Principles of fluoride toxicity and the cellular response: a review

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Abstract

Fluoride is ubiquitously present throughout the world. It is released from minerals, magmatic gas, and industrial processing, and travels in the atmosphere and water. Exposure to low concentrations of fluoride increases overall oral health. Consequently, many countries add fluoride to their public water supply at 0.7–1.5 ppm. Exposure to high concentrations of fluoride, such as in a laboratory setting often exceeding 100 ppm, results in a wide array of toxicity phenotypes. This includes oxidative stress, organelle damage, and apoptosis in single cells, and skeletal and soft tissue damage in multicellular organisms. The mechanism of fluoride toxicity can be broadly attributed to four mechanisms: inhibition of proteins, organelle disruption, altered pH, and electrolyte imbalance. Recently, there has been renewed concern in the public sector as to whether fluoride is safe at the current exposure levels. In this review, we will focus on the impact of fluoride at the chemical, cellular, and multisystem level, as well as how organisms defend against fluoride. We also address public concerns about fluoride toxicity, including whether fluoride has a significant effect on neurodegeneration, diabetes, and the endocrine system.

Keywords

fluoride; toxicity; metal; cell stress

I. Chemical properties of fluoride

The element fluorine has the highest electronegativity and the second highest electron affinity, making it highly reactive. At room temperature, fluorine exists as the gas F_2, which reacts explosively with many elements. Fluorine is so reactive that it can form complexes with noble gases, most notably xenon (Holloway 1966). Due to its low stability, isolated fluorine is never found in nature. Instead, fluorine is either found as a complex or in its ionized form, fluoride.

Fluoride interacts with many cations, including hydrogen and a wide variety of metals. It is the only halide with a positive pK_a (3.2), and therefore exists in acidic environments as its

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protonated form (HF). HF, commonly released as industrial or volcanic fumes, turns gaseous above 20. Fluoride is most toxic in its protonated form, and vertebrates that reside in areas near HF production often show symptoms of lung damage and fluoride toxicity.

Fluoride readily associates with metals. This affinity is driven by three factors: a negative Gibbs free energy for formation, a high stability constant, and poor solubility of the metal-fluoride complex. The energy state and stability of metallo-fluoride is much greater than that of other metallo-halides, and as such fluoride can often displace metal interacting partners in nature. The most favorable metal reactions with fluoride are aluminum, calcium, and magnesium, while the most stable are aluminum, iron, and beryllium. Overall, the most favorable metal to bind fluoride is aluminum (Skelton 1971; Martin 1996). Aluminum is also the most abundant metal on the earth's crust, and only micromolar levels of aluminum are necessary to complex with fluoride (Mullenix 2014; Berger *et al.* 2015). In terms of biological relevancy, calcium and magnesium are at much higher abundance and form complexes with fluoride *in vivo* (Marier 1980; Spencer *et al.* 1980). Interestingly, both of these metals are highly insoluble when complexed with fluoride. As calcium- or magnesium-fluoride precipitates out of solution, these complexes do not readily associate with other elements.

II. The abundancy of fluoride in nature

Fluoride gradually accumulates in the environment from volcanic emissions, dissolution of minerals, and industrial byproducts. Fluoride is present at roughly 300–900 ppm throughout the earth's crust, and is estimated to be 200-fold more abundant in the mantle (Weinstein and Davison 2004; Koga and Rose-Koga 2018). Therefore, magma and magmatic gas contain fluoride, both unbound and complexed with hydrogen, silicon, and ammonium (Das and Behera 2008). These molecules settle as ash into the nearby soil and groundwater. Monitoring of fluoride particles in ash from the 1991 eruption of Mt. Hudson in Chile found that fluoride was highest in ash deposited furthest from the volcano; fluoride completely dissolved out of the ash and seeped into the surrounding soil and water sources after the first rainfall (Rubin *et al.* 1994). Historically, organisms in areas with high volcanic activity have displayed signs of fluoride poisoning (Olsen and Fruchter 1986; Witham *et al.* 2005; Flueck and Smith-Flueck 2013). Today, volcanic regions with high fluoride include Sicily, New Zealand, Iceland, the East African Rift, China, and South India (Cronin *et al.* 2003; Bellomo *et al.* 2006; D'Alessandro 2006).

Fluorine can also be found in 296 different species of minerals, the most abundant of which are fluorspar (CaF_2), fluorapatite ($Ca_5(PO_4)_3F$), topaz ($Al_2(SiO_4)(F,OH)_2$), and cryolite (Na_3AlF_6) (Mineralogy Database; Garcia and Borgnino 2015). These occur as vein deposits associated with igneous rocks, especially metallic minerals. Some of the largest deposits of fluoro-minerals are found in China, Mexico, Mongolia, South Africa, and Russia; this occurrence correlates strongly with areas whose populations suffer from endemic fluoride toxicity (Fig. 1) (Kilgore and Pelham 1987; Gupta and Ayoob 2016). Endemic fluoride toxicity in humans occurs from the chronic consumption of over 1.5 ppm (75 μM) fluoride.

Demineralization leads to an increased concentration of fluoride in the soil and water. Over time, minerals naturally break down through weathering and erosion. Most fluoride-metal complexes, especially fluorspar and fluorapatite, are poorly soluble in water (Moreno *et al.* 1974; Pan and Darvell 2007). Without further dissolution, the fluoride-metal complexes distribute into the soil where they are absorbed by plants and microbes. However, alkaline groundwater (pH >8) can solubilize fluoride from minerals. The East African Rift Valley has several alkaline lakes near volcanic sites; there fluoride concentrations in groundwater range from 2.1–9.0 ppm (100–500 μM) (Malago *et al.* 2017). Fluoride dissolution from minerals is also favored under conditions with geothermal water, low concentrations of calcium and other metals, and areas with strong evaporation (Jha *et al.* 2013; Bouzourra *et al.* 2015; Batabyal and Gupta 2017). Acid rain has similar effects. The acidity releases sodium bicarbonates into the water, which in turn react with and dissolve fluoride minerals (Nath and Dutta 2010; Salifu *et al.* 2012). Over time, fluoride eventually deposits into the ocean. Oceanic water naturally contains fluoride, which cycles in and out of the atmosphere through the water cycle (Carpenter 1969). This fluoride is thought to be from the breakdown of marine sediments, such as phosphate rock. Free fluoride is present in the ocean at 1.2–1.4 ppm (60–80 μM) (WHO 2004).

Fluoride is also released into the air, water, and soil during mining and industrial processes. Fluoride is either used in a reaction, such as in aluminum smelting or glass production, or released as a byproduct, such as in phosphate fertilizer production, ceramic production and coal burning (Monfort *et al.* 2008; Gouider *et al.* 2010; Seixas *et al.* 2010). From these processes, excess fluoride is released into the environment as fumes or in groundwater (Roy *et al.* 2017). Phosphate fertilizer is a particular problem for widespread fluoride groundwater contamination, as fluoride is both released during the breakdown of phosphate ore, and makes up an additional 1.5–3% of the final fertilizer, which then enters the environment (McLaughlin *et al.* 1996; Li *et al.* 2015). For fluoride released as fumes, coal burning poses a substantial health risk. Fluoride is emitted as fumes upon coal combustion. In areas with indoor coal usage, many individuals show signs of fluoride toxicity (Li and Cao 1994; Ando *et al.* 2001). In 1997, an estimated 31 million people in China showed signs of fluoride toxicity from coal intake (Ando *et al.* 2001). Fluoride travels efficiently through the air before precipitating back into the soil and water, so fluoride emissions can lead to higher fluoride exposure in areas far from the original industrial site (Walna *et al.* 2013).

Fluoride exposure can also vary based on diet. All foods contain fluoride. Some of the highest concentrated food sources occur from fluoride-accumulating plants, such as tomato, spinach, grapes, tea, and elderberry. The Hitchcock lab found that tomatoes grown in 10 mM NaF (190 ppm) accumulated 900 ppm fluoride (dry weight) in their leaves (Jacobsen *et al.* 1966). However, the amount of fluoride in plants depends strongly on the fluoride concentration in the soil; U.S.-grown tomatoes typically contain around 0.02 ppm fluoride. Plants such as tea accumulate fluoride with age. In populations with heavy tea consumption, fluoridated tea is believed to be the primary mechanism of adult fluoride toxicity (Cao *et al.* 2003). In many developed countries, a large majority of fluoride exposure comes from government-instituted fluoridated water (0.7–1.5 ppm), toothpaste (typically 1,000–2,500 ppm), dental gel (12,300 ppm) and varnish (23,000 ppm) (Cappelli and Mobley 2008; Pretty 2016).

III. Benefits of water fluoridation

The practice of adding fluoride to the public water supply has, in general, increased overall dental health. The first documentation of fluoride's effects began with three independent reports from Italy, America, and the U.K. in the early 1900s noting that individuals with mottled, brown-stained teeth had lower incidences of dental caries (Eager 1901; McKay 1912; Ainsworth 1933). H.V. Churchill, who originally attributed the effect to high aluminum exposure, later found fluoride to be the causative agent preventing caries and triggering the skeletal defect later known as "fluorosis" (Churchill 1931). Fluoride was first added to the public water supply in Grand Rapids, Michigan in 1945 as a method of caries prevention. As of 2019, 24 countries participate in water fluoridation. An estimated 30–60% of caries incidences have been reduced because of water fluoridation (Armfield 2010). Because of this advancement, the CDC named water fluoridation one of the top ten greatest public health achievements of the twentieth century (CDC 1999).

Controlled exposure to fluoride increases overall teeth quality through enamel replacement and the killing of plaque-causing bacteria. Fluoride reacts with tooth enamel because of its high affinity to metals. Tooth enamel is comprised mostly of the mineral hydroxyapatite $(Ca_{10}(PO_4)_6(OH)_2)$. Fluoride readily displaces the hydroxide to form fluorapatite $(Ca_5(PO_4)_3)F)$ (Fig. 2). Under normal conditions bacteria like *Streptococcus mutans* ferment along the enamel, producing acid that gradually dissociates the hydroxyapatite (critical pH 5.5) (Featherstone 2008). When fluoride is present, it scavenges excess phosphate and calcium in the saliva for partial tooth remineralization, as well as displaces the hydroxyl group in remaining enamel (Featherstone 1999; Amaechi and van Loveren 2013). The resulting fluorapatite is more resistant to acidity (critical pH 4.5) than normal enamel, and as such, individuals exposed to fluoride have less tooth decay and more enamel than individuals with no fluoride exposure (Slade *et al.* 2018). For this reason, fluoride is added to the water sources of many countries.

IV. Toxicity from the interaction of fluoride with metals

While low doses of fluoride are beneficial for overall teeth integrity, high doses of fluoride lead to a myriad of toxicity phenotypes. Broadly, fluoride triggers oxidative stress, cell cycle arrest, and apoptosis. While the exact mechanism of fluoride toxicity is unknown, the stress phenotypes are generally attributed to the inhibition of proteins, organelle disruption, altered pH, and electrolyte imbalance (Adamek *et al.* 2005; Barbier *et al.* 2010; Agalakova *et al.* 2011). These four mechanisms occur to varying degrees depending on the concentration of fluoride, its route of administration in multicellular organisms, and each cell's surrounding environment. There is not a complete consensus over which downstream stress phenotype is linked to each mechanism, as each stressor can independently trigger oxidative stress and apoptosis. However, most papers attribute the primary mechanism of fluoride toxicity with its ability to inhibit metalloproteins (Adamek *et al.* 2005; Agalakova *et al.* 2011). A common practice in the study of fluoride toxicity is to look for reduced activity in an organelle or pathway under fluoride exposure, and then find essential metalloproteins within that organelle/pathway. However, many pathways show altered rates at lower fluoride exposure

than that needed to inhibit essential proteins *in vitro*. This suggests either a still unidentified primary target, or that fluoride initiates toxicity through a combination of many targets.

Inhibition of metalloproteins by ionized fluoride

Fluoride's properties as a protein inhibitor were established prior to the discovery of its properties in preventing dental caries. Starting in the late 1800s, sodium fluoride was identified as a lipase inhibitor (Tappeiner 1889; Loevenhart and Pierce 1906). Later fluoride was found to inhibit a range of phosphatases, kinases, hydrolases, and other metalloproteins *in vitro*. As of 2018, the protein databank has collected over 700 proteins crystallized in complex with fluoride (RCSB). Over 100 of these have had separate, independent enzymatic studies for fluoride inhibition *in vitro* (Adamek *et al.* 2005). While fluoride has been reported in a few select cases to bind directly to amino acids or to displace essential hydroxides, the majority of noted protein interactions are through either (1) fluoride binding to an essential metal in a metalloprotein's active site, or (2) the complexing of fluoride with metal and acting as a substrate mimic (Edwards *et al.* 1984; Adamek *et al.* 2005; Schenk *et al.* 2008).

An estimated 30–50% of proteins require metal; consequently, there are thousands of potential targets for fluoride inhibition (Ascone *et al.* 2003). Fluoride is negatively charged, and associates with positive sites on proteins. In the case of metalloproteins, fluoride interacts with the essential metals, forming a highly stable, often insoluble complex. Fluoride can also form ternary complex with metal and phosphate, which has even greater stability than metallo-fluoride (Qin *et al.* 2006). Many pathways, particularly glycolysis, nutrient transport, and cellular respiration are inhibited during fluoride exposure, presumably through metalloprotein inhibition (Feig *et al.* 1971; Fina *et al.* 2014; Rogalska *et al.* 2017).

One of the most cited protein targets for fluoride inhibition is enolase. This enzyme catalyzes the penultimate step of glycolysis, converting 3-phosphoglycerate to phosphoenolpyruvate. The ability of fluoride to bind enolase was discovered accidentally by Warburg and Christian (Warburg and Christian 1942). Their laboratory in Germany happened to use water high in fluoride while working to crystallize enolase, and during structural analysis they found fluoride bound to phosphate and magnesium at enolase's active site. Later enzymatic analysis found that enolase has one of the lowest K_D's for fluoride, at 1–10 mM (20–200 ppm) depending on species (Cimasoni 1972; Shahed *et al.* 1980; Maurer and Nowak 1981; Qin *et al.* 2006). Although enolase inhibition is often referenced as the primary target of fluoride, several papers have offered data against this hypothesis. Fluoride-resistant species of bacteria had no significant change in enolase activity or sequence compared to wild type (Van Loveren *et al.* 2008; Mitsuhata *et al.* 2014; Liao *et al.* 2015). RNA-Seq analyses of mammalian and plant cells exposed to fluoride reported no changes in enolase expression (Li *et al.* 2017; Pereira *et al.* 2018). Overexpression of enolase in fluoride-sensitized yeast produced no change in fluoride resistance (Johnston and Strobel 2019). Clinical researchers, using fluoride to inhibit glucose consumption in stored blood, found that metabolic activity decreased only after the induction of stress signaling from high fluoride (Montagnana and Lippi 2017). One

alternative explanation is that metabolism is inhibited during fluoride exposure as a side effect of stress induction, such as oxidative stress or intracellular acidification. In agreement with this hypothesis, several studies found that acidification alone reduced glycolysis to a greater degree than fluoride treatment (Boink *et al.* 1994; Belli *et al.* 1995; Gambino *et al.* 2009). Regardless of the mechanism, glycolysis is consistently inhibited across organism models during fluoride toxicity.

Inhibition of proteins by metallo-fluoride substrate mimics

Fluoride toxicity is greatly enhanced when complexed with metal. Among the most toxic (and most studied) complexes are aluminum- ($AlF_{3,4}$) and beryllium-fluoride ($BeF_{3,4}$). These compounds are isomorphous to phosphate, and consequently able to inhibit phosphoryl-transfer enzymes (Chabre 1990). Nonetheless, the affinity for each metallo-fluoride to various enzymes depends on the pH and interaction with essential positively-charged amino acids (Schlichting and Reinstein 1999; Strunecka *et al.* 2002). Over 100 enzymes have been crystallized with aluminum- or beryllium-fluoride, the majority of which are classified as either ATPases, GTPases, or kinases (Berman *et al.* 2000).

Aluminum- and beryllium-fluoride can also alter the phosphorylation state of various proteins, particularly through GTP mimicry. Many proteins, including G-proteins, actin, and microtubules, are regulated by GTP binding; GDP-bound proteins are in an "off" state, while GTP-bound proteins are "on." In the case of aluminum- or beryllium-fluoride exposure, both metallo-fluorides bind GDP to mimic the bound -phosphate of GTP (Bigay *et al.* 1987; Antonny and Chabre 1991). AlF_x and BeF_x stabilize the transition state for the "on" conformation. Consequently, proteins activated by metallo-fluoride are more stable than those activated naturally, and stay in the "on" conformation (Li 2003). Because of the far-reaching roles of GTPase regulation, their non-selective activation by metallo-fluorides leads to wide dysregulation of functions including cell signaling, transport, and cytoskeleton integrity (Muhlrad *et al.* 1994; Loweth *et al.* 1996; Li 2003; Agalakova and Gusev 2011).

Free fluoride can also activate protein and pathway activity by altering phosphorylation states. Rho GTPase-binding proteins have been shown to bind to GTPases and stabilize the transition state in a fluoride-, but not aluminum-dependent manner (Antonny *et al.* 1993; Vincent *et al.* 1998). Instead, in both cases magnesium was found to be essential, and could theoretically form a magnesium-fluoride phosphate mimic.

V. Fluoride-induced pH and electrolyte imbalance

At both the single- and multi-cellular level, fluoride exposure causes acidification and electrolyte imbalance. The exact mechanism is unknown. Prolonged exposure of vertebrates to high fluoride results in the loss of calcium and magnesium from the plasma, and an excess of potassium (Dalamaga *et al.* 2008). Complementary to this finding, fluoride exposure in single cells results in an influx of calcium and magnesium, and a loss of potassium (Johnston and Strobel 2019). This effect has been proposed to be due to either downstream stress signaling, or the binding of fluoride to metals (Boink *et al.* 1994; Giachini and Pierleoni 2004). Regardless of mechanism, the imbalance of electrolytes in organisms from fluoride exposure has far reaching implication in cell homeostasis and signaling disruption.

Fluoride exposure is also associated with a drop in intracellular pH. Fluoride is a weak acid that enters cells as HF and dissociates, thus releasing one proton per fluoride. Consequently, the more fluoride that enters a cell, the more acidic the cytoplasm becomes. However, fluoride causes intracellular acidification to a larger degree than can be explained by proton shuttling (Kawase and Suzuki 1989; Belli *et al.* 1995; Guha-Chowdhury *et al.* 1996; Marquis *et al.* 2003; Gassowska *et al.* 2013). Many hypotheses have been put forward to explain this observation. Among these, the most common are metabolic disruption, perturbation of the mitochondria, transmembrane protein inhibition, and induction of stress signaling (Belli *et al.* 1995; Marquis *et al.* 2003; Gassowska *et al.* 2013).

Fluoride inhibits metabolism through an unclear mechanism, the downstream effects include reduction in intracellular ATP and damage to the mitochondria. ATP is reduced both in cells containing mitochondria – which display signs of permanent damage and reduced respiration after fluoride exposure – and in erythrocytes (red blood cells), which lack mitochondria and produce ATP solely through anaerobic glycolysis (Feig *et al.* 1971; Agalakova and Gusev 2012; Fina *et al.* 2014). ATP depletion leads to the hydrolysis of ATP into ADP and AMP along with the release of protons, consequently leading to intracellular acidification. Damage to the mitochondria releases free radicals, resulting in oxidative stress. This in turn causes DNA damage, metabolic disruption, ATP hydrolysis, protein inhibition, and intracellular acidification (Boonstra and Post 2004). However, just as free radicals are known to disrupt metabolism, acidify the cell, and activate stress signaling, each of these phenotypes also activate the release of free radicals from the mitochondria (Chen *et al.* 2003; Liu *et al.* 2003; Berezhnov *et al.* 2016). As such, the order in which each known phenomenon occurs is difficult to determine.

Another potential explanation for the drop in pH is the inhibition of transmembrane proton transporters by fluoride. Key among them are the Na^+/H^+ symporters, ATPases, and G-coupled proteins. Fluoride generally inhibits transmembrane proteins by changing the protein's confirmation to its transitional state, such as metallo-fluoride forcing Na^+/K^+-ATPase into its E2P state (Montes *et al.* 2015; Faraj *et al.* 2019). Na^+/H^+ symporters, ATPases, and G-coupled proteins regulate a wide range of cellular processes, including pH homeostasis (Loweth *et al.* 1996; Palmgren *et al.* 2010; Syrovatkina *et al.* 2016). In the case of the Na^+/H^+ symporter, inhibition by fluoride would theoretically accumulate extracellular protons. However, the opposite occurs (Kawase and Suzuki 1989). Kawase and Suzuki proposed that this extracellular proton buildup would result in increased fluoride protonation, which would therefore increase overall fluoride toxicity and stress signaling. It could also be that while this particular protein inhibition raises extracellular pH, the overall inhibition of transmembrane transporters leads to a net accumulation of protons in the cytoplasm.

VI. Organelle disruption by fluoride

Prolonged exposure to high levels of fluoride leads to widespread organelle damage. This damage is both time- and concentration-dependent. While the sensitivity of each organelle to fluoride varies slightly by organism, in general, fluoride disrupts the cell surface, mitochondria, endoplasmic reticulum, Golgi, and nucleus (Fig. 3).

Fluoride can irreversibly damage the cell surface. Organisms that have calcium-pectate in their cell walls, such as plants, are prone to calcium depletion upon fluoride exposure (Tsunoda and Yu 1985). The plasma membrane during fluoride exposure is prone to lipid peroxidation and cytoskeleton rearrangement (Wang *et al.* 2004; Liang *et al.* 2015; Chen *et al.* 2017). This effect is not specific to fluoride; many stressors are known to induce membrane remodeling (Farah *et al.* 2011; Brandao *et al.* 2014; Westman *et al.* 2019). While most membrane changes are attributed to either oxidative stress, apoptotic signaling or lipid peroxidation, free or metallo-fluoride can also directly bind actin and change its polymerization (Combeau and Carlier 1989; Allen *et al.* 1996). Regardless of membrane damage, cell cycle arrest during fluoride exposure is enhanced with an increased uptake in fluoride. This suggests that fluoride's principal mechanism of toxicity at high doses is intracellular (Ji *et al.* 2014).

The organelle most widely reported to be inhibited by fluoride is the mitochondria. Many mitochondrial proteins are metalloproteins and have been linked to fluoride inhibition, including respiratory complexes I-IV and F-ATPase (Batenburg and van den Bergh 1972; Sutton *et al.* 1987; Fina *et al.* 2014; Zhao *et al.* 2019). Either directly or indirectly, fluoride exposure damages mitochondrial membrane integrity (Yan *et al.* 2015). This reduces the overall activity of the mitochondria, inhibits cellular respiration, and triggers leakage of free radicals and cytochrome c, ultimately inducing oxidative stress (Miller and Miller 1974; Anuradha *et al.* 2001; Jothiramajayam *et al.* 2014). Not surprisingly, the addition of antioxidants partially rescues from fluoride toxicity (Basha and Sujitha 2011). The last few years have seen renewed interest in mitochondrial inactivation under fluoride exposure as the primary mechanism of fluoride toxicity, as much of the known adverse effects can be attributed to free radicals (Farrugia and Balzan 2012; Yan *et al.* 2015; Lu *et al.* 2017).

Fluoride-exposed cells undergo intensive DNA damage, presumably through free radical oxidation. Fluoride triggers both single- and double-stranded DNA damage following oxidative stress (Podder *et al.* 2015). Due to DNA damage, high fluoride eventually leads to S-phase cell cycle arrest (Wang *et al.* 2004). Typically, studies that expose mice or rats to high concentrations of fluoride for weeks to months report DNA damage. The necessity of a long incubation time suggests that DNA damage is a downstream stress effect rather than a direct target of fluoride.

The endoplasmic reticulum (ER) activates stress signaling under fluoride exposure. The mechanism by which fluoride causes ER stress is unclear. Prolonged fluoride exposure damages the ER membrane, as well as induces the unfolded protein response pathway (Matsuo *et al.* 2000; Kubota *et al.* 2005). Fluoride also disrupts Ca^{2+} homeostasis in the ER (Borke and Whitford 1999; Zhang *et al.* 2016). These phenotypes could either be through the direct scavenging of Ca^{2+} by fluoride, the interaction of fluoride with ER proteins such as Ca^{2+}-ATPase, or the oxidative stress signaling pathway, which involves the release of Ca^{2+} into the cytoplasm (Murphy and Coll 1992; Borke and Whitford 1999; Ermak and Davies 2002). In support of the hypothesis that ER stress is linked with oxidative stress, the addition of ER stress inhibitors reduced DNA damage upon fluoride exposure (Kubota *et al.* 2005).

Fluoride inhibits Golgi stacking, although this effect is reversible. Aluminum-fluoride (AlF_4^-) has been shown to activate G-proteins by mimicking the gamma-phosphate of GTP (Bigay *et al.* 1987). As G-proteins function in signal transduction, their activation has far-reaching implications, especially in vesicle-mediated exocytosis and Golgi function. Activation of plasma membrane G-proteins by aluminum-fluoride leads to a calcium-dependent activation of exocytosis (Elferink *et al.* 1980). G-proteins are also found on the Golgi, and are essential for Golgi stacking (van Hook 2015). The Rothman lab demonstrated that addition of aluminum-fluoride to a cell-free system inhibited protein transport between Golgi stacks (Melancon *et al.* 1987). Aluminum-fluoride also alters the protein coating assembly along the Golgi (Finazzi *et al.* 1994; Tomas *et al.* 2010). Later studies in neuroendocrine cells found that free fluoride also inhibited Golgi stacking by interfering with matrix and cisternae protein assembly, and that this effect was reversed within two hours of recovery (Back *et al.* 2004). However, the fluoride may still have been in a metal complex; a study by the Lowe lab demonstrated that in the absence of aluminum, fluoride will combine with magnesium to bind G-proteins (Graham *et al.* 1999). As a whole, fluoride reversibly inhibits Golgi function, although this might only happen when fluoride is complexed with metal.

Fluoride exposure inhibits protein synthesis, an effect linked to both stress-signaling and ribosome inhibition. Stress pathways, particularly oxidative stress, are known to inhibit protein synthesis (Sheikh and Fornace 1999; Liu and Qian 2014). This effect also occurs during ATP depletion, a well-established fluoride effect (Freudenberg and Mager 1971). Unsurprisingly then, fluoride exposure has been shown to reduce the turnover of new proteins (Hongslo and Holland 1979, and Sharma *et al.* 2008). Additionally, fluoride exposure *in vitro* causes the reversible inhibition of polyribosome formation, and "NaF ribosomes": ribosomes thought to either contain an extra 40S subunit, or a deacetylated tRNA^Met bound to the 40S complex (making a 43S subunit) (Ravel *et al.* 1966; Bishop 1968; Culp *et al.* 1970; Sameshima *et al.* 1972; Godchaux III and Atwood IV 1975; Holland 1979). Overall, the exposure to fluoride results in inhibition of polypeptide chain initiation (Ravel *et al.* 1966; Mosteller *et al.* 1967; O'Rourke and Godchaux III 1975). The research on fluoride's interaction with ribosomes effect was mostly conducted in the 1960–1970s; consequently, the direct action of fluoride on ribosomes has been largely forgotten. Fluoride has been widely recorded to halt protein translation, but this effect is generally linked to apoptotic signaling. Recent RNA-Seq data have reported strongly repressed ribosomal subunit expression during fluoride exposure, indicating that this underappreciated phenotype may play a role in fluoride toxicity (Melo *et al.* 2017; Li *et al.* 2017; Johnston and Strobel 2019).

VII.   Fluoride exposure and toxicity in vertebrates

Most vertebrates are exposed to fluoride through their diet. Fluoride is present in food and water, where after ingestion fluoride passes along the gastrointestinal tract and into the plasma. Over time, fluoride accumulates in soft tissue such as the spleen, kidney, and especially the bone, leading to potential chronic toxicity.

Acute fluoride toxicity has only been reported in individuals exposed to very high concentrations of fluoride. Typically, symptoms include nausea, diarrhea, headaches, and gastric pain after ingestion of 5–8 mg/kg bodyweight (Ullah *et al.* 2017). Chronic exposure to high fluoride has been correlated with inflammatory bowel disease, disruption of tissue lining, and nerve damage along the gastrointestinal tract (Das *et al.* 1994; Follin-Arbelet and Moum 2016; de Oliveira *et al.* 2017; Melo *et al.* 2017).

Exposure of tissues to fluoride depends on the surrounding pH. Protonated fluoride (HF) readily passes through biological membranes, and can spread to tissues far from the gastrointestinal tract. Consequently, organisms in contact with HF gas can experience cellular toxicity in tissues far from the original area of exposure (Sheridan *et al.* 1995). The mouth, esophagus, and upper stomach are at neutral pH (6.5–7.5, 5.0–7.0 and 4.0–6.5, respectively) (Fallingborg 1999; Tutuian and Castell 2006; Baliga *et al.* 2013). However, the lower stomach is highly acidic (pH 1.5–3.5). Given that fluoride has a pK_a of 3.2, 67–96% of fluoride forms HF and can dissociate in equilibrium across the stomach lining. Nonetheless, only 25% of ingested fluoride is thought to be absorbed by the stomach and pass on to other cells; the rest of the fluoride is moved to the intestines (Kanduti *et al.* 2016). The majority of ingested fluoride is absorbed by the small intestine. Just ~10% of consumed fluoride is never absorbed by the body and excreted through feces (Whitford 1994; Buzalaf *et al.* 2015). Fluoride has been hypothesized to pass through the small intestines as its anion form F^- (Nopakun and Messer 1990). Because of fluoride's reactivity, the nutrients ingested along with fluoride, especially those rich in calcium, influence the amount of fluoride absorbed by the intestines (Whitford 1994). Once absorbed, fluoride travels throughout the body via the blood stream, before being filtered by the kidney and excreted in urine. Approximately 45–60% of ingested fluoride is excreted in urine, with the rest re-circulated into the plasma or deposited into the bone (Buzalaf *et al.* 2015).

Dental fluorosis

The risk for developing adverse symptoms from fluoride exposure is dose dependent. Typically, the concentration of fluoride in water is regulated to be between 0.7–1.2 ppm (40–60 μM) (USDHHS 2015). Chronic toxicity to low doses of fluoride occurs after prolonged exposure to >1.5 ppm (75 μM) fluoride. This toxicity, known as dental fluorosis, is characterized by mottled and discolored teeth. Fluorosis generally correlates with a high degree of fluorapatite formation. Fluorapatite is mechanically weaker than hydroxyapatite, and the replacement of natural enamel with fluorapatite increases the brittleness of teeth (DenBesten and Li 2011). Fluoride most readily forms fluorapatite as fresh enamel is developing; as such, vertebrates are most sensitive to dental fluorosis during tooth development. For humans, juveniles up to 8 years of age (with particular sensitivity during the first 2 years) are susceptible to fluorosis (Hong *et al.* 2006; Bhagavatula *et al.* 2016).

Ameloblast toxicity

Although it is well established that fluoride triggers fluorapatite formation in teeth and that excess fluoride leads to fluorosis, the full mechanism of dental fluorosis is unknown. Several mechanisms besides remineralization have been put forward, most notably that fluorosis is caused by ameloblast damage (Fejerskov *et al.* 1977). Ameloblasts, or cells that function in

depositing enamel, are present in the gums of organisms developing new teeth. Fluoride exposure to ameloblasts in cell culture causes endoplasmic reticulum stress, DNA fragmentation, cytoskeleton defects, protein synthesis inhibition, reduced secretion of enamel matrix proteins, and eventual apoptosis (Kubota *et al.* 2005; Li *et al.* 2005; Hassanuma *et al.* 2007; Bronckers *et al.* 2009). However, it is important to note that these studies are typically conducted at much higher fluoride (100–150 ppm) than levels causing dental fluorosis in humans (>1.5 ppm).

Skeletal fluorosis

Skeletal fluorosis is the most severe form of chronic fluoride toxicity. The underlying mechanics of skeletal fluorosis are similar to dental fluorosis: high fluoride exposure leads to a change in mineral formation, as well as stress to surrounding cells. The uptake of fluoride into bone results in the conversion of the bone mineral hydroxyapatite into fluorapatite, which alters the general bone lattice and reduces its overall strength (Grynpas 1990). As such, vertebrates exposed to high fluoride have mechanically weaker bones (Evans and Wood 1976; Sogaard *et al.* 1995). This increased brittleness is associated with skeletal dysmorphia and higher risk of fracture.

The onset of skeletal fluorosis in humans varies by individual. The reported dosage required for developing skeletal fluorosis varies widely, and is affected by age, metabolic rate, genetic disposition, and overall health of the individual (Marier *et al.* 1963; Krishnamachari 1986; USDHHS 1991; Pramanik and Saha 2017). It is estimated that for the average person, 6–10 mg per day of fluoride ingestion for at least 10 years leads to skeletal fluorosis (Whitford 1996). Children are the most susceptible, and accumulate fluoride at a greatly accelerated rate compared with adults (Teotia *et al.* 1998). Patients with kidney disorders are also at high risk for skeletal fluorosis, given that kidneys are essential for filtering fluoride (Gerster *et al.* 1983; Krishnamachari 1986; Bansal and Tiwari 2006).

Skeletal fluorosis affects surrounding osteoblasts. Low fluoride concentrations (5–1000 μM, or 0.1–20 ppm *in vitro*), stimulates osteoblasts (Hall 1987; Lau and Baylink 1998; Qu *et al.* 2008). The exact mechanism by which fluoride causes osteoblast proliferation is unclear, although it has been shown that fluoride causes an uptake in phosphate, increased alkaline phosphatase activity, decreased acid phosphatase activity, and an activation of the MAPK pathway (Farley *et al.* 1983; Lau *et al.* 1989; Selz *et al.* 1991; Lau *et al.* 2002). The stimulation of osteoblasts triggers new bone formation and an increase in overall bone density in patients with over 0.2% fluoride content in bones (Aaron *et al.* 1991). However, the bone matrix is disrupted by fluoride, and the improved bone mass is of lower quality and more likely to fracture. Fluoride above 1 mM (20 ppm) results in osteoblast cytotoxicity, particularly to the nucleus and endoplasmic reticulum (Qu *et al* 2008; Zhou *et al.* 2013; Liu *et al.* 2015). As such, there is an overall bimodal trend of low fluoride stimulating bone mass, while high fluoride decreases bone mass.

Neurotoxicity

There are many claims as to fluoride's long-term toxicity in the human body that lack thorough scientific studies, but hold a sympathetic belief in the public. Probably the most

well-known is the concern that fluoride exposure leads to neuronal damage, including Parkinson's disease, Alzheimer's disease, and a reduced IQ. While there have been several studies to explore this connection, the evidence is insufficient or even anti-correlative.

For a molecule to cause neurotoxicity, it must pass through the blood brain barrier. The blood brain barrier only allows passive diffusion of small, uncharged, lipid soluble compounds, or molecules that can pass through selective channels, pumps, or vesicles (Wong *et al.* 2013). In the case of fluoride, the ion is highly electronegative and would only pass through the blood brain barrier as HF. As blood has a pH between 7.3–7.45, the majority of fluoride would remain in its unprotonated form (Kellum 2000). Consequently, the levels of fluoride detected in brain tissue are typically much lower than the concentration of fluoride in serum, and generally lower than all other tissues in the body (Whitford 1996).

Many studies have found correlations between low IQ and high fluoride exposure. Over 60 studies have been conducted in areas with high fluoridated groundwater, and most of these studies have reported a lower average IQ in the children of those regions compared with children in areas with normal fluoride exposure (Tang *et al.* 2008; Aravind *et al.* 2016; Green *et al.* 2019). Several other studies have been conducted and found either no correlation between fluoride and IQ, or that high fluoride correlated with higher IQs (Spittle *et al.* 1998; He *et al.* 2010; Li *et al.* 2010; Soto-Barreras *et al.* 2019). Each of these studies do not fully take into consideration socioeconomic factors, unconscious biases, as well as other possible toxicants in groundwater that could be affecting IQ. The areas of interest – primarily China, India, and Mexico – frequently have groundwater high in other neurotoxicants such as arsenic and mercury (Wang *et al.* 2007; UNEP 2013). In a systematic review of fluoride-IQ studies, the Grandjean lab concluded that if fluoride is a neurotoxicant, it would be over 1,000-times less potent than other known neurotoxicants (Choi *et al.* 2012). Because of each of these factors, the correlation of high fluoride exposure and lower IQ does not necessarily imply causation.

In a laboratory setting, high doses of fluoride are toxic to neuronal cells. Neuronal cell lines exposed to 60 ppm (3 mM) NaF undergo DNA damage, oxidative stress, mitochondrial agglutination, and cytoskeleton damage (Zhang *et al.* 2008; Chen *et al.* 2017; Tu *et al.* 2018). Because the primary function of neurons is synaptic signaling, membrane defects from fluoride exposure reduces the overall activity of neurons. Rats and mice exposed to high fluoride (50 ppm, or 2.5 mM) showed decreased nicotinic acetylcholine receptor expression, lowered acetyl cholinesterase activity, and damaged myelin and microtubules (Long *et al.* 2002; Basha and Sujitha 2012; Niu *et al.* 2015; 2018). Nonetheless, tests done with 50 ppm (2.5 mM) fluoride showed no significant brain damage (Varner *et al.* 1998; Shivarajashankara *et al.* 2002).

Several studies have argued that fluoride damages neuronal cells through an indirect mechanism. The Rigalli lab suggested that the alteration of glucose and insulin homeostasis in the blood during fluoride exposure could lead to nutrient depletion and downstream stress in the nervous system (Lombarte *et al.* 2016). However, the Nowak lab reported an enhancement in glucose uptake in brain tissue upon treatment of rats with 50 ppm (2.5 mM) fluoride (Rogalska *et al.* 2017). Others have argued that the release of free radicals by either

the intake of 60 ppm (3 mM) NaF, or the intragastric injection of 20 mg/kg body weight NaF – about four times more concentrated than that needed to cause acute gastrointestinal toxicity in humans – could have negative effects on neuronal tissue, particularly the hippocampus (Bhatnagar *et al.* 2002; Pan *et al.* 2015; Shanmugam *et al.* 2018). Again, these studies rely on exposure to high doses of fluoride, well above that typically found in fluoridated water.

While free fluoride is unlikely to pass through the blood brain barrier, metallo-fluoride, specifically aluminumfluoride, can cause neurotoxicity. Aluminum fluoride (AlF_3 or AlF_4) are phosphate mimics, and could theoretically cross the blood brain barrier through phosphate transporters (Strunecka *et al.* 2002). Rats exposed to 10 ppm NaF and 100 ppm $AlCl_3$ in combination for 30 days showed neuronal shrinkage and inhibition of acetylcholinesterase activity (Akinrinade *et al.* 2015). Furthermore, aluminum fluoride was found to cause more histopathological changes to brain tissue than sodium fluoride alone, particularly in the neocortex and hippocampus (Varner *et al.* 1998; NRC 2006). However, it is also possible that the neurotoxicity is due to free aluminum. Several studies of brain defects found the highest association with aluminum exposure, not fluoride levels (Forbes *et al.* 1991; Kraus and Dorbes 1992; Jacqmin *et al.* 1994). Free aluminum acts as a neurotoxicant, disrupting the cell membrane integrity of the blood brain barrier, activating the innate immune response, and potentially increasing dementia (Banks and Kastin 1989; Armstrong *et al.* 1996). As such, aluminum could be at least partially responsible for fluoride neurotoxicity.

Diabetes

Given that fluoride alters cellular metabolism, there have been concerns about the effect of fluoride on sugar homeostasis and diabetes. Prolonged exposure to high fluoride inhibits glycolysis and ATP production. Cells respond to fluoride stress, as well as to many acids, by increasing glucose uptake (Hay and Paul 1967; Rogalska *et al.* 2017). The direct intraperitoneal injection of mammals with high fluoride, resulting in at least 0.1 mg/L total fluoride in the blood, can lead to higher blood glucose (McGown and Suttie 1977; Suketa *et al.* 1985; NRC 2006). However, studies with rats fed 15–50 ppm (0.8–3 mM) fluoride showed either no change, or a decrease in serum glucose levels (Lupo *et al.* 2011; Lobo *et al.* 2015; Malvezzi *et al.* 2019).

Fluoride exposure has also been demonstrated to alter insulin concentrations in the blood. Fluoride reversibly inhibits insulin secretion, leading to an overall reduced insulin concentration in the serum (Rigalli *et al.* 1990; Menoyo *et al.* 2005). However, exposure to low (10 ppm, or 0.5 mM) doses of fluoride enhances insulin sensitivity (Lobo *et al.* 2015). In all, chronic exposure to high fluoride may partially contribute to diabetes, while low fluoride exposure may be protective against diabetes.

Endocrine disruption

Fluoride was officially classified by the National Research Council in 2006 as an endocrine disruptor for its ability to inhibit the thyroid at high concentrations. However, its mechanism of action remains unknown (NRC 2006). Exposure to fluoride can lead to a decrease in the

thyroid hormones triiodothyronine (T3) and thyroxine (T4). Studies that have reported a significant decrease of T3 and T4 by fluoride typically involve either patients with dental fluorosis, or mammals exposed to 30–80 ppm (2–4 mM) fluoride for   2 months (Bobek *et al.* 1976; Jiang *et al.* 2015). On a cellular level, the thyroid undergoes DNA damage, membrane disruption, mitochondrial and endoplasmic reticulum stress, and oxidative stress signaling during fluoride exposure (Sundström 1971; Liu *et al.* 2016, Abdelaleem *et al.* 2018). Cells exposed to both excessive fluoride and iodide showed synergistic triggering of endoplasmic stress, IRE1 signaling, and DNA damage (Liu *et al.* 2014; Jiang *et al.* 2015; Liu *et al.* 2016). The resulting hypothyroidism alters the body's ability to regulate temperature, metabolism, and heart rate, which has far-reaching implications for patients with severe fluorosis. Given that fluoride does not accumulate in the thyroid, this toxicity is attributed to high fluoride in the blood (Galetti and Joyet 1958; Singh *et al.* 2014).

In 2015, the public nominated fluoride for investigation by the United States National Toxicology Program (NTP) for its potential role in non-thyroidal endocrine disruption, cancer, and neurological disorders. After conducting a systematic review, the NTP cited insufficient research on low doses of fluoride to investigate the real risk of these effects (NTP 2017). Research conducted with high dosages of fluoride reports direct and indirect inhibitory effects on the endocrine system, independent of the thyroid. The most notable targets are the pineal gland, adrenal glands, and the parathyroid. The pineal gland lies outside of the blood brain barrier and has the highest calcification rate of any organ in the body (Tan *et al.* 2018). Fluoride has been shown to gradually accumulate in the pineal gland along with calcium (Luke 2001; Kalisinska *et al.* 2014). Few studies have been conducted on whether fluoride accumulation significantly disrupts pineal gland activity. The key study, conducted in 1997, reported that prepubescent gerbils fed 40 ppm (2 mM) NaF daily had lower melatonin production by the pineal gland than the controls, but that melatonin was restored to normal concentrations when the gerbils reached adulthood (Luke 1997). Patients with skeletal fluorosis (the most severe form of fluoride toxicity), are sometimes found to have secondary hyperparathyroidism (Teotia and Teotia 1973). This has been suggested as the body's attempt to restore calcium and phosphate homeostasis, rather than the direct disruption of the parathyroid by fluoride (Faccini 1969; Krishnamachari 1986).

VIII.   Resistance to fluoride

Many organisms have evolved defense mechanisms against fluoride. The first discovery of a fluoride-specific defense pathway occurred in 2012, when the Breaker lab identified a region of RNA, known as a riboswitch, conserved in many bacteria and archaea (Baker *et al.* 2012). Riboswitches are located on some mRNA and control downstream gene expression upon binding to ligands. In the case of the fluoride riboswitch, the RNA coordinates with three magnesium ions to bind fluoride at a K_D of 60 uM (1.1 ppm) (Ren *et al.* 2012). Over 2,000 examples of the fluoride riboswitch were identified, and found to control expression of many genes linked to fluoride resistance (Baker *et al.* 2012). Included in this list were genes functioning in oxidative stress, DNA repair, and intracellular acidification. There were also genes for proteins known to be inhibited by fluoride, including enolase, Na^+/H^+ antiporters, and pyrophosphatase. Thirdly, there were two newly discovered fluoride channels: EriC^F and Fluc.

The transporters EriC[F] and Fluc confer significant fluoride resistance. EriC[F] is a member of the Clc family of membrane proteins, and acts as a F^-/H^+ antiporter (Lim *et al.* 2013). Fluc is believed to be a channel whose driving force for fluoride efflux is the electrochemical gradient of the bacterial plasma membrane (Stockbridge *et al.* 2013; Ji *et al.* 2014). Fluoride exporters are an essential part in mediating fluoride's toxic effects, conferring a 200-fold increase in resistance to bacterial growth arrest. In yeast this resistance is even more pronounced, with over 1000-fold increased resistance to fluoride (Li *et al.* 2013). While the fluoride transporter is conserved across many species of eukaryotes and prokaryotes, no homolog has been identified in vertebrates. Nonetheless, fluoride sensitivity varies across species and tissue-type, suggesting there is an as yet undiscovered mechanism of defense.

Microbes have evolved multiple mechanisms of fluoride resistance. These resistance factors are generally found by either isolating organisms from areas with high fluoride, or exposing cells to fluoride in a laboratory. Several studies identified fluoride-resistant bacteria that express higher copies of fluoride transporters, as well as higher copies of known fluoride targets (Liao *et al.* 2015; 2016; Liu *et al.* 2017). In one of the few reports that did not find an altered fluoride channel, fluoride-resistant *S. mutans* adjusted their composition of fatty acids and had enhanced general acid resistance (Zhu *et al.* 2012). This correlation of fluoride-resistance with acid resistance has been widely observed, although the mechanism has never been found (Sheng and Liu 2000; Marquis *et al.* 2003). A DNA microarray of a resistant strain of *A. ferrooxidans* offered one of the most complete pictures of fluoride tolerance, showing a change in expression of genes related to metabolism, protein synthesis, and cell membrane maintenance (Ma *et al.* 2016). As a whole, microbes appear to gain fluoride resistance by increasing expression of fluoride transporters and protein targets of fluoride inhibition, most notably ATPases and glycolytic enzymes (Fig. 4).

Mammalian cells are also capable of gaining fluoride resistance. A study comparing mice that were either resistant or sensitive to developing fluorosis showed differences on chromosomes 2 and 11, although these differences were comprised of nearly 2,000 genes (Everett *et al.* 2009). An RNA-Seq study investigating gene expression differences in a fluoride-resistant mouse adipose cell line showed increased expression in genes related to general stress response, protein synthesis, and cell membrane maintenance (Ran *et al.* 2017). Studies on mammalian resistance to fluoride have not yet found a link with overexpressing glycolytic enzymes, although RNA-Seq of rats after 20–60 days exposure to 50 ppm (3 mM) NaF showed an increased expression of genes related to glucose uptake (Pereira *et al.* 2018). Continued investigation into fluoride toxicity and the concurrent mechanism of resistance across species is a much-needed avenue for fully understanding the biological effects of fluoride exposure.

CONCLUDING REMARKS

The issue of whether fluoride is safe depends on the sensitivity of the organism, the concentration of fluoride, and the conditions by which fluoride is administered. Fluoride has both positive and negative effects. Low fluoride levels decrease cavities and partially restore the minerals in teeth. High levels of fluoride lead to protein inhibition, a release of free

radicals, disruption of metal homeostasis, and tissue damage. The question then becomes how much fluoride an organism will encounter during their lifetime.

Most organisms are in regions with low to mid-range fluoride, and are at low risk to experience fluoride toxicity. Consequently, the majority of the world's population have no visible signs of fluorosis. However, the inhabitants of certain regions around the world, including in India, China, and Africa, have to be particularly aware of the fluoride levels to which they are exposed. In these areas, there is emerging interest on the safe removal of fluoride from the groundwater and air.

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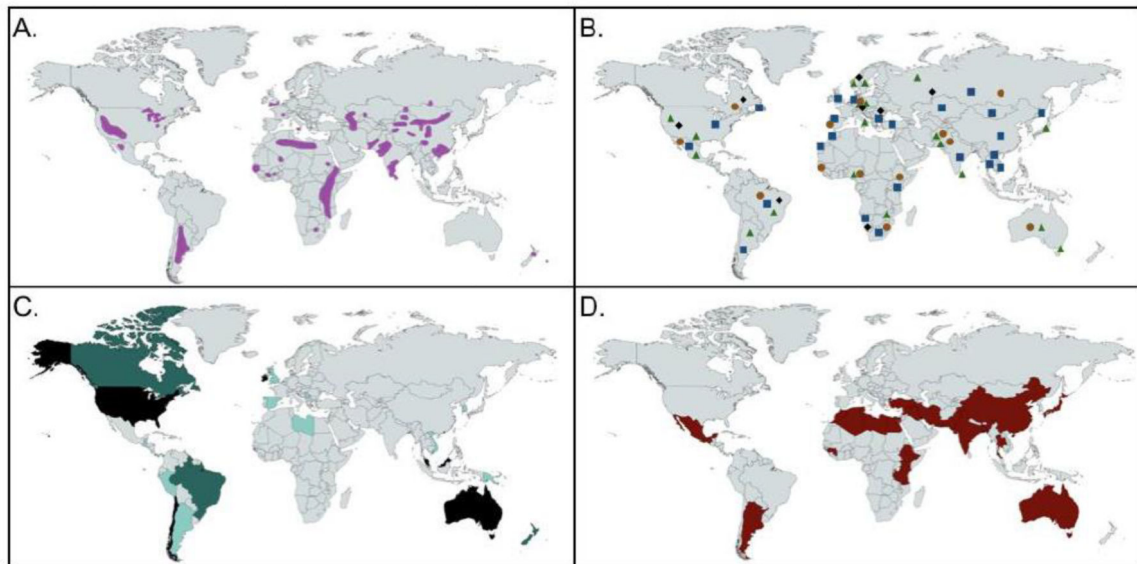
Figure 1:

Global distribution of fluoride. Shown here are areas with (A) <1.5 ppm fluoride in the groundwater (purple), (B) mining of fluoro-compounds (blue squares -fluorspar, brown circles – fluorapatite, green triangles - topaz, and black diamonds - cryolite), (C) percent of the population given government-regulated fluoridated water (light teal – 0–33%, teal – 34–67%, black – 68–100% of population), and (D) endemic fluoride toxicity (red) [Unicef 1999, WHO 2004, Gupta and Ayoob 2016, British Geological Survey, and the USGS Database].
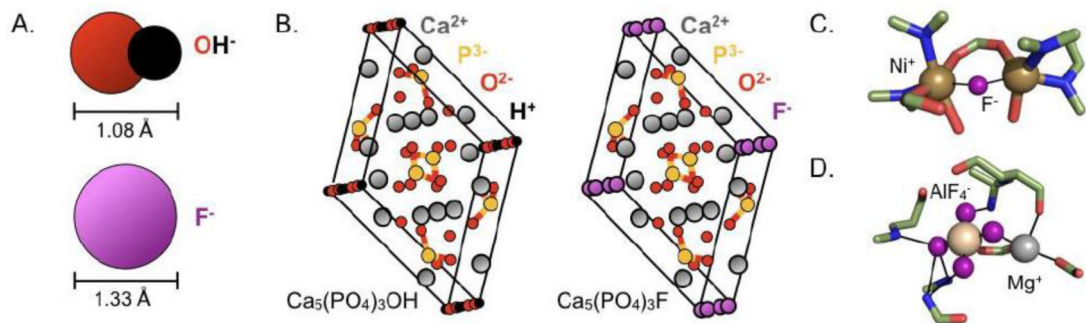
Figure 2:
Fluoride interactions *in vitro.* (A) Fluoride is the most similar in terms of size and charge to hydroxide, but has a much higher affinity for metals. (B) unit cells of crystallized hydroxyapatite (Ca5(PO4)3OH) and fluorapatite (Ca5(PO4)3F) [Minerology Database]. (C) Crystal structures of fluoride bound to urease and (D) phosphoserine phosphatase. Structures were generated on PyMOL using RCSB PDB (C) 4GOA and (D) 1L7N.
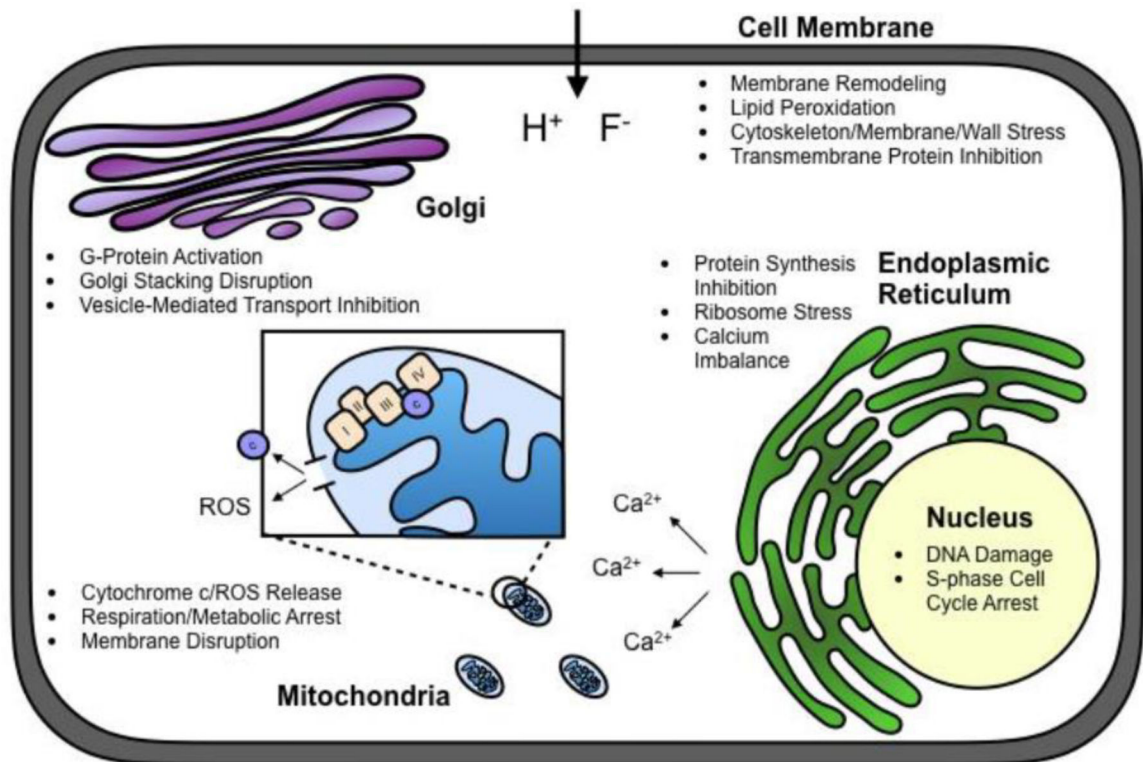
Figure 3:
Intracellular fluoride toxicity. General scheme of downstream organelle damage after prolonged exposure to high fluoride, conserved across eukaryotes.
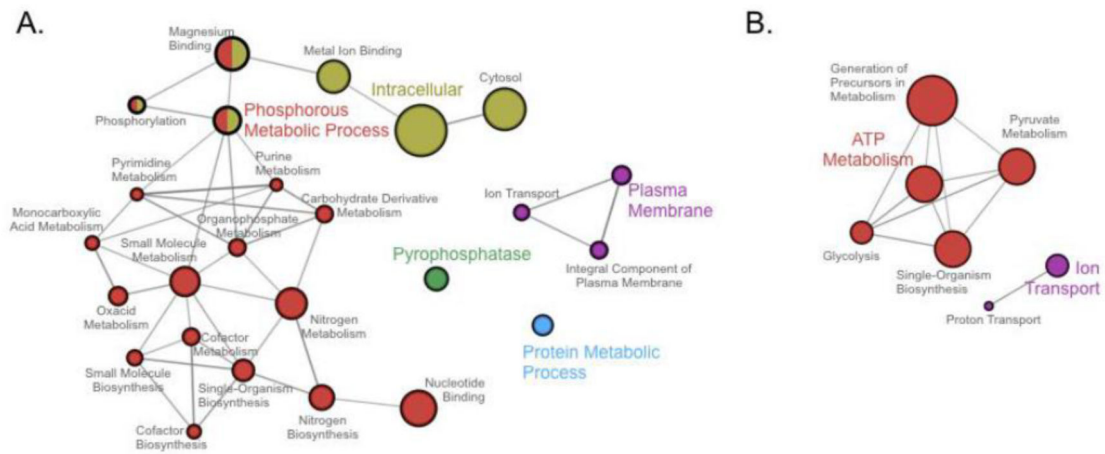
Figure 4:
Global network of cellular processes involved in the resistance of bacteria to fluoride. (A) Conserved molecular functions and cellular components of (A) genes regulated by the fluoride riboswitch, as reported by Weinberg *et al.* 2010, and (B) altered genes in fluoride resistant bacteria (Zhu *et al.* 2012; Liao *et al.* 2015; 2016; Ma *et al.* 2016; Liu *et al.* 2017). Genes were converted to their *E. coli* homologs, and duplicates were discarded. Data was analyzed on Cytoscape using ClueGO. Node size corresponds with the number of genes per category, and similar colored nodes denote a similar cluster in function.