

Research



Cite this article: Abbas HA, Nasr RA, Abu-Zurayk R, Al Bawab A, Jamil TS. 2020 Decolourization of crystal violet using nano-sized novel fluorite structure $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ photocatalyst under visible light irradiation. *R. Soc. open sci.* **7**: 191632. <http://dx.doi.org/10.1098/rsos.191632>

Received: 16 October 2019

Accepted: 11 February 2020

Subject Category:

Chemistry

Subject Areas:

nanotechnology/environmental chemistry/
environmental engineering

Keywords:

$\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$, fluorite phase, wastewater treatment, nanomaterials, crystal violet dye, photocatalytic degradation

Author for correspondence:

Rabab A. Nasr

e-mail: rababelsheikh@yahoo.com

This article has been edited by the Royal Society of Chemistry, including the commissioning, peer review process and editorial aspects up to the point of acceptance.

Electronic supplementary material is available online at <https://doi.org/10.6084/m9.figshare.c.4880334>.



Decolourization of crystal violet using nano-sized novel fluorite structure $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ photocatalyst under visible light irradiation

H. A. Abbas¹, Rabab A. Nasr², Rund Abu-Zurayk³,
Abeer Al Bawab³ and Tarek S. Jamil²

¹Inorganic Chemistry Department, and ²Water Pollution Control Department, National Research Center, El Behouth Street, PO Box 12622, Dokki, Cairo, Egypt

³Chemistry Department School of Science, The University of Jordan, Hamdi Mango Centre for Scientific Research, 11942, Amman, Jordan

HAA, 0000-0002-6387-224X; RAN, 0000-0003-2613-7032; RA-Z, 0000-0002-1072-1624; AAB, 0000-0003-2131-1791; TSJ, 0000-0002-1856-8046

Fluorite-type Zr-based oxides with the composition $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ ($x=0, 0.05, 0.1, 0.15$ and 0.2) were prepared using the citrate technique. Appropriate characterizations of all prepared materials were carried out. X-ray diffraction clarified that the undoped and W-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$ samples were crystallized in the cubic fluorite phase structure. The average particle size of the samples was in the range of 3–8 nm. The lowest band gap (1.7 eV) and the highest surface area ($124.3 \text{ m}^2 \text{ g}^{-1}$) were recorded for $\text{Ga}_2\text{Zr}_{0.85}\text{W}_{0.15}\text{O}_7$. The photocatalytic impacts of the prepared systems were studied by removal of crystal violet (CV) dye employing visible light illumination and taking into consideration the initial dye concentrations, duration of visible irradiation treatment, catalysts dose and the dopant concentration. The obtained results showed higher dye removal with the boost of the catalyst dosage. W doping shifted the absorption to the visible light range by decreasing the band gap from 4.95 eV for parent $\text{Ga}_2\text{Zr}_2\text{O}_7$ to 1.7 eV for 15 mol% tungsten-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$ enhancing the photocatalytic decolourization of CV from 4.2% to 83.6% for undoped and 15 mol% W-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$, respectively, at optimum operating conditions (pH 9, 1 g l^{-1} catalyst dose and 300 min) while heavily doped W sample containing 20 mol% W showed lower removal than 15 mol% W-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$. Complete CV degradation using 15 mol% W-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$ was attained with the assistance of 25 mmol l^{-1} hydrogen peroxide. The reaction is aligned to

pseudo-first-order kinetics. Different scavengers were introduced to decide the significance of the reactive species in CV degradation. $O_2^{\bullet-}$ and h^+ had the major role in the degradation of CV by $Ga_2Zr_{2-x}W_xO_7$ system compared with HO^{\bullet} .

1. Introduction

Crystal violet (CV) dye is triphenylmethane cationic dye (figure 1). It is used in textile and paper dye industries as well as navy blue and black inks for printing, ball-point pens and inkjet printers. It is also used to colourize diverse products such as fertilizers, antifreeze, detergents and leather. CV is also used as a histological stain, particularly in Gram staining for classifying bacteria [1,2].

Disposal of dyes in wastewater is a source of water contamination and disturbance of aquatic life [3]. Therefore, a suitable and efficient method is critically required to treat the wastewater containing dyes such as CV [4,5] for its proven carcinogenic and mutagenic properties in animals [6,7] and in humans [8].

Conventional techniques such as biodegradation, coagulation, adsorption, physical deposition conventional oxidants and coagulants were inefficient for CV treatment [9,10]. On the other hand, advanced oxidation processes (AOPs) such as microwave catalysis, photocatalysis, membrane technique and advanced oxidants [11–13] are promising in CV decolourization.

The essential defect of physical treatment involves only moving the dyes from the liquid to the solid state which is not easy to decontaminate [14]. Therefore, chemical treatment using AOPs, especially the heterogeneous photocatalysis, received attention for degrading such pollutants [15]. In heterogeneous AOPs, the metal oxides produce some powerful non-selective hydroxyl radicals (HO^{\bullet}) that dissociate a wide range of organic contaminants [16] into short-chain aliphatic acid that is easier to be completely degraded [17]. In the UV–visible light, the electron–hole pair mechanism is demanded in order to introduce intermediate organic compounds that might be completely mineralized at the surface of metal oxides attaining green end products [17]. From the economic side, novel nano-sized photocatalysts' response to visible light received valuable consideration since it is cost-effective compared with UV light [18–20].

Decolourization of CV has been studied using different oxidants such as nanosphere TiO_2 [20], Mn-doped and PVP-capped ZnO NPs [21], Ag-modified Ti-doped- Bi_2O_3 [22], ZnS NPS [23], CeO_2 – TiO_2 nanocomposite [24], AgBr–ZnO nanocomposite [25], grafted sodium alginate/ZnO/graphene oxide [26]. Afterwards, the performance of the prementioned oxides will be compared with the results of the present study.

$A_2B_2O_7$ oxides (where A and B abbreviate trivalent lanthanides elements and tetravalent D and F groups elements, respectively) have either a pyrochlore-type or a defect fluorite-type structure. They have attractive physical and chemical properties, such as high melting point, high thermal expansion coefficient, low thermal conductivity, high thermal stability, high radiation stability and high electrical conductivity. Consequently, they are used in several applications such as solid electrolytes, thermal barrier coating materials, nuclear waste host materials and high-temperature heating elements [27]. The electrical properties of the pyrochlores vary from highly insulating through semiconducting to metallic behaviour [28]. Many studies prepared various pyrochlore metal oxides such as $La_2Zr_2O_7$ (that acts as thermal barrier coating), [29] $Y_2Sn_2O_7$ (that acts as excellent host matrices for photoluminescence) [30] and $Gd_2Zr_2O_7$ (that acts as a proper host material for fixation of some of the nuclear waste products) [31]. Several pyrochlore-type oxides such as $K_2Ta_2O_6$ [32], $Na_2Ta_2O_6$ [33], $Pb_2Sn_2O_6$ [34], $KAl_{0.33}W_{1.67}O_6$ [35], Ag/Sn-doped $KSbTeO_6$ [36], Ag/Sn-doped $KSbTeO_6$ [37], $Na_2Ta_2O_6$ [38] and $ASbO_3$ [39] were used for the decolourization of dyes such as acid red G, Congo red, methyl orange, methylene blue and rhodamine B. The luminescence properties of $Ln_2Ce_2O_7$ fluorite-type is also studied [40]. There is a deficiency in the literature regarding the preparation of fluorite-type structure for various applications. Besides, the photocatalytic activity of $Ga_2Zr_{2-x}W_xO_7$ fluorite-type system for CV degradation has not been reported yet. In this frame, the present work aims to prepare and characterize nano-sized $Ga_2Zr_{2-x}W_xO_7$ fluorite-type system using a reliable, cost-effective, eco-friendly and easy method to optimize the shape and grain size of the nano-sized metal oxides (the Pechini method [41]). Tungsten as a dopant is selected due to the difference in the oxidation state and ionic radii of W^{6+} and Zr^{4+} that will permit studying the impact of doping on both the structural and photocatalytic activity of $Ga_2Zr_2O_7$. Furthermore, W is used for the reduction of the band gap in order to use $Ga_2Zr_{2-x}W_xO_7$ system in visible light irradiation. Finally, the photocatalytic activity has been studied for the prepared systems in CV dye removal in visible light and the reaction operation conditions were adopted (reaction time, pH, catalyst dose and initial pollutant concentrations).

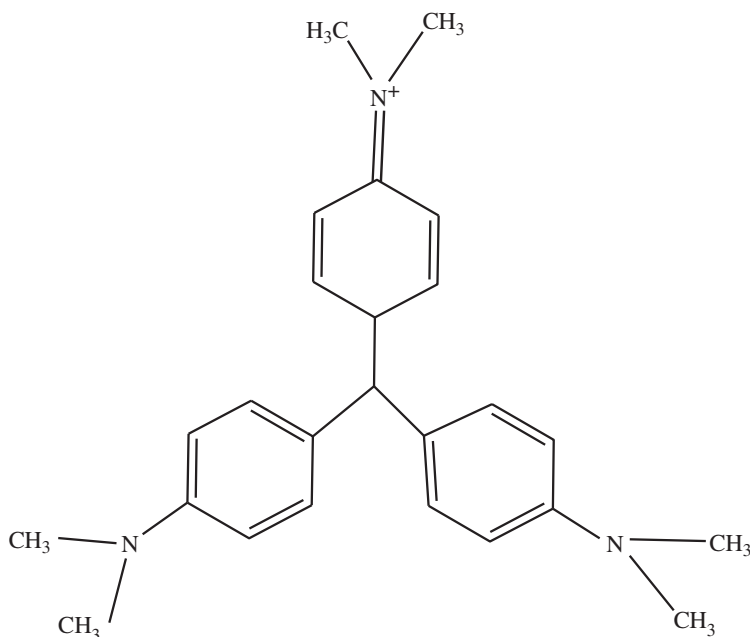


Figure 1. Chemical structure of crystal violet.

2. Experimental

2.1. Preparation and characterization of the prepared materials

$\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system was prepared; where $x=0, 0.05, 0.1, 0.15$ and 0.2 using the citrate technique (Pechini method) which is a wet-chemical method based on polymeric precursor [41] that was used to prepare several metal oxides [42–46].

In this method, α -hydroxy acid (citric acid) is used to chelate the cations forming a polybasic acid. Polyhydroxy alcohol (ethylene glycol) reacts with these chelates forming ester and water. Heating the mixture leads to polyesterification and after the evolution of nitrous oxide and water, the gel is obtained. The thermal decomposition of this gel results in a chemically homogeneous powder containing the desired stoichiometry [42,43].

Zirconium (IV) oxynitrate hydrate (Sigma-Aldrich), tungsten (VI) chloride (Sigma-Aldrich), gallium (III) nitrate (Silverton, San Diego), ethylene glycol (Sandycroft, Deeside, Clwyd) and citric acid anhydrous extra pure (LobaChemie) are used as starting materials. All chemicals were reagent grade and used as received without any modification.

$\text{Ga}_2\text{Zr}_2\text{O}_7$ was prepared using the Pechini method as follows: aqueous zirconium oxynitrate and gallium nitrate solutions were mixed, considering the desired stoichiometry of the metal oxides in the final ceramic powder solution (A). The citric acid (CA) was then added to the solution (A) to chelate metal cations at the CA:Me molar ratio of 4:1. Me denotes Ga^{3+} , Zr^{4+} in the final ceramic powder. After dissolving the CA, ethylene glycol (EG) was added into the solution at a CA:EG molar ratio of 1:1.5. The solution was then heated at 140°C and kept under stirring to promote the esterification and polymerization reactions. After elimination of nitrous oxides and water, a gel was obtained. The gel was charred gradually up to 300°C then heated in the muffle furnace at 300°C for 2 h. The charred gel thus produced was ground and calcined for 2 h at 500°C , then ground and calcined for 2 h at 600°C . $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ systems where $x=0.05, 0.1, 0.15$ and 0.2 were prepared using the same sequence. For the preparation of $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ samples, tungsten chloride was dissolved in ethanol and then added to the solution (A). Flowchart of the preparation of $\text{Ga}_2\text{Zr}_2\text{O}_7$ powder is presented in the electronic supplementary material, S.1. The samples' identification, as well as their composition, are presented in table 1.

X-ray diffraction (XRD) is the standard technique for determination of the crystal structure of a solid. XRD is used to identify the crystal structure, to determine the lattice parameters. The XRD measurements were carried out using 7000 Shimadzo (Japan) 2 kW model X-ray spectrophotometer with a nickel-filtered Cu radiation ($\text{CuK}\alpha$) with $\lambda = 1.54056 \text{ \AA}$. The scanning 2θ range was 5–80 with a step size of

Table 1. The composition determined by ICP compared to the expected composition of the prepared samples.

sample	sample compositions	expected (wt%)			experimental (wt%)		
		Ga	Zr	W	Ga	Zr	W
ZG	Ga ₂ Zr ₂ O ₇	32.14	42.05	0	32.26	42.13	0
ZGW1	Ga ₂ Zr _{1.95} W _{0.05} O ₇	31.80	40.57	2.10	32.91	40.43	2.13
ZGW2	Ga ₂ Zr _{1.9} W _{0.1} O ₇	31.48	39.11	4.15	31.39	39.18	4.18
ZGW3	Ga ₂ Zr _{1.85} W _{0.15} O ₇	31.14	37.69	6.16	31.19	37.63	6.20
ZGW4	Ga ₂ Zr _{1.80} W _{0.2} O ₇	30.82	36.30	8.13	30.77	36.37	8.15

0.2. The lattice parameters were determined using a program called UnitCellWin [47]. FTIR spectra were recorded in the frequency range 400–4000 cm⁻¹ with a resolution of 4 cm⁻¹ using FTIR 6100 Jasco (Japan) spectrum equipment.

Diffuse reflectance measurements were performed to study the optical properties of the prepared samples using Shimadzu UV-3600 (Japan). The free radicals created (EPR signals) were recorded at room temperature by X-band EMX spectrometer (Bruker, Germany) using a standard rectangular cavity of ER 4102 operating at 100 kHz field modulation. The microstructures were studied by transmission electron microscope (TEM, JEOL JEM2100, Japan). The specific surface area of the prepared samples was determined by NOVA surface area analyser from Thermo Pascal 140 mercury porosimetry under a pressure range of 0.1–200 MPa. Mercury surface tension of 480 dyne cm⁻¹ and the contact angle of 141.3° were used. Elemental analysis was carried out using inductive coupled plasma-atomic spectrometry (Agilent ICP-OES).

2.2. Photodegradation activity

CV was obtained from Sigma-Aldrich Chemical Company. All solutions were prepared in double-distilled water. Photocatalytic experiments were carried out with CV dye solution using all the prepared catalysts under visible irradiation. Irradiation was carried out by commercial visible metal halide lamp (HQI-T250/Daylight, OSRAM GmbH, Germany) with a luminous efficacy of 82 lm W⁻¹ and luminous flux of irradiation 20 000 lm.

A stirred slurry composed of dye solution and catalyst was placed in the dark for 30 min in order to establish equilibrium between adsorption and desorption phenomenon of dye molecule on the photocatalyst surface. Then the lamp was turned on and the slurry was magnetically stirred for homogeneous distribution of catalyst in the solution. At specific time intervals, an aliquot (5 ml) was collected and centrifuged for 2 min at 3500 r.p.m. to remove catalyst particles from aliquot to assess the extent of decolourization. The absorption spectra recorded 588 nm as λ_{\max} on the double-beam UV-visible spectrophotometer (Cary-100). The desired pH of the solution was adjusted by the addition of previously standardized 0.050 M H₂SO₄ and 1.0 M NaOH solutions. Performance efficiency was calculated as

$$\% \text{efficiency} = \frac{C_0 - C}{C_0} \times 100, \quad (2.1)$$

where C and C_0 are initial and final dye concentration, respectively, for reaction time t .

2.3. Evaluation of active species

To check the influence of some active species on the catalytic activity of catalyst trapping experiments were carried out for difference species. In these experiments, 1 mmol l⁻¹ of three scavengers were used which are isopropyl alcohol (IPA), ethylene diamine tetra acetic acid (EDTA) and benzoquinone for HO[•], h⁺ and O^{2•} species, respectively.

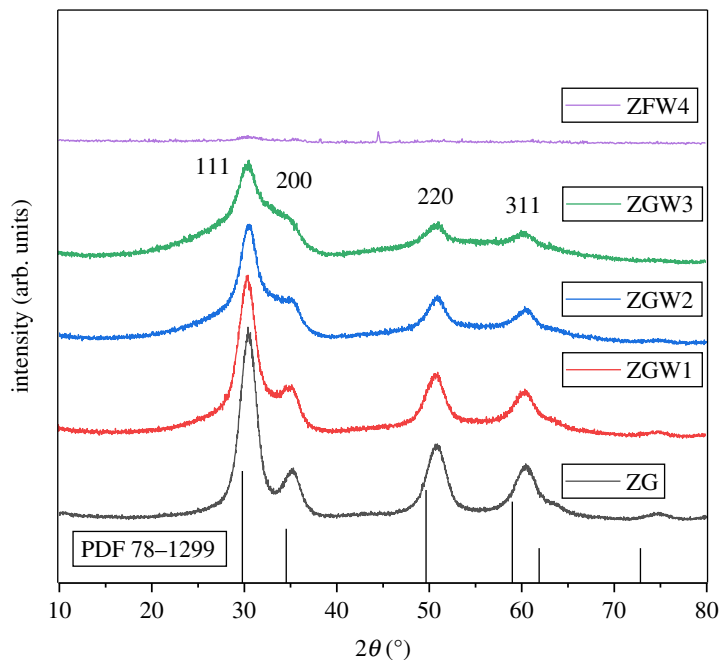


Figure 2. Powder X-ray diffraction pattern of $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system calcined at $600^\circ\text{C}/2\text{ h}$ ($x=0, 0.05, 0.1, 0.15$ and 0.2).

3. Results and discussion

3.1. Characterization of the prepared materials

The crystallization of pyrochlore or fluorite phases for the mixed oxides $\text{A}_2\text{B}_2\text{O}_7$ depends on the radius ratio of A and B cations (r_A/r_B) in addition to the conditions of samples processing [48–50]. $\text{A}_2\text{B}_2\text{O}_7$ crystallizes in the stable pyrochlore structure when the r_A/r_B is in the range of 1.46–1.78 [51] depending on the coordination number. The defect fluorite structure (cubic, $\text{Fm}\bar{3}\text{m}$) is obtained for the lower or upper limits of the previously mentioned range of r_A/r_B . In the present study, the radius ratio r_A/r_B for $\text{A}=\text{Ga}^{3+}$ (ionic radius = 62 pm) and $\text{B}=\text{Zr}^{4+}$ (ionic radius = 72 pm [52]) ions was found to be 0.85 which is lower than the above-mentioned range. In this frame, it is predicted that $\text{Ga}_2\text{Zr}_2\text{O}_7$ will be crystallized in the fluorite structure. This is confirmed by the XRD pattern of $\text{Ga}_2\text{Zr}_2\text{O}_7$ (ZG sample) calcined at 600°C for 2 h (figure 2) where ZG sample has the cubic fluorite phase structure (PDF 78–1299 for $\text{Er}_{0.5}\text{Zr}_{0.5}\text{O}_{1.75}$, which is the best-matched card that can be used where there is no card for the novel $\text{Ga}_2\text{Zr}_2\text{O}_7$ material). The shift to lower 2θ value is due to the difference in the ionic radius between Er^{3+} (ionic radius = 89 pm [52]) and Ga^{3+} (ionic radius = 62 pm [52]) ions. The peaks at about 14° , 28° , 37° , 45° 2θ corresponding to (1 1 1), (3 1 1), (3 3 1), (5 1 1) planes [53], respectively, characteristic for the pyrochlore structure, do not exist. Accordingly, ZGW1, ZGW2 and ZGW3 samples have the cubic fluorite phase structure (figure 2). Traces of cubic fluorite phase structure were detected for ZGW4 sample, which means that this sample may need further calcination in order to improve its crystallinity, which is not in alliance with the calcination temperature of all the prepared samples in the manuscript.

According to the ionic radius of Zr^{4+} (ionic radius = 72 pm [52]) and W^{5+} (ionic radius = 62 pm [52]) or W^{6+} (ionic radius = 60 pm [52]) ions, it is predicted that the cubic lattice parameter and unit cell volume will decrease as the W concentration increases because ionic radius of $\text{W}^{5,6+}$ ion is less than that of Zr^{4+} and also Ga^{3+} ions. Surprisingly, it was found that as the W concentration raised, the cubic lattice parameter and unit cell volume are increased (table 2). This might be attributed to the substitution of Zr^{4+} ions by $\text{W}^{5,6+}$ ions creating a distorted coordination environment and leading to unit cell expansion [54].

ICP was used to determine the chemical composition of the prepared materials by determining the wt% of Zr, Ga and W in all the prepared samples. The experimental wt% of Zr, Ga and W in all the prepared samples were in alliance with those of the expected wt% (table 1) indicating that the prepared materials have the exact proposed chemical compositions shown for ZG, ZGW1, ZGW2, ZGW3 and ZGW4.

The FTIR spectra of the prepared samples are shown in figure 3. For the parent ZG sample, the broad absorption peak at about 3422 cm^{-1} is due to the stretching vibration of OH group in water molecule. The absorption band at about 1634 cm^{-1} is characteristic to the bending vibration of the water molecules [55]. The band at about 460 cm^{-1} is due to Zr–O vibration [56]. The peaks at 1040 and 1080 cm^{-1} might be

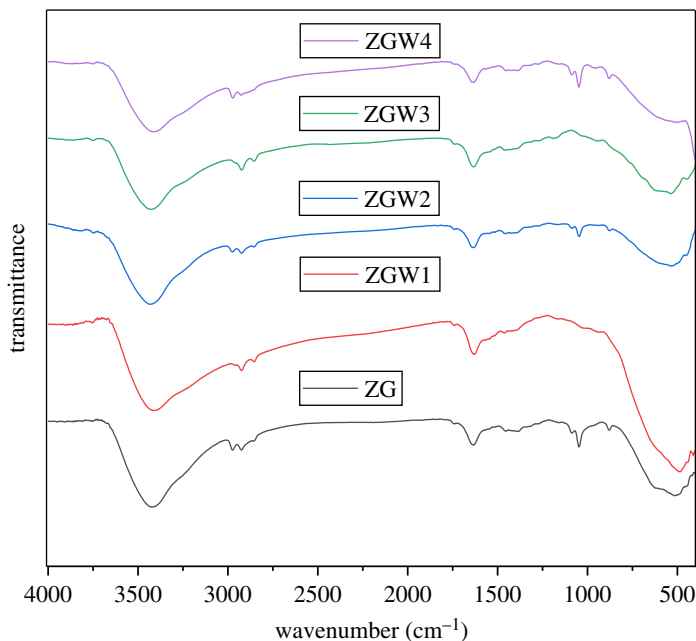


Figure 3. FTIR spectra of $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system where $x=0, 0.05, 0.1, 0.15$ and 0.2 .

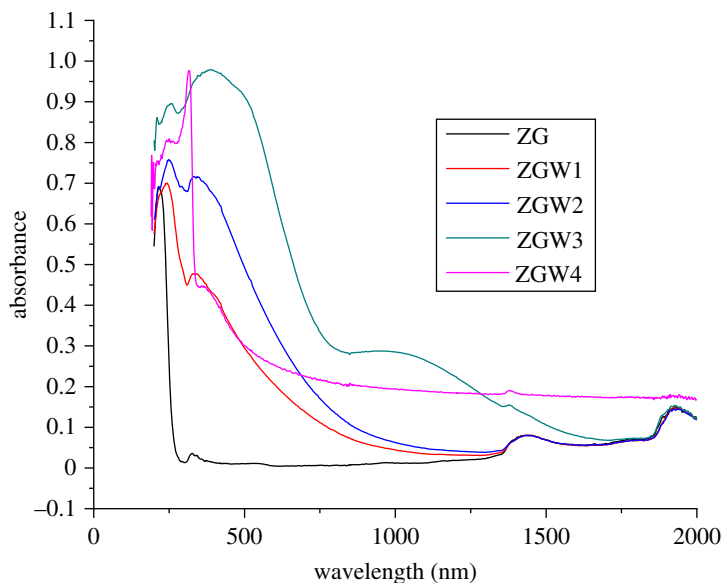


Figure 4. The diffuse reflectance spectra of ZG, ZGW1, ZGW2, ZGW3 and ZGW4 samples.

associated with stretching vibrations of Zr–O terminals [57]. The peak at about 880 cm^{-1} corresponds to the bending vibration of hydroxyl groups bounds to zirconium oxide [55]. The bands at about 510 and 620 cm^{-1} are assigned to Ga–O stretching and Ga–O–Ga torsion movements [58]. For the doped samples (ZGW1, ZGW2, ZGW3 and ZGW4), a little shift in the peak positions and peak intensities were observed with the increasing in the W concentration. The region $600\text{--}900\text{ cm}^{-1}$ might correspond to O–W–O stretching modes [59]. The peak at about 620 cm^{-1} is assigned to W–O_{inter}–W bridging vibration of the corner-sharing WO₆ octahedron [60].

Diffuse reflectance spectroscopy (DRS) was used to determine the system optical properties. The absorption edge of ZG sample is about 247 nm . As shown in figure 4, W doping (till 15 mol%, ZGW3 sample) shifted the absorption to the visible light range (red shift). For ZGW4 sample ($\text{Ga}_2\text{Zr}_{0.18}\text{W}_{0.2}\text{O}_7$) the tungsten doping shifted the absorption to lower wavelength. The band gap is a very important parameter for the photocatalytic performance of the photocatalyst where it indicates the range at which the photocatalyst will be active (UV–visible light ranges).

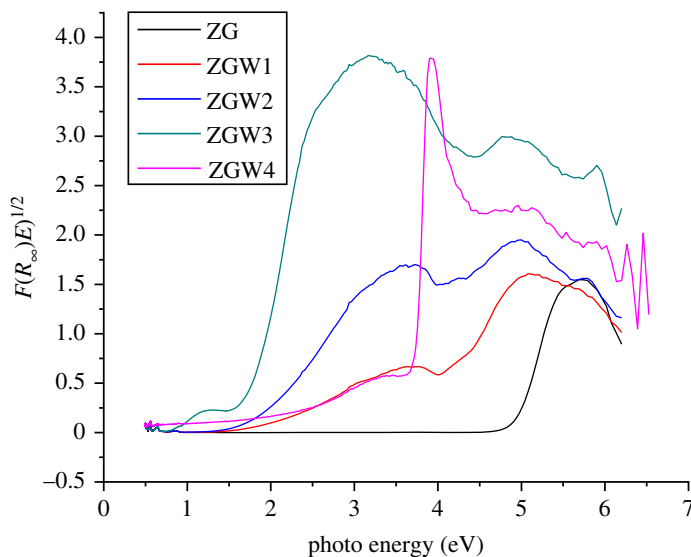


Figure 5. The plots of $F(R_{\infty})E^{1/2}$ versus photo energy for the estimation of the band gap energy for ZG, ZGW1, ZGW2, ZGW3 and ZGW4 samples.

Table 2. Microstructural parameters, the estimated band gap from DRS, the TEM particle size range and surface area for $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system.

sample	a (Å)	V (Å ³)	TEM particle size range (nm)	band gap (eV)	surface area $\text{m}^2 \text{g}^{-1}$
ZG	5.08693	131.6337	4–7	4.95	91.3
ZGW1	5.08706	131.6437	4–8	3.88	96.5
ZGW2	5.08824	131.7358	3–5	1.81	99.9
ZGW3	5.09405	132.1876	3–4	1.7	124.3
ZGW4	—	—	—	2.66	90

The Kubelka–Munk function $F(R)$ is correlated to the diffuse reflectance R according to the following equation:

$$F(R) = \frac{(1 - R^2)}{2R}. \quad (3.1)$$

Figure 5 shows the plots of $F(R_{\infty})E^{1/2}$ versus photo energy for the estimation of the band gap energy for ZG, ZGW1, ZGW2, ZGW3 and ZGW4 samples. The band gap value was determined by drawing $(F(R).h\nu)^{1/2}$ against photo energy and extrapolating the linear part of the curve to $(F(R).h\nu)^{1/2} = 0$ according to Kubelka–Munk using linear fit [61] (electronic supplementary material, S.2).

Table 2 shows that the band gap of ZG is 4.98 eV. The band gap is decreased from 3.88 to 1.7 by increasing the W concentration from 5 to 15 mol% while at 20 mol% W, the band gap increased again to be 2.66 eV.

For the understudied novel $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system, the band gap of $\text{Ga}_2\text{Zr}_2\text{O}_7$ (4.95 eV) is closer to that of ZrO_2 (5 eV) [62] which renders its absorption of visible light accordingly, it is predicted to have a limited photocatalytic activity under visible light. The band gaps of ZGW2 and ZGW3 samples are smaller than that of ZrO_2 [63], $\text{Sm}_2\text{Zr}_2\text{O}_7$ (2.86 eV) [63] and $\text{Nd}_2\text{Zr}_2\text{O}_7$ (2.67 eV) [63]. The band gap of ZGW4 sample is similar to that of $\text{Nd}_2\text{Zr}_2\text{O}_7$.

For $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system in this study, the conduction band (CB) is composed of Zr 4d orbitals whereas the valence band (VB) consists of the O 2p bands.

Electron paramagnetic resonance (EPR) is one of the most important tools used to describe the defects in solids because most of the defects contain unpaired electrons. Intrinsic or extrinsic point defects exist in solids. The intrinsic defects are present in the solid itself without introducing any impurity; when

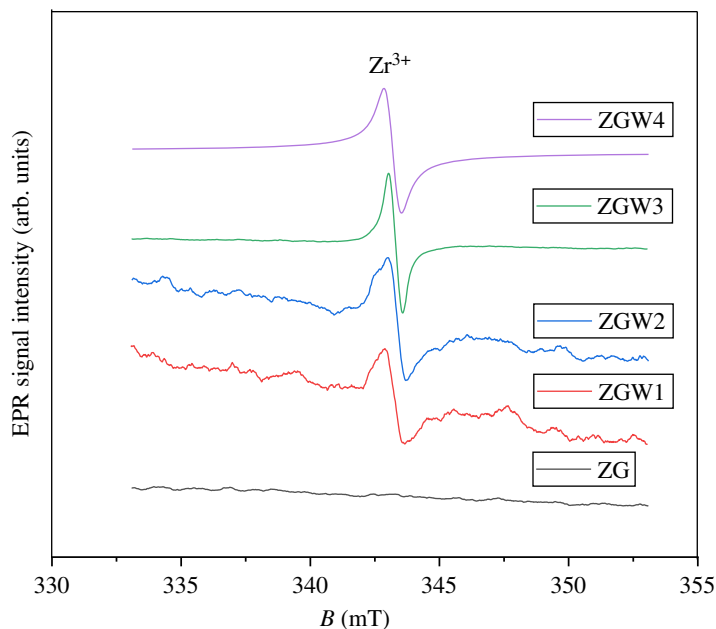


Figure 6. EPR spectra for ZG, ZGW1, ZGW2, ZGW3 and ZGW4 samples.

Table 3. The spin numbers for the prepared samples.

sample	spin numbers
ZG	0
ZGW1	4.83×10^{16}
ZGW2	4.33×10^{17}
ZGW3	1.89×10^{18}
ZGW4	1.25×10^{17}

chemical impurity is introduced, the extrinsic defects are produced. Recently, the impurities (dopants) have been used in photocatalysis to modulate the optical properties to improve the photocatalytic performance of the catalyst [64]. EPR spectra for the prepared samples are presented in figure 6, where no EPR spectra are detected for the undoped $\text{Ga}_2\text{Zr}_2\text{O}_7$ (ZG) sample. For W-doped samples (ZGW1, ZGW2, ZGW3 and ZGW4 samples), an EPR signal is detected. For W-doped samples, Zr^{4+} ion is substituted by $\text{W}^{5+,6+}$ ion which results in the creation of free electrons for electroneutrality, these electrons are trapped to Zr^{4+} ion forming Zr^{3+} ion which corresponds to the detected EPR signal [65]. No signals were detected for the oxygen vacancy. Table 3 shows the spin number (free radicals) for the prepared samples. As the W concentration increases, the spin number increases up to 15 mol% W (ZGW3 sample) and for 20 mol% sample (ZGW4 sample) the spin number decreases which is in accordance with DRS results. DRS results demonstrated that, Zr^{3+} point defect might introduce a new energy level between the CB and the VB resulting in decreasing the band gap with increasing the W concentration up to 15 mol% W (ZGW3 sample). Increasing the band gap from 1.7 eV for ZFW3 sample to 2.66 eV for heavily doped ZGW4 sample might be due to the donor electrons of W filling the lowest level of the CB (the Burstein–Moss effect) [66]. The schematic energy level diagram for the prepared $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system and its corresponding charge separation towards dye degradation under visible light illumination is presented in electronic supplementary material, S.3.

The TEM micrographs of samples are presented in the electronic supplementary material, S.4, S.5. Small quasi-spherical particles, which agglomerate into denser aggregates were observed. The TEM particle size range for all samples is presented in table 2. The lattice fringes with an interplanar distance of 0.25 nm which could be assigned to (2 0 0) plane were detected for the cubic phase of $\text{Ga}_2\text{Zr}_2\text{O}_7$ for ZGW3 sample (electronic supplementary material, S.6).

To manifest the surface area of the prepared oxides, the Barrett–Joyner–Halenda (BJH) nitrogen adsorption tests were used (figure 7). A type-IV isotherm demonstrates the typical mesopore materials

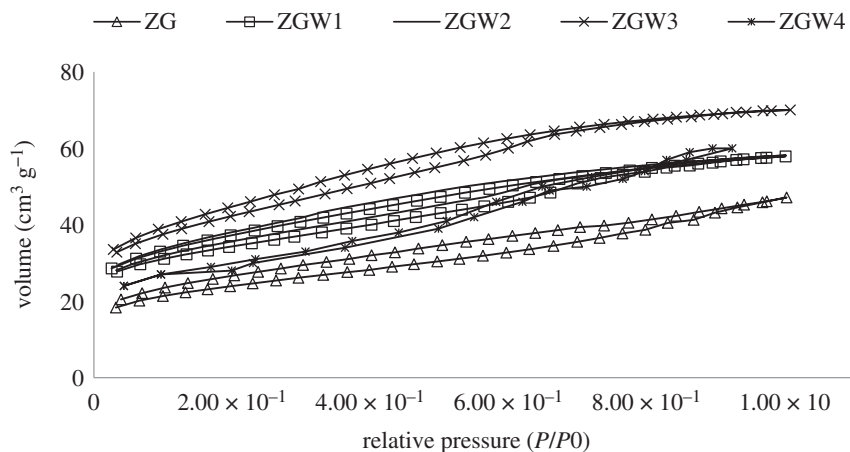


Figure 7. N_2 adsorption and desorption isotherms for prepared materials sample.

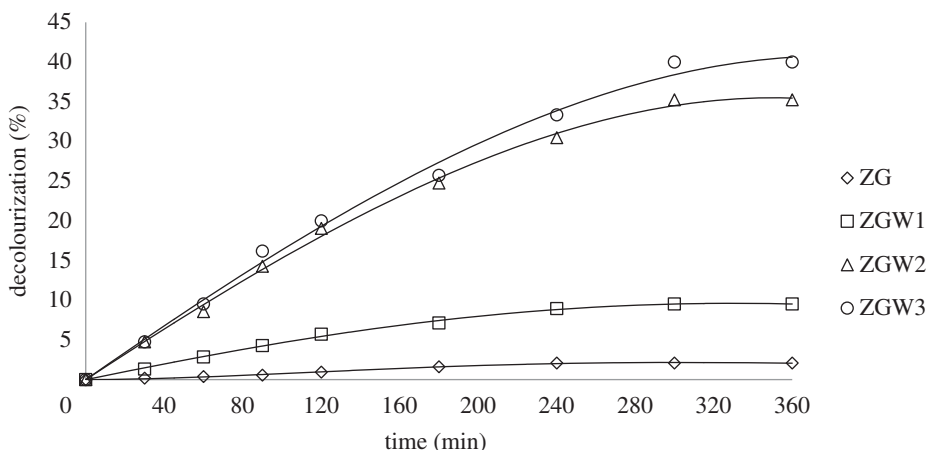


Figure 8. Photocatalytic degradation of 10 mg l^{-1} CV under visible light by 0.75 g l^{-1} of all the prepared materials as a function of time at pH 7.

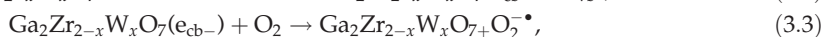
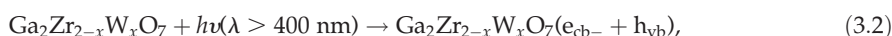
that are related to aggregates presented in mesopores, and the little-marked uptake over a range of high P/P_0 . [61,67]. Specific surface areas were significantly increased with the increment of tungsten concentration in the order of 91.3 , 96.51 , 99.9 and $124.3 \text{ m}^2 \text{ g}^{-1}$ for ZG, ZGW1, ZGW2 and ZGW3 samples, respectively, and decreased to be $90 \text{ m}^2 \text{ g}^{-1}$ for ZGW4 sample.

3.2. Photocatalytic activity of $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system

3.2.1. Degradation time influence

Figure 8 demonstrates the influence of reaction time on photodegradation of CV by using $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system under visible irradiation for 6 h. The degradation rate was increased with the increment of time till 5 h then it became stable after that, concluding that 5 h is the optimum reaction time. The decolourization % of CV dye after 5 h was recorded 2.1%, 9.5%, 35.23% and 40% for ZG, ZGW1, ZGW2 and ZGW3, respectively, at initial CV concentration of 10 mg l^{-1} at pH 7 and 0.75 g l^{-1} catalyst dose. Concluding that the order of photocatalytic activity was in accordance with the reduction in their band gaps (table 2).

Moreover, O_2 and HO groups on the surface were converted to $\text{O}_2^{\bullet-}$ and HO^{\bullet} , respectively. Both of them can assist in CV degradation in the following equations:



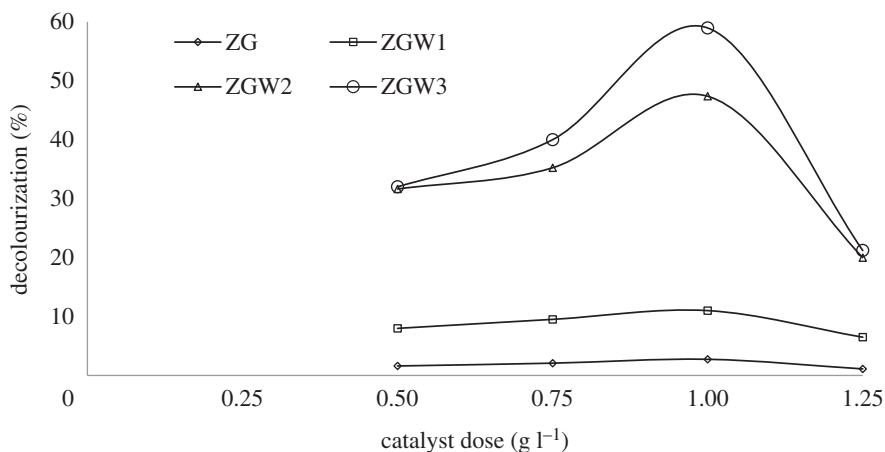


Figure 9. Photocatalytic decolourization under visible light for all the prepared materials as a function of catalysts dose (CV concentration 10 mg l^{-1} at pH 7).

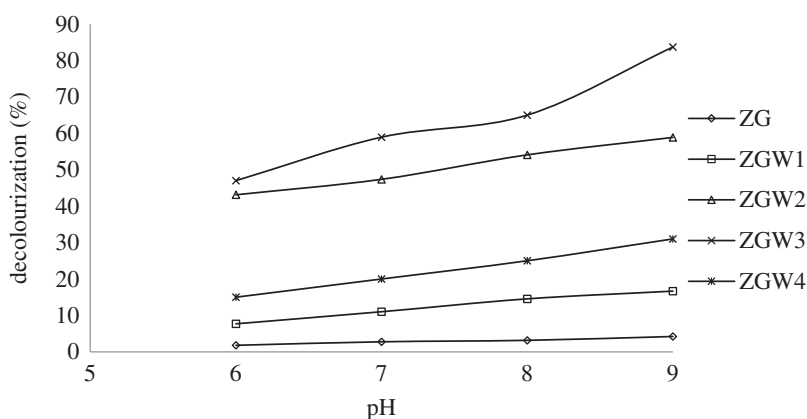


Figure 10. Photocatalytic degradation under visible light for all the prepared materials as a function of pH (CV concentration 10 mg l^{-1} at 1 g l^{-1} catalyst dose).

3.2.2. Catalyst load impact

The degradation of 10 mg l^{-1} CV at optimum reaction time (5 h) and neutral pH were studied at different catalysts loads ($0.5\text{--}1.25 \text{ g l}^{-1}$).

Figure 9 displays the direct relation between the % decolourization and the catalyst's loads for both undoped and W-doped GZ catalysts up to 1 g l^{-1} due to intense numbers of catalytically active sites by increasing of catalysts load [68], that raises the rate of $\text{O}_2^{\bullet -}$ and HO^{\bullet} creation [19]. Reduction of decolourization rate after 1 g l^{-1} was noted and was owed to the excess catalyst amount impedes the light penetration [20,68,69]. Accordingly, 1 g l^{-1} was elected as the optimum dose for CV degradation.

3.2.3. Influence of pH

Different pH values (6–9) under pre-optimized time and catalyst load were employed to decolourize 10 mg l^{-1} CV. The % decolourization was directly proportional to pH value up to pH 9 with % decolourization of 4.2%, 16.6%, 58.88%, 83.7% and 31.2% for GZ, GZW1, GZW2, GZW3 and ZGW4, respectively (figure 10). In acidic medium, very low photodegradations of CV were spotted due to its hard deposition on the catalyst surface [70–73]. On the other hand, in alkaline medium, the amount of hydroxide ions increased and their availability to be converted to HO^{\bullet} increased leading to an acceleration of the degradation rate [72].

Additionally, the impact of pH 10–pH 12 on dye decolourization were examined but not taken into consideration because at $\text{pH} > 9$, a colourless CV molecule occurred without illumination [73]. Consequently, pH 9 was elected to be the optimum pH.

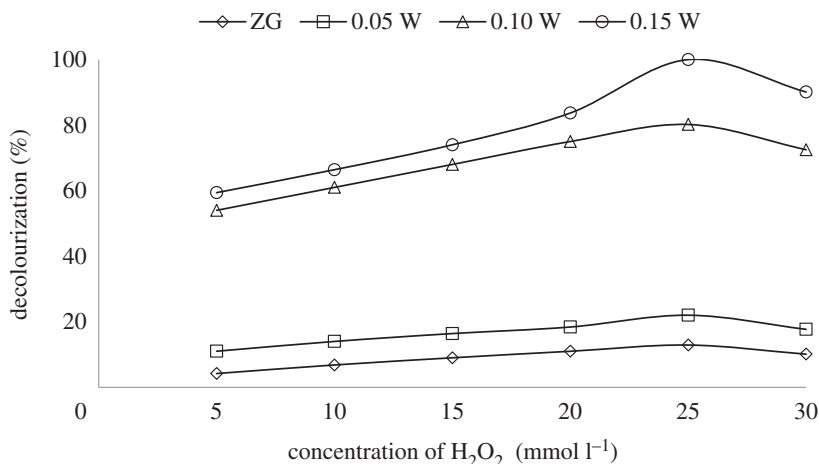


Figure 11. The effect of H₂O₂ doses on % decolourization of CV for ZGW3 (CV concentration 10 mg l⁻¹ at 1 g l⁻¹ catalyst dose and pH 9).

3.2.4. Effect of doping on the photocatalytic performance

The effect of W doping was explained in several parts in the manuscript including XRD, DRS, EPR as well as the photocatalytic efficiency. The positive role of doping for certain concentration might be attributed to (i) formation of new energy levels between the VB and the CB; these levels act as effective charge carrier traps, and (ii) improving the adsorption of the molecules of the pollutant on the surface of the catalyst by altering the catalyst surface acid–base properties [74]. For the prepared samples, the photocatalytic degradation of CV increases as W concentration increases, as shown in figure 10, which is in accordance with their band gap values in the DRS part. ZGW3 sample has high CV removal compared with the undoped ZG sample. This might be attributed to the substitution of Zr⁴⁺ by W^{5+,6+} in the Ga₂Zr₂O₇ lattice, which is reflected in the cubic lattice parameter and unit cell volume values for ZG and ZGW3 sample and introduction of new Zr³⁺ level between the VB and the CB of Ga₂Zr₂O₇ (electronic supplementary material, S.3), decreasing the band gap as well as increasing the spin numbers detected by EPR. After visible light illumination, the electron is promoted from the VB to the CB through the Zr³⁺ level and the hole is formed. The electron and hole react with the adsorbed oxygen and hydrogen peroxide forming O₂^{•-} and HO[•] which are responsible for CV photodegradation [75]. For the highly W-doped sample (ZGW4), the decrease of the photocatalytic degradation of CV as compared with the other W-doped samples might be due to filling of the lowest level of CB by W donor electrons [66] which is matched with the band gap values calculated by DRS.

3.2.5. Effect of H₂O₂

Hydrogen peroxide is counted as one of the most fundamental photo-oxidants in water decontamination, hydrogen peroxide is used in the visible radiation to assess HO[•] radicals generation, which is the essential promoter for the destruction of toxic organic compounds as represented in the following equation [76]:



H₂O₂ doses of 0–30 mmol l⁻¹ were used to study the effect of H₂O₂ on % decolourization of 10 ppm CV at pH 9 by 1 g l⁻¹ catalyst as presented in figure 11. Increasing H₂O₂ dose from 0 to 25 mmol l⁻¹ is directly proportional to the % CV degradation attaining complete degradation for ZGW3 after 5 h due to the increase in the amount of HO[•] promoting the degradation rate. Further addition of H₂O₂ to 30 mmol l⁻¹ decreased the % decolourization of CV due to hydroxyl radical and hole scavenging effects (equation (3.9)) [77]. Hence the optimum dose of H₂O₂ was elected as 25 mmol l⁻¹ [78,79].



3.2.6. Influence of initial dye concentration

The influence of different initial CV concentrations was demonstrated under prementioned optimum conditions with ZGW3. As affirmed in figure 12 which was set up on the linear relationship, the

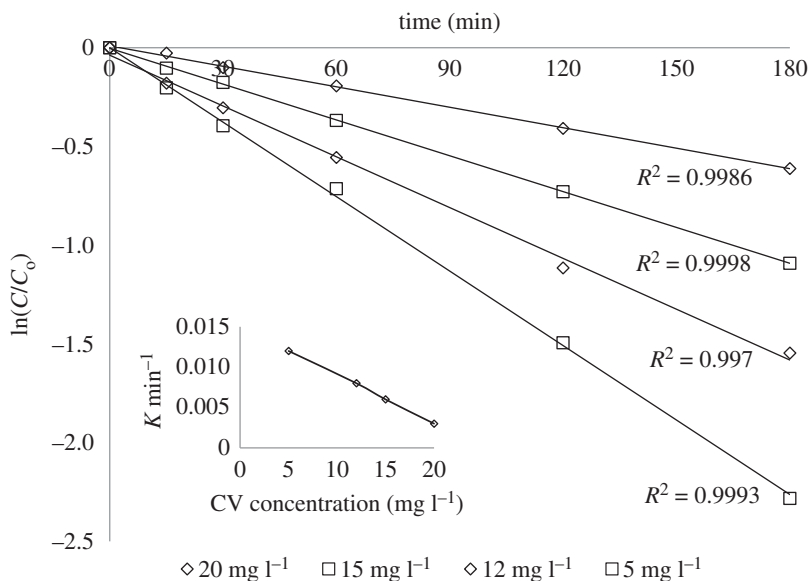


Figure 12. Pseudo-first-order kinetics for different CV doses for ZGW3 (CV concentration 10 mg l⁻¹ at 1 g l⁻¹ catalyst dose and pH 9). Inset: the effect of different CV doses on the rate of the degradation reaction for ZGW3.

decolourization of CV arraigned by the pseudo-order kinetics in the following equation:

$$\ln \frac{C}{C_0} = K_{app} t \quad (3.10)$$

where K_{app} is the spotted rate constant, C_0 and C are the concentrations of CV at zero time and at a certain time, respectively.

Using ZGW3, the photocatalytic rate has increased with decreasing dye concentrations in the tested solutions which attained 0.012, 0.008, 0.006 and 0.003 min⁻¹ at CV concentration 5, 12, 15 and 20 mg l⁻¹, respectively (figure 12). For the adequately low initial CV concentration (5 mg l⁻¹), the dye was completely degraded after the photocatalytic time of 3 h. Additionally, increment in initial CV concentrations, photocatalytic degradation time increased to attain the nearly complete decolourization of CV (figure 12). This might be credited to the augmentation of optical densities of the CV dye solutions with the increment of dye concentrations, which may act as a filter to the incident light [72] and consecutive possible restrict of irradiation penetration to the catalysts' surfaces in all of the test solutions. In this manner only, fewer photons can arrive at the catalyst surface, and therefore the creation of HO[•] free radical on the surface of the catalyst declined, since the available effective sites of catalyst become covered by the crowded dye ions. This results in decolourization rate reduction [68,72,73,76,77,80,81].

3.2.7. The degradation pathway of crystal violet

UV-visible spectra of CV dye solution as a function of reaction time for ZGW3 are depicted in figure 13. As noted from these spectra, at 0 time of the experiment prior to the oxidation reaction, the absorption spectrum of CV in water was distinguished by one main peak in the visible region ($\lambda = 584$ nm) and by two other peaks in the UV region ($\lambda = 250$ and 300 nm). The peaks at 250 and 300 nm were related to aromatic structures in the molecule, and that at 584 nm originated from the chromophore [82]. The gradual decay of the visible peaks with time was owing to the cleavage of the aromatic rings by oxidation. In addition to this rapid decolourization effect, the decrease of the absorbance at 250 or 300 nm was considered as an index of aromatic fragment degradation of the dye molecule and its intermediates [83,84].

GC-MS study was conducted to further recognition of the intermediate products formed at the end of the photocatalytic reaction figure 14. Based on the results and previous studies [82-85], figure 15 proposed initial degradation pathways that start with *N*-de-methylation followed by an attack of the oxidizing species on the central carbon portion of the CV to form 4-(*N,N*-dimethylamino)-4'-(*N,N'*-dimethylamino) benzophenone [82,85]. Then the central carbon was successively attacked by the active radicals [85]. Finally, the gradual cleavage of the aromatic intermediates would lead up to the

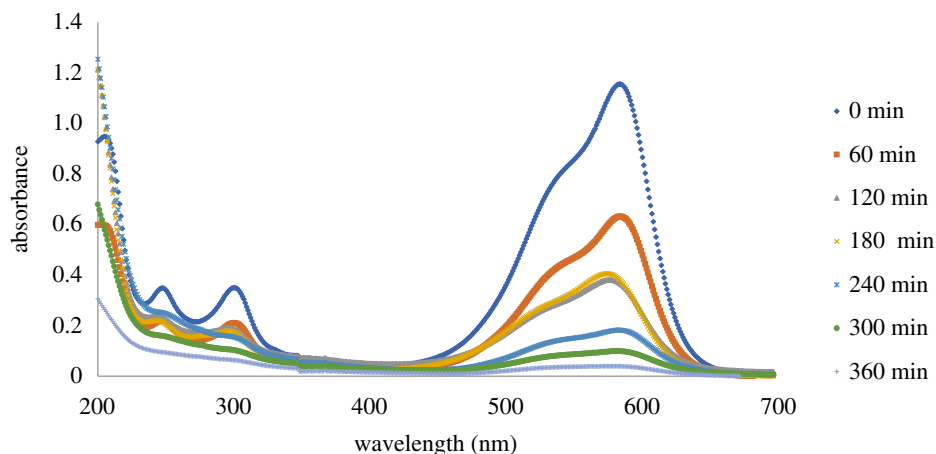


Figure 13. UV–visible spectra study of CV dye as a function of reaction time for ZGW3 (CV concentration 10 mg l^{-1} at 1 g l^{-1} catalyst dose and pH 9).

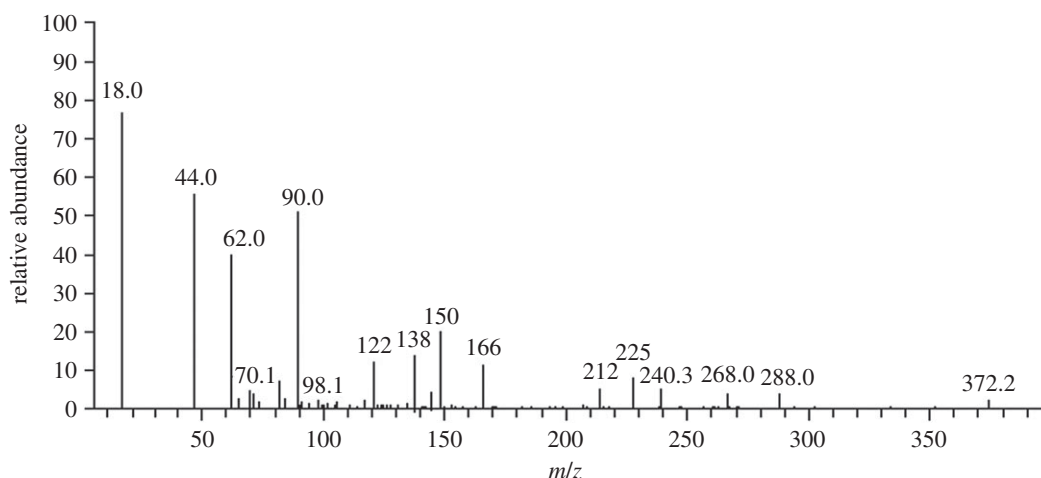


Figure 14. The GC–MS mass spectra of photocatalytic degradation for CV using ZGW3 under optimum operating condition (CV concentration 10 mg l^{-1} at 1 g l^{-1} catalyst dose and pH 9).

formation of carboxylic acids before transformation into carbon dioxide and water. The treated water was safe to be used for water remediation since it was non-toxic for *Vibrio fischeri* organism according to the test performed by Microtox analyser 500 [86].

The Photocatalytic efficiency of the prepared $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ for CV dye degradation in comparison with that of various other photocatalysts is presented in table 4. The prepared nano-sized cubic fluorite $\text{Ga}_2\text{Zr}_{1.85}\text{W}_{0.15}\text{O}_7$ oxide showed promising photocatalytic activity for decolourization of the harmful CV dye under visible light irradiation (which is more applicable from an economic view) compared with modified TiO_2 .

3.2.8. The reusability for ZGW3 sample

The reusability of the catalyst is one of the main obstacles to the application of photocatalyst in water treatment. In order to examine the reusability, 10 cycles for CV decolonization over ZGW3 sample were accomplished under the pre-optimized operating conditions. The catalyst was deposited settling the solution for enough time. After detaching the supernatant, the catalyst had been introduced for another cycle. The variation in % CV removals with various cycles is presented in figure 16. The trivial diminishment in photocatalytic adequacy (100–94%) pointed to satisfactory results obtained with increasing the number of runs up to 10 affirming that the prementioned sample can be reused without losing the profitable synergy activity.

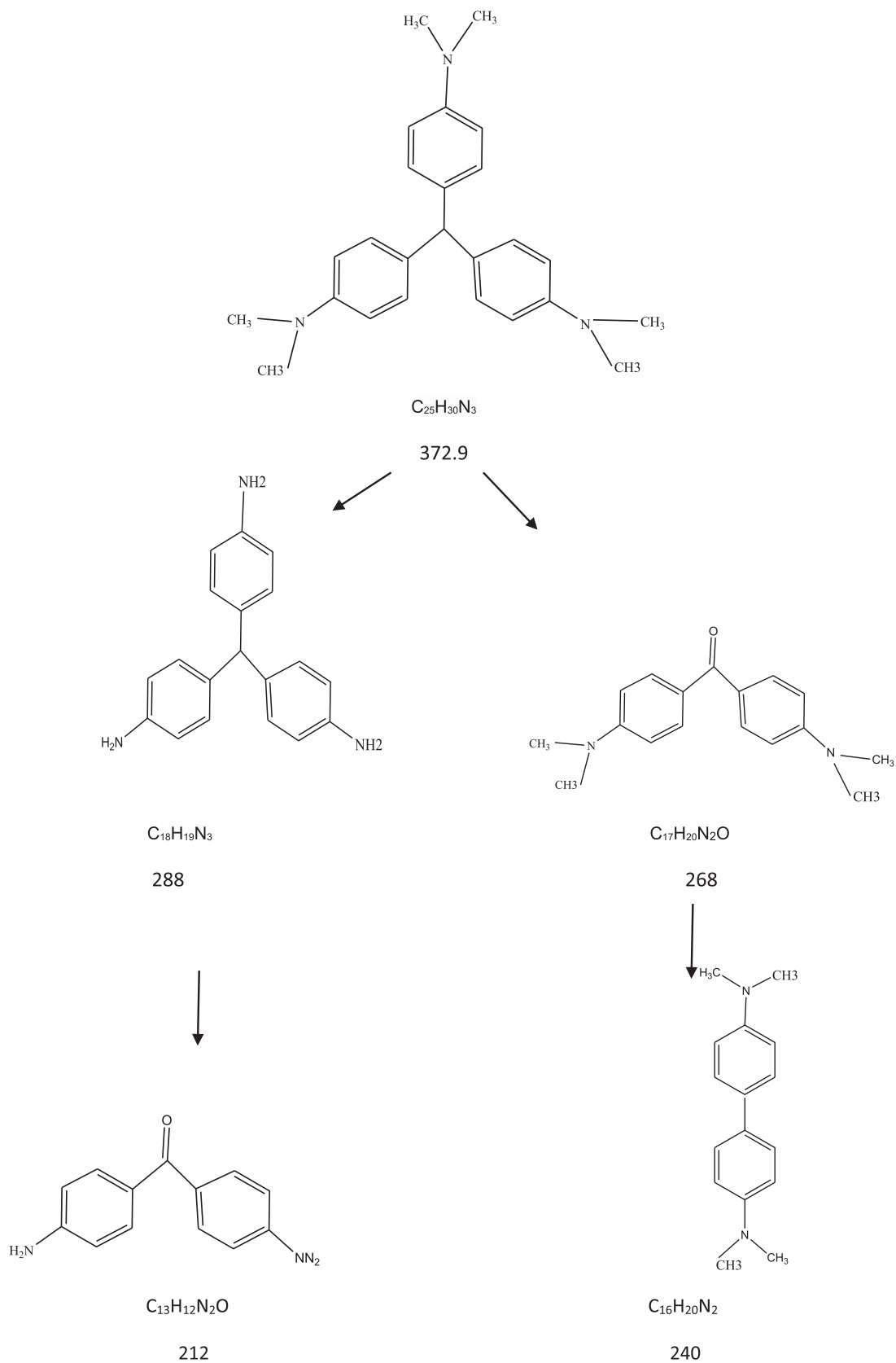


Figure 15. The proposed pathway for photocatalytic degradation for CV.

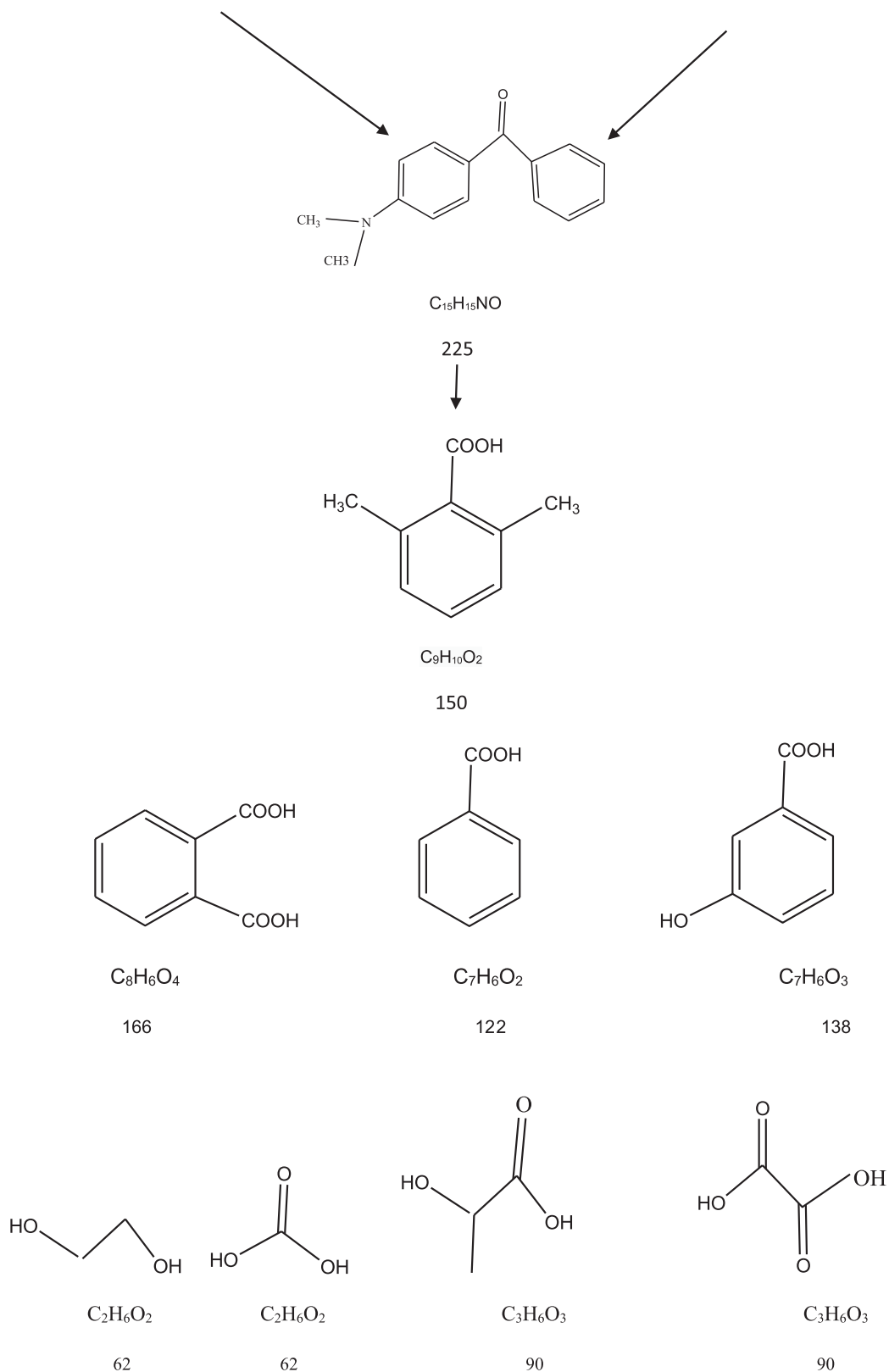


Figure 15. (Continued.)

3.2.9. Evaluation of active species

Free radicals trapping experiments were conducted for ZGW3, as the highest catalytic activity sample, to explore the significant contributor in the photodegradation reaction under optimum operating condition (CV concentration 10 mg l^{-1} at 1 g l^{-1} catalyst dose and pH 9). When 1 mmol l^{-1} IPA was introduced as

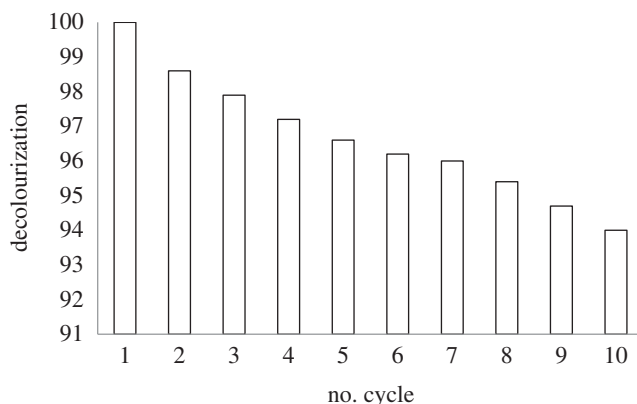


Figure 16. Number of cycles for CV over ZGW3 sample under visible light at optimum conditions (CV concentration 10 mg l^{-1} at 1 g l^{-1} catalyst dose and pH 9).

Table 4. Comparison of photocatalysts used for the degradation of CV dye.

catalyst	efficiency %	irradiation source	time	cited by
$\text{BiOCl}/\text{H}_2\text{O}_2$	100	visible	8 h	[19]
anatase nanosphere TiO_2	99	UV	6 h	[20]
Mn-doped and PVP-capped ZnO NPs	100	UV–visible irradiation	3 h	[21]
Ag-modified Ti-doped- Bi_2O_3	65	UV	90 min	[22]
TG-capped ZnS NPs	87	UV–visible irradiation	3 h	[23]
AgBr–ZnO/ H_2O_2 nanocomposite	86.93	visible light	50 min	[25]
grafted sodium alginate/ZnO/graphene oxide	94	sun light	5 h	[26]
$\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7/\text{H}_2\text{O}_2$	100	visible light	5 h	present work

HO^\bullet scavenger, the degradation of CV was not clearly influenced (CV % removal accomplished 80%), which demonstrated that the HO^\bullet was not the major reactive species. However, the degradation process could be hindered proficiently when 1 mmol l^{-1} benzoquinone (BQ) was added, since CV removal was decreased to 45%. It indicated that the O_2^\bullet played a demonstrating role in the catalysis process. When 1 mmol l^{-1} EDTA was added, the catalytic degradation of CV could be extremely inhibited (CV removal accomplished to 21%), indicating the h^+ also played a major role in the catalytic process. Consequently, h^+ and O_2^\bullet had the major contribution to photocatalytic degradation of CV while, HO^\bullet had a minor contribution [87].

4. Conclusion

Nano-sized $\text{Ga}_2\text{Zr}_{2-x}\text{W}_x\text{O}_7$ system is prepared successfully in the cubic fluorite phase using the Pechini method where $x = 0, 0.05, 0.1, 0.15$ and 0.2 . XRD, IR, EPR, XPS, TEM, BET, ICP and diffuse reflectance are used for the characterization of the prepared samples. The undoped in addition to W-doped $\text{Ga}_2\text{Zr}_2\text{O}_7$ has cubic fluorite phase structure. According to XRD, it was found that the samples are in the nano-sized range (3–4 nm). The band gap of the $\text{Ga}_2\text{Zr}_2\text{O}_7$ (4.95 eV) is close to that of ZrO_2 (5 eV). W doping decreased the band gap so that the band gap of $\text{Ga}_2\text{Zr}_{1.9}\text{W}_{0.1}\text{O}_7$ (1.81 eV) and $\text{Ga}_2\text{Zr}_{1.85}\text{W}_{0.15}\text{O}_7$ (1.7 eV) samples were found to be smaller than that of pyrochlore $\text{Sm}_2\text{Zr}_2\text{O}_7$ (2.86 eV) and $\text{Nd}_2\text{Zr}_2\text{O}_7$ (2.67 eV), while $\text{Ga}_2\text{Zr}_{1.8}\text{W}_{0.2}\text{O}_7$ has band gap matched with $\text{Nd}_2\text{Zr}_2\text{O}_7$ (2.67 eV). Full degradation for the CV dye (at 300 min, 25 mmol l^{-1} H_2O_2) is reached for $\text{Ga}_2\text{Zr}_{1.85}\text{W}_{0.15}\text{O}_7$ sample with 15 mol% W doping while lower removal was observed for 20 mol% W-doped sample which is in accordance with their band gaps obtained by DRS as well as the amount of free radical obtained from EPR analysis.

The CV dye photocatalytic degradation followed the pseudo-first-order kinetics. UV–visible and GC–MS studies were conducted to identify the by-products at the end of the reaction of decolourization. GC–MS study indicated that the degradation processes might include *N*-de-methylation followed by aromatic ring rupture. $\text{Ga}_2\text{Zr}_{1.85}\text{W}_{0.15}\text{O}_7$ can be used as a promising photocatalyst to purify recalcitrant complicated structure dye for textile water decontamination.

Data accessibility. Data are available from the Dryad Digital Repository: <https://doi.org/10.5061/dryad.zpc866t54> [88].
Authors' contributions. A.A.B. and R.A.-Z. carried out the preparation of nano materials, H.A.A. carried out the characterization of the prepared material, T.S.J. and R.A.N. carried out applications of the prepared materials for photocatalytic degradation of crystal violet dye as model compound of textile wastewater. H.A.A. and R.A.N. wrote the manuscript, and T.S.J. and A.A.B. critically revised the manuscript. R.A.N. submitted it.

Competing interests. We declare we have no competing interests.

Funding. This work was funded by the Science and Technology Development Fund in Egypt and the Scientific Research Support Fund in Jordan for financing that work through collaborative project no. 21734 in Egypt and no. EGY-Jor/1/01/2015 in Jordan.

Acknowledgements. Authors acknowledge the Science and Technology Development Fund in Egypt and the Scientific Research Support Fund in Jordan for financing that work through collaborative project no. 21734 in Egypt and no. EGY-Jor/1/01/2015 in Jordan.

References

- Zahoor M. 2012 Removal of crystal violet from water by adsorbent prepared from Turkish coffee residue. *Tenside Surfactants Deterg.* **49**, 107–113. (doi:10.3139/113.110171)
- Thairu Y, Nasir IA, Usman Y. 2014 Laboratory perspective of gram staining and its significance in investigations of infectious diseases. *Sub-Saharan Afr. J. Med.* **1**, 168. (doi:10.4103/2384-5147.144725)
- Edokpayi JN, Odiyo JO, Durowoju OS. 2017 Impact of wastewater on surface water quality in developing countries: a case study of South Africa. In *Water quality* (ed. H Tutu), pp. 401–416. Vienna, Austria: INTECH.
- Song Y, Fang H, Xu H, Tan X, Chen S. 2016 Treatment of wastewater containing crystal violet using walnut shell. *J. Residuals Sci. Technol.* **13**, 243–249. (doi:10.12783/issn.1544-8053/13/4/1)
- Mashkour F, Nasar A, Asiri AM. 2018 Exploring the reusability of synthetically contaminated wastewater containing crystal violet dye using *Tectona grandis* sawdust as a very low-cost adsorbent. *Sci. Rep.* **8**, 8314. (doi:10.1038/s41598-018-26655-3)
- Mani S, Bharagava RN. 2016 Exposure to crystal violet, its toxic, genotoxic and carcinogenic effects on environment and its degradation and detoxification for environmental safety. In *Reviews of environmental contamination and toxicology volume 237*, pp. 71–104. Berlin, Germany: Springer.
- Sun M, Ricker K, Osborne G, Marder ME, Schmitz R. 2018 *Evidence on the carcinogenicity of gentian violet*. Office of Environmental Health Hazard Assessment, California Environmental Protection Agency.
- Wallace RB, Oria ME. 2010 Appendix E: The U.S. Food and Drug Administration and imported food safety. In *Enhancing food safety: the role of the Food and Drug Administration* (eds RB Wallace, ME Oria), pp. 451–492. Washington, DC: The National Academic Press.
- Zhang X, Yang Y, Lv X, Wang Y, Liu N, Chen D, Cui L. 2019 Adsorption/desorption kinetics and breakthrough of gaseous toluene for modified microporous–mesoporous UiO-66 metal organic framework. *J. Hazard. Mater.* **366**, 140–150. (doi:10.1016/j.jhazmat.2018.11.099)
- Yang Y, Ding Q, Yang M, Wang Y, Liu N, Zhang X. 2018 Magnetic ion exchange resin for effective removal of perfluorooctanoate from water: study of a response surface methodology and adsorption performances. *Environ. Sci. Pollut. Res.* **25**, 29 267–29 278. (doi:10.1007/s11356-018-2797-1)
- Wei L, Sun H, Yang T, Deng S, Wu M, Li Z. 2018 Iron carbide encapsulated by porous carbon nitride as bifunctional electrocatalysts for oxygen reduction and evolution reactions. *Appl. Surf. Sci.* **439**, 439–446. (doi:10.1016/j.apsusc.2018.01.056)
- Zhang X, Yang Y, Huang W, Yang Y, Wang Y, He C, Liu N, Wu M, Tang L. 2018 g-C₃N₄/UiO-66 nanohybrids with enhanced photocatalytic activities for the oxidation of dye under visible light irradiation. *Mater. Res. Bull.* **99**, 349–358. (doi:10.1016/j.materresbull.2017.11.028)
- Liu N, Huang W, Tang M, Yin C, Gao B, Li Z, Tang L, Lei J, Cui L, Zhang X. 2019 In-situ fabrication of needle-shaped MIL-53 (Fe) with 1T-MoS₂ and study on its enhanced photocatalytic mechanism of ibuprofen. *Chem. Eng. J.* **359**, 254–264. (doi:10.1016/j.cej.2018.11.143)
- Bhayani RB. 2014 Color removal of dyes wastewater by coagulation and microfiltration processes. Bachelor of Engineering in Civil Engineering, Maharaja Sayajirao University, Vadodara, India.
- Sillanpää M, Ncibi MC, Matilainen A. 2018 Advanced oxidation processes for the removal of natural organic matter from drinking water sources: a comprehensive review. *J. Environ. Manage.* **208**, 56–76. (doi:10.1016/j.jenvman.2017.12.009)
- Wang B, Zhang H, Wang F, Xiong X, Tian K, Sun Y, Yu T. 2019 Application of heterogeneous catalytic ozonation for refractory organics in wastewater. *Catalysts* **9**, 241. (doi:10.3390/catal9030241)
- Matavos-Aramyan S, Moussavi M. 2017 Advances in Fenton and Fenton based oxidation processes for industrial effluent contaminants control: a review. *Int. J. Environ. Sci. Nat. Resour.* **2**, 1–18.
- Qiao Y. 2018 Preparation, characterization, and evaluation of photocatalytic properties of a novel NaNbO₃/Bi₂WO₆ heterostructure photocatalyst for water treatment. Master of Applied Science degree in Chemical Engineering, University of Ottawa, Canada.
- Pare B, Swami D, Thapak TR, Qureshi T. 2011 Photocatalytic degradation of environmentally hazardous crystal violet dye using bismuth oxychloride as photocatalyst. *Int. J. Chem. Sci.* **9**, 1183–1193.
- Jadhav VV, Dhabbe RS, Sabale SR, Nikam GH, Tamhankar BV. 2013 Degradation of dyes using high temperature stable anatase nanosphere TiO₂ photocatalyst. *Univers. J. Environ. Res. Technol.* **3**, 667–676.
- Mittal M, Sharma M, Pandey OP. 2014 Photocatalytic studies of crystal violet dye using Mn doped and PVP capped ZnO nanoparticles. *J. Nanosci. Nanotechnol.* **14**, 2725–2733. (doi:10.1166/jnn.2014.8615)
- Zhang L, Niu J, Li D, Gao D, Shi J. 2014 Preparation and photocatalytic activity of Ag modified Ti-doped-Bi₂O₃ photocatalyst. *Adv. Condens. Matter Phys.* **2014**, 1–6.
- Sharma M, Jain T, Singh S, Pandey OP. 2012 Photocatalytic degradation of organic dyes under UV–visible light using capped ZnS nanoparticles. *Sol. Energy* **86**, 626–633. (doi:10.1016/j.solener.2011.11.006)

24. Zahoor M *et al.* 2018 Enhanced photocatalytic performance of CeO₂-TiO₂ nanocomposite for degradation of crystal violet dye and industrial waste effluent. *Appl. Nanosci.* **8**, 1091–1099. (doi:10.1007/s13204-018-0730-z)
25. Abdel-Khalek AA, Mahmoud SA, Zaki AH. 2018 Visible light assisted photocatalytic degradation of crystal violet, bromophenol blue and eosin Y dyes using AgBr-ZnO nanocomposite. *Environ. Nanotechnol. Monit. Manag.* **9**, 164–173. (doi:10.1016/j.enmm.2018.03.002)
26. Mohamed SK, Hegazy SH, Abdelwahab NA, Ramadan AM. 2018 Coupled adsorption-photocatalytic degradation of crystal violet under sunlight using chemically synthesized grafted sodium alginate/ZnO/graphene oxide composite. *Int. J. Biol. Macromol.* **108**, 1185–1198. (doi:10.1016/j.jbiomac.2017.11.028)
27. Liu ZG, Ouyang JH, Zhou Y, Xia XL. 2010 Electrical conductivity and thermal expansion of neodymium-ytterbium zirconate ceramics. *J. Power Sources* **195**, 3261–3265. (doi:10.1016/j.jpowsour.2009.11.135)
28. Chon MP, Tan KB, Zainal Z, Taufiq-Yap YH, Tan PY, Khaw CC, Chen SK. 2016 Synthesis and electrical properties of Zn-substituted bismuth copper tantalate pyrochlores. *Int. J. Appl. Ceram. Technol.* **13**, 718–725. (doi:10.1111/ijac.12547)
29. Hatnean MC, Lees MR, Balakrishnan G. 2015 Growth of single-crystals of rare-earth zirconate pyrochlores, Ln₂Zr₂O₇ (with Ln = La, Nd, Sm, and Gd) by the floating zone technique. *J. Cryst. Growth* **418**, 1–6. (doi:10.1016/j.jcrysgro.2015.01.037)
30. Nigam S, Sudarsan V, Vatsa RK. 2013 Effect of annealing temperature on the structural and photoluminescence properties of Y₂Sn₂O₇:Eu nanoparticles. *Eur. J. Inorg. Chem.* **2013**, 357–363. (doi:10.1002/ejic.201200804)
31. Mandal BP, Tyagi AK. 2010 Pyrochlores. Potential multifunctional material. *BARC Newsletter* **313**, 6–13.
32. Zhang G, Jiang W, Yu S. 2010 Preparation, characterization and photocatalytic property of nanosized K-Ta mixed oxides via a sol-gel method. *Mater. Res. Bull.* **45**, 1741–1747. (doi:10.1016/j.materresbull.2010.06.052)
33. Fowler P, Homan A, Atkins D, Whitwell J, Lloyd M, Bradford R. 2016 The utility of the *in vitro* micronucleus test for evaluating the genotoxicity of natural and manmade nanoscale fibres. *Mutat. Res. Genet. Toxicol. Environ. Mutagen.* **809**, 33–42. (doi:10.1016/j.mrgentox.2016.09.002)
34. Wang W, Bi J, Wu L, Li Z, Wang X, Fu X. 2008 Hydrothermal synthesis and performance of a novel nanocrystalline Pb₂Sn₂O₆ photocatalyst. *Nanotechnology* **19**, 505705. (doi:10.1088/0957-4484/19/50/505705)
35. Ravi G, Veldurthi NK, Palla S, Velchuri R, Pola S, Reddy JR, Vithal M. 2013 Synthesis, characterization and photocatalytic activity of KAl_{0.33}W_{1.67}O₆ and Sn_{0.5}Al_{0.33}W_{1.67}O₆·xH₂O. *Photochem. Photobiol.* **89**, 824–831. (doi:10.1111/php.12079)
36. Jitta RR, Guje R, Veldurthi NK, Prathapuram S, Velchuri R, Muga V. 2015 Preparation, characterization and photocatalytic studies of N, Sn-doped defect pyrochlore oxide KTi_{0.5}W_{1.5}O₆. *J. Alloys Compd.* **618**, 815–823. (doi:10.1016/j.jallcom.2014.08.157)
37. Guje R, Shrujana P, Veldurthi NK, Gundeboina R, Kappera NR, Muga V. 2014 Synthesis, characterization and photocatalytic activity of Ag⁺- and Sn²⁺-substituted K₂TeO₆. *Chem. Pap.* **69**, 269–278. (doi:10.1515/chempap-2015-0013)
38. Kanhere P, Tang Y, Zheng J, Chen Z. 2013 Synthesis, photophysical properties, and photocatalytic applications of Bi doped NaTaO₃ and Bi doped Na₂Ta₂O₇ nanoparticles. *J. Phys. Chem. Solids* **74**, 1708–1713. (doi:10.1016/j.jpcs.2013.06.013)
39. Singh J, Uma S. 2009 Efficient photocatalytic degradation of organic compounds by bilmenite AgSbO₃ under visible and UV light irradiation. *J. Phys. Chem. C* **113**, 12 483–12 488.
40. Malathi M, Sreenu K, Ravi G, Kumar PV, Reddy CS, Guje R, Velchuri R, Vithal M. 2017 Low temperature synthesis of fluorite-type Ce-based oxides of composition Ln₂Ce₂O₇ (Ln = Pr, Nd and Eu): photodegradation and luminescence studies. *J. Chem. Sci.* **129**, 1193–1203. (doi:10.1007/s12039-017-1321-3)
41. Pechini M. 1967 *Method for preparing lead and alkaline earth titanates and niobates and coating method using the same to form a capacitor*. Patent no. US3330697A.
42. Carrillo AJ, Serrano DP, Pizarro P, Coronado JM. 2016 Design of efficient Mn-based redox materials for thermochemical heat storage at high temperatures. In *AIP Conf. Proc., 21st SolarPACES Conf., Cape Town, South Africa, 13–16 October*, p. 50009. AIP Publishing.
43. Ribeiro PC, De Melo Da Costa ACF, Kiminami RHGA, Sasaki JM, Lira HL. 2013 Synthesis of TiO₂ by the Pechini method and photocatalytic degradation of methyl red. *Mater. Res.* **16**, 468–472. (doi:10.1590/S1516-14392012005000176)
44. Danks AE, Hall SR, Schnepf Z. 2016 The evolution of ‘sol-gel’ chemistry as a technique for materials synthesis. *Mater. Horizons* **3**, 91–112. (doi:10.1039/c5mh00260e)
45. Sunde TOL, Grande T, Einarssrud M-A. 2016 Modified Pechini synthesis of oxide powders and thin films. In *Handbook of sol-gel science and technology* (eds L Klein, M Aparicio, A Jitianu), pp. 1–30. Cham, Switzerland: Springer.
46. Zhang L, Yang J, Li J. 2014 A novel composite cathode for intermediate temperature solid oxide fuel cell. *J. Power Sources* **269**, 723–726. (doi:10.1016/j.jpowsour.2014.07.076)
47. Holland TJB, Redfern SAT. 1997 Unit cell refinement from powder diffraction data; the use of regression diagnostics. *Miner. Mag.* **61**, 65–77. (doi:10.1180/minmag.1997.061.404.07)
48. Mandal BP, Garg N, Sharma SM, Tyagi AK. 2006 Preparation, XRD and Raman spectroscopic studies on new compounds RE₂Hf₂O₇ (RE = Dy, Ho, Er, Tm, Lu, Y): pyrochlores or defect-fluorite? *J. Solid State Chem.* **179**, 1990–1994. (doi:10.1016/j.jssc.2006.03.036)
49. Catchen GL, Rearick TM. 1995 O-anion transport measured in several R₂M₂O₇ pyrochlores using perturbed-angular-correlation spectroscopy. *Phys. Rev. B* **52**, 9890–9899. (doi:10.1103/PhysRevB.52.9890)
50. Mandal BP, Banerji A, Sathe V, Deb SK, Tyagi AK. 2007 Order-disorder transition in Nd_{2-x}Gd_xZr₂O₇ pyrochlore solid solution: an X-ray diffraction and Raman spectroscopic study. *J. Solid State Chem.* **180**, 2643–2648. (doi:10.1016/j.jssc.2007.07.007)
51. Whittle KR, Granswick LMD, Redfern SAT, Swainson IP, Lumpkin GR. 2009 Lanthanum pyrochlores and the effect of yttrium addition in the systems La_{2-x}Y_xZr₂O₇ and La_{2-x}Y_xHf₂O₇. *J. Solid State Chem.* **182**, 442–450. (doi:10.1016/j.jssc.2008.11.008)
52. Shannon RD. 1976 Revised effective ionic radii and systematic studies on inter-atomic distances in halides and chalcogenides. *Acta Cryst. A* **32**, 751–767. (doi:10.1107/S0567739476001551)
53. Zhang Y, Guo L, Zhao X, Ye F. 2014 Effects of non-stoichiometry on the mechanical properties of Nd_{2-x}Zr_{2+x}O_{7+x/2} (x = 0, 0.1, 0.2, 0.3, 0.4, 0.5) ceramics. *Mater. Lett.* **136**, 157–159. (doi:10.1016/j.matlet.2014.08.065)
54. Hao CK, Lee CS. 2013 Metal-doped pyrochlore as novel electrode materials for intermediate temperature solid oxide fuel cell. *ECS Trans.* **58**, 165–173. (doi:10.1149/05803.0165ecst)
55. Yakout SM, Hassan HS. 2014 Adsorption characteristics of sol gel-derived zirconia for cesium ions from aqueous solutions. *Molecules* **19**, 9160–9172. (doi:10.3390/molecules19079160)
56. Gurusantha K, Anantharaju KS, Nagabhushana H, Sharma SC, Vidya YS, Shivakumara C, Nagaswarupa HP, Prashantha SC, Anilkumar MR. 2015 Facile green fabrication of iron-doped cubic ZrO₂ nanoparticles by *Phyllanthus acidus*: structural, photocatalytic and photoluminescent properties. *J. Mol. Catal. A Chem.* **397**, 36–47. (doi:10.1016/j.molcata.2014.10.025)
57. Badenes JA, Vicent JB, Llusar M, Tena MA, Monros G. 2002 The nature of Pr-ZrSiO₄ yellow ceramic pigment. *J. Mater. Sci.* **37**, 1413–1420. (doi:10.1023/A:1014537000690)
58. Krehula S, Ristić M, Kubuki S, Iida Y, Fabián M, Musić S. 2015 The formation and microstructural properties of uniform α-GaOOH particles and their calcination products. *J. Alloys Compd.* **620**, 217–227. (doi:10.1016/j.jallcom.2014.09.134)
59. Rougier A, Portemer F, Quédè A, El Marssi M. 1999 Characterization of pulsed laser deposited WO₃ thin films for electrochromic devices. *Appl. Surf. Sci.* **153**, 1–9. (doi:10.1016/S0169-4332(99)00335-9)
60. Orel B, Grošelj N, Krašovec UO, Ješe R, Georg A. 2002 IR spectroscopic investigations of gasochromic and electrochromic sol-gel—derived peroxotungstic acid/ormosil composite and crystalline WO₃ films. *J. Sol-Gel Sci. Technol.* **24**, 5–22. (doi:10.1023/A:1015147530846)
61. Jamil TS, Mansor ES, Nasr RA. 2016 Degradation of Lindane using two nanosized BiOXs and their heterojunction under visible light. *Desalin. Water Treat.* **57**, 14 750–14 761. (doi:10.1080/19443994.2015.1063461)
62. Sayama K, Arakawa H. 1994 Effect of Na₂CO₃ addition on photocatalytic decomposition of liquid water over various semiconductor catalysis. *J. Photochem. Photobiol. A Chem.*

- 77, 243–247. (doi:10.1016/1010-6030(94)80049-9)
63. Sayama K, Arakawa H 1996 Effect of carbonate addition on the photocatalytic decomposition of liquid water over a ZrO_2 catalyst. *J. Photochem. Photobiol. A Chem.* **94**, 67–76.
64. Chiesa M, Giamello E, Livraghi S, Paganini MC, Polliotto V, Salvadori E. 2019 Electron magnetic resonance in heterogeneous photocatalysis research. *J. Phys. Condens. Matter* **31**, 444001. (doi:10.1088/1361-648X/ab32c6)
65. Li X, Mao X, Feng M, Xie J, Jiang B, Zhang L. 2016 Optical absorption and mechanism of vacuum-sintered ZrO_2 -doped Y_2O_3 ceramics. *J. Eur. Ceram. Soc.* **36**, 4181–4184. (doi:10.1016/j.jeurceramsoc.2016.05.046)
66. Venkatesan M, Stamenov P, Dorneles LS, Gunning RD, Bernoux B, Coey JMD. 2007 Magnetic, magnetotransport, and optical properties of Al-doped $Zn_{0.95}Co_{0.05}O$ thin films. *Appl. Phys. Lett.* **90**, 100–103. (doi:10.1063/1.2748343)
67. Sharaf El-Deen SE, Ammar NS, Jamil TS. 2016 Adsorption behavior of Co(II) and Ni(II) from aqueous solutions onto titanate nanotubes. *Fuller. Nanotub. Carbon Nanostructure* **24**, 455–466. (doi:10.1080/1536383X.2016.1179287)
68. Vasic MB, Radelović MS, Momčilović MZ, Matović B, Zarubica AR. 2016 Degradation of crystal violet over heterogeneous TiO_2 -based catalysts: the effect of process parameters. *Process. Appl. Ceram.* **10**, 189–198. (doi:10.2298/PAC1603189V)
69. Pare B, Singh P, Jonnalagadda SB. 2010 Visible light driven photocatalytic degradation and mineralization of neutral red dye in slurry photoreactor. *Indian J. Chem. Technol.* **17**, 391. (doi:10.1080/02726351.2016.1168893)
70. Ksibi M, Ben AS, Cherif S, Elaloui E, Houas A, Elaloui M. 2003 Photodegradation of lignin from black liquor using a UV/ TiO_2 system. *J. Photochem. Photobiol. A Chem.* **154**, 211–218. (doi:10.1016/S1010-6030(02)00316-7)
71. Alex S, Santhosh U, Das S. 2005 Dye sensitization of nanocrystalline TiO_2 : enhanced efficiency of unsymmetrical versus symmetrical squaraine dyes. *J. Photochem. Photobiol. A Chem.* **172**, 63–71. (doi:10.1016/j.jphotochem.2004.11.005)
72. Verma NK, Gurjar L, Bhardwaj S. 2011 Use of Ta_2O_5 as a photocatalyst for degradation of crystal violet using solar energy. *Der Chem. Sin.* **2**, 258–264.
73. Suhail FS, Mashkour MS, Saeb D. 2015 The study on photo degradation of crystal violet by polarographic technique. *Int. J. Basic Appl. Sci.* **15**, 12–21.
74. Devi LG, Kumar SG. 2012 Exploring the critical dependence of adsorption of various dyes on the degradation rate using Ln^{3+} - TiO_2 surface under UV/solar light. *Appl. Surf. Sci.* **261**, 137–146. (doi:10.1016/j.apsusc.2012.07.121)
75. Abbas HA, Jamil TS, Hammad FF. 2016 Synthesis, characterization and photocatalytic activity of nano sized undoped and Ga doped $SrTi_{0.7}Fe_{0.3}O_3$ for 2,4,6-trichlorophenol photodegradation. *J. Environ. Chem. Eng.* **4**, 2384–2393. (doi:10.1016/j.jece.2016.04.019)
76. Karim A, Soltani RD, Khataee A, Soltani B, Arefi-Oskoui S. 2015 Sonochemical synthesis of Pr-doped ZnO nanoparticles for sonocatalytic degradation of Acid Red 17. *Ultrason. Sonochem.* **22**, 371–381. (doi:10.1016/j.ultrsonch.2014.05.023)
77. Subash B, Krishnakumar B, Swaminathan M, Shanthi M. 2013 Highly efficient, solar active, and reusable photocatalyst: Zr-loaded Ag–ZnO for reactive red 120 dye degradation with synergistic effect and dye-sensitized mechanism. *Langmuir* **3**, 939–949. (doi:10.1021/la303842c)
78. Dong C, Ji J, Shen B, Xing M, Zhang J. 2018 Enhancement of H_2O_2 decomposition by the co-catalytic effect of WS_2 on the Fenton reaction for the synchronous reduction of Cr(VI) and remediation of phenol. *Environ. Sci. Technol.* **52**, 11 297–11 308. (doi:10.1021/acs.est.8b02403)
79. Yi Q, Ji J, Shen B, Dong C, Liu J, Zhang J, Xing M. 2019 Singlet oxygen triggered by superoxide radicals in a molybdenum cocatalytic Fenton reaction with enhanced REDOX activity in the environment. *Environ. Sci. Technol.* **53**, 9725–9733. (doi:10.1021/acs.est.9b01676)
80. Liu Y, Yu H, Lv Z, Zhan S, Yang J, Peng X, Ren Y, Wu X. 2012 Simulated-sunlight-activated photocatalysis of Methylene Blue using cerium-doped SiO_2/TiO_2 nanostructured fibers. *J. Environ. Sci.* **24**, 1867–1875. (doi:10.1016/S1001-0742(11)61008-5)
81. Kumar R, Rashid J, Barakat MA. 2015 Zero valent Ag deposited TiO_2 for the efficient photocatalysis of methylene blue under UV-C light irradiation. *Colloids Interface Sci. Commun.* **5**, 1–4. (doi:10.1016/j.colcom.2015.05.001)
82. Palma-Goyes RE, Guzmán-Duque FL, Peñuela G, González I, Nava JL, Torres-Palma RA. 2010 Electrochemical degradation of crystal violet with BDD electrodes: effect of electrochemical parameters and identification of organic by-products. *Chemosphere* **81**, 26–32. (doi:10.1016/j.chemosphere.2010.07.020)
83. Zhang H, Wu J, Wang Z, Zhang D. 2010 Electrochemical oxidation of crystal violet in the presence of hydrogen peroxide. *J. Chem. Technol. Biotechnol.* **85**, 1436–1444. (doi:10.1002/jctb.2447)
84. He H, Yang S, Yu K, Ju Y, Sun C, Wang L. 2010 Microwave induced catalytic degradation of crystal violet in nano-nickel dioxide suspensions. *J. Hazard. Mater.* **173**, 393–400. (doi:10.1016/j.jhazmat.2009.08.084)
85. Guzman-Duque F, Pétrier C, Pulgarin C, Peñuela G, Torres-Palma RA. 2011 Effects of sonochemical parameters and inorganic ions during the sonochemical degradation of crystal violet in water. *Ultrason. Sonochem.* **18**, 440–446. (doi:10.1016/j.ultrsonch.2010.07.019)
86. Jamil TS, Abbas HA, Nasr RA, El-Kady AA, Ibrahim MIM. 2017 Detoxification of aflatoxin B1 using nano-sized Sc-doped $SrTi_{0.7}Fe_{0.3}O_3$ under visible light. *J. Photochem. Photobiol. A Chem.* **341**, 127–135. (doi:10.1016/j.jphotochem.2017.03.023)
87. Chiu YH, Chang TFM, Chen CY, Sone M, Hsu YJ. 2019 Mechanistic insights into photodegradation of organic dyes using heterostructure photocatalysts. *Catalysts* **9**, 430. (doi:10.3390/catal9050430)
88. Abbas HA, Nasr RA, Abu-Zurayk R, Al Bawab A, Jamil TS. 2020 Data from: Decolorization of crystal violet using nano-sized novel fluorite structure $Ga_2Zr_{2-x}W_xO_7$ photocatalyst under visible light irradiation. Dryad Digital Repository. doi:10.5061/dryad.zpc866t54.