

Islands in the desert: environmental distribution modelling of endemic flora reveals the extent of Pleistocene tropical relict vegetation in southern Arabia

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> **• Background and Aims** Southern Arabia is a global biodiversity hotspot with a high proportion of endemic desert-adapted plants. Here we examine evidence for a Pleistocene climate refugium in the southern Central Desert of Oman, and its role in driving biogeographical patterns of endemism.

> **• Methods** Distribution data for seven narrow-range endemic plants were collected systematically across 195 quadrats, together with incidental and historic records. Important environmental variables relevant to arid coastal areas, including night-time fog and cloud cover, were developed for the study area. Environmental niche models using presence/absence data were built and tuned for each species, and spatial overlap was examined.

> **• Key Results** A region of the Jiddat Al Arkad reported independent high model suitability for all species. Examination of environmental data across southern Oman indicates that the Jiddat Al Arkad displays a regionally unique climate with higher intra-annual stability, due in part to the influence of the southern monsoon. Despite this, the relative importance of environmental variables was highly differentiated among species, suggesting that characteristic variables such as coastal fog are not major cross-species predictors at this scale.

> **• Conclusions** The co-occurrence of a high number of endemic study species within a narrow monsoon-influenced region is indicative of a refugium with low climate change velocity. Combined with climate analysis, our findings provide strong evidence for a southern Arabian Pleistocene refugium in Oman's Central Desert. We suggest that this refugium has acted as an isolated temperate and mesic island in the desert, resulting in the evolution of these narrowrange endemic flora. Based on the composition of species, this system may represent the northernmost remnant of a continuous belt of mesic vegetation formerly ranging from Africa to Asia, with close links to the flora of East Africa. This has significant implications for future conservation of endemic plants in an arid biodiversity hotspot.

> **Key words:** Arabian Peninsula, Central Desert, coastal fog desert, desert endemic, ecological niche modelling, important plant areas, Oman, Pleistocene, refugia.

INTRODUCTION

Southern Arabia is part of the Horn of Africa global biodiversity hotspot, and is one of only two hotspots that are entirely arid [\(Mittermeier](#page-10-0) *et al.*, 2005; [Mallon, 2013\)](#page-10-1). The flora of southern Arabia arises from the relatively late separation of Arabia from Africa and Asia around 25 million years ago [\(Raven and Axelrod,](#page-10-2) [1974](#page-10-2); [Delany, 1989](#page-9-0)). During the Miocene, Arabia supported palaeo-tropical vegetation with swamps and open savannah grassland [\(Whybrow and Mcclure, 1981](#page-11-0)). This was progressively replaced by more drought-adapted vegetation through the Pliocene, with mesic elements of the flora persisting only in climatically favourable refugia [\(Kürschner, 1998](#page-10-3); Jolly *et al.*[, 2009\)](#page-10-4). The environment of southern Arabia subsequently oscillated between climatic extremes throughout the Quaternary period [\(Fleitmann](#page-10-5) [and Matter, 2009;](#page-10-5) [Parker, 2010;](#page-10-6) [Jennings](#page-10-7) *et al.*, 2015). These

oscillations, combined with the relative stability of localized climatic refugia, may have contributed to the high degree of species endemism [\(Patzelt, 2015](#page-10-8); [Sandel](#page-10-9) *et al.*, 2011).

The biogeographic consequences of contraction and expansion from glacial refugia have been well described in the temperate zones of Europe and North America [\(Bennett](#page-9-1) *et al.*, 1991; [Comes and Kadereit, 1998](#page-9-2); [Birks and Willis, 2008](#page-9-3); [Keppel](#page-10-10) *et al.*, [2012;](#page-10-10) Wang *et al.*[, 2014\)](#page-11-1). In comparison, these processes are poorly known in the arid environments of the Arabian Peninsula ([Ghazanfar, 1998](#page-10-11); [Meister](#page-10-12) *et al.*, 2006; [Patzelt, 2015\)](#page-10-8). Therefore, establishing the spatio-temporal distribution of past climate refugia in southern Arabia is likely to have important implications for future conservation planning [\(Al-Abbasi](#page-9-4) *et al.*, 2010), building evolutionary resilience under climate change (Sgrò *et al.*[, 2011](#page-11-2); [Keppel](#page-10-10) *et al.*, 2012) and even interpreting the history of hominid dispersal out of Africa [\(Jennings](#page-10-7) *et al.*, 2015; [Gandini](#page-10-13) *et al.*, 2016).

A key centre for plant endemism in southern Arabia is Oman's Central Desert [\(White and Léonard, 1990;](#page-11-3) [Miller and](#page-10-14) [Nyberg, 1991](#page-10-14); [Ghazanfar, 2004](#page-10-15); [Patzelt, 2014](#page-10-16)). The Central Desert is a hyper-arid region, characterized by scarce precipitation often $<$ 100 mm year⁻¹ with high inter-annual variability and temperatures ranging from 6 \degree C to >50 \degree C (Stanley-Price *et al.*[, 1988;](#page-11-4) [Fisher and Membery, 1998;](#page-10-17) [Almazroui](#page-9-5) *et al.*, 2013). Provisionally divided into 'northern' and 'southern' systems, the Central Desert has relatively low species diversity, but the highest proportion of range-restricted endemic and regionally endemic plants in Oman [\(Patzelt, 2015](#page-10-8)). This represents an ideal study system in which to test for evidence of climatic refugia and their influence on the floral biogeography of southern Arabia.

Despite significant progress in documenting the flora of Oman [\(Miller and Cope, 1996](#page-10-18); [Ghazanfar, 1998,](#page-10-11) [2004](#page-10-15); [Patzelt, 2009,](#page-10-19) [2014](#page-10-16); [Brinkmann](#page-9-6) *et al.*, 2011; [Ghazanfar and Fisher, 2013](#page-10-20)) high resolution plant diversity and distribution data are limited or lacking for many areas, hindering our ability to test these biogeographic hypotheses and identify putative refugia. To address this knowledge gap, here we report the results of a systematic botanic survey of the southern Central Desert. We focus on seven high conservation value narrow-range endemic desert plants identified through development of the *Oman Red Data Book* [\(Patzelt, 2014](#page-10-16)) [\(Table 1;](#page-1-0) [Fig. 1](#page-2-0)). All selected species are readily field identifiable, encompass broad taxonomic diversity and are restricted to the coastal belt and the adjacent escarpment.

To enable a transition from survey data to predictive distribution maps, we use an environmental niche modelling (ENM) approach. ENMs are a suite of methods used to establish the relationship between a species and a set of environmental variables [\(Elith and Leathwick, 2009](#page-10-21); [Peterson](#page-10-22) *et al.*, 2015). In principle, ENMs evaluate the environmental conditions in grid cells known to be occupied by a species and identify additional cells that represent similar environmental conditions ([Merow](#page-10-23) *et al.*[, 2013\)](#page-10-23). The species' niche can then be projected across the study area to predict its spatial distribution and identify environmental variables that contribute significantly to model performance [\(Searcy and Shaffer, 2016\)](#page-11-5). In this study, we benefit from both presence and absence survey data and a stratified

survey design, which negates many of the biases common in environmental niche models ([Jiménez-Valverde](#page-10-24) *et al.*, 2008; [Warren and Seifert, 2011\)](#page-11-6).

To distinguish amongst alternative environmental drivers for local endemism, selection of appropriate environmental modelling variables in arid environments is important (Dilts *et al.*[, 2015](#page-9-7); [Title and Bemmels, 2018\)](#page-11-7). In other coastal desert systems, endemic plant distributions are strongly influenced by the presence of cloud shade and fog, which causes condensation on leaves and stems that trickles down to root systems ([Cereceda](#page-9-8) *et al.*, [2008;](#page-9-8) [Fischer](#page-10-25) *et al.*, 2009). The Central Desert is thought to lie at the northern limit of a regional monsoon system known as the Khareef. The Khareef provides significant seasonal horizontal precipitation to Dhofar from July to September, resulting in localized cloud forest [\(Hildebrandt](#page-10-26) *et al.*, 2007). Given the extreme temperatures and low precipitation of the Central Desert, the presence of coastal fog and the cooling effect of the prevailing wind have been hypothesized to be a key driver of plant distributions ([Miller, 1994](#page-10-27); [Stanley-Price](#page-11-4) *et al.*, 1988; [Patzelt, 2015\)](#page-10-8).

The presence of coastal fog is supported by data from the northern Central Desert in 1984, where water collected from fog collectors at Jiddat Al Harasisi ranged from 0.08 L m⁻² in January to 3.6 L m⁻² in October. A total of 93 nights with fog moisture were recorded across the year, with fog moisture at ground level coinciding with reduced night-time temperatures, increased humidity and a wind speed <15 km h^{-1} ([Stanley-](#page-11-4)Price *et al.*[, 1988\)](#page-11-4). In a subsequent study (Fisher and Membery, [1998\)](#page-10-17), a monthly maximum of 4.0 L m^{-2} during March and a minimum 2.5 L m⁻² during January, May, June and December was recorded. To our knowledge, no empirical data are available on coastal fog from the southern Central Desert.

Validation of environmental variables across a large, arid and sparsely populated study area such as the Central Desert is also exacerbated by the paucity of weather stations (meaning that global climate models are highly interpolated) and the fact that where weather stations do exist, cloud cover and coastal fog are rarely recorded. To address these shortfalls, we make use of the newly available WorldClim 2 data set that incorporates high accuracy remotely sensed maximum and minimum land surface temperature

TABLE 1. Summary of the seven study taxa, including Red List status from [Patzelt \(2014\)](#page-10-16).

ID Species*		Family	Habitat description			
AA	Wadithamnus artemisioides subsp. Batharitica (A.G.Mill. $&$ J.A.Nyberg) T.Hammer & R.W.Davis.	Amaranthaceae	Coastal areas including dry cliffs, rocky slopes and wadi banks on limestone, $20-300$ m.			
HG	Hyoscyamus gallagheri A.G. Mill. & Biagi.	Solanaceae	Low sand dunes and gravel desert, 0–250 m.	VU		
OH	Ochradenus harsusiticus A.G. Mill.	Resedaceae	Open Acacia tortilis–Prosopis cineraria woodland, in sandy and gravelly depressions of the central desert, 100–210 m.	VU		
PI	Polycarpaea jazirensis R.A. Clement.	Caryophyllaceae	Coastal areas of Sahil al Jazir in sandy depressions and on limestone cliffs, $100 - 210$ m.	CR		
PP	Pulicaria pulvinata E. Gamal-Eldin	Asteraceae	Limestone cliffs and central limestone plateau on sand and gravel, 0–300 m.	NT		
SG	Searsia gallagheri Ghaz.	Anacardiaceae	Stony wadis and depressions, 130–205 m.	VU		
SH	Salvia aff. hillcoatiae R.A. Clement.	Lamiaceae	Dry limestone plateau of the central desert, 0–300 m.	VU		

*Voucher specimens are deposited at the Oman Botanic Gardens, accession numbers: OH, CDBE-1; PP, CDBE-2; SH, CDBE-3; PJ, CDBE-4; SG, CDBE-5, 6. † All assessments are global assessments which have been submitted to IUCN to be included to the Global Red List of Threatened Species, with the exception of a national assessment for *S. gallagheri*. The relationship between *S. gallagheri* (formerly *Rhus*) and *S. vulgaris* awaits more detailed taxonomic investigation; thus, a national assessment is reported.

Fig. 1. Images of the endemic Central Desert plant species included in this study. (A) *Wadithamnus artemisoides* subsp. *batharitica*; (B) *Hyoscyamus gallagheri*; (C) *Ochradenus harsusiticus*; (D) *Polycarpaea jazirensis*; (E) *Pulicaria pulvinata*; (F) *Salvia aff. hillcoatiae*; (G) *Searsia gallagheri*; (H) a typical Central Desert landscape.

[\(Fick and Hijmans, 2017\)](#page-10-28), together with remotely sensed cloud cover data [\(Wilson and Jetz, 2016](#page-11-8)). We also incorporate remotely sensed fog and putative fog covariates, to explore whether these variables contribute significantly to Central Desert species distributions. Specifically, we derive night-time fog intensity from MODIS data [\(MODIS Characterization Support Team, 2018](#page-10-29)) using the approach of [Chaurasia](#page-9-9) *et al.* (2011), as well as topography (including elevation, slope, aspect and terrain roughness, as locally higher areas may catch more moisture; [Schemenauer](#page-11-9) *et al.*[, 1987](#page-11-9)), wind speed and night-time land surface temperature (LST) (Wan *et al.*[, 2018\)](#page-11-10). As an additional line of evidence in assessing the importance of coastal fog, we also survey physiological fog capture adaptations across our study species using the approach of [Larraín-Barrios](#page-10-30) *et al.* (2018).

Here, building on novel systematic survey data from the Central Desert, we aim to address three main questions. First, we use newly available climate data to examine the influence of the southern monsoon and coastal fog influx on the Central Desert climate. Secondly, we model the distributions of seven high priority narrow-range endemics, and test the hypothesis that the same environmental variables are consistently important across taxa. Thirdly, we consider evidence that the current floral biogeography of the Central Desert has been strongly influenced by a southern Arabian Pleistocene refugium. We consider these data in the context of conserving rare desert endemics across southern Arabia, a global biodiversity hotspot.

MATERIALS AND METHODS

Study area

The southern Central Desert is dominated by the Jiddat Al Arkad, a meandering escarpment 50–100 m high dissected by extensive wadi systems, depressions and runnels which discharge into the Sahil Al Jazir coastal plain. Surface flows are only present following heavy rains. Geologically, the study area is dominated by Oligocene–Miocene white bioclastic limestone with coral debris flow deposits and laminated dolomitic limestone [\(Patel, 1992\)](#page-10-31). Soil is predominantly Calciorthids – gravelly sandy loam on alluvial fans and stream terraces and torriorthents (rock outcrops) weakly developed or undeveloped, low in organic matter and moderately calcareous [\(Dregne,](#page-9-10) [1976](#page-9-10)).

The vegetation is classified into three units as per [Patzelt](#page-10-8) [\(2015\)](#page-10-8). (1) *Acacia tortilis*–*Prosopis cineraria* open woodland. Common grass and shrub species include the endemic shrubs *Convolvulus oppositifolius* and *Ochradenus harsusiticus*, and endemic grass *Stipagrostis sokotrana.* (2) Xeromorphic dwarf shrubland intermixed with grasses and annual species. The dwarf palm *Nannorrhops ritchieana* and *Acacia ehrenbergiana* are common in sandy depressions close to the coastal escarpments. (3) Xeromorphic dwarf shrubland with *Searsia gallagheri* and *O. harsusiticus*. In addition to flowering plants, several species of corticolous and saxicolous lichens and epilithic cyanobacteria occur here, most of which are restricted to the putatively fog-affected zones ([Ghazanfar and Gallagher,](#page-10-32) [1998](#page-10-32)). The seven study species ([Fig. 1\)](#page-2-0) are considered a part of, though not restricted to, the xeromorphic dwarf shrubland community. A description of their known habitat and conservation status is provided in [Table 1](#page-1-0).

Field surveys and plant morphology

Fieldwork was conducted during the period 13–24 January 2017, from Ras Madrakah, across the Sahil Al Jazer (coastal plain) to the southern extent of the Jiddat Al Arkad, as part of the Central Desert Botanic Expedition 2017. A stratified survey strategy was designed orientating ten 20 km transects at 315° (NW) along a 270 km portion of coastline, at 30 km intervals. This approach was designed to cross multiple environmental gradients that frequently run perpendicular to the coastline, achieving both geographical and environmental representativity. Stratified quadrat locations at 1 km intervals were plotted prior to field work and their co-ordinates uploaded to hand-held GPS units (Garmin, Oregon). Due to the remote nature of the study area, with several deep wadis bisecting these transects and inhibiting access, some portions of these tran sects were not surveyed. When moving between transects, we opportunistically sampled additional quadrats at 5 km intervals and recorded incidental observations of target species to maxi mize data collection. These additional quadrats were positioned via random number generation to determine distance and bearing from the vehicle. A significant portion of travel was away from roads but, where roads (mostly gravel tracks) were present, quadrat positioning began >100 m from the road to mitigate disturbance bias in the vegetation recorded.

The following data were recorded for each quadrat: location, soil texture, soil pH, soil electrical conductivity (EC) (following the method of [Zhang](#page-11-11) *et al.*, 2005), total vegetation cover, maximum vegetation height, elevation, topography description and the pres ence/absence and count of the seven study species. Voucher speci mens were collected for subsequent analysis, and are deposited in the Oman Botanic Garden herbarium (OBG) [\(Table 1](#page-1-0)). Summary statistics of quadrats were calculated in R software V3.1.2, implemented in RStudio [\(R Development Core Team, 2014;](#page-10-33) [RStudio](#page-10-34) [Team, 2016](#page-10-34)). A checklist of fog moisture capture and water use efficiency functional traits commonly observed in xerophytic plants was also collated, following the approach of [Larraín-Barrios](#page-10-30) *et al.* [\(2018\).](#page-10-30) Each study species was examined and scored for their pres ence/absence and degree of development (see [Supplementary data](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data) [Tables S1](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data) and [S2\)](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data). In addition to observations recorded during this field study, historical records were cleaned and included from relevant national and international collections, specifically; Oman Botanic Garden Herbarium (OBG), Sultan Qaboos University (SQUH), the Oman Natural History Museum National Herbarium (ON) and the Royal Botanic Garden Edinburgh (E) (summarized in [Table 2](#page-3-0); [Supplementary data Fig. S1](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)).

Preparation of environmental variables

To ensure that we captured important environmental variation, we collated 54 bioclimatic variables covering the study area at 1 km resolution [\(Supplementary data Tables S3](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data) and [S4](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)) reflecting annual trends, seasonality and limiting environmental factors. In addition to WorldClim 2 and Bioclim variables [\(Fick and Hijmans,](#page-10-28) [2017\)](#page-10-28), we generated a complementary set of bioclimatic layers that may better characterize arid environments using the 'ENVIREM' package [\(Title and Bemmels, 2018\)](#page-11-7). We also estimated mean nighttime fog intensity, an important candidate variable in determining plant distributions in this region [\(Stanley-Price](#page-11-4) *et al.*, 1988). Specifically, we followed the approach of [Chaurasia](#page-9-9) *et al.* (2011) and classified fog based on the brightness temperature difference (ΔBT) of the 3.9 and 10.75 μm bands (channels 22 and 31) of the MODIS satellite. The emissive properties of these two bands differ for fog water droplets which are typically small, and do not ex cite the 3.9 μm band, whereas emissivity for both cloud and fog

Feature classes: linear, L; quadratic, Q; hinge, H; product, P; and threshold, T. Feature classes: linear, L; quadratic, Q; hinge, H; product, P; and threshold, T.

RM = regularization multiplier. RM = regularization multiplier.

droplets is approximately the same for the 10.75 μm band [\(Hunt,](#page-10-35) [1973](#page-10-35)). Twice nightly images at 1 km resolution were collated for the period 2001–2017 (MODIS product: MOD021KM) from the LAADS database [\(MODIS Characterization Support Team, 2018\)](#page-10-29). Raw radiance values were converted to brightness temperature using Planck's function implemented in ENVI software (Harris Geospatial) and the difference calculated. In contrast to [Chaurasia](#page-9-9) *et al.* [\(2011\)](#page-9-9), high-quality real-time ground truth data are not available for our region, therefore we did not apply a fog classification threshold; instead we retained the data as a continuous variable, with higher (less negative) values considered more likely to represent smaller fog water droplets.

At a fine spatial scale, other variables may also interact with fog moisture and influence the local ecology [\(Rastogi](#page-10-36) *et al.*, [2016;](#page-10-36) [Chung](#page-9-11) *et al.*, 2017), thus we also incorporated several relevant fog proxies or covariates. Cloud cover data (period 2001– 2015) were extracted from the global high-resolution cloud cover data set generated by [Wilson and Jetz \(2016\).](#page-11-8) Roughness, Terrain Ruggedness Index (TRI), slope and aspect were generated from a digital elevation model (GTOPO30) using the R package 'Raster' ([Hijmans, 2017](#page-10-37)). Aspect was cosine transformed and rotated to align with the prevailing wind direction (30°N). Night-time LST was obtained from the MODIS satellite mission at 1 km resolution (Wan *et al.*[, 2018\)](#page-11-10). The complete suite of variables is reported in [Supplementary data Table S3.](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)

To compare and characterize the range of environmental conditions across the Central Desert and other regional centres of endemism, we randomly sampled all 54 environmental variables for *n* cells in each region (with *n* being proportional to the area of the sampled polygon) and performed principal components analysis (PCA). We report variable loadings of the first and second principal components (PCs) in [Supplementary data Fig. S2](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data). To provide an additional line of evidence for local climatic stability, we also compared interpolated Worldclim 2 data (mean for 1970–2000) with more recent independent meteorological records from four contemporary weather stations (data period 1999–2017) in the Central Desert. Additional mapping and data visualization were performed using the R packages 'ggplot2' [\(Wickham, 2009](#page-11-12)), 'rgeos' [\(Bivand](#page-9-12) *et al.*[, 2018](#page-9-12)) and 'RasterVis' [\(Lamigueiro, 2018](#page-10-38)).

Variable selection and environmental niche modelling

The suite of environmental layers retained for modelling was refined in three stages. First, layers that had low variability at the spatial scale of our study area were removed (e.g. soil). Secondly, correlated environmental variables across the study area $(r > 0.7)$ were grouped, with a single variable from each group considered to be most relevant to arid environment plant ecology retained. Thirdly, we performed an iterative selection procedure by removing variables with the highest variance inflation factor (VIF) using the package 'usdm' [\(Naimi](#page-10-39) *et al.*, [2014](#page-10-39)), with an upper threshold of VIF \leq 2.5. This resulted in a reduction from 54 to nine retained variables [\(Table 3\)](#page-5-0).

Environmental niche modelling was performed with Maxent v3.3.3 [\(Phillips](#page-10-40) *et al.*, 2006), implemented in the packages 'Dismo' (Hijmans *et al*., 2011) and 'ENMeval' [\(Muscarella](#page-10-41) *et al.*, [2014](#page-10-41)). To minimize model overfitting, species data (including historic and incidental observations) were geographically rarefied to a 3 km bin size and examined across environmental space.

Models were individually run and tuned for each study species over a study area encompassing the southern system, with quadrat surveys providing both presence and true absence data ([Table 2](#page-3-0)). Due to low sample sizes, data were partitioned using a jackknife approach where the number of model runs is equal to the number of occurrence localities, with a single data point excluded from each run for testing. Runs were performed iteratively across the full range of feature classes, with regularization multiplier values increasing from 1 to 4 in 0.5 increments.

Models were evaluated based on Akaike's information criterion corrected for small sample sizes (i.e. $\Delta AICc = 0$), which penalizes models that employ a greater number of parameters to describe the data ([Warren and Seifert, 2011](#page-11-6); [Muscarella](#page-10-41) *et al.*, [2014](#page-10-41)). We report AUC_{TEST} averaged over all iterations, with higher values reflecting better model discrimination of presence locations from absences. To quantify model overfitting, we use two metrics. First, we report the mean difference in AUC between training and test data (AUC_{DIFF}) ; this is expected to be higher where models are overfit to training data ([Muscarella](#page-10-41) *et al.*[, 2014\)](#page-10-41). Secondly, we report the proportion of testing localities with predicted habitat suitability values lower than the training locality with the lowest reported value (OR_{MTP}) . For each species, the best performing model was projected across the study area, and a maximum training sensitivity plus specificity (MaSS) logistic threshold, which balances the trade-off between omission and commission errors (Lobo *et al.*[, 2008;](#page-10-42) [Liu](#page-10-43) *et al.*[, 2016](#page-10-43)), was employed to estimate binary habitat area. For two species (*Ochradenus harsusiticus* and *Searsia gallagheri*) where model evaluation indicated evidence of weak overfitting, we used a minimum training presence (MTP) threshold to ensure that all training observations are included within the predicted suitable habitat area.

In an effort to understand the abiotic drivers of the resulting distributions, several studies have shown that ranking variable contributions successfully captures biologically important factors ([Kearney and Porter, 2009;](#page-10-44) [Searcy and Shaffer, 2016](#page-11-5)). To assess relative variable importance across species we compare ranked permutation importance using Kendall's *W*, corrected for ties, implemented in the package 'irr' [\(Gamer](#page-10-45) *et al.*, 2012). Secondly, we use linear regression to assess the relationship between the contribution of mean annual fog to model performance, and the species' trait score (see [Supplementary data Tables S1](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data) and [S2](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)). Finally we calculated niche overlap across study species using the method of [Warren](#page-11-13) *et al.* (2008) and then combined thresholded species distribution classifications to identify areas of spatial overlap and co-occurrence of multiple species. RGB images were obtained from Sentinel 2 (Copernicus Sentinel data 2015, processed by ESA, accessed from <https://remotepixel.ca/> on 20 December 2018) and plotted with increased contrast. Surface wind direction data, averaged for the months June to August (2015–2017), were obtained from the Global Forecasting System, via the package 'rWind' ([Fernández-López, 2018\)](#page-10-46).

RESULTS

Evaluation of regional climate identifies a weak influence of the southern monsoon system on the Jiddat Al Arkad of the southern system [\(Fig. 2](#page-5-1)). Concurrently, increased summer fog incidence in the southern Central Desert coincides with the warmest temperatures of the summer months, which appears

Environmental variable	AA	HG	OH	PP	SG	SH	Mean
Elevation		40.6	33.1	37.7		2.5	19.0
Thornthwaite aridity index	44.5		4.2		82.3	54.2	30.9
Annual mean temperature	0.5	22.6	31.3	30.4		17.3	17.0
Terrain roughness index	24		7.6	21.1	12.4	11.6	12.8
Aspect			10.2	1.6		10.2	3.7
Cloud cover seasonality	0.5	31.2		2.9	0.4	0.2	5.9
Mean annual fog		3.4	13.7	2.5	0.9		4.8
Night LST	22.7			.3	4.1	1.6	5.0
Mean annual cloud cover	0			2.4		2.5	1.2

Table 3*. Percentage contribution of environmental variables to environmental niche models across study species*

Predictors are ordered by mean value.

Fig. 2. (A) Elevation map of Oman, with annual offshore cloud cover percentage identifying the major southern and Indian monsoon climate patterns. Cloud cover over land is not shown, but is negligible for the Central Desert. Northern and southern Central Desert systems have previously been treated as separate floristic units, and are denoted by shaded polygons. Field surveys reported in this study were undertaken in the southern system. (B) Quarterly mean temperature across the Central Desert. (C) Quarterly night-time fog intensity (change in brightness temperature) across the Central Desert. Higher values (less negative) are indicative of greater fog intensity.

to result in cooler coastal night-time temperatures. This pattern is not observed in the northern system. The PCA of abiotic variables clearly differentiated the major regions of endemism ([Fig. 3\)](#page-6-0). Overall temperature-related variables were the major contributors to PC1, with climate seasonality and precipitation differentiating PC2 ([Supplementary data Fig. S2](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)). The two

Fig. 3. Principal component analysis of all environmental variables across principal ecoregions of Oman. Overall temperature-related variables were the major contributors to PC1, with climate seasonality and precipitation differentiating PC2. Grey points denote a random background sample from across Oman. The five major centres of plant endemism comprise (1) the Hajar Mountains; (2–3) the Dhofar Mountains comprising the Jabal Samhan and Jabal Al Qamar/Qara centres of endemism, combined for the purposes of this figure; (4) the northern Central Desert consisting of the Jiddat Al Harasis/Huqf; and (5) the southern Central Desert comprising the Sahil Al Jazir/Jiddat Al Arkad. The Empty Quarter is plotted (yellow) for reference as it comprises a significant proportion of Oman's land area, but is not considered a centre of endemism. The candidate refugium (red) is a sub-set of the southern system of the Central Desert.

Central Desert systems are found to be differentiated, but partially overlapping, with 88.5 and 62 % of points representing a unique climate variables for the northern and southern systems, respectively.

Fieldwork surveys assessed 195 quadrats throughout the Southern System and successfully located all seven regional endemics. Study species were recorded in 41 % of quadrats. In addition, 288 incidental and 68 historical observations were collated. After spatial filtering and data cleaning, 177 records were retained [\(Table 2;](#page-3-0) [Supplementary data Fig. S1](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)). Mean pH and EC across all quadrats was 7.73 (s.d. $= 0.44$) and 14.99 $(s.d. = 27.13)$, respectively. No significant difference in pH or EC values was identified across species [analysis of variance $(ANOVA) P > 0.05$. In our assessment of fog and drought functional trait adaptation, *Polycarpea pulvinata* scored highest, and *S. gallagheri* scored lowest, with low stature, low leaf area and hairs the most frequent adaptations.

Environmental niche modelling and variable importance

After filtering, we retained nine environmental variables for modelling [\(Supplementary data Table S3](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)). Ecological niche model evaluation statistics are reported in [Table 2.](#page-3-0) Modelling was not performed for *Polycarpaea jazirensis* due to insufficient data. AUC_{TEST} ranged from 0.76 (*P. pulvinata*) to 0.93 (*S. gallagheri*). AUC values are often lower for more widespread species, which may be the case for *P. pulvinata* and *O. harsusiticus* ([Jiménez-Valverde](#page-10-24) *et al.*, 2008). Model logistic habitat suitability projections are plotted in [Fig. 4](#page-6-1), with

FIG. 4. Environmental niche models for each of six study species across the southern Central Desert. Higher values are indicative of greater modelled habitat suitability. The dashed line denotes the southern Central Desert system.

the percentage contribution of variables reported in [Table 3](#page-5-0). Binawry threshold maps are provided in Supplementary data [Fig. S3.](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data) The most important variables varied substantially, with no evidence of consistent rank importance across species ($Wt = 0.079$, $P = 0.87$). Annual mean fog did not appear to rank highly for any species, and was not significantly associated with fog adaptation trait scores ($F_{14} = 0.42$, $P = 0.55$, R^2 = 0.10). Niche overlap was high in all pairwise comparisons (median = 0.84; [Supplementary data Table S5](http://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcz085#supplementary-data)).

Identification of climate refugia

Combined binary species distributions identified a key area where all study species are predicted to co-occur [\(Fig. 5A\)](#page-7-0). High predicted habitat suitability across all models was localized to the southern Jiddat Al Arkad. Satellite imagery shows the region, seasonal cloud cover and the prevailing summer wind direction in [Fig. 5B](#page-7-0). Independent contemporary weather station records provide an additional line of evidence. Whilst the three northern stations show elevated maximum daily temperatures (period 2002–2017) compared with the Worldclim 2 reference (1970–2000), Shalim station – close to our putative coastal fog- and cloud-affected area – shows summer maximum temperatures below the Worldclim 2 reference ([Fig. 6\)](#page-8-0).

DISCUSSION

In this study, we present evidence for a southern Arabian Pleistocene refugium in Oman's Central Desert ([Figs 2](#page-5-1) and [5](#page-7-0)). As shown by Sandel *et al.* [\(2011\),](#page-10-9) the negative relationship between endemism and the increasing velocity of changing climate is strongest in poorly dispersing species such as plants. Therefore, co-occurrence of a high number of endemic study species within a narrow monsoon-influenced region is indicative of a refugium with low climate change velocity [\(Sandel](#page-10-9) *et al.*, 2011; [Abellán and Svenning, 2014](#page-9-13); [Harrison and Noss, 2017\)](#page-10-47). Climate analysis identified cooler mean annual temperatures in the study area and the influence of coastal cloud and fog [\(Fig. 2](#page-5-1)), which, combined with novel survey data and environmental distribution modelling, suggests that the vegetation of the southern Central Desert is a relict of an earlier, more mesic period. This is further supported by the biogeography of genera such as *Aerva*, *Searsia* and *Ochradenus* which have global distributions from Africa to South-East Asia, yet with endemic species restricted to Arabia ([POWO, 2018](#page-10-48)), indicating support for a refugium further back into the Neogene. Thus, this system may represent the northernmost remnant of a continuous belt of mesic vegetation formerly ranging from Africa to Asia, with close links to the flora of East Africa ([Kürschner,](#page-10-3) [1998;](#page-10-3) [Patzelt, 2011\)](#page-10-49).

The relictual distributions that we observe appear to be driven by the interaction of climate and topographic factors, in particular the influence of the southern monsoon. It had been speculated that parts of the Central Desert may be at the fringe of the monsoon-affected area ([Patzelt, 2015](#page-10-8)), thus benefiting from occasional low clouds, cool winds and coastal fog during the southern monsoon, but previously this could not be tested because of the lack of climate stations. Here, using evidence from remote sensing, we demonstrate that the southern monsoon does indeed influence the southern system of the Central Desert, with patterns of night-time fog detected via the MODIS satellite also consistent with the limited reports available. This putatively places the southern Arabian coastal fog-influenced Central Desert together with other coastal fog deserts such as the Namib and Peruvian lomas [\(Cereceda](#page-9-8) *et al.*, 2008; [Henschel and Seely, 2008\)](#page-10-50), though, based on limited fog adaptation traits in the flora, fog intensity may be lower.

By applying these climate data to systematic field records of endemic plants, we develop a suite of models characterizing

FIG. 5. Identification of endemic species co-occurrence in the southern Central Desert. (A) Composite map of the binary distributions of six study species. The dashed line identifies a region of high diversity with potential as a candidate IPA. (B) False colour Sentinel 2 image of the high diversity area. The red line shows a primary road crossing the study area. The inset rose diagram shows the prevailing wind direction during the Khareef (Southern monsoon). Cloud cover mean is shown in blue.

Fig. 6. Locations of Omani weather stations contributing to interpolated WorldClim climate variables used in this study (1970–2000), as well as four independent validation weather stations from the Central Desert (1999–2017). Maximum daily temperature recorded at these stations is reported, with the black line denoting the historic Worldclim 2 average for this period. Recording periods for these stations vary; therefore, data points are coloured by year. Northern and southern systems are depicted in blue and orange, respectively.

each species' environmental niche. We show that despite small sample sizes, it is possible to generate robust niche models, incorporating true absence data, which identify important areas of plant diversity. Surprisingly, the relative importance of retained climatic and topographic variables differed substantially across study species. Therefore, we conclude that it is not a single set of environmental variables contributing to the distribution of this unique flora. For example, whilst overall, aridity and mean annual temperature are unsurprisingly important predictors in an arid environment, almost all retained variables are important across specific taxa. Therefore, our analysis does not support the hypothesis that it is predominantly fog that influences the distribution of this endemic flora; rather a range of factors appear to be important, consistent with the diverse traits and phylogenetic provenance of the species. We note, however, that on a finer spatial scale, factors such as microrefugia and fog hydrology may have greater importance ([Mclaughlin](#page-10-51) *et al.*, 2017).

Despite being recognized for its global importance, the arid Horn of Africa biodiversity hotspot is one of the most severely degraded, with <5 % of habitat considered to be in pristine condition [\(Mittermeier](#page-10-0) *et al.*, 2005; [Mallon, 2013](#page-10-1)). However, in our study we did not observe evidence for variation in land use that could explain the heterogeneous distribution of plant diversity. Key long-term threats to the Central Desert include overgrazing ([Ghazanfar, 2004](#page-10-15)) and climate change ([Almazroui](#page-9-5) *et al.*, 2013), with mean annual temperature for the Arabian Peninsula increasing at 0.6 °C per decade and a significant decreasing trend in annual rainfall ([Almazroui](#page-9-5) *et al.*[, 2013](#page-9-5)). It is also concerning that climate change has been associated with historic shifts in intensity and northward extent of the monsoon ([Fleitmann and Matter, 2009](#page-10-5)) and, elsewhere, a contemporary decline in coastal fog frequency [\(Johnstone and Dawson, 2010\)](#page-10-52), with strong implications for persistence of endemic flora. In our study area, a relatively minor shift in the northward extent of the monsoon could have significant implications for regional climate.

Refugia have been suggested as priority sites to conserve global biodiversity under climate change precisely because of their demonstrated ability to facilitate species survival under adverse conditions ([Keppel](#page-10-10) *et al.*, 2012). Based on previous studies, it is also likely that refugial populations harbour the highest genetic diversity across the species' distribution [\(Meister](#page-10-53) *et al.*, 2005), helping to build future evolutionary resilience (Sgrò *et al.*[, 2011](#page-11-2)). This may be particularly important in the Central Desert, where many endemic species have been restricted to only a single refugial location, reducing potential for subsequent population admixture (Petit *et al.*[, 2003\)](#page-10-54). The southern coast of the Arabian peninsula has also been predicted to contain a significant proportion of unassessed at-risk vascular plant species ([Pelletier](#page-10-55) *et al.*, 2018). In recent years, an Important Plant Area (IPA) programme has been initiated for the Middle East, which highlights the value of an ecological and evolutionary process-based view in identifying candidate conservation sites [\(Al-Abbasi](#page-9-4) *et al.*, 2010). Therefore, the site outlined here $(Fig. 5)$ $(Fig. 5)$, covering approx. 880 km², may be a candidate for further evaluation and consideration as an IPA.

In conclusion, this study makes an important contribution to our understanding of southern Arabian climate refugia, and the biogeographical origins of the endemic flora of Oman's Central Desert. In the future, we highlight the value of a network of detectors to characterize coastal fog across the landscape, particularly in the southern Central Desert. These would better enable an assessment of how coastal fog covaries with other readily available data sets such as topography, cloud cover and land surface temperature to enable higher resolution predictions of the influence of coastal fog on species distributions. More generally, we emphasize the value of predictive modelling in the region to advance beyond initial presence–absence grids, both to identify the drivers of biogeographic patterns and to prioritize sites for the conservation. In the future, further identification and characterization of southern Arabian climate refugia may be a useful strategy to support conservation in a global biodiversity hotspot.

SUPPLEMENTARY DATA

Supplementary data are available online at [https://academic.](https://academic.oup.com/aob) [oup.com/aob](https://academic.oup.com/aob) and consist of the following. Table S1: functional traits associated with desert fog moisture capture and drought tolerance. Table S2: fog and drought functional trait adaptation scores for desert endemic study plants from the Central Desert. Table S3: collated and retained environmental variables for environmental niche modelling of desert endemic plants. Table S4: environmental variable correlation matrix. Table S5: niche overlap of endemic Central Desert study species, estimated using Warren's I. Figure S1: species observation data for the

southern Central Desert. Figure S2: top 15 variable loadings of the first and second principal component axes. Figure S3: binary threshold maps based on MESS and MTP values for six study species in the Central Desert.

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