Review Article

The Cardiovascular Effects of Noise

Omar Hahad, Swenja Kröller-Schön, Andreas Daiber, Thomas Münzel

Summary

<u>Background</u>: Traffic noise can induce chronic stress reactions and thereby elevate the risk of cardiovascular disease. The responsible pathophysiological mechanisms are as yet unclear.

<u>Methods</u>: This review is based on publications retrieved by a selective search in PubMed for epidemiological and experimental studies (2007–2018) on the relation between noise and the risk of cardiovascular disease. The search terms were "noise AND cardiovascular effects" and "noise cardiovascular effects."

Results: Epidemiological studies have shown that noise caused by air, road, and rail traffic has a dose-dependent association with elevated cardiovascular morbidity and mortality. A current meta-analysis commissioned by the World Health Organization concludes that road-traffic noise elevates the incidence of coronary heart disease by 8% per 10 dB(A) increase starting at 50 dB(A) (95% confidence interval [1,01; 1,15]). Traffic noise at night causes fragmentation of sleep, elevation of stress hormone levels, and oxidative stress. These factors can promote the development of vascular dysfunction (endothelial dysfunction) and high blood pressure, which, in turn, elevate the cardiovascular risk.

<u>Conclusion</u>: Traffic noise, and air-traffic noise in particular, is an important cardiovascular risk factor that has not been sufficiently studied to date. Preventive measures are needed to protect the population from the harmful effects of noise on health.

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he Nobel Prize laureate Robert Koch is said to have predicted as early as 1910 that "One day man will have to fight noise as fiercely as plague and cholera" (1). The spectrum of disease has changed greatly over the past few decades, with the result that nontransmissible chronic illnesses such as cardiovascular disorders make up a considerable part of the global disease burden (2). Although research efforts in recent years have focussed primarily on classic risk factors such as high blood pressure, smoking, and diabetes mellitus, recently published results point to the influence of environmental factors such as noise in the development of cardiovascular disease (3). Insights from epidemiological studies show that exposure to traffic noise (from aircraft, road vehicles, and trains) is associated with increased cardiovascular morbidity and mortality (4, 5). In Europe, traffic noise is responsible for 18 000 premature deaths, 1.7 million cases of hypertension, and 80 000 hospitalizations each year (6). The World Health Organization (WHO) has stated that the following years of healthy life are lost annually in Western Europe (7):

- 61 000 due to ischemic heart disease
- 45 000 due to cognitive impairments in childhood
- 903 000 due to sleep disturbance
- 22 000 due to tinnitus
- 654 000 due to noise annoyance.

Altogether, illness induced by traffic noise results in over 1 million disability-adjusted life years (DALYs; i.e. loss of healthy life years) annually in Western Europe. Research into the effects of noise has made significant advances in recent years, particularly with regard to the pathophysiological mechanisms involved in the development of traffic noise–induced cardiovascular disease. A review of the recent literature has thus become necessary.

Methods

We carried out a selective survey of the epidemiological and experimental studies added to the PubMed database in the period 2007–2018. The search terms were "noise AND cardiovascular effects" and "noise cardiovascular effects." We included large primary studies and systematic meta-analyses.

Noise effects model

The definition of noise is sound perceived as annoying. Noise has an objectively (physically) quantifiable dimension in the shape of sound levels or decibels (dB[A]) (*Figure 1*) and a subjective (psychological) dimension determined by the listener's individual assessment of the sound source (8). According to



Examples of sound levels from various sources. The dB(A) scale is a logarithmic scale, so that an increase/decrease of 3 dB(A) means doubling/halving of the sound level. An increase/decrease of 10 dB(A) is perceived as doubling/halving of loudness. In circumstances of long-term exposure to over 55 dB(A) weighted day–evening–night levels (L_{den}), with an increment of 5 dB(A) for the evening and 10 dB(A) for the night, the risk of cardiovascular disease rises (13, 20). (Modified from Münzel et al. [39]. Reproduced with the permission of the publisher. Copyright © 2017, Oxford University Press)

Babisch's noise effects model (9), apart from the direct auditory effects of noise on the organism (damage to the acoustic apparatus by very loud sounds), noise's non-auditory effects are primarily responsible for triggering a stress cascade (Figure 2). Long-term or shortterm environmental noise can cause stress reactions/ noise annoyance reactions, mediated by disturbances of sleep, communication, and activity together with the associated cognitive and emotional reactions. These affect the autonomic (sympathetic) and endocrine systems (e.g., elevated catecholamine and cortisone concentrations) and may as a result lead to changes in lipd metabolism, glucose metabolism, and blood pressure regulation that then contribute to an increased risk of cardiovascular illness. The noise-related stress is attributed an intermediary role in the development of noise-induced cardiovascular disease, i.e., the effects of noise on health are determined partly by the extent to which the listener perceives the noise as annoying (10). Data from the large, population-based Gutenberg Health Study (N = 15 010 persons resident in Mainz and the Mainz-Bingen region) show that the noiserelated stress reaction to various sources of noise during the day and while sleeping at night, including aircraft (odds ratio [OR] 1.09, 95% confidence interval [1.05; 1.13]), road traffic (OR 1.15 [1.08; 1.22]), and rail traffic (OR 1.13 [1.04; 1.22] at night and per one point increase in noise annoyance) is associated dosedependently with an increased risk of atrial fibrillation (11). Moreover, in the context of this study it was shown that the risk of depression (prevalence ratio [PR] 1.97 [1.62; 2.39]) and anxiety disorder (PR 2.14 [1.71; 2.67]), comparing no noise annovance with extreme noise annoyance, rose with increasing total noise annoyance, which in turn may impact negatively on the cardiovascular system (11, 12). However, here only the subjectively assessed noise annoyance was documented, not the objective noise exposure (sound levels).

Epidemiological evidence of the association between noise and cardiovascular disease

Because of the recent increase in traffic, scientific evaluation of the health-related effects of noise is receiving increasing attention. Epidemiological research into the impact of noise has increased considerably in both quantity and quality in recent years. Methods for calculation/standardization of noise exposure have improved, larger populations have been studied, and there has been better correction for confounding variables. Furthermore, phenomena such as habituation, sensitization, conditioning, and exhaustion are adequately represented by the reproduction of daily living conditions in epidemiological studies. The most powerful studies are on the topics of coronary heart disease, myocardial infarction, stroke, and hypertension *(eTable)*.

Coronary heart disease

Three meta-analyses have shown that traffic noise leads to a significant increase in coronary heart disease (13-15). The meta-analysis published by Vienneau et al. in 2015 covered studies on the link between noise from aircraft and road traffic and the incidence of coronary heart disease (13). For traffic noise, they found a pooled relative risk (RR) of 1.06 [1.03; 1.09] starting at 50 dB(A) and per increase of 10 dB(A) L_{den} (weighted day [7 a.m. to 7 p.m.]-evening [7 to 11 p.m.]—night [11 p.m. to 7 a.m.] noise level over a period of 24 h, with an increment of 5 dB[A] added for the evening period and 10 dB[A] for the night). This effect persisted after correction for airborne toxins (mostly nitrogen oxides [NO_X] or nitrogen dioxide [NO2], documented in a subgroup of studies) and exclusion of studies with no data on smoking behavior. The meta-analysis by Babisch came to a similar conclusion, with RR 1.08 [1.04; 1.13]-starting at 52 dB(A) and per increase of 10 dB(A) L_{den} road traffic noise (weighted day-night level for a period of 24 h, with an increment of 10 dB[A] added for the night) (14). In a WHO study published in 2018, the investigators analyzed longitudinal studies and found that road traffic noise, starting at 50 dB(A) and per increase of 10 dB(A), raised the incidence of coronary heart disease by 8% [1.01; 1.15] (15).

Because increasing traffic not only represents a noise problem but also leads to greater emission of air pollutants, it is necessary to ascertain the independent effects of these two variables. Despite the high correlation of the two parameters, the assumption is that noise and air pollutants have independent negative impacts on the cardiovascular system (16).

Myocardial infarction

Cohort studies that controlled for air pollutants, socioeconomic status, and lifestyle factors identified a significant association between road traffic noise and the occurrence of myocardial infarctions (17, 18). The data from a large Danish cohort (N = 50 744) showed that exposure to road traffic noise (10-year mean), independent of the air NO₂ concentration, was associated with a hazard ratio (HR) of 1.12 [1.03; 1.21] per interquartile range L_{den} (17). The findings were similar in a Swedish cohort (OR 1.38 [1.11; 1.71] for road traffic noise \geq 50 vs. <50 dB(A) L_{Aeq.24h} [unweighted 24-h level] after exclusion of persons with hearing loss or exposure to other sources of noise (18).

In a nationwide cohort study in Switzerland (N = 4 415 206), the authors investigated the influence of noise from aircraft, road traffic, and rail traffic on cardiovascular mortality (19). The strongest associations were observed for death from myocardial infarction, with an HR of 1.038 [1.019; 1.058] for road traffic noise, 1.018 [1.004; 1.031] for noise from rail traffic, and 1.026 [1.004; 1.048] for aircraft noise, starting at 30 dB(A) (for rail traffic and aircraft noise) or 35 dB(A) (for road traffic) and per increase of 10 dB(A) L_{den} with correction for air concentrations of NO₂.

Analysis of a Danish cohort (N = 57 053) showed that road traffic noise, starting at 42 dB(A) (N = 57 053) and per increase of 10 dB(A) L_{den}, increases the incidence of myocardial infarction by 12% (incidence rate ratio [IRR] [1.02; 1.22]) independent of air NO_X concentration, smoking behavior, education, and style of nutrition (20).

Stroke

In a large cohort study (N = 57 053), Sørensen et al. found that, starting at 55 dB(A) and per increase of 10 dB(A) L_{den}, road traffic noise increases the risk of stroke by 14% (IRR [1,03; 1,25]), independent of air NO_X concentration, smoking behavior, style of nutrition, and alcohol consumption (21). However, this association was not observed in younger persons (<64.5 years) (IRR 1.02 [0.91; 1.14]). A large study including 3.6 million inhabitants of the area around Heathrow Airport near London, UK demonstrated that aircraft noise, after controlling for age, sex, and lifestyle factors, was dose-dependently associated with increased hospitalization due to stroke both during the day (L_{Aeq,16h}, level over a period of 16 h, 7 a.m. to 11 p.m.) and at night



The noise effect model of Babisch et al. (9) adjusted according to Münzel et al. (4). Both persistent and acute exposure to noise leads to increased development of cardiovascular risk factors (raised blood pressure, increases in blood lipids and blood sugar, greater cardiac output, increased blood viscosity, and activation of blood coagulation), mediated by mental and physiological stress reactions. In the long term, this leads in turn to the clinical manifestation of cardiovascular diseases such as hypertension, coronary heart disease, heart failure, arrhythmia, and stroke. (Modified from Münzel et al. [4]. Reproduced with the permission of the publisher. Copyright © 2014, Oxford University Press)

(L_{night}, from 11 p.m. to 7 a.m.). The risk of hospitalization was greater for aircraft noise at night (RR 1.29 [1.14; 1.46]) than during the day (RR 1.24 [1.08; 1.43]) when comparing high and low noise exposure (night: >55 vs. \leq 50 dB[A]; day: >63 vs. \leq 51 dB[A]) (22). The results for stroke-related mortality in this study were similar even after taking account of particulates (PM₁₀, particle diameter less than 10 µm). This pattern of findings suggests that particularly aircraft noise at night is associated with greater stress on the cardiovascular system. Another study carried out in London, this time including 8.6 million members of the population, revealed that after controlling for particulate pollution (PM_{2.5}), road traffic noise (>60 vs. <55 dB[A]) during

MEDICINE



Dose–response relationships for noise and cardiovascular disease. L_{den} is the weighted day–evening–night sound level, with an increment of 5 dB(A) for the evening and 10 dB(A) for the night. CVD, cardiovascular disease; Hyp, hypertension. (Modified from Münzel et al. [4]. Reproduced with the permission of the publisher. Copyright © 2014, Oxford University Press)

the day ($L_{Aeq,16h}$) was associated with an elevated risk of stroke-related hospitalization in persons aged $\geq 25-74$ years (RR 1.05 [1.02; 1.09]) and in those ≥ 75 (RR 1.09 [1.04; 1.14]) (23). Moreover, road traffic noise (>60 vs. <55 dB[A]) during the day was significantly associated with overall mortality (RR 1.04 [1.00; 1.07]), although the association with cardiovascular mortality did not attain significance (RR 1.03 [0.98; 1.07]).

Hypertension

A meta-analysis of 24 studies by van Kempen and Babisch revealed that road traffic noise is associated with an elevated risk of the occurrence of high blood pressure (OR 1.034 [1.011; 1.056], starting at 45 dB(A) and per increase of 5 dB(A) $L_{Aeq,16h}$) (24). This analysis was restricted to cross-sectional studies, so the power of any conclusions about the causal role of noise is limited. However, large prospective studies have demonstrated that both aircraft noise and road traffic noise exposure are associated with an increased risk of hypertension (25–27), with stronger effects for aircraft noise at night (OR 2.63 [1.21; 5.71] per increase of 10 dB(A) L_{night}) than for road traffic noise (RR 1.03 [0.99; 1.07] per increase of 10 dB(A) L_{den}) (27).

Other endpoints

The investigators involved in the NORAH study (Noise-Related Annoyance, Cognition, and Health) of persons resident in the Rhine–Main region (in and around Darmstadt, Mainz, and Worms) found that traffic noise was associated with an increased risk of heart failure or hypertensive heart disease (greatest increase in risk, of 3.1% for rail traffic noise per increase of 10 dB(A) $L_{Aeq,24h}$ [2.2; 4.1%]) and of myocardial infarction (greatest increase in risk, of 2.8% for road traffic noise per increase of 10 dB[A] $L_{Aeq,24h}$ [1.2; 4.5%]) (28, 29). Analyses of the risk of hypertension, however, showed no significant associations with noise exposure from aircraft, road traffic, or rail traffic, although it must be pointed out that the study suffered from methodological deficiencies in blood pressure measurement and limited inclusion of confounding factors (30).

In addition, a Danish prospective study (N = 57 053) showed that road traffic noise is associated with an increased incidence of atrial fibrillation (IRR: 1.06 [1.00; 1.12] starting at 40 dB(A) and per increase of 10 dB(A) L_{den}); however, this effect was not independent of the air concentration of NO_X or NO₂ (after controlling for NO_X, IRR: 1.04 [0.96; 1.11]; after controlling for NO₂, IRR: 1.01 [0.94; 1.09]) (31).

In conclusion, it can be stated that large studies of high methodological quality that controlled for airborne pollutants, lifestyle factors, and other important confounding factors were able to demonstrate a systematic association of exposure to traffic noise with the risk of cardiovascular disease (*Figure 3*).

Noise-induced pathophysiological effects on the cardiovascular system

The pathophysiological mechanisms by which noise causes more cardiovascular illness have not yet been fully elucidated and are a topic of ongoing research (Box). The assumption is that long-term noise leads to stress reactions or noise annoyance reactions, which in turn, via activation of the sympathetic nervous system and increased excretion of stress hormones (5, 9), favors the development of oxidative stress and inflammatory processes, resulting in disruption of vascular and endothelial function (32-34). As well as the emergence of endothelial dysfunction and the associated negative impact on blood pressure regulation, the noise-related development of further risk factors such as elevated blood pressure, rises in blood lipids and blood sugar, increased cardiac output, raised blood viscosity, and activation of blood coagulation leads to a higher risk of cardiovascular disease (4).

Summary and conclusion

Epidemiological studies have shown that noise from aircraft, road traffic, and rail traffic represents a significant risk factor for the development of cardiovascular disease. The research involved in these studies has identified important pathophysiological factors that will enable development of drug treatment strategies to minimize the harmful effects of noise on the human organism. Particularly on the basis of the new WHO noise guidelines (3), which, in summary, state an RR of 1.08 [1.01; 1.15] for the occurrence of coronary heart disease, starting at 50 dB(A) and per increase of 10 dB(A) L_{den} in connection with recommendations for much lower mean noise levels during the day and at night (aircraft noise: 45 dB[A] L_{den} and 40 dB[A] L_{night} ; road traffic: 53 dB[A] L_{den} and 45 dB[A] L_{night} ; rail traffic: 54 dB[A] L_{den} and 44 dB[A] L_{night}), it is urgently necessary to introduce measures for adequate protection of the population from the negative health effects of environmental noise.

Limitations

Our goal in writing this article was to give a selective overview of the most important recent research findings, reviewing systematic analyses and large cohort studies in the light of our own experience. This approach entails limitations, particularly due to the selective inclusion of primary studies and meta-analyses that vary with regard to quality of results, controlling for confounding factors, operationalization of endpoints, and statistical methods, making it difficult to compare their findings. Nevertheless, overall the results of the studies presented here reinforce the contention that noise significantly increases the occurrence of cardiovascular disease.

Conflict of interest statement

Prof. Daiber coordinated third-party funding for a research project from the Boehringer-Ingelheim Foundation and the Mainzer Herz Foundation. The remaining authors declare that no conflict of interest exists.

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BOX

Noise-related endothelial dysfunction

Only a small number of studies have explicitly explored the association between noise and endothelial function. In two recent studies. Schmidt et al. demonstrated that simulated aircraft noise at night (30 or 60 flights overhead, peak noise level 60 dB[A], mean level 43 or 46 dB[A]) led to deterioration of vascular function (measured as flow-mediated vasodilation), elevated levels of stress hormone (epinephrine), and lower sleep quality both in healthy probands and in patients with coronary heart disease (32, 33). In two recently published animal studies in mice, the investigators uncovered, for the first time, the molecular mechanisms of noise-induced vascular function disorder (35, 36). Münzel et al. showed that simulated aircraft noise (recorded previously and played over a loudspeaker system) led within 4 days to a marked increase in the stress hormone concentration, elevated blood pressure, alteration of gene expression in the vascular wall, and disturbance of vascular function that was attributable particularly to increased formation of free radicals and thus oxidative stress (35). The authors identified two radical-forming enzymes that played a crucial part in induction of the vascular function disorders: nicotinamide adenine dinucleotide phosphate oxidase (NADPH oxidase) and nitric oxide synthase (NOS). The noise-related dysregulation of these factors facilitated the formation of reactive oxygen species, which led directly to decreased vascular bioavailability of nitric oxide (NO), resulting in deterioration of endothelial function. A follow-up study by Kröller-Schön et al. went on to demonstrate that the vascular injuries are accompanied by cerebral changes. Notably, these were found after exposure of the mice to noise in their sleeping phase, but not during the waking phase (36). A newly published WHO meta-analysis shows that traffic noise significantly increases the risk of sleep disorders, to the point where noise-induced reduction and fragmentation of sleeping time may represent an important factor in the mediation of cardiovascular disease (37). Findings from the SAPALDIA study (Swiss Cohort Study on Air Pollution and Lung and Heart Diseases in Adults) also showed that the frequency of traffic noise events, especially during the night, is associated with increased arterial stiffness, an important marker of vascular dysfunction (38).

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Key Messages

- Epidemiological studies show that traffic noise raises the risk of cardiovascular diseases such as hypertension, coronary heart disease, myocardial infarction, heart failure, and stroke.
- Sustained noise causes stress reactions that bring about activation of the autonomic and endocrine system, resulting per se in increased formation of cardiovascular risk factors.
- At the pathophysiological level, particularly traffic noise at night causes oxidative stress and inflammatory processes in the brain, which in turn impact negatively on the vascular system and result in endothelial dysfunction and hypertension.
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eTABLE

Selected primary studies on the association between noise and cardiovascular endpoints

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First author, year	Exposure	Study design, location	Overall population and endpoint (selected)	Control variables	Results (selected)
O. Hahad, 2018 (11)	Noise from various sources (including aircraft, road traffic, and rail traffic) during the day and while sleeping at night	Cross-sectional, data from the Gutenberg Health Study, Germany	N = 15 010, 35–74 years, AF diagnosis or diagnosis from study ECG	Age, sex, SES, shift work, depression, medication intake, diabetes, hypertension, smoking, obesity, dys- lipidemia, family history of MI/stroke	OR per 1-point increase in exposure during sleeping hours Aircraft noise: 1.09 [95% CI: 1.05; 1.13] Road traffic noise: 1.15 [95% CI: 1.08; 1.22] Rail traffic noise: 1.13 [95% CI: 1.04: 1.22]
N. Roswall, 2017 (17)	Road traffic noise (L _{den})	Longitudinal, data from the Diet, Cancer and Health study, Denmark	N = 50 744, 50–64 years, incidental MI	Age, sex, smoking, alcohol, fruit and vegetable consumption, exercise, education, SES, calendar year, rail traffic noise, NO ₂	HR per IQR L _{den} Road traffic noise: 1.12 [95% CI: 1.03; 1.21]
J. Selander, 2009 (18)	Road traffic noise (L _{Aeq,24h})	Population-based case–control study, Sweden	N = 5452 (2 246 cases, 3 206 controls), 45–70 years, MI diagnosis	Age, sex, smoking, physical inactivity, diabetes, NO ₂ , noise exposure in the workplace	OR L _{Aeq,24h} <50 vs. ≥ 50 dB(A):
H. Heritier, 2017 (19)	Road traffic, rail traffic, and aircraft noise (L _{den})	Longitudinal, data from the Swiss National Cohort study, Switzerland	N = 4 415 206, >30 years, Mortality due to MI	Sex, SES, relationship status, NO ₂ , native language, nationality	HR per 10 dB(A) L _{den} Road traffic noise: 1.038 [95% CI: 1.019; 1.058] Rail traffic noise: 1.018 [95% CI: 1.004; 1.031] Aircraft noise: 1.026 [95% CI: 1.004; 1.048]
M. Sørensen, 2012 (20)	Road traffic noise (L _{den})	Longitudinal, data from the Diet, Cancer and Health study, Denmark	N = 57 053, 50–64 years, incidental MI	Age, sex, smoking, fruit and vegetable consump- tion, education, alcohol, BMI, exercise, calendar year, rail traffic and aircraft noise, NO _x	IRR per 10 dB(A) L _{den} At time of diagnosis: 1.12 [95% CI: 1.02; 1.22] L _{den} 5 years before diagnosis: 1.12 [95% CI: 1.02; 1.23]
M. Sørensen, 2011 (21)	Road traffic noise (L _{den})	Longitudinal, data from the Diet, Cancer and Health study, Denmark	N = 57 053, 50–64 years, hospitalization due to incidental stroke	Smoking, fruit and vegetable consumption, coffee consumption, alcohol, BMI, exercise, education, income, rail traffic and aircraft noise, NO _x	IRR per 10 dB(A) L _{den} : 1.14 [95% CI 1.03; 1.25] 1.02 [95% CI: 0.91; 1.14] for persons <64.5 years 1.27 [95% CI: 1.13; 1.43] for persons ≥ 64.5 years
AL. Hansell, 2013 (22)	Aircraft noise (L _{Aeq,16h} , L _{night})	Longitudinal, small area study, England	N = 3.6 million, hospitalization due to incidental stroke	Age, sex, ethnicity, deprivation, smoking	Significant linear association between hospitalization and L _{Aeq,16h} / L _{night} RR L _{night} > 55 vs. ≤ 50 dB(A): 1.29 [95% CI: 1.14; 1.46]

First author, year	Exposure	Study design, location	Overall population and endpoint (selected)	Control variables	Results (selected)
Halonen, 2015 (23)	Road traffic noise (L _{Aeq,16h} , L _{night})	Longitudinal, small area study, England	N = 8.6 million, ≥ 25 and ≥ 75 years, hospitalization due to incidental stroke	Age, sex, deprivation, ethnicity, smoking, PM _{2.5} , topographical structure of neighborhood	RR L _{Aeq.16h} >60 vs. <55 dB(A): 1.05 [95% CI: 1.02; 1.09] for persons ≥ 25–74 years RR L _{night} 55–60 vs. <55 dB(A): 1.05 [95% CI: 1.01; 1.09]
Eriksson, 2007 (25)	Aircraft noise (L _{Amax})	Longitudinal, data from the Stockholm Diabetes Preventive Program, Sweden	N = 2754 men, 35–56 years, incidental hypertension	Age, BMI	for persons \geq 75 years RR L _{Amax} <70 vs. \geq 70 dB(A): 1.20 [95% CI: 1.03; 1.40] After exclusion of persons with tobacco consumption before blood pressure measurement: 1.32 [95% CI: 1.12; 1.55]
Dimakopoulou, 2017 (26)	Aircraft noise (L _{night} , L _{Aeq,16h}) Road traffic noise (L _{Aeq,24h})	Longitudinal, data from the HYENA study, Greece	N = 420, 45–70 years, incidental hypertension	Sex, age, smoking, BMI, education, exercise, alcohol, salt consumption	OR per 10 dB(A) L _{night} Aircraft noise: 2.63 [95% Cl: 1.21; 5.71] OR per 10 dB(A) L _{Aeq,16h} Aircraft noise: 1.46 [95% Cl: 0.89; 2.39] OR per 10 dB(A) L _{Aeq,24h} Road traffic noise: 1.18 [95% Cl: 0.92; 1.52]
Fuks, 2017 (27)	Road traffic noise (L _{den})	Longitudinal, data from the ESCAPE study, Norway, Sweden, Denmark, Germany, Spain	N = 41 072, 47–71 years (mean), incidental hypertension	Age, sex, education, working situation, BMI, smoking, alcohol, exercise, family history of hypertension, SES	RR per 10 dB(A) L _{den} Self-reported hypertension: 1.03 [95% CI: 0.99; 1.07] Measured hypertension: 0.99 [95% CI: 0.94; 1.04]
Seidler, 2016 (28)	Aircraft, road traffic, and rail traffic noise (L _{Aeq.24h})	Population-based case–control study, data from the NORAH study, Germany	N = 854 366 (19 632 cases, 834 734 controls), >40 years, MI diagnosis	Age, sex, social status, education, occupation	OR per 10 dB(A) L _{Aeq.24h} Aircraft noise: 0.993 [95% CI: 0.966; 1.020] Road traffic noise: 1.028 [95% CI: 1.012; 1.045] Rail traffic noise: 1.023 [95% CI: 1.005; 1.042]
Seidler, 2016 (29)	Aircraft, road traffic, and rail traffic noise (L _{Aeq,24h})	Population-based case_control study, data from the NORAH study, Germany	N = 758 857 (104 145 cases, 654 172 controls), >40 years, diagnosis of heart failure or hypertensive heart disease	Age, sex, social status, education, occupation	OR per 10 dB(A) L _{Aeq.24h} Aircraft noise: 1.016 [95% Cl: 1.003; 1.030] Road traffic noise: 1.024 [95% Cl: 1.016; 1.032] Rail traffic noise: 1.031 [95% Cl: 1.022; 1.041]
Zeeb, 2017 (30)	Aircraft, road traffic, and rail traffic noise (L _{Aeq,24h})	Population-based case–control study, data from the NORAH study, Germany	N = 493 168 (137 577 cases, 355 591 controls), >40 years, newly diagnosed hypertension	Age, sex, SES, education, occupation	OR per 10 dB(A) L _{Aeq,24h} Aircraft noise: 0.997 [95% Cl: 0.985; 1.010] Road traffic noise: 1.003 [95% Cl: 0.995; 1.011] Rail traffic noise: 1.003 [95% Cl 0.994; 1.011]

Epidemiological primary studies were selected for this tabular presentation. AF, atrial fibrillation; BMI; body mass index; CI, confidence interval; HR, hazard ratio; IQR, interquartile range; IRR, incidence rate ratio;

weighted day (7 a.m. to 7 p.m.)-evening (7 to 11 p.m.)-night (11 p.m. to 7 a.m.) sound level over a 24-h period, with an increment of 5 dB(A) for the evening and 10 dB(A) for the night; $\mathsf{L}_{\mathsf{den}}$

- L_{night} noise level at night (11 p.m.) rooming (1 or 1 p.m.) rooming (1 or 1 p.m.) rooming (1 or 1 p.m.); L_{Ang,24/16h} Unweighted level over a period of 24 h or 16 h (usually 7 a.m. to 11 p.m.); maximum level over a certain period

L_{Amax} maximum level over a certain period MI, myocardial infarction; NO₂, nitrogen dioxide; NO_x, nitrogen oxides; OR, odds ratio; PM_{2.5}, particulate matter with diameter less than 2.5 μm; SES, socioeconomic status