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Carbon (C) mineralization of crop residues is an important process occurring in soil which is helpful in predicting CO_2 emission to the atmosphere and nutrient availability to plants. A laboratory experiment was conducted in which C mineralization of residues of rice (*Oryza*)

placement in an Alfisols of Northwest India

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Carbon mineralization in soil as influenced by crop residue type and

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A laboratory experiment was conducted in which C mineralization of residues of rice (*Oryza sativa*), wheat (*Triticum aestivum*), maize (*Zea mays*), mungbean (*Vigna radiata*) and their mixtures was applied to the soil surface or incorporated into an Alfisols from Northwest India. C mineralization was significantly affected by residue placement and type and their interactions. Rice residue had a higher decomposition rate (k = 0.121 and $0.076 day^{-1}$) than wheat (0.073 and $0.042 day^{-1}$) and maize residues ($0.041 day^{-1}$) irrespective of placements. Higher decomposition rates of rice and wheat were observed when placed on soil surface than incorporated in the soils. Additive effects of the contribution of each residue type to C mineralization of the residue mixture were observed. When mungbean residue was added to the rice/wheat or maize/wheat mixture, decomposition of the residue mixture was enhanced. Crop residues with low N and high C/N ratio such as maize, wheat, rice and their mixtures can be applied on the soil surface for faster C and N mineralization, thereby helping to manage high volumes of residues under conservation agriculture-based practices in northwest India.

Introduction

ABSTRACT

Rice–wheat is a major crop rotation in the Indo-Gangetic Plains (IGP) of South Asia, spread over 13.5 million ha in Bangladesh, India, Nepal and Pakistan [1]. Productivity of the double-grain system cannot be sustained unless the declining trend in soil fertility resulting from nutrient mining by these crops is reversed [2] through healthy and well-designed supplements. Decreasing sources of organic manure (such as farmyard manure) along with escalating fertilizer prices and limited input availability have aggravated the undernourishment of agricultural fields.

Returning crop residues to soil is the most economical and ecological means of conserving soil and water for sustaining crop production [3]. Crop residues serve as a source of nutrients, help in maintaining soil organic carbon (SOC) and subsequently improve overall soil quality [4]. Release of essential nutrient elements in soil as well as emission of CO_2 into the atmosphere are controlled by the mineralization of organic matter in soil [5–7], making it the most important process occurring in soil. In India, to improve the productivity of the rice–wheat cropping system there is no doubt that an adequate and balanced supply of essential nutrients is of critical concern. Indian farmers, mostly marginal and small, do not apply the recommended doses of fertilizers. Crop residues represent a locally available source that has been often shown to enhance the efficiency of and reduce the requirement for chemical fertilizers [8]. Burning of these crop residues leads to loss of organic matter as well as plant nutrients such as nitrogen (N), phosphorus (P) and sulfur (S). The best alternative is to recycle the crop residues into agricultural land for improving soil health and crop productivity [4].

In northwest India, particularly Punjab and Haryana, conservation agriculture (CA) is widely adopted by farmers due to energy and water saving which ultimately leads to lower production costs and higher profitability [9]. In CA the soil is less disturbed and residues typically remain on its surface. A major question about these expanding practices is to what extent the decomposition and nutrient release of these residues is favored or

KEYWORDS

C mineralization; CO₂ emission; C/N ratio; decomposition rate; residue type and placement; residue mixture



retarded by its retention on the surface as opposed to its incorporation into the soil and how this may vary with different crop species and mixtures.

For better understanding of organic matter dynamics in soil, carbon and or nitrogen mineralization studies are essential. CA also includes a cover crop in the rotation where the residue of the cover crop acts as mulch on the soil surface. The fresh organic matter derived from the crop residues decomposes and is subsequently mineralized in soil. C mineralization of these crop residues results in the release of CO₂ to the atmosphere and nutrient elements for soil fertility improvement. The dynamics of C and N mineralization is directly linked with the decomposition of fresh plant material.

The initial placement of crop residues in soils can modify physical, biological and chemical conditions for decomposition and long-term soil conditions as well [10]. But there are contradictory reports on the effects of residue location on C mineralization. A non-significant effect of residue placement on C mineralization is also reported [11]. Other studies have shown that decomposition of crop residues was faster when they were incorporated into the soil rather than placed on the surface [10,12]. Residue type and nutrient availability in soil might have played an important role in such inconsistencies [10,13]. However, there is lack of information on the interactive effects of residue type and placement. The mineralization of crop residues returned to soils is controlled by many factors [10,14], of which residue type has long been identified as an important one [15,16].

In the CA system, rice and wheat are widely cultivated crops in the northwest IGP, and in many places rice is being replaced by maize and legumes, particularly mungbean, as a cover crop are included in the rotation. When crops are grown in association, different crop residues usually mix together and decompose simultaneously within the same soil [17]. Therefore, the present laboratory experiment was designed to examine C mineralization patterns of these four residues and their mixtures when placed on the soil surface as well as incorporated into soils.

Materials and methods

Soils and crop residues

Soils and crop residues used for the incubation study were collected from the seed production field (latitude 29°43'N, longitude 76°58'E, altitude 245m mean sea level) located at Indian Council of

Table 1. Initial characteristics of the soil used for the incubation study.

Soil property	Mean
pH (1:2 soil:water)	8.06
EC (dS m^{-1} ; 1:2 soil:water)	0.21
Total carbon (%)	0.65
Oxidizable organic carbon (%)	0.45
Available N (kg ha^{-1})	130
Available P (kg ha ^{-1})	15.73
Available K (kg ha ^{-1})	183.4
Total N (%)	0.14
Bulk density (Mg m ⁻³)	1.67

Agricultural Research (ICAR)-Central Soil Salinity Research Institute, Karnal, Haryana, India. Carbon mineralization was determined by a laboratory incubation experiment in this study. The studied soil corresponds to the Zarifa Viran soil series. It is a member of the fine-loamy mixed hyperthermic family of Typic Natrustalf [18].

Soil samples were collected from 0–15 cm depth in October 2013 after rice harvesting using a metal core sampler. After removal of the visible crop and residue debris, the soil samples were sieved through a 2-mm sieve and divided into two subsamples. One subsample was used to conduct the incubation experiment and another subsample was dried under shade and used for chemical analyses. The air-dried sample was used for analysis of pH, electrical conductivity (EC), oxidizable organic carbon, total C, available N, P and K, and sand, silt and clay contents. The values of the different soil parameters are presented in Table 1.

Crop residues were collected during the harvest in March (wheat) and May (mungbean), 2013 and October, 2013 (maize and rice). Triplicate samples were collected per crop type. The crop residues were mixed together for each replicate and air-dried. A portion of the samples was oven-dried at 60°C to determine the water content and ground prior to determination of the concentrations of total C and N. Other samples were also ground (<0.25 mm) and stored in sealed containers for the incubation experiment. The chemical compositions of the residue mixtures were calculated based on the mass ratios of the individual components (Table 2).

Soil analyses

Total C and N of the soil and crop residues were estimated using a CHNS Vario El III analyzer (Elementar, Germany). Soil pH (soil:water ratio 1:2) and EC were determined by following standard methods [19]. The soil available nitrogen was determined by the alkaline permanganate method [20], available phosphorus by the method of Olsen

Table 2. Initial chemical characteristics of the crop residues used for the study.

		•	
Residue type	Total C (%)	Total N (%)	C/N ratio
Rice	$26.86 \pm 0.64^{\circ}$	$0.617 \pm 0.03^{\circ}$	43.59 ± 1.07 ^d
Wheat	38.75 ± 0.16^{a}	$0.619 \pm 0.02^{\circ}$	62.67 ± 2.62^{a}
Maize	39.32 ± 0.05^{a}	$0.766 \pm 0.03^{\circ}$	$51.38 \pm 1.65^{\circ}$
Mungbean	39.04 ± 0.09^{a}	2.517 ± 0.01^{a}	15.51 ± 0.01^{e}
$Rice + wheat^{\dagger}$	32.81 ± 0.45^{b}	$0.62 \pm 0.01^{\circ}$	53.13 ± 1.50 ^c
$Maize + wheat^{\dagger}$	39.03 ± 0.40^{a}	$0.69 \pm 0.01^{\circ}$	57.02 ± 1.04^{b}
$Rice + wheat + mungbean^\dagger$	34.88 ± 0.52^{ab}	1.25 ± 0.02^{b}	40.59 ± 0.85^{d}
$Maize + wheat + mungbean^{\dagger}$	39.03 ± 0.28^{a}	1.30 ± 0.03^{b}	43.19 ± 0.70^{d}

[†]The residues were mixed in equal proportions on a dry weight basis (values followed by different lower-case letters are significantly different at P < 0.05 according to Duncan's multiple range tests for separation of mean). For all variables, $n = 3 \pm$ standard error of mean.

et al. [21], and available potassium with a flame photometer using neutral normal ammonium acetate [19]. Soil textural analysis was performed by following the International Pipette Method [22]. Soil bulk density was measured by the core method using metal cores 5 cm long with 5 cm internal diameter by placing the core in the middle of the surface soil layer [23].

Incubation experiment

In the incubation experiment, 15 treatments were included with three replicates. Treatments were applied as follows. Surface application of rice (R_s), maize (M_s) and wheat (W_s) residue was done. They were also incorporated in the soil, designated as rice (R_i), maize (M_i) and wheat (W_i). Mixtures of these residues were also used in the experiment, as surface application and by incorporation into the soil, of rice + wheat (RW_s and RW_i), maize + wheat (MW_s and MW_i), rice + wheat + mungbean ($RWMb_s$ and $RWMb_i$), maize + wheat + mungbean ($MWMb_s$ and $MWMb_i$), and Control (only soil). The ratio of the residues in the mixed residue treatments was equal on a mass basis.

Carbon mineralization was measured under controlled conditions using a modified laboratory incubation method [24]. Briefly, fresh soils were preincubated at 25°C for 24 hours to allow the microorganisms to adapt to the laboratory incubation conditions. Residues and their mixtures were incorporated into the soil homogeneously (equivalent to 20 g oven-dried soil) or placed as a single layer on the soil surface (soil:residue ratio 10:1) in a conical flask (500 mL in volume, 12 cm in height). The present experiment used a very high proportion of residue because in northwest India, farmers have been practicing conservation agriculture with zero tillage and residue retention for the last 15–20 years. Every year about 10–15 t ha^{-1} of residue has been added and this practice will be continued in the years to come. So this study has tried to capture the changes in C mineralization

under different residue placements and types with higher rates of residue addition.

The conical flasks containing soils with or without residues were incubated at a constant temperature (25 °C) and water content of soil (60% of water-holding capacity) for 106 days. A doublelayered high-quality parafilm cover was used to make the conical flasks air-tight, and they were placed in an incubator. The moisture content of the soil in the conical flasks was checked daily by measuring weight loss, and deionized water was added whenever needed. Gas samples were collected periodically (after 1, 3, 8, 15, 29, 43, 57, 78, 106 days) from the top of the conical flask using a 50-mL polypropylene syringe with a three-way leur lock and transferred immediately into a pre-evacuated and labeled 30-mL capacity vial. These vials were evacuated with an evacuator suction pump and labeled with the date and treatment. Collected samples were analyzed for CO₂ using a gas chromatograph (GC model: Varion Green House Gas Analyzer 450) equipped with a thermal conductivity detector (TCD for CO₂), with heater temperature settings of 200 °C with column oven temperature 50 °C. Zero air, hydrogen, nitrogen, 5% methane-argon and helium were used as carrier gas for the GC. The carrier gas helium was used for TCD with a flow rate of 20 mL min^{-1} . Concentrations of gases were calculated by comparing relative peak areas against the curves prepared from known concentrations of standard gases (100 ppm and 500 ppm for CO₂) obtained from Linde Specialty Gases, North America. To address the issue of GC drift, the GC was calibrated using these standards before running each batch of samples. Fluxes of CO₂ were calculated by linear regression of gas concentration against time of sampling. All data were checked for linearity by visual inspection during data analysis. After collection of gas samples, air was allowed into the conical flasks by opening them for 2 hours and the moisture content of soil was adjusted with distilled water on a weight basis. The C mineralization was calculated as the difference between the amount of CO₂ produced from

the samples containing residues and those without residues. It was assumed that the effect of external residue inputs on the mineralization of native soil organic matter was negligible and therefore was ignored due to non-availability of isotope labelling facilities. This is the total CO₂ evolved from decomposition of crop residues and native soil organic matter (i.e. priming effect). The amount of CO₂ evolved after 1, 3, 8, 15, 29, 43, 57, 78, and 106 days was added together to determine the cumulative C mineralization for the 106-day incubation period. The values of CO₂ evolved were divided by the mass of the soil samples (on an oven-dried weight soil basis) and were expressed as mg CO₂-C g⁻¹ soil. The cumulative C mineralization data were fitted into CurveExpert Professional version 2.2.0 software (a comprehensive data analysis software system for Windows, Mac and Linux) to find the trend of observed and predicted values of C mineralization.

Kinetics model and statistical analyses

The decomposition kinetics was fitted with a firstorder exponential model [25]. The first-order kinetic model is expressed as:

$$C_m = C_0 (1 - e^{-kt})$$
 (1)

where $C_m = \text{cumulative CO}_2\text{-C}$ emission, $C_0 = \text{potentially mineralizable C (mg g⁻¹), } k = \text{decomposition rate constant (day⁻¹), and t = time of incubation (days).$

Model evaluation

The model was evaluated using the goodness-offit test to compare observed data with predictions [26]. The mathematical expressions that describe these measures of analysis are the mean residual error (ME), mean relative error (MRE), the root mean square error (RMSE), modelling efficiency (EF), coefficient of determination (CD) and coefficient of residual mass (CRM). For a perfect fit between observed and simulated data, values of ME, MRE, RMSE, CRM, CD and EF should equal 0.0, 0.0, 0.0, 0.0, 1.0 and 1.0, respectively.

n

Mean residual error (ME):

$$\mathsf{ME} = \frac{\sum_{i=1}^{n} (O_i - P_i)}{n} \tag{2}$$

Mean relative error (MRE):

$$\mathsf{MRE} = \frac{\sum\limits_{i=1}^{n} |O_i - P_i|}{nO_i} \tag{3}$$

Root mean square error (RMSE):

$$\mathsf{RMSE} = \left[\frac{\sum\limits_{i=1}^{n}\left(P_{i} - O_{i}\right)^{2}}{n}\right] \times \frac{n}{\bar{O}} \tag{4}$$

Coefficient of determination (CD):

$$CD = \frac{\sum_{i=1}^{n} (O_i - \bar{O})^2}{\sum_{i=1}^{n} (P_i - \bar{O})^2}$$
(5)

Modelling efficiency (EF):

$$EF = 1 - \frac{\sum_{i=1}^{n} (P_i - O_i)^2}{\sum_{i=1}^{n} (O_i - \bar{O})^2}$$
(6)

Coefficient of residual mass:

$$\mathsf{CRM} = \frac{\sum_{i=1}^{n} (O_i - P_i)^2}{n\bar{O}}$$
(7)

where:

 P_i = Predicted values O_i = Observed values O = Mean of the observed data n = Number of samples

Results

Chemical properties of the crop residues

Among the four crop species, mungbean residues had the highest total nitrogen (TN) concentration and the lowest C/N ratio, whereas rice residues showed the lowest total C and wheat residue showed the highest C/N ratio (P < 0.05; Table 1). Rice and wheat residues had similar TN concentrations, whereas maize, wheat and mungbean showed similar total C concentrations. Averages for these values were used to calculate the properties of the residue mixtures.

C mineralization

Significant variation was observed with both residue type and placement, and mineralization was always lowest in unamended soils throughout the incubation period (Figures 1 and 2). Among the treatments, where residues and their mixtures were incorporated into the soil, RWMb_i showed consistently higher cumulative C mineralization up to 43 days (Figure 1). A steep increase was observed in W_i after 15 days, and cumulative C



Figure 1. Changes in cumulative carbon mineralization (μ g C g⁻¹ soil) on different days under different residues incorporated in soil. Vertical bars indicate \pm standard error of the mean of the observed values.

 R_{i} : Rice, incorporated into the soils. W_{i} : Wheat, incorporated into the soils. M_{i} : Maize, incorporated into the soils. RW_{i} : Rice + wheat, incorporated into the soils. RWM_{i} : Rice + wheat + mungbean, incorporated into the soils. RWMb_i: Rice + wheat + mungbean incorporated into the soils. Control: only soil.



Figure 2. Changes in cumulative carbon mineralization (μ g C g⁻¹ soil) on different days under different residues surfaceapplied to the soil. Vertical bars indicate ± standard error of the mean of the observed values. R_i: Rice, incorporated into the soils. W_i: Wheat, incorporated into the soils. M_i: Maize, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. RWM_i: Rice + wheat + mungbean, incorporated into the soils. MWMb_i: Maize + wheat + mungbean incorporated into the soils. Control: only soil.

mineralization exceeded the value of RWMb_i at 57 days of incubation and showed its highest value (666.8 μ g C g⁻¹ soil) after 106 days of incubation (Figure 1). Treatment R_i showed a consistent increase in cumulative C mineralization up to 29 days of incubation, after which no more increase was observed and it became constant and parallel to the x-axis. After 106 days of incubation, the highest (666.8 μ g C g⁻¹ soil) cumulative C mineralization was observed under W_i followed by RWMb_i (640.4 μ g C g⁻¹ soil), RW_i (610 μ g C g⁻¹ soil), MWMb_i (600

 μ g C g⁻¹ soil), M_i (556.6 μ g C g⁻¹ soil), MW_i (549 μ g C g⁻¹ soil) and R_i (413.7 μ g C g⁻¹ soil) (Figure 1).

Among the surface-applied residue treatments, MWMb_s produced the highest (682.8 μ g C g⁻¹ soil) cumulative C mineralization followed by MW_s (659.4 μ g C g⁻¹ soil), RWMb_s (605.3 μ g C g⁻¹ soil), M_s (574.4 μ g C g⁻¹ soil), W_s (548.6 μ g C g⁻¹ soil), RW_s (538.9 μ g C g⁻¹ soil), and Control (366 μ g C g⁻¹ soil) after 106 days of incubation (Figure 2). MWMb_s showed consistently higher C mineralization than the other treatments throughout the

Treatment		Decomposition kinetics model			
	C mineralization (μ g C g ⁻¹ soil)	$C_0 (mg g^{-1})$	k (day ⁻¹)	R ²	
R _s	366.06 ± 15.47	0.340 ± 0.038	0.121 ± 0.009	0.94	
R _i	413.70 ± 12.01	0.403 ± 0.030	0.076 ± 0.002	0.98	
Ŵ,	548.61 ± 13.75	0.501 ± 0.027	0.073 ± 0.005	0.94	
Wi	666.77 ± 10.87	0.664 ± 0.023	0.042 ± 0.001	0.97	
Ms	574.39 ± 9.05	0.560 ± 0.021	0.041 ± 0.002	0.98	
Mi	556.56 ± 8.45	0.528 ± 0.019	0.041 ± 0.002	0.97	
RWs	538.89 ± 11.80	0.488 ± 0.008	0.065 ± 0.003	0.93	
RW _i	610.02 ± 9.29	0.605 ± 0.042	0.040 ± 0.001	0.96	
MWs	659.64 ± 11.94	0.622 ± 0.034	0.050 ± 0.002	0.96	
MWi	549.12 ± 7.89	0.515 ± 0.003	0.044 ± 0.001	0.88	
RWMb,	605.27 ± 16.09	0.554 ± 0.012	0.081 ± 0.004	0.93	
RWMb _i	640.41 ± 16.89	0.599 ± 0.052	0.075 ± 0.008	0.94	
MWMBs	682.75±13.58	0.544 ± 0.021	0.065 ± 0.004	0.96	
MWMb _i	600.10 ± 13.89	0.568 ± 0.042	0.070 ± 0.009	0.94	
T15 [Control (only soil)]	121.59 ± 14.75	0.123 ± 0.024	0.029 ± 0.001	0.97	
P values					
Туре	<0.001				
Placement	<0.001				
Time	<0.001				
Type $ imes$ placement	<0.001				
Time × type	<0.001				
Time \times placement	<0.001				
$Time \times type \times placement$	<0.001				

Table 3. Cumulative C mineralization (μ g C g⁻¹ soil) in soils amended with different residues after 106 days of incubation.

Estimated mean values of the model parameters are presented. C₀, potentially mineralizable C and k, decomposition rate.

Rs: Rice, surface-applied. R_i: Rice, incorporated into the soils. Ws: Wheat, surface-applied. W_i: Wheat, incorporated into the soils. Ms: Maize, surface-applied. M_i: Rice + wheat, incorporated into the soils. RWs: Rice + wheat, surface-applied. RW, Rice + wheat, incorporated into the soils. MWs: Maize + wheat, incorporated into the soils. RWMs: Rice + wheat, surface-applied. RWMbs: Rice + wheat + mungbean, surface-applied. RWMbs; Rice + wheat + mungbean, surface-applied. RWMbs; Rice + wheat + mungbean, surface-applied. RWMbs; Maize +

incubation period. Up to 15 days of incubation, there was an increase in cumulative C mineralization in all the treatments, but after that Control showed a declining trend and became constant and parallel to the x-axis (Figure 2).

Results show that cumulative C mineralization was significantly (both P < 0.001) affected by residue type and placement, whereas significant interactions between them were observed for cumulative C mineralization (P < 0.001; Table 3). Sampling time and its interactive effects with residue placement were also significant for C mineralization (both P < 0.001). There was significant interaction between sampling time and residue type on C mineralization, and the interactive effects of sampling time, residue type and placement were also statistically significant (Table 3).

The carbon mineralization data were well fitted to a first-order exponential model ($R^2 = 0.88 - 0.99$; Table 3 and Figure 6). Rice residue had a higher decomposition rate constant (k values) than did wheat and maize residue irrespective of their placements (Table 3). The k values for all the treatments where rice, wheat and mung bean residues and their mixtures were placed on the soil surface were higher than those of the same residues incorporated into the soils, except the treatments with maize residue alone and a mixture of maize, wheat, and mung bean residues, where the opposite trend was observed (Table 3). This indicates that the mixed residues supported soil microbial activities more readily when were applied on the soil surface. Significant differences in cumulative C mineralization of the residue mixtures were found between the observed and predicted values for the two placements (P > 0.05, Figure 6), which suggests that mere arithmetic prediction is neither sufficient nor scientific for guessing the micro-biochemical treatments that the residue mixtures receive in the soil but the individual residues do not.

C mineralization rate

Rates of C mineralization varied significantly with incubation time under both residue placements. Initially all the treatments where residue was incorporated showed an increase in C mineralization rate up to the third day of incubation, except MWMb_i which showed an inconsistent rate up to 29 days (Figure 3). An initially higher rate of C mineralization was observed under RWMb_i treatment compared with other treatments, up to 15 days, and then it sharply decreased until the end of the incubation period. The highest increase (0.0040 µg C g^{-1} soil d^{-1}) in C mineralization rate was observed under W_i at 29 days of incubation (Figure 3). After 29 days, R_i showed the highest decrease in mineralization rate, which was lower than the control treatment at 57 days of incubation. After 43 days, the highest (0.0028 μ g C g⁻¹



Figure 3. Changes in rate of carbon mineralization (μ g C g⁻¹ soil) on different days under different residues incorporated into the soil. Vertical bars indicate ± standard error of the mean of the observed values. R_i: Rice, incorporated into the soils. W_i: Wheat, incorporated into the soils. M_i: Maize, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. RWMb_i: Rice + wheat + mungbean, incorporated into the soils. MWMb_i: Maize + wheat + mungbean incorporated into the soils. Control: only soil.

soil d⁻¹) C mineralization rate was observed under treatment RW_i followed by MW_i (0.0020 μ g C g⁻¹ soil d⁻¹). At 78 days of incubation, MW_i showed the highest (0.0023 μ g C g⁻¹ soil d⁻¹) increase in C mineralization rate followed by W_i (0.0016 μ g C g⁻¹soil d⁻¹), and after that a decrease in mineralization rate was observed at 106 days of incubation (Figure 3).

All the treatments where residue was applied to the soil surface showed an increase in C mineralization rate up to 8 days of incubation, except W_s and RWs treatments which showed a slight decrease in mineralization rate (Figure 4). The highest C mineralization rate (0.0033 μ g C g⁻¹ soil d⁻¹) was observed under the MWMb_s treatment followed by M_s (0.0028 μ g C g⁻¹ soil d⁻¹), MW_s (0.0026 μ g C g⁻¹ soil d⁻¹), RWMb_s (0.0024 μ g C g^{-1} soil d^{-1}) = RW_s (0.0024 µg C g^{-1} soil d^{-1}) and W_s (0.0023 μ g C g⁻¹ soil d⁻¹) at 15 days of incubation (Figure 4). Treatments RWMb_s and R_s showed a decrease in mineralization rate at 15 days of incubation. The highest decrease in mineralization rate was observed under the Rs treatment which went lower than the Control at 29 days of incubation, and then increased steeply at 43 days where other treatments showed a sharp decrease in

mineralization rate. At 57 days of incubation, all the treatments showed a decline in the mineralization rate, except RW_s which recorded an increase in the mineralization rate. After that, $RWMb_s$ and W_s showed a slight increase in C mineralization rate and the others were parallel to the x-axis (Figure 4).

Discussion

C mineralization as influenced by residue placement

Higher cumulative C mineralization was observed for residues incorporated into soils than placed on the soil surface, except for treatments where maize residue was applied. Li *et al.* [25] also observed higher cumulative C mineralization with maize residues placed on soil surface than incorporated into soils. The effect of residue placement was significant (P < 0.001). The decomposition rate constant (k) values for rice and wheat residues applied on surface soil were higher than those of the same residues incorporated into the soil (Table 3), whereas a similar k value was observed with maize residue placed on the surface as well as



Figure 4. Changes in rate of carbon mineralization (μ g C g⁻¹ soil) on different days under different residues surfaceapplied to the soil. Vertical bars indicate ± standard error of the mean of the observed values. R_i: Rice, incorporated into the soils. W_i: Wheat, incorporated into the soils. M_i: Maize, incorporated into the soils. RW_i: Rice + wheat, incorporated into the soils. MW_i: Maize + wheat, incorporated into the soils. RWMb_i: Rice + wheat + mungbean, incorporated into the soils. MWMb_i: Maize + wheat + mungbean incorporated into the soils. Control: only soil.

incorporated into soil. MWMb_i showed a higher k value than the surface application of the same residues (Table 3). Giacomini et al. [10] observed a higher C mineralization rate under residue incorporation than surface application. Sarma et al. [7] observed higher cumulative CO2 emission with the addition of organic amendments, and the highest increase was associated with vermicompost incorporated into soil. A higher C mineralization rate was observed under MWMb_i treatment than surface application. Treatments where maize residues were surface-applied (M_s) and incorporated into soil (M_i) showed similar C mineralization rates, which is in accordance with the findings of Abiven and Recous [11] who did not find any significant differences in C mineralization rate between surface-applied and soil-incorporated residues. The initial location of high N content and residues rich in soluble compounds in the soils did not affect C mineralization [12]. But Sarma et al. [27] found that organic substances such as fresh crop residues, farmyard manure, and composts with less humified organic amendments are more vulnerable to

mineralization than stable structured organics (biochar, C-rich pyrolyzed product of biomass). Soil physical, chemical and biological properties were modified by the contact between soils and residues, ultimately affecting the decomposition of the residues [28]. Many studies showed that closer contact between soil and residue facilitated the decomposition of the residues [12,29,30]. The chemical characteristics (soluble compounds) and morphological features of the residues, particle size [24], and location in the soil [31,32] influenced the residue-soil contact. Moisture content and temperature also affect the C mineralization of the residues incorporated or surface-applied [14,33]. But in the present case temperature and moisture content were controlled throughout the experiment. When residues are applied on the soil surface or incorporated, microorganisms act on the residues. Oxygen availability to the microorganisms is an important factor affecting the decomposition of residues [34]. When residues are incorporated into the soil, gas diffusivity might be lower, thereby reducing the residue

decomposition. Residues applied on the soil surface produced more CO₂ than did residues incorporated into soils [25]. In this study, this is true for the treatments where maize residues were applied singly or in combination with wheat or mungbean residues. Incorporation of maize residues in soil inhibited decomposition, which might be attributable to a lack of oxygen availability to the microorganisms [25]. This might also have happened with the treatments MWMb_i, M_s, and M_i, where lower or on-par values of the decomposition constant were observed. Except for the maize residue treatments, all treatments showed higher cumulative C mineralization when incorporated into the soil than when surface-applied. This might be attributable to the closer contact between soil and residues incorporated into the soil, leading to faster decomposition of residues [12,29,30], although the rate of decomposition was higher in treatments where residues were surface-applied.

For decomposition of crop residues in soil, microorganisms need nitrogen. So, the availability of N in soil is an important factor determining decomposition of residues in soil. The rate of C and N mineralization in soil is regulated by the C/ N ratio of the crop residues [35]. The N mineralization in soil upon residue addition was not studied. But in general, when residues are applied in soil, immobilization of N takes place. In this study, in the treatments where mungbean residues were applied in combination with rice, wheat or maize residues, higher C mineralization was observed. This might be attributable to higher N content as well as a lower C/N ratio of mungbean leading to higher C mineralization in soil (Table 2). Li et al. [25] also observed greater N immobilization upon maize residue incorporation in soil than with surface application. Abiven and Recous [11] conducted a study where they maintained the mineral N content in the soil at 80 ppm through supplemental N to check N limitations during the process of decomposition. The present study did not use any N supplement, which might have caused nitrogen limitation for the microorganisms during the decomposition process of the residues. This might have resulted in different residue placement effects on C mineralization in the present study.

Residue-type effect on C mineralization

Crop residues with high N concentration and low C/ N ratio showed higher initial C mineralization as compared to residues with low N and high C/N

ratios (Table 2; Figures 1 and 2). Many researchers have reported similar observations [5,25,36]. The concentration of readily assimilable substrate and availability of nitrogen in soil determine the activity of the microorganisms [37]. Microorganisms require nitrogen for their metabolism and cell synthesis; normally 1 part N is required for every 10 parts C assimilation [38]. Decomposition of residue releases nitrogen into the soil which serves as an N source for the microorganisms, therefore favoring a stimulating effect on C mineralization [25]. Therefore, C mineralization of the residues might have been affected by the initial N concentration. Because mungbean is a legume, its residue decomposed faster than rice, wheat and maize residues irrespective of placement (Table 3), as it contains higher N and has a lower C/N ratio (Table 2). In the treatments where mungbean residue was mixed with rice, wheat and maize, higher C mineralization was observed than without mungbean residue, irrespective of their placement (Figures 1, 2 and 5). Many studies showed that addition of high C/N ratio residues to soil caused immobilization of N during the process of decomposition [39,40]. In this experiment, a high C/N ratio was observed for rice, wheat and maize residues as well as their combinations (Table 2). Moreover, the soil used for the experiment contained low available N (130kg N ha⁻¹) which might have caused immobilization of N, and ultimately there was a limitation of N for the microorganisms for decomposition of rice, wheat and maize residues in soil.

Mixed effects of residues on C mineralization

As the C and N concentrations of different residues were not similar (Table 2), the C mineralization of



Figure 5. Changes in cumulative C mineralization in soils amended with mixed residues of rice, wheat, maize and mungbean after the 106-day incubation.

Table 4. Statistical parameters to evaluate the observed and predicted cumulative C mineralization.

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Treatment	MRE	ME	RMSE	CD	CRM	EF
Rs	0.130	2.25	25.418	0.959	0.009	0.949
R _i	0.081	3.98	17.414	0.908	0.014	0.984
Ŵ,	0.132	8.40	36.919	0.859	0.024	0.951
Wi	0.151	12.73	32.766	0.855	0.032	0.978
Ms	0.153	7.99	27.580	0.873	0.024	0.978
Mi	0.164	10.24	30.378	0.860	0.033	0.970
RŴ,	0.156	9.70	42.821	0.856	0.029	0.931
RWi	0.181	12.85	40.502	0.838	0.036	0.958
MW,	0.157	12.28	38.486	0.852	0.031	0.966
MWi	0.248	17.27	54.649	0.778	0.054	0.889
RWMb,	0.151	11.09	41.315	0.831	0.028	0.947
RWMb _i	0.144	11.82	40.598	0.828	0.028	0.957
MWMB,	0.233	76.49	84.729	1.05	0.177	0.846
MWMb _i	0.159	12.37	39.205	0.812	0.032	0.955
Control (only soil)	0.214	3.16	6.713	0.821	0.049	0.969

CD: Coefficient of determination. CRM: Coefficient of residual mass. EF: Modeling efficiency. ME: Mean residual error. MRE: Mean relative error. RMSE: Root mean square error.

Rs: Rice, surface-applied. R_i: Rice, incorporated into the soils. Ws: Wheat, surface-applied. W_i: Wheat, incorporated into the soils. Ms: Maize, surface-applied. M_i: Maize, incorporated into the soils. RWs: Rice + wheat, surface-applied. RW_i: Rice + wheat, incorporated into the soils. MWs: Maize + wheat, incorporated into the soils. RWMs: Rice + wheat, surface-applied. RW_i: Rice + wheat + mungbean, surface-applied. RWMb_i: Rice + wheat + mungbean, surface-applied. RWMb_i: Rice + wheat + mungbean, surface-applied. RWMb_i: Maize + wheat + mungbean, surface-applied. RWMb_i



Figure 6. Graphical presentation of observed and predicted cumulative C mineralization under different residues applied to soil.

those residues in single or mixed forms was different (Figure 5). A synergistic effect of mixed residues was found when mungbean residue was added with rice/wheat and maize/wheat residues irrespective of placement (Figure 5). When rice or maize residue was added with wheat residue, there was a decrease in C mineralization irrespective of placement, except in the case of maize + wheat placed on the soil surface (Figure 5). A higher decrease was observed when the residues were incorporated into soil, which might be due to lower O_2 gas diffusivity causing O_2 limitation to decomposer microorganisms [25]. A similar negative effect of residue mixtures was also observed by Robinson *et al.* [41]. But in the case of the maize + wheat residue mixture placed on the soil surface, about 20% higher C mineralization than with only wheat surface-applied was observed (Figure 5). At the soil surface, higher O_2 availability to microorganisms might have facilitated the decomposition of maize residues. The higher N and lower C/N ratio of mungbean residues facilitated the



Figure 6. Continued.

decomposition of mixed residues of rice/wheat or maize/wheat through higher activities of decomposer microorganisms. A similar synergistic effect of residue mixtures on C mineralization was reported by McTiernan et al. [42]. Interestingly, the highest rates of C mineralization on the 29th, 57th and 78th days of incubation were recorded with the treatments W, RW and MW, respectively, where wheat residue was the sole component or one of the components. This may indicate that rice and maize residues are more readily mineralized than wheat (an initial acceptability sequence can be drafted as R > M > W) due to which the peak shifts toward the right from W to RW and MW, but after some initial difficulty, wheat residue becomes the most acceptable substrate for the microbiota in the soil (Figures 3 and 4).

Cumulative C mineralization as well as the decomposition rate constant in all the treatments except rice and wheat were higher when residues as well as their mixtures were placed on the soil surface rather than incorporated into soil. This implies higher mineralization of organic carbon leading to more nutrient availability to plants under CA-based practices with a maize-based cropping system. But higher mineralization also

produces higher CO₂ to the atmosphere. Therefore, in situ collection of CO₂ as well as GHGs under different CA-based practices deserves further experimental study. As this study used a much higher rate of residue addition to soil, the potential C mineralization in all the treatments except the control were lower than the cumulative C mineralization. This indicates the process of mineralization was still going on even after the last observation on the 106th day. As a result, higher cumulative C mineralization was found under rice/ wheat residue when incorporated into soil singly or in combination, but the decomposition rate constant was higher for the surface-applied treatments (Table 3). Therefore, the incubation experiment should have been continued for another 2-3 months to achieve complete mineralization.

Prediction of C mineralization

For prediction of the cumulative C mineralization under different residues and their mixtures, the observed cumulative C mineralization was fitted to a first-order exponential model using the curveexpert software (Figure 6). From there the potential C mineralization was found as well as the decomposition rate constant (Table 3). This method provided an excellent visual interpretation of the differences between observed and predicted values of C mineralization for all the treatments with different residues and their mixtures placed on the soil surface or incorporated into the soil, in terms of bias and point scatter. Moreover, the results from the graph were evaluated by calculating the MRE, ME, RMSE, CD, CRM and EF, which gave a measurement of bias and scatter. The values of six statistics for the observed and predicted cumulative C mineralization are presented in Table 4. The ME and MRE represent the worst-case performance of the model, whereas the RMSE value shows how much the predictions overestimate or underestimate the measurements. In all the graphs, the dark-colored area indicates the confidence interval and the light-colored area indicates the prediction band. For all treatments, the MRE values were very low, but the treatments except rice incorporated into soil where wheat or maize, singly or in combination, was applied at the soil surface showed the lowest value and best fit the model. But the reverse took place with the treatments of mixed residues of rice/ wheat/maize/mungbean that were incorporated into the soil. In the case of ME, surface-applied treatments were best fitted with the model with lower values, except the MWMb_i treatment which was well fitted with the model. The lower value of MRE and the higher positive values of RMSE indicate a considerable scatter, and a very low value of CRM shows a tendency to overestimate. Except for the M_s and MW_s treatments, lower values of RMSE were observed in treatments where residues and their mixtures were incorporated and best fitted to the model. Treatments where residues and their mixtures were applied on the soil surface showed the lowest values of CRM, except maize + wheat + mungbean treatment showing lower CRM under incorporation. Higher values of CD were observed under surfaceapplied treatments indicating best fit to the model. Except for the M_s and MW_s treatments, higher modelling efficiency was observed under treatments where residues and their mixtures were incorporated. Overall, the statistics imply that the residues and their mixtures placed at the soil surface produced a satisfactory fit to the model compared with the residues incorporated into the soil.

Conclusion

Maize residues had a lower decomposition rate constant than rice and wheat residues due to low N content. Additive effects were observed when

mungbean residue was added to both rice/wheat and maize/wheat mixtures, due to the low C/N ratio of mungbean. Higher C mineralization was observed when the residues as well as their mixtures were placed on the soil surface as compared to incorporated into the soil. A significant interaction between residue type and placement on C mineralization was observed. Residues decompose faster when surface-applied, thereby helping to manage a high volume of residues and sustainable intensification/diversification through inclusion of mungbean, which provides further opportunities to manage residues. Faster decomposition may release nutrients faster, which may lead to a higher crop yield, and hence the emission intensity will be lower. Therefore, CA with surface retention of mixed crop residue may serve as the best alternative and most efficient solution for eliminating residue burning. When conducting incubation experiments with higher rates of residue addition, the duration of the incubation should be more than 6 months for complete mineralization of the residues.

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