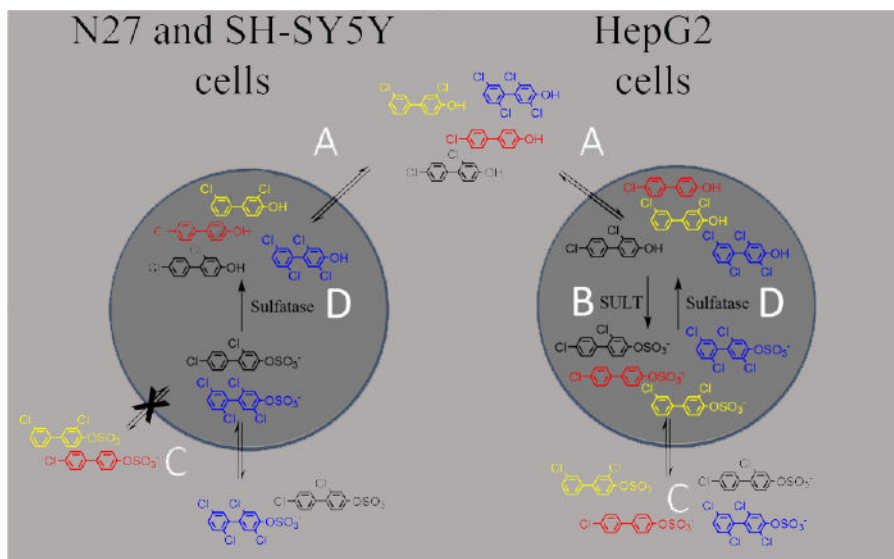
**Fig. 4.**

The distribution of: A) 4-OH-PCB 8 and B) 4-PCB 8 sulfate, in N27, SH-SY5Y, and HepG2 cells as determined by HPLC analysis. Cells were treated with 25 $\mu$ M (2.5nmol) compound for 24h, and subjected to analysis of the extracellular media and intracellular contents. The treatment compound is annotated with an arrow. The values shown are the means  $\pm$  SE, n=3.



**Fig. 5.**

The distribution of: A) 4-OH-PCB 52 and B) 4-PCB 52 sulfate, in N27, SH-SY5Y, and HepG2 cells as determined by HPLC analysis. Cells were treated with 25 $\mu$ M (2.5nmol) compound for 24h, and subjected to analysis of the extracellular media and intracellular contents. The treatment compound is annotated with an arrow. The values shown are the means  $\pm$  SE, n=3.



**Scheme 1.**

A summary of the observed cell-type specific differences in PCB metabolite permeability and potential for further metabolic transformations. Colors indicate OH-PCB and PCB sulfate metabolites of PCB 3 (red), PCB 8 (black), PCB 11 (yellow), and PCB 52 (blue). A) OH-PCBs were able to enter all three cell lines tested, where B) some congeners were susceptible to sulfation in the hepatic HepG2 cells. n=3. C) All PCB sulfates were able to enter HepG2 cells, while uptake in the neural cell lines was more selective. D) Both the HepG2 and N27 cell lines exhibited sulfatase activity for some congeners, thereby resulting in *in situ* production of OH-PCBs.



**Table 1.**

Summary of EC<sub>50</sub> values for viability of N27, SH-SY5Y, and HepG2 cells as determined by the MTT assay following 24 h exposure to PCBs or their metabolites

PCB or Metabolite	N27 EC <sub>50</sub> (μM)	SH-SY5Y EC <sub>50</sub> (μM)	HepG2 EC <sub>50</sub> (μM)
PCB 3	>50	>50	>50
PCB 8	>50	>50	>50
PCB 11	>50	>50	>50
PCB 52	28.5 ± 2.4	>50	>50
4'-OH-PCB 3	31.4 ± 4.2	>50	>50
4-OH-PCB 8	13.8 ± 2.1 <sup>b</sup>	16.2 ± 0.1	27.0 ± 0.7 <sup>b,c</sup>
4-OH-PCB 11	15.2 ± 0.2 <sup>a</sup>	20.7 ± 0.1 <sup>a</sup>	>50
4-OH-PCB 52	7.9 ± 0.1 <sup>a,b</sup>	6.7 ± 0.2 <sup>a,c</sup>	13.0 ± 0.3 <sup>b,c</sup>
4'-PCB 3 sulfate	>50	>50	>50
4-PCB 8 sulfate	>50	>50	>50
4-PCB 11 sulfate	>50	>50	>50
4-PCB 52 sulfate	14.3 ± 0.6 <sup>a,b</sup>	32.7 ± 2.2 <sup>a</sup>	28.8 ± 3.9 <sup>b</sup>

<sup>a</sup>Significant difference between N27 and SH-SY5Y (p<0.05)

<sup>b</sup>Significant difference between N27 and HepG2 (p<0.05)

<sup>c</sup>Significant difference between SH-SY5Y and HepG2 (p<0.05)

**Table 2.**

Summary of EC<sub>50</sub> values for viability of N27, SH-SY5Y, and HepG2 cells as determined by LDH release following 24 h exposure to PCBs or their metabolites

PCB or Metabolite	N27 EC <sub>50</sub> (μM)	SH-SY5Y EC <sub>50</sub> (μM)	HepG2 EC <sub>50</sub> (μM)
PCB 3	>50	>50	>50
PCB 8	>50	>50	>50
PCB 11	>50	>50	>50
PCB 52	>50	>50	>50
4'-OH-PCB 3	>50	>50	>50
4-OH-PCB 8	18.5 ± 0.4 <sup>a,b</sup>	14.0 ± 0.5 <sup>a,c</sup>	25.1 ± 2.6 <sup>b,c</sup>
4-OH-PCB 11	22.2 ± 0.9 <sup>b</sup>	16.8 ± 1.1 <sup>c</sup>	39.8 ± 4.1 <sup>b,c</sup>
4-OH-PCB 52	8.3 ± 0.6 <sup>a,b</sup>	10.2 ± 0.5 <sup>a,c</sup>	15.3 ± 0.5 <sup>b,c</sup>
4'-PCB 3 sulfate	>50	>50	>50
4-PCB 8 sulfate	>50	>50	>50
4-PCB 11 sulfate	>50	>50	>50
4-PCB 52 sulfate	>50	33.3 ± 2.5	29.2 ± 1.3

<sup>a</sup>Significant difference between N27 and SH-SY5Y (p<0.05)

<sup>b</sup>Significant difference between N27 and HepG2 (p<0.05)

<sup>c</sup>Significant difference between SH-SY5Y and HepG2 (p<0.05)