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## Prenatal Exposure to Per- and Polyfluoroalkyl Substances and Maternal and Neonatal Thyroid Function in the Project Viva Cohort: a mixtures approach

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SKS, TFW, EO, ENP, and SLRS carried out conceptualization of the study; EVP, TFW, BCH, MDM, and CG designed the statistical analyses/models; EVP implemented the formal analysis; EO, ENP, AMC, and SLRS were responsible for data generation, curation, and review; ENP and AFF provided guidance on outcome measures and interpretation; all authors were involved in the investigation throughout the study; EVP wrote the manuscript and all authors reviewed and revised the manuscript.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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## Abstract

**Background**—Maternal and neonatal thyroid function is critical for growth and neurodevelopment. Exposure to individual per- and polyfluoroalkyl substances (PFAS) can alter circulating thyroid hormone levels, but few studies have investigated effects of combined exposure to multiple PFAS.

**Objectives**—Estimate associations of exposure to multiple PFAS during early pregnancy with maternal and neonatal thyroid function.

**Methods**—The study population consisted of 726 mothers and 465 neonates from Project Viva, a Boston, Massachusetts area longitudinal pre-birth cohort. We measured six PFAS [perfluorooctanoate (PFOA), perfluorooctane sulfonate (PFOS), perfluorononanoate (PFNA), perfluorohexane sulfonate (PFHxS), 2-(N-ethyl-perfluorooctane sulfonamido)acetate (EtFOSAA), and 2-(N-methyl-perfluorooctane sulfonamido)acetate (MeFOSAA)] and thyroxine (T<sub>4</sub>), Free T<sub>4</sub> Index (FT<sub>4</sub>I), and thyroid stimulating hormone (TSH) in maternal plasma samples collected during early pregnancy, and neonatal T<sub>4</sub> in post-partum heel sticks. We estimated individual and joint effects of PFAS exposure with thyroid hormone levels using weighted quantile sum (WQS) regression and Bayesian kernel machine regression (BKMR), and evaluated potential non-linearity and interactions among PFAS using BKMR.

**Results**—Higher concentrations of the PFAS mixture were associated with significantly lower maternal FT<sub>4</sub>I, with MeFOSAA, EtFOSAA, PFOA, and PFHxS contributing most to the overall mixture effect in BKMR and WQS regression. In infants, higher concentrations of the PFAS mixture were associated with lower T<sub>4</sub> levels, primarily in males, with PFHxS and MeFOSAA contributing most in WQS, and PFHxS contributing most in BKMR. The PFAS mixture was not associated with maternal T<sub>4</sub> or TSH levels. However, in maternal BKMR analyses, ln-PFOS was positively associated with T<sub>4</sub> levels ( 25<sup>th</sup> to 75<sup>th</sup> percentile:0.21 µg/dL; 95% credible interval: -0.03, 0.47) and ln-PFHxS was associated with a non-linear effect on TSH levels.

**Conclusions**—These findings support the hypothesis that there may be combined effects of prenatal exposure to multiple PFAS on maternal and neonatal thyroid function, but the direction and magnitude of these effects may vary across individual PFAS.

## Keywords

PFAS; thyroid; pregnancy; endocrine disrupting chemicals; chemical mixtures

## 1. INTRODUCTION

Proper maternal and neonatal thyroid function is critical for growth and neurodevelopment. During early pregnancy, the fetus is completely dependent on maternal thyroid hormones until fetal thyroid hormone production increases around 18 to 20 weeks gestation (Fisher 1997). After birth, normal thyroid function is necessary to regulate continued growth and neurodevelopment during childhood, as well as metabolism and other major bodily systems throughout adulthood (Miller et al. 2009). Alterations in maternal thyroid function during pregnancy are associated with adverse fetal outcomes including preterm delivery, insufficient fetal growth, and neurodevelopmental deficits (de Escobar et al. 2004).

Additionally, neonatal thyroid dysfunction has been associated with impaired cognition and neurodevelopment (Lyall et al. 2016; Rose et al. 2006; Simic et al. 2009a).

Recent epidemiological studies suggest that prenatal exposure to per- and polyfluoroalkyl substances (PFAS) may alter maternal and neonatal thyroid function (Ballesteros et al. 2016). PFAS are a group of fluorinated synthetic compounds commonly used in consumer and industrial products such as stain-resistant and non-stick coatings, firefighting foams, food packaging, upholstery, and carpeting (Lindstrom et al. 2011). Because of the strength of the carbon-fluorine bond, PFAS are environmentally persistent, and many also have relatively long elimination half-lives in humans (Olsen et al. 2007). Exposure to PFAS contaminated drinking water is an emerging issue of concern, particularly in areas near industrial contamination or use of aqueous film forming foam (Barton et al. 2019; Daly et al. 2018; Hu et al. 2019; Sunderland et al. 2019). In addition, substantial exposure of the general population to PFAS also occurs through diet and the indoor environment (Gebink et al. 2015; Hu et al. 2016; Makey et al. 2017; Sunderland et al. 2019). Widespread exposure to PFAS has resulted in ubiquitous detection of PFAS in human serum in the U.S. general population (Calafat et al. 2007).

Toxicology studies have demonstrated that PFAS can disrupt thyroid function in animals, often leading to a hypothyroid-like effect, with reduced levels of circulating total thyroxine ( $T_4$ ) or free  $T_4$ , with or without an increase in thyroid stimulating hormone (TSH) (Boas et al. 2012; Zoeller 2010). Previous epidemiologic studies assessing associations between PFAS exposure and thyroid hormone levels in pregnant women and neonates have been inconsistent, but have generally found associations between higher PFAS serum or plasma concentrations and lower free or total  $T_4$  levels and/or increased TSH levels (Ballesteros et al. 2016). In our previous work in the Project Viva cohort located in the Boston, Massachusetts area, we reported that prenatal plasma concentrations of some PFAS were inversely associated with maternal Free  $T_4$  Index (FT<sub>4</sub>I) levels and neonatal  $T_4$  levels, when assessing PFAS individually (Preston et al. 2018). However, the majority of previous epidemiologic studies, including our own, have assessed each PFAS-thyroid hormone association individually, without evaluating possible combined effects of exposure to multiple PFAS.

There is growing recognition of the need to investigate health effects of exposure to chemical mixtures, understanding that individuals are exposed to many different environmental chemicals (Braun et al. 2016; CDC 2017; Rosofsky et al. 2017). To address this need, multiple novel statistical methods have been developed to assess associations between mixtures of correlated exposures and health outcomes (Taylor et al. 2016), including weighted quantile sum (WQS) regression and Bayesian kernel machine regression (BKMR).

The goal of this study was to assess individual and combined effects of exposure to multiple PFAS with maternal and neonatal thyroid hormones in a prospective cohort of pregnant women from the Boston, MA area, using both WQS regression and BKMR, and to compare the results across these novel statistical methods.

## 2. MATERIALS & METHODS

### 2.1 Study participants

The study population is described in detail in Preston et al. (2018). Briefly, the study population consisted of a subset of pregnant women and their neonates enrolled in the Project Viva prospective pre-birth cohort study between 1999 and 2002. Project Viva enrolled women at their first prenatal visit (median 9.6 weeks gestation), collecting plasma samples and detailed participant information. Of the 2,128 singleton births in Project Viva, 768 had both plasma PFAS and maternal thyroid hormone measurements, and 505 had both plasma PFAS and neonatal T<sub>4</sub> data. We excluded women from our analysis if they were using thyroid-altering medication and/or had a prior or current diagnosis of thyroid disease at enrollment (maternal, n=36; neonatal, n=25) or had missing covariate information (maternal, n=6; neonatal, n=15). The final study population consisted of 726 pregnant women and 465 neonates.

The Centers for Disease Control and Prevention (CDC) laboratory's involvement did not constitute engagement in human subjects research. The Institutional Review Boards of all other participating institutions approved all study protocols and all participating women provided written informed consent.

### 2.2 PFAS quantification

PFAS quantification is described in detail in Preston et al. 2018. Briefly, maternal plasma samples, collected at a median 9.6 weeks gestation, were analyzed at the Division of Laboratory Sciences at the CDC (Atlanta, GA) for concentrations of eight PFAS [perfluorooctanoate (PFOA), perfluorooctane sulfonate (PFOS), perfluorononanoate (PFNA), perfluorohexane sulfonate (PFHxS), 2-(N-ethyl-perfluorooctane sulfonamido) acetate (EtFOSAA; also known as Et-PFOSA-AcOH), 2-(N-methyl-perfluorooctane sulfonamido) acetate (MeFOSAA; also known as Me-PFOSA-AcOH), perfluorodecanoate (PFDA; also known as PFDeA), perfluorooctane sulfonamide (FOSA)] (Kato et al. 2011). PFAS were detected in 99–100% of plasma samples with the following limits of detection (LOD): PFOS, 0.2 ng/mL; all other PFAS, 0.1 ng/mL. PFDA and FOSA were detected in <50% of samples and were not included in further analyses. Concentrations of all other PFAS below the LOD were replaced with the value of the LOD/ 2 for analysis

### 2.3 Thyroid hormone measures

As described in Preston et al. (2018), we quantified levels of TSH, total thyroxine (T<sub>4</sub>), and triiodothyronine (T<sub>3</sub>) resin uptake (T<sub>3</sub>U) in the same maternal blood samples used for PFAS quantification, using the Bayer Advia Centaur assay (Centaur; Bayer Diagnostics, Tarrytown, NY) at the Boston University School of Medicine. We calculated free T<sub>4</sub> index (FT<sub>4</sub>I) using total T<sub>4</sub> and T<sub>3</sub>U levels, which is a standard clinical estimate of circulating free T<sub>4</sub> levels, accounting for potential changes in thyroid binding protein levels such as increased levels of thyroid binding globulin (TBG) during pregnancy and is therefore less prone to bias than traditional immunoassay methods (Lee et al. 2009). TSH levels below the method LOD (0.01 mIU/L; n=7) were replaced with values of 0.01/ 2 for this analysis and an additional 8 women with missing TSH data from TSH models only.

As previously described (Oken et al. 2009; Preston et al. 2018), Project Viva obtained neonatal T<sub>4</sub> data from post-partum heel sticks as part of the New England Newborn Screening Program (NENSP).

## 2.4 Statistical Analysis

We calculated summary statistics and distributions of PFAS concentrations and thyroid hormone levels in mothers and neonates and used Spearman rank correlation coefficients to calculate the degree of correlation among PFAS. Building upon our previous research estimating individual PFAS-hormone associations using multivariable linear regression models in (Preston et al. 2018), we used WQS regression and BKMR (described below) to examine individual (BKMR) and joint (WQS and BKMR) associations of exposure to all six PFAS with maternal and neonatal thyroid hormones. All models were adjusted for potential confounding covariates. Maternal models included: maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, and fish intake; neonatal models included: maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, child sex, age at heel stick, gestational age at birth, and cesarean delivery. Descriptions of variable selection and inclusion are described in detail in Preston et al. (2018). We modeled thyroid hormone levels as continuous outcomes and ln-transformed FT<sub>4</sub>I and TSH levels in all models.

**2.4.1 WQS regression**—Using WQS regression, we estimated joint associations of exposure to all six PFAS with each thyroid hormone. A detailed description of WQS regression can be found in Carrico et al. (2015). Briefly, WQS regression simultaneously estimates empirical weights for each PFAS concentration based on their association with the outcome using a bootstrap step in the training subset of the data. The weighted mean of the bootstrap weights is then applied to plasma concentration quantiles and summed to create an index of the combined exposure, the weighted quantile sum. The exposure index is then used to assess the association between the WQS index and the outcome of interest using the hold-out validation subset of the data with the following model:

$$g(\mu) = \beta_0 + \beta_1 \left( \sum_{i=1-c} w_i * PFASq_i \right) + z' \phi \quad [\text{Equation 1}]$$

Where  $\beta_0$  is the intercept,  $z'$  is a vector of covariates,  $\phi$  is a vector of corresponding regression coefficients, and  $g$  represents the link function, the identity function in our models, linking the predictor variables to the continuous outcome mean,  $\mu$ .

$\sum_{i=1-c} w_i * PFASq_i$  represents the weighted index of all six ( $c$ ) PFAS for each individual, where  $w_j$  is the weight for each exposure and  $PFASq_j$  is the variable for the  $j$ th PFAS scored into quantiles. Weights for all PFAS are constrained from 0  $w_j$  1, and must sum to 1 ( $\sum_{i=1-c} w_i = 1$ ).  $\beta_1$  is the regression coefficient for the weighted index and can be interpreted as the combined effect of exposure to PFAS mixtures on the outcome, while the individual exposure weights ( $w_j$ ) can be interpreted as the relative contribution of that PFAS to the association with the outcome. We used a threshold of  $(100\%/n)$  for weights, where  $n$  represents the number of exposure variables, as a guide to determine which PFAS were the greatest contributors to the overall mixture effect, as well as qualitatively discussing their contributions to the overall mixture effect in relation to each other. WQS regression assumes

linear exposure-outcome associations across quantiles of each exposure and is a first-order approximation of non-additivity. Additionally, WQS regression focuses inference in a single direction at a time with constrained optimization of the beta parameter. Effects can be estimated for both directions individually by conducting separate analyses constraining the analyses in positive and negative directions.

We created WQS indexes based on concentrations of all six PFAS categorized into deciles. We split the data into training (40%) and validation (60%) datasets, estimating empirical weights in the training dataset, averaging weights for each PFAS over 500 bootstrapped samples, and testing the association between the resulting WQS index and our outcome in the validation dataset. We scaled WQS  $\beta_1$  estimates by the interquartile range (IQR) of the WQS index for greater ease of interpretation. Based on our previous work (Preston et al. 2018), we hypothesized that PFAS would be negatively associated with maternal T<sub>4</sub>, FT<sub>4</sub>I, and neonatal T<sub>4</sub> levels, and positively associated with maternal TSH levels. However, we created WQS indexes in both positive and negative directions for all thyroid hormone outcomes in order to account for potential disparate directions of effect across PFAS.

In our previous work modeling PFAS individually, we found differential effects on neonatal T<sub>4</sub> based on infant sex (Preston et al. 2018). Therefore, we assessed heterogeneity in effects by child sex by estimating sex-specific weights within the full neonatal dataset, rather than stratifying our sample to optimize the sample size when splitting the dataset for testing and validation. We did this by creating female and male exposure variables (e.g. “PFOA\_male” and “PFOA\_female” variables), which were used to simultaneously estimate weights for each PFAS variable (Brunst et al. 2017). The results included twelve sex-specific PFAS weights, six male and six female, all constrained from 0  $w_j$  1, and to sum to 1 ( $\sum_{i=1}^c w_i = 1$ ), and an overall beta estimate for the combined effect of exposure to the PFAS mixture in all neonates based on the weighted exposures.

**2.4.2 BKMR analyses**—BKMR is a statistical method that can be used to flexibly model the individual and joint effects of exposure to mixtures of chemicals by using a kernel function (Bobb et al. 2018; Bobb et al. 2015). Unlike WQS, BKMR allows the user to visualize individual exposure-response functions, while accounting for the other exposures and allowing potential non-linear relationships and/or differential directions of effect among exposures. This allows the user to identify independent effects of individual PFAS in addition to the overall combined mixture effect. In addition, BKMR allows the user to assess potential interactions among exposures.

We created separate BKMR models for all maternal thyroid hormone outcomes, and sex-stratified models for neonatal T<sub>4</sub>, based on the model below:

$$Y_i = h(\text{PFOS}_i, \text{PFOA}_i, \text{PFHxS}_i, \text{PFNA}_i, \text{EtFOSAA}_i, \text{MeFOSAA}_i) + \beta z_i + e_i \quad [\text{Equation 2}]$$

Where  $Y_i$  is the continuous thyroid hormone outcome,  $h()$  is the exposure-response function, which can incorporate both non-linear relationships and interactions among exposures,  $z_i$  is a vector of covariates. PFAS were ln-transformed and scaled for BKMR analyses. BKMR

can also be used as a variable selection tool, identifying exposure variables that are important to the overall effect of the exposure mixture by estimating posterior inclusion probabilities (PIPs) for each exposure variable. Similar to the interpretation of the WQS weights, PIPs can be used to identify the relative importance of individual exposure variables to the overall mixture effect. However, unlike WQS weights, PIPs are not constrained to sum to 1.

Results from these models were then be used to 1) provide a visual representation of individual PFAS exposure-response functions,  $h()$ , and their uncertainty, while holding all other PFAS at a given value (e.g. median); 2) calculate specific point estimates for the difference in outcome levels for a change in individual PFAS concentration between the 25<sup>th</sup> and 75<sup>th</sup> percentile, with a corresponding 95% credible interval; 3) Estimate the overall joint effect of exposure to the chemical mixture by providing an estimate of the difference in outcome levels holding concentrations of all six PFAS at various percentiles, compared to holding all PFAS at their median concentrations. Interactions between PFAS can also be assessed by estimating the change in outcome level associated with a change in individual PFAS concentrations, at varying levels (e.g. 25<sup>th</sup>, 50<sup>th</sup>, 75<sup>th</sup> percentile) of one or more additional PFAS. Further details on post-estimation visualizations and statistics from BKMR analyses are discussed in more detail in Bobb et al. 2018.

We used SAS (version 9.4; SAS Institute, Inc) to calculate summary statistics and used R (version 3.5.2; R Development Core Team) for all other analyses including WQS (gWQS package) and BKMR (bkmr package).

### 3. RESULTS

#### 3.1 Participant characteristics

Study population characteristics are described in detail in Preston et al. (2018). Briefly, maternal participants were primarily white (77%), nonsmoking (70%), college graduates (74%), with a mean age of 32.5 years. Neonates were evenly split between males (51%) and females (49%) and had a mean gestational age of 39.5 weeks. Neonatal heel sticks were performed at a mean of 2.0 days post-partum.

#### 3.2 PFAS concentrations and thyroid hormone levels

Table 1 summarizes the distributions of prenatal plasma PFAS concentrations, as well as maternal and neonatal thyroid hormone levels. Prenatal plasma PFAS concentrations were comparable in maternal and neonatal subsets (data not shown).

Figure 1 illustrates the Spearman rank correlation coefficients among prenatal plasma PFAS concentrations. PFAS were moderately to highly correlated with each other ( $r_s$  range: 0.19–0.74;  $p < 0.0001$ ), with PFOS and PFOA having the highest correlation ( $r_s = 0.74$ ,  $p < 0.0001$ ).

#### 3.3 WQS regression

**3.3.1 Maternal thyroid hormones**—Figure 2 summarizes the results of the WQS regression analysis of covariate-adjusted associations between maternal plasma PFAS concentrations and thyroid hormone levels. The results include the beta estimate and

corresponding 95% CI and p-value for an IQR increase in WQS index ( $WQS_{\text{pfas}}$ ) and the corresponding weighted mean of the estimated weights across the 500 bootstrap samples for each individual PFAS. An IQR increase in the  $WQS_{\text{PFAS}}$  index was not associated with maternal  $T_4$  levels (Fig 2a) or TSH levels (Fig 2c), in either the positive or negative direction, indicating that exposure to the PFAS mixture was not associated with maternal  $T_4$  or TSH levels. We observed a significant inverse association between an IQR increase in the  $WQS_{\text{PFAS}}$  index and lower maternal  $\ln\text{-FT}_4\text{I}$  ( $WQS_{\text{PFAS}} \beta = -0.03$ ; 95% CI:  $-0.05, -0.003$ ) when constraining the analysis in the negative direction (Fig 2b). Based on the weighted mean empirical weights for each PFAS, MeFOSAA (49%), EtFOSAA (20%), and to a lesser extent PFOA (15%) and PFHxS (11%), contributed to the overall association, while PFOS (0%) and PFNA (4%) did not appear to contribute.

**3.3.2 Neonatal  $T_4$** —Figure 3 summarizes the results of the WQS regression analysis of the covariate-adjusted association between prenatal plasma PFAS concentrations and neonatal  $T_4$  levels, estimating sex-specific weights for each PFAS. An IQR increase in the  $WQS_{\text{PFAS}}$  index was significantly inversely associated with lower neonatal  $T_4$  levels ( $WQS_{\text{PFAS}} \beta = -1.17 \mu\text{g/dL}$ ; 95% CI:  $-1.88, -0.45$ ). Male infants received the majority of the weights (69%) associated with the decreased  $T_4$  levels, with PFHxS (38%) and MeFOSAA (22%) receiving the highest weights. However, MeFOSAA in female infants also received a contributing weight (13%).

Results from the WQS regression analysis of infants without sex-specific weights were similar. An IQR increase in the  $WQS_{\text{PFAS}}$  index was associated with a  $-0.80 \mu\text{g/dL}$  (95% CI:  $-1.53, -0.06$ ) decrease in neonatal  $T_4$  levels, with MeFOSAA (46%), PFOS (15%), PFHxS (14%), and PFOA (14%) receiving the highest weights (Table S1).

### 3.4 BKMR Analyses

**3.4.1 Maternal thyroid hormones**—Figure 4 shows the results of covariate-adjusted BKMR analyses for maternal thyroid hormones. Figures 4A–C show the univariate exposure-response relationship and 95% credible intervals for select PFAS and maternal thyroid hormones based on the estimated kernel function, controlling for all other PFAS by holding them at their median concentrations. Other PFAS not included in the figure were not individually associated with thyroid hormone levels (Table S2). Figures 4D–F represent the overall effects of the PFAS mixture, showing the estimated differences in thyroid hormone levels and 95% credible intervals when all PFAS concentrations are held at a certain percentile compared to when all PFAS are held at their median concentrations. We did not observe an overall effect of exposure to the PFAS mixture on maternal total  $T_4$  levels (Fig 4D) or  $\ln\text{-TSH}$  levels (Fig 4F). However, BKMR did identify an individual U-shaped association between PFHxS concentrations and  $\ln\text{-TSH}$  levels (Fig 4C), as well as an individual positive association between PFOS concentrations and total  $T_4$  levels (Fig 4A), while holding all other PFAS at their median concentrations. A change in PFOS concentrations from the 25<sup>th</sup> to 75<sup>th</sup> percentile was associated with a  $0.21 \mu\text{g/dL}$  increase in total  $T_4$  levels (95% credible interval:  $-0.03, 0.47$ ).



We observed a significant inverse association between exposure to the entire PFAS mixture and  $\ln$ -FT<sub>4</sub>I (Fig 4E). Holding all PFAS at the 75<sup>th</sup> percentile compared to the 50<sup>th</sup> percentile was associated with a -0.01 decrease in  $\ln$ -FT<sub>4</sub>I (95% credible interval: -0.02, -0.003). Based on the estimated PIPs, BKMR identified PFOA (0.07), PFHxS (0.05), EtFOSAA (0.03), and MeFOSAA (0.02) as being important contributors to the overall association. However, only PFOA was independently associated with  $\ln$ -FT<sub>4</sub>I (Fig 4B); a change in PFOA concentration from the 25<sup>th</sup> to 75<sup>th</sup> percentile was associated with a -0.02 decrease in  $\ln$ -FT<sub>4</sub>I (95% credible interval: -0.04, 0.01). We saw no evidence of interactions among PFAS in any of our maternal BKMR analyses (data not shown).

**3.4.2 Neonatal T<sub>4</sub>**—Figure 5 shows the results of BKMR our sex-stratified analyses for total T<sub>4</sub> in male infants. We observed a non-significant inverse association between exposure to the overall PFAS mixture and T<sub>4</sub> levels in male infants (Fig 5C); associations were null in female infants (Table S3). Based on the PIPs, PFHxS (0.97) was the primary contributor to the overall association, while the other PFAS did not seem to contribute to the association (PIPs: EtFOSAA=0.30, PFOS=0.25, PFOA=0.22, PFNA=0.20, MeFOSAA=0.20). When evaluating individual associations, we found a significant inverse association between PFHxS concentrations and T<sub>4</sub> levels (Fig 5A). A change in  $\ln$ -PFHxS concentrations from the 25<sup>th</sup> to 75<sup>th</sup> percentile was associated with a -0.89  $\mu$ g/dL decrease in T<sub>4</sub> levels (95% credible interval: -1.64, -0.15), holding all other PFAS at their medians. Conversely, BKMR also identified a suggestive positive linear association between EtFOSAA concentrations and T<sub>4</sub> levels in male infants (Fig 5B), with an estimated 0.68  $\mu$ g/dL increase in T<sub>4</sub> levels (95% credible interval: -0.15, 1.51) associated with a change in EtFOSAA concentrations from the 25<sup>th</sup> to 75<sup>th</sup> percentile, while holding all other PFAS at their medians. We saw no evidence of interactions among PFAS in our neonatal analysis (data not shown).

Results from BKMR in all infants were similar in direction of effect to those in male infants. We observed a non-significant trend of higher levels of the overall PFAS mixture associated with lower neonatal T<sub>4</sub> levels (Figure S1) and a suggestive individual association between a change in  $\ln$ -PFHxS concentrations from the 25<sup>th</sup> to 75<sup>th</sup> percentile and lower neonatal T<sub>4</sub> levels (-0.37  $\mu$ g/dL; 95% CI: -0.88, 0.13) (Table S3).

## 4. DISCUSSION

The aims of this study were to assess the individual and combined effects of prenatal exposure to multiple PFAS on maternal and neonatal thyroid hormone levels, using two different statistical approaches. In this relatively large cohort of pregnant women and their neonates, the PFAS mixture was inversely associated with maternal FT<sub>4</sub>I and neonatal T<sub>4</sub> levels in male infants, using both BKMR and WQS regression. However, the PFAS mixture was not associated with maternal T<sub>4</sub> or TSH levels in either WQS regression or BKMR analyses. Additionally, our BKMR results identified multiple individual PFAS-hormone associations, while accounting for exposures to the other PFAS. In moms, higher PFOS concentrations was associated with higher total T<sub>4</sub> levels, while higher PFOA concentrations were associated with lower FT<sub>4</sub>I, and PFHxS concentrations were non-linearly associated with TSH levels. In infants, higher PFHxS concentrations were associated with lower T<sub>4</sub>

levels and higher EtFOSAA concentrations were suggestively associated with higher T<sub>4</sub> levels in male, but not female infants.

Overall, results from BKMR and WQS regression analyses were relatively consistent. Table 2 summarizes the results of the maternal FT<sub>4</sub>I and male neonatal T<sub>4</sub> analyses from BKMR, WQS, and our previous linear regression analysis (Preston et al. 2018). Both WQS and BKMR identified associations of the PFAS mixture with decreased maternal FT<sub>4</sub>I and neonatal T<sub>4</sub>, primarily in male infants. Neither method identified combined effects of the PFAS mixture with maternal total T<sub>4</sub> or TSH levels or neonatal T<sub>4</sub> in females. When comparing which individual PFAS were identified as important contributors to these combined effects, BKMR and WQS regression identified PFOA, PFHxS, EtFOSAA, and MeFOSAA as contributing to the negative joint association of PFAS exposure with maternal FT<sub>4</sub>I. Conversely, WQS identified both PFHxS and MEFOSAA exposure in males as contributing to the negative joint association of PFAS exposure with neonatal T<sub>4</sub> levels, while BKMR only identified PFHxS as an important contributor to the suggestive negative association with T<sub>4</sub> levels. However, in our WQS regression analysis we were able to create sex-specific weights while still utilizing the entire neonatal dataset, while for BKMR analyses we analyzed male and female infants separately, which could have led to slightly different results. Additionally, the directionality assumption of WQS may have been violated in our neonatal analysis, as BKMR identified disparate directions of effect for PFHxS and EtFOSAA on T<sub>4</sub> levels in male infants, which likely contributed to the slightly different results from WQS and BKMR analyses. While the PFAS mixture was associated with neonatal T<sub>4</sub> levels and maternal FT<sub>4</sub>I, it was not associated with maternal total T<sub>4</sub> levels. The differences between neonatal and maternal hormone analyses could be due to increased variability in maternal T<sub>4</sub> levels during pregnancy, which could have limited our ability to detect subtle associations.

In our previous analysis modeling individual PFAS-hormone associations in the Project Viva cohort, we observed individual inverse associations between PFOA, PFHxS, and MeFOSAA with maternal FT<sub>4</sub>I (Preston et al. 2018), which were all also identified here in both WQS and BKMR as contributing to the joint effect of PFAS exposure on maternal FT<sub>4</sub>I levels (Table 2). However, only PFOA was individually associated with FT<sub>4</sub>I in BKMR analyses. The consistency across traditional regression and novel mixtures methods indicates that there was not significant confounding among PFAS of the association with individual PFAS and maternal FT<sub>4</sub>I. In our previous neonatal analyses, we observed individual inverse associations between PFOS, PFOA, and PFHxS with T<sub>4</sub> levels in male infants (Preston et al. 2018). However, in our WQS regression and BKMR analyses, PFOS and PFOA did not contribute to the overall joint associations with neonatal T<sub>4</sub> levels, and were also not individually associated with neonatal T<sub>4</sub> in BKMR analyses. MeFOSAA was not associated with neonatal T<sub>4</sub> levels in our previous analyses, but was an important contributor to the overall mixture effect in both WQS and BKMR analyses. Differences in findings across the analyses may be due to the relatively high correlations among these exposures, and/or confounding among PFAS in our previous single- PFAS analyses.

While we did not see consistent associations between exposures to multiple PFAS and T<sub>4</sub> or TSH in our analyses, reductions in maternal FT<sub>4</sub>I (free T<sub>4</sub>) alone have been consistently

associated with impaired fetal growth and neurodevelopment (de Escobar et al. 2004; Henrichs et al. 2013; Morreale de Escobar 2001). Effects of low neonatal T<sub>4</sub> on subsequent growth and development are less well characterized, but altered neonatal thyroid function has been associated with reductions in IQ, attention, and neurocognitive tests (American Academy of Pediatrics et al. 2006; Lyall et al. 2017; Simic et al. 2009b).

While there is growing evidence from both the epidemiology and toxicology literature that exposure to PFAS alters thyroid function, the mechanism(s) of effect and potential differences across PFAS remain unclear. Proposed mechanisms include reduced responsiveness to the hypothalamic-pituitary-thyroid axis, increased hepatic clearance of T<sub>4</sub>, increased conversion of T<sub>4</sub> to T<sub>3</sub> by type 1 deiodinase, and competitive binding to thyroid hormone binding proteins (Long et al. 2013; Weiss et al. 2009; Yu et al. 2009).

Toxicology studies investigating effects of PFAS exposure on thyroid function have primarily investigated effects of PFOS or PFOA, but less is known about the toxicological effects of other PFAS and even less is known about exposure to PFAS mixtures. Studies including a range of PFAS have demonstrated varying levels of effect based on PFAS structure (Long et al. 2013; Ren et al. 2016; Ren et al. 2015; Weiss et al. 2009). For example, PFAS binding affinity to the thyroid hormone transport protein, transthyretin, varies based on PFAS alkyl chain length and functional group (Ren et al. 2016; Weiss et al. 2009). Similarly, Ren et al. (2015) demonstrated that there is a structure-dependent binding relationship between PFAS and the human thyroid receptor (TR), where longer-chain PFAS with acid end groups had the highest binding affinity to TR. Long et al. (2013) showed differences in effects on the thyroid hormone system using the T-screen assay, across seven PFAS. While the specific differences in action across PFAS described above may not directly explain why we observed that only certain PFAS were responsible for the joint effects on maternal FT<sub>4</sub>I and neonatal T<sub>4</sub> levels, they demonstrate the potential for different toxicities and mechanisms of action across different PFAS. There is a critical lack of data on the joint effects of PFAS on thyroid function. However, limited evidence from studies investigating other endpoints (e.g. cytotoxicity) have reported both additive and more complex interaction effects between exposure to multiple PFAS (Ding et al. 2013; Hoover et al. 2019).

Results from previous studies of associations between PFAS exposure and maternal and neonatal thyroid hormone levels have been inconsistent, likely due to the extreme heterogeneity in study design, population demographics, PFAS and thyroid hormone analysis techniques, statistical analysis, and PFAS concentration distributions (Ballesteros et al. 2016). The most consistent findings have been a positive association between PFAS and TSH in both maternal and neonatal analyses, with some studies also reporting an inverse association between PFAS and maternal and/or neonatal free or total T<sub>4</sub> levels (Ballesteros et al. 2016; Berg et al. 2015; Berg et al. 2016; Wang et al. 2014; Wang et al. 2013). The inverse association between PFAS and maternal FT<sub>4</sub>I was also reported (ft4) in single-PFAS models in a Taiwanese cohort (Wang et al. 2014) and in TPOAb positive women in Canadian (Webster et al. 2014) and Japanese (Itoh et al. 2019) cohorts. Inverse associations between PFAS concentrations and neonatal T<sub>4</sub> levels have also been reported in single-PFAS models in previous studies (Kim et al. 2011; Wang et al. 2014). Previous studies have also reported

sex-specific effects of PFAS exposure on thyroid hormones. However, the mechanism behind these effects remains unclear. Male infants may be more susceptible to alterations in thyroid function due to their generally lower levels of T<sub>4</sub> compared to female infants (Chan et al. 2011; Herbstman et al. 2008; Kuppens et al. 2011; Preston et al. 2018). While in the opposite direction of our findings, a recent Chinese cohort reported positive associations between multiple PFAS and free T<sub>4</sub> levels among male, but not female infants (Aimuzi et al. 2019). Conversely, Wen et al. (2013) observed a similar sex-specific association as seen in our study, reporting higher serum PFHxS concentrations associated with lower free T<sub>4</sub> levels in male, but not female adult participants in NHANES.

Few previous studies have examined the combined effects of exposure to multiple PFAS, and those that have done so have primarily used traditional regression models containing multiple PFAS in a single model (Chan et al. 2011; Shah-Kulkarni et al. 2016). Berg et al. (2016) used principle components analysis and hierarchical clustering methods to assess associations between multiple pollutants, including several PFAS, and maternal thyroid hormone levels, due to high correlations among the chemicals. However, these methods group exposures based on correlations with each other rather than with the outcome. Thus, individual exposures within a component or correlated group may vary in respect to their association with the outcome, without a clear method of disentangling these relationships within the components. A recent Chinese study (Aimuzi et al. 2019) used sparse partial least squares (SPLS) regression, a novel method proposed to identify the dominant exposures related to the outcome of interest in a group of highly correlated exposures (Lenters et al. 2015), to assess associations of prenatal PFAS concentrations with neonatal thyroid hormone levels. However, the majority of these methods do not estimate the overall joint effects of exposure to multiple PFAS. While toxicological data on the joint effects of PFAS is extremely limited, additive and even non-additive interactions between PFAS have been reported (Ding et al. 2013; Hoover et al. 2019). Because we know that individuals are ubiquitously exposed to multiple PFAS, understanding the joint effects of these exposures on thyroid function and other endpoints is critically important and should be considered in future studies.

The major strength of the present study was the use of BKMR and WQS regression, which allowed us to examine both the joint and individual effects of exposure to multiple PFAS on maternal and neonatal thyroid hormone levels, accounting for moderately to highly correlated PFAS concentrations. Additionally, the use of BKMR allowed us to evaluate potential non-linear exposure-response functions and interactions among PFAS. Other strengths of the present study include the relatively large sample size, which allowed us to assess these complex relationships while controlling for multiple potential confounding factors and stratifying by neonatal sex, the prospective nature of our neonatal analysis, and the collection of maternal blood samples during early pregnancy. Unlike many previous studies, we collected maternal blood samples during the first trimester of pregnancy, a period when offspring are particularly susceptible to the potential adverse neurodevelopmental effects of alterations in maternal thyroid hormone levels (de Escobar et al. 2004; Henrichs et al. 2013). Physiological changes such as plasma volume expansion and increased glomerular filtration rate, which can affect both plasma PFAS concentrations and thyroid hormone levels, are less likely to be significant confounding factors during early pregnancy compared

to later pregnancy. As previously reported in the Project Viva cohort, controlling for these factors using plasma albumin and estimated GFR levels, in addition to gestational week at blood draw, did not affect the observed associations and were therefore not included in the present analyses (Preston et al. 2018).

There are several potential limitations to the present study. WQS regression assumes a linear relationship between concentration quantiles and the outcome of interest. We were able to assess the linearity of these relationships in our BKMR analysis. While most exposure-response associations appeared to be relatively linear, we observed a non-linear association between PFHxS concentrations and maternal TSH levels, indicating that WQS may not have been an appropriate method to assess the joint effects of PFAS on this hormone. WQS requires the exposures to all act in the same direction of effect on the outcome for a given analysis. However, users can perform separate analyses in both directions to assess disparate directions of effect across exposures, as done in the present study. The direction of associations did vary for some PFAS in our analyses. For example, in our BKMR analysis, we observed a significant inverse association between PFHxS and T<sub>4</sub> levels, but a suggestive positive association between EtFOSAA and T<sub>4</sub> levels in male neonates. WQS does not directly account for differences in concentration ranges across exposures as exposure concentrations are categorized into quantiles. In addition, WQS assumes that there are no greater than additive interactions among exposure variables, but we were able to validate this assumption in our BKMR analysis.

There are several additional limitations of the present study. The Project Viva study population is of higher socioeconomic status and less racially diverse than the general population, potentially limiting the generalizability of our study results. However, concentrations of PFAS in Project Viva were comparable to those in NHANES during the corresponding sampling time period (Sagiv et al. 2015). In both maternal and neonatal analyses, we did not have a direct measure of iodine status, an important potential confounder, but we did have information on dietary intake of iodine rich foods. Our maternal analysis was cross-sectional, limiting our ability to determine temporality of exposures in relation to the outcomes, but is potentially offset by the relatively long half-lives of many PFAS in plasma, measured in years. In neonatal analyses, we only had data on T<sub>4</sub> levels and were not able to explore associations with other neonatal thyroid hormone levels. While we were able to assess potential confounding by numerous demographic and physiological factors as well as PFAS evaluated, we cannot rule out potential residual confounding from uncontrolled factors such as demographic factors or other exposures to additional thyroid disrupting chemicals. We also cannot rule out potential interactions between PFAS and other unmeasured chemical exposures.

## 5. CONCLUSION

We examined the effect of exposure to multiple PFAS on maternal and neonatal thyroid function using two novel statistical methods, WQS regression and BKMR. We found that exposure to the PFAS mixture was associated with lower maternal FT<sub>4</sub>I and lower neonatal T<sub>4</sub>, particularly in male infants, with PFOA, PFHxS, EtFOSAA, and MeFOSAA primarily contributing to the joint effect on maternal FT<sub>4</sub>I, and PFHxS and MeFOSAA primarily

contributing to the joint effect on T<sub>4</sub> levels in male infants. To our knowledge, this is one of the first epidemiologic studies to use WQS regression and/or BKMR to assess the joint and individual effects of exposure to multiple PFAS on maternal and neonatal thyroid hormone levels. Further studies are needed to confirm these results in other populations and to better understand the toxicological mechanism(s) behind the observed differential effects of individual PFAS.

## Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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The findings and conclusions in this manuscript are those of the authors and do not necessarily represent the official position of the Centers for Disease Control and Prevention (CDC). Use of trade names is for identification only and does not imply endorsement by the CDC, the Public Health Service, or the US Department of Health and Human Services.

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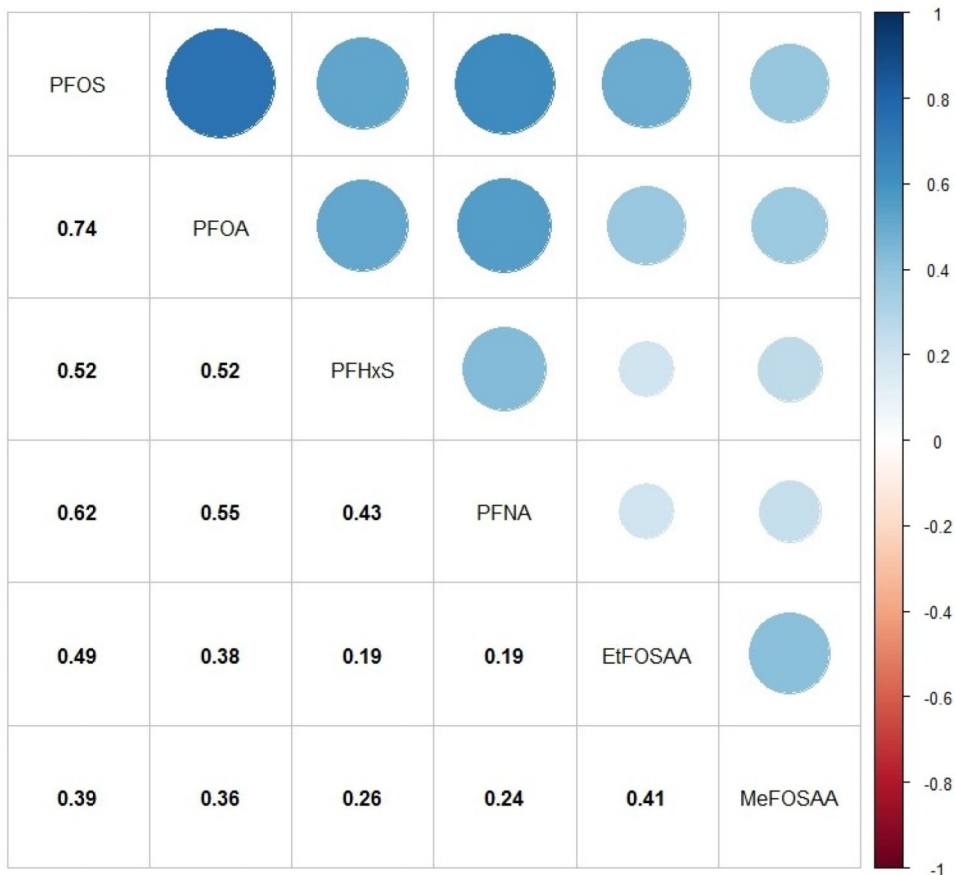
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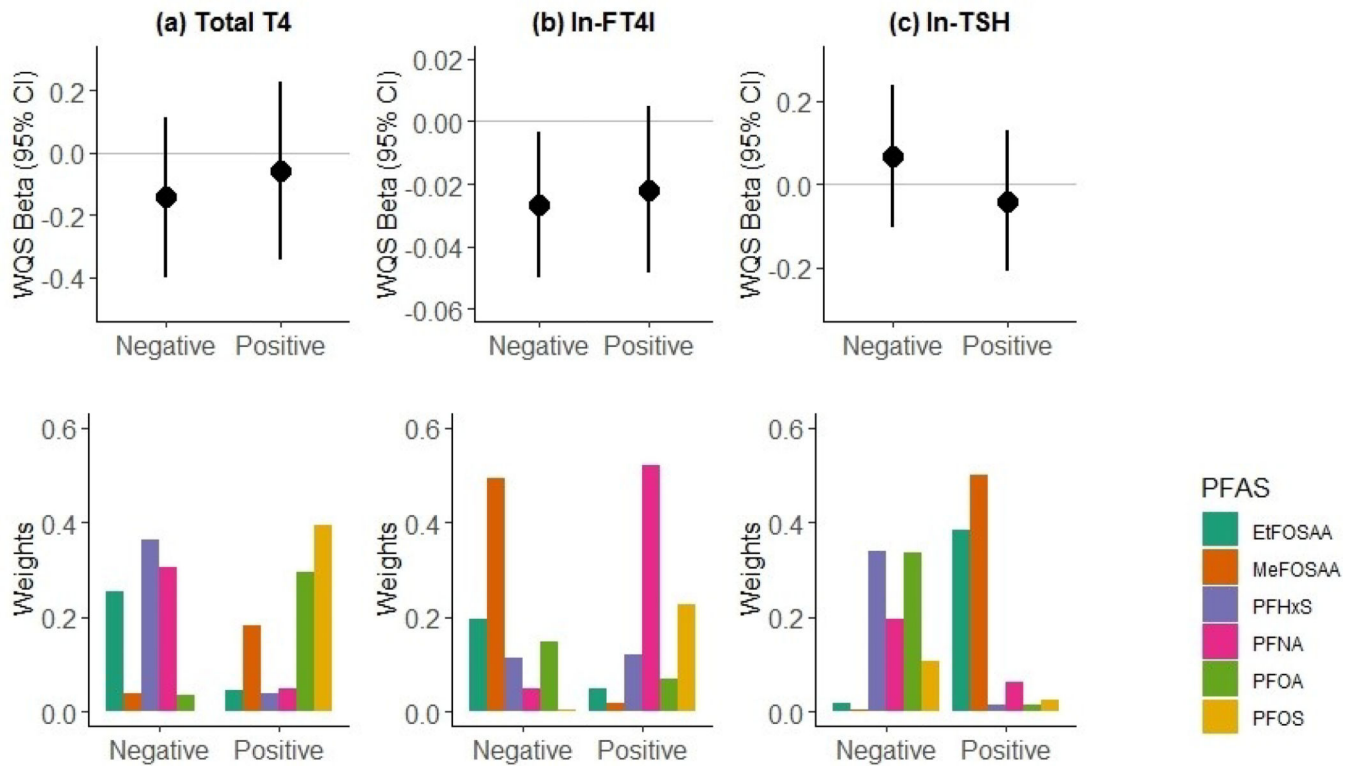
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**HIGHLIGHTS**

- We examined associations of PFAS mixtures with maternal and neonatal thyroid function
- Results from WQS regression and BKMR analyses were relatively comparable
- The PFAS mixture was inversely associated with maternal FT<sub>4</sub>I and neonatal T<sub>4</sub>
- PFOA, PFHxS, EtFOSAA, and MeFOSAA primarily contributed to the joint effect on FT<sub>4</sub>I
- PFHxS and MeFOSAA primarily contributed to the joint effect on neonatal T<sub>4</sub> levels

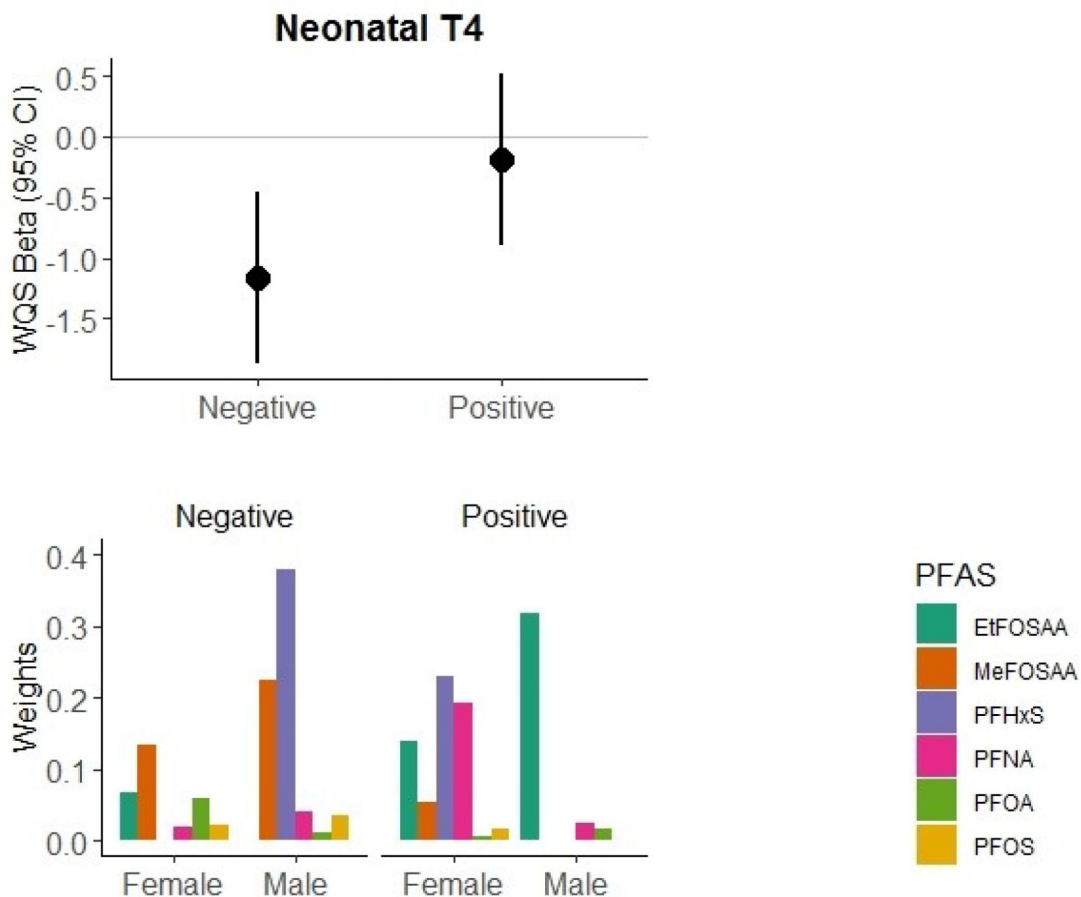


**Figure 1.** Spearman rank correlation matrix for maternal PFAS concentrations (ng/mL) measured in early pregnancy plasma samples (n=726). All p-values <0.0001. Abbreviations: PFAS, per- and polyfluoroalkyl substances; PFOS, perfluorooctane sulfonate; PFOA, perfluorooctanoic acid; PFHxS, perfluorohexane sulfonate; PFNA, perfluorononanoate; EtFOSAA, 2-(N-ethyl-perfluorooctane sulfonamide) acetate; MeFOSAA, 2-(N-methyl-perfluorooctane sulfonamide) acetate.

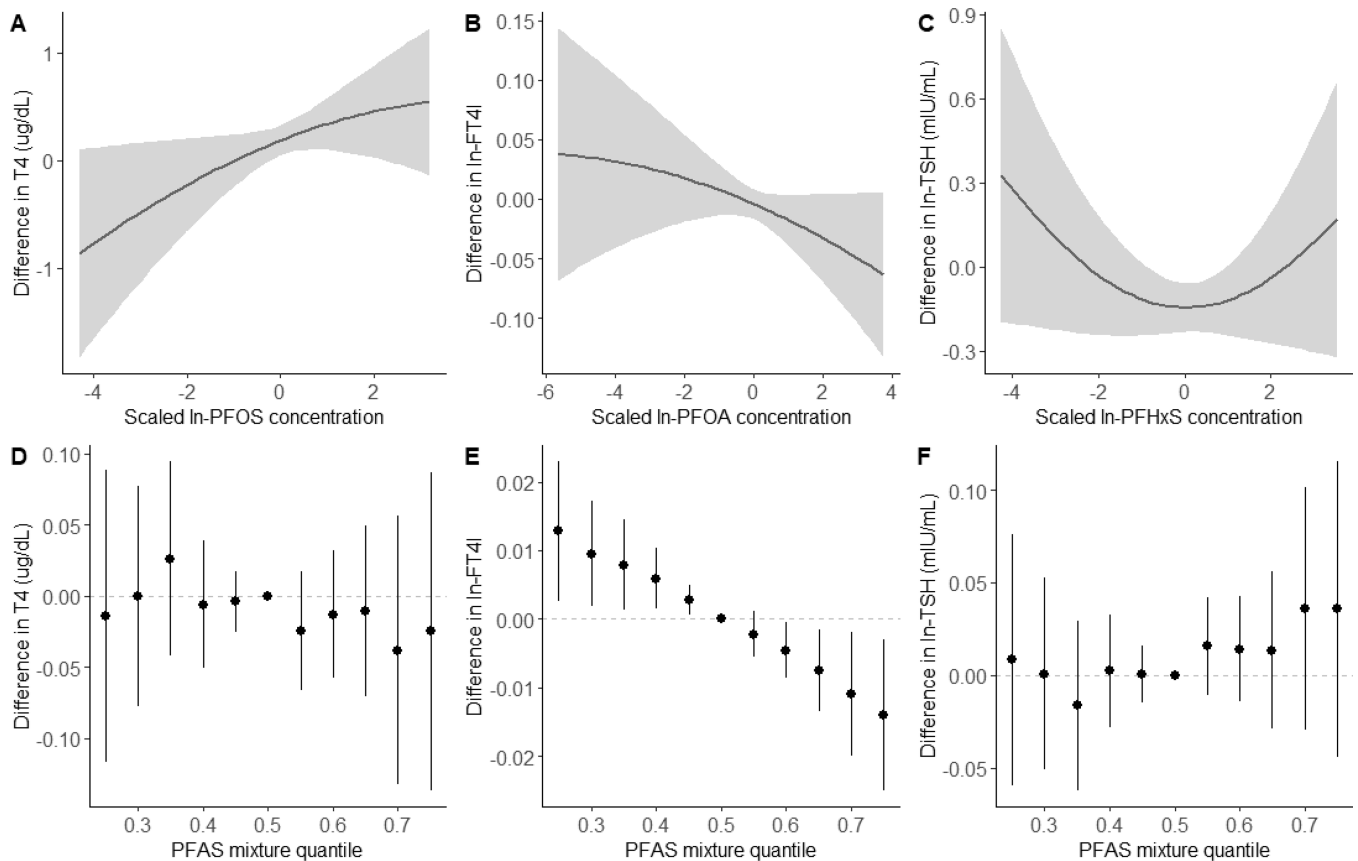


**Figure 2.**

Associations of combined PFAS exposure with maternal thyroid hormone levels based on weighted quantile sum (WQS) regression analysis ( $n=726$ ), adjusting for maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, and fish intake. We calculated two separate WQS indices for associations with each hormone, one in the positive direction of effect and one in the negative direction of effect. Results of each analysis include the WQS beta and 95% confidence interval for the effect of combined exposure to the PFAS mixture on (a) total T<sub>4</sub> ( $\mu\text{g/dL}$ ), (b) ln-FT<sub>4</sub>I, and (c) ln-TSH levels (mIU/mL), with corresponding weights for the contributions of individual PFAS to the overall effect.

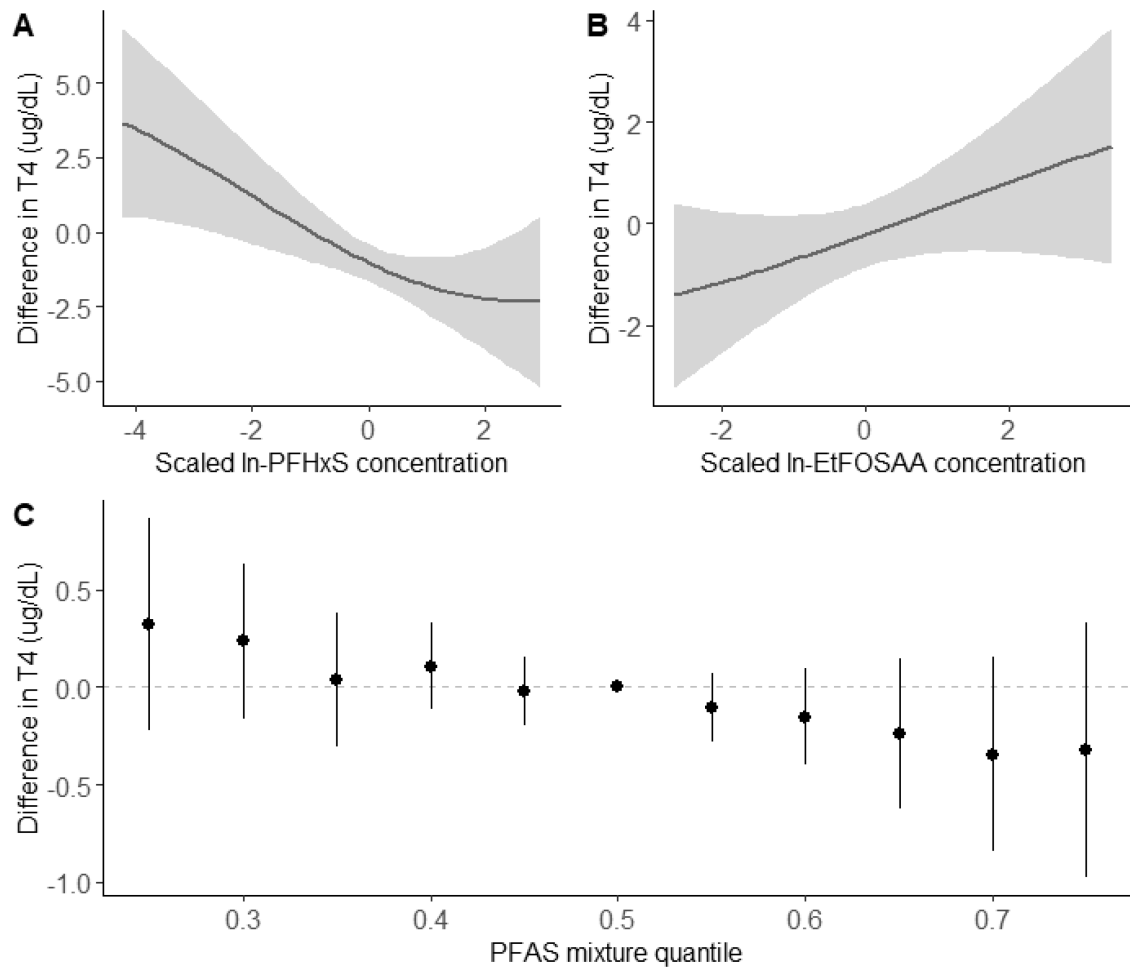


**Figure 3.** Associations of combined PFAS exposure with neonatal T<sub>4</sub> levels based on weighted quantile sum (WQS) regression analysis (n=465), adjusting for maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, age at heel stick, gestational age at birth, and cesarean delivery. We calculated two separate WQS indices for associations with each hormone, one in the positive direction of effect and one in the negative direction of effect. Results of each analysis include the WQS beta and 95% confidence interval for the effect of combined exposure to the PFAS mixture on T<sub>4</sub> levels (µg/dL), with corresponding sex-specific weights for the contributions of individual PFAS to the overall effect.



**Figure 4.**

Combined effects of the PFAS mixture on maternal thyroid hormone levels (n=726) estimated by Bayesian Kernel Machine Regression (BKMR), adjusting for maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, and fish intake. Univariate exposure-response function and 95% confidence bands for (A) ln-PFOS and total T<sub>4</sub> (µg/dL), (B) ln-PFOA and ln-FT<sub>4</sub>I, and (C) ln-PFHxS and ln-TSH levels (mIU/mL) holding all other PFAS at the median. Overall effect of the PFAS mixture on (D) total T<sub>4</sub>, (E) ln-FT<sub>4</sub>I, and (F) ln-TSH levels. These plots show the estimated difference in hormone levels and 95% credible intervals when all PFAS concentrations are held at a certain percentile compared to when PFAS concentrations are held at their medians, representing the exposure-response relationship of the PFAS mixture with hormone levels.



**Figure 5.**

Combined effects of the PFAS mixture on total T<sub>4</sub> levels (μg/dL) in male infants (n=236) estimated by Bayesian Kernel Machine Regression (BKMR), adjusting for maternal age, race/ethnicity, smoking habits, parity, gestational week at blood draw, age at heel stick, gestational age at birth, and cesarean delivery). Univariate exposure-response functions and 95% confidence bands for (A) ln-PFHxS and (B) ln-EtFOSAA holding all other PFASs at their medians. (C) Overall effect of the PFAS mixture on T<sub>4</sub> levels in male infants. This plot shows the estimated difference in total T<sub>4</sub> levels and 95% credible intervals when all PFAS concentrations are held at a certain percentile compared to when all PFAS concentrations are held at their medians, representing the exposure-response relationship of the PFAS mixture with T<sub>4</sub> levels.



**Table 1.**

Prenatal plasma PFAS (ng/mL) and maternal (n=726) and neonatal (n=465) thyroid hormone distributions

Analyte	Detection frequency <sup>a</sup>	Percentile				Max
		Min	25%	50%	75%	
Prenatal plasma PFAS (ng/mL)						
PFOS	100	2.8	17.6	23.9	32.6	115.0
PFOA	100	0.3	3.9	5.6	7.7	36.7
PFHxS	98.5	<LOD	1.6	2.4	3.7	43.2
PFNA	98.6	<LOD	0.5	0.6	0.8	6.0
EtFOSAA	99.6	<LOD	0.7	1.1	1.7	33.6
MeFOSAA	100	0.1	1.2	1.8	2.9	29.7
Thyroid hormones						
Maternal (n=726)						
Total T <sub>4</sub> (µg/dL)		3.9	8.7	9.9	11.2	24.4
Free T <sub>4</sub> Index		1.3	1.9	2.1	2.3	6.0
TSH (mIU/mL) <sup>b</sup>		<LOD	0.7	1.2	1.9	19.3
Neonatal (n=465)						
Total T <sub>4</sub> (µg/dL) <sup>c</sup>		3.73	14.8	17.3	20.2	35.7

Abbreviations: PFAS, per- and polyfluoroalkyl substances; PFOS, perfluorooctane sulfonate; PFOA, perfluorooctanoic acid; PFHxS, perfluorohexane sulfonate; PFNA, perfluorononanoate; EtFOSAA, 2-(N-ethyl-perfluorooctane sulfonamide) acetate; MeFOSAA, 2-(N-methyl-perfluorooctane sulfonamide) acetate

<sup>a</sup>Limits of detection (LOD) were 0.2 ng/mL for PFOS and 0.1 ng/mL for all other PFAS

<sup>b</sup>n=718, 8 participants with missing TSH values

**Table 2.**

Summary of results from statistical methods estimating individual and joint associations of PFAS concentrations with maternal FT<sub>4</sub>I and male neonatal T<sub>4</sub> levels

Analysis	PFAS Mixture	PFOS	PFOA	PFHxS	PFNA	EtFOSAA	MeFOSAA
<b>Maternal FT<sub>4</sub>I</b>							
Linear regression <sup>a</sup>	NA		↓	↓			↓
BKMR	↓		↓	x		x	x
WQS	↓		x	x*		x	x
<b>Neonatal T<sub>4</sub> – males</b>							
Linear regression <sup>a</sup>	NA	↓	↓	↓			
BKMR	↓			↓		↑	
WQS	↓			x			x

Note: Arrows represent individual associations with the outcome; “x” represents PFAS identified as contributing to overall mixture effect with high PIPs (BKMR) or high weights (WQS)

<sup>a</sup>Results from Preston et al. 2018 – modeling single PFAS-hormone associations using individual covariate-adjusted linear regression models, where PFAS concentrations were modeled as quartiles.

\* WQS weight < 100%/n

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