

CONTEMPORARY REVIEW

The Impact of Environmental Chemicals on the Gut Microbiome

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ABSTRACT

Since the surge of microbiome research in the last decade, many studies have provided insight into the causes and consequences of changes in the gut microbiota. Among the multiple factors involved in regulating the microbiome, exogenous factors such as diet and environmental chemicals have been shown to alter the gut microbiome significantly. Although diet substantially contributes to changes in the gut microbiome, environmental chemicals are major contaminants in our food and are often overlooked. Herein, we summarize the current knowledge on major classes of environmental chemicals (bisphenols, phthalates, persistent organic pollutants, heavy metals, and pesticides) and their impact on the gut microbiome, which includes alterations in microbial composition, gene expression, function, and health effects in the host. We then discuss health-related implications of gut microbial changes, which include changes in metabolism, immunity, and neurological function.

Key words: environmental chemicals; gut microbiome; health risks.

Humans are exposed to hundreds of chemicals as evidenced by the fact that more than 300 environmental chemicals or their metabolites have been measured in human biological samples (CDC, 2017). Human exposure to these environmental chemicals is constant, and some of these chemicals have long half-lives in the body and environment. Chemicals such as bisphenols, phthalates, pesticides, persistent organic pollutants (POPs), and heavy metals have endocrine-disrupting effects that can alter hormonal metabolism. Many of these environmental chemicals are associated with adverse health outcomes, including male and female reproductive and developmental defects, type 2 diabetes, cardiovascular dysfunction,

liver disease, obesity, thyroid disorders, and immune dysfunction (Chiang and Flaws, 2019; Hannon and Flaws, 2015; Jaishankar et al., 2014; Nagy et al., 2019; Rochester, 2013). The gut microbiome influences host metabolism, and thus, it may mediate some of the toxic effects of environmental chemicals (Martin et al., 2019). With chronic exposure to a variety of environmental chemicals, it is vital to understand how the gut microbial community changes in response to environmental chemical exposures and the implications of such changes on health outcomes.

In this review, we summarize the known impact of environmental chemical exposure on the gut microbiome, with a

specific focus on the effects of bisphenols, phthalates, POPs, heavy metals, and pesticides on the gut microbiome. These chemical classes were included in the review because exposure to these chemicals is frequent in humans, and increasing evidence suggests they alter gut microbiota composition. In addition to summarizing the impact of these chemicals on altering the gut microbiome, we also summarize alterations in gene expression, functional changes, and health implications associated with these gut microbiome changes.

BISPHENOLS AND THE GUT MICROBIOME

Bisphenol A (BPA), a plastic monomer and high-volume industrial chemical, is one of the most widely studied endocrine-disrupting chemicals. Extensive research has shown that BPA is a reproductive toxicant, a neurotoxicant, and an obesogen, likely due to its affinity to multiple hormone receptors in the body (Patisaul, 2020; Vom Saal et al., 2012; Ziv-Gal and Flaws, 2016).

Recent animal studies of developmental and adult exposure to BPA demonstrate that BPA can alter the gut microbiota of a variety of species in a sex-specific manner (Table 1). For example, male mice exposed to BPA have increased Prevotellaceae (Javurek et al., 2016), bacteria that are known to be involved in the mucosal barrier function (Wright et al., 2000). However, female mice have no change in the relative abundance of Prevotellaceae (Javurek et al., 2016). BPA exposure also upregulated the levels of Akkermansia and Methanobrevibacter in the gut microbiome of males. This is of concern because Akkermansia, which plays a role in butyrate production, is also elevated in human cancers (Baxter et al., 2014; Weir et al., 2013), implying that BPA-induced gut microbiome alterations may increase the risk of carcinogenesis. However, other species of Akkermansia have been shown to be beneficial for intestinal immunity, glucose metabolism, and lipid metabolism (Naito et al., 2018). Furthermore, Methanobrevibacter is shown to heighten the host's ability to metabolize exogenous fuels, resulting in upregulated host energy intake and weight gain (Samuel and Gordon, 2006; Sweeney and Morton, 2013). This raises a strong possibility that BPA-induced weight gain is caused at least partially by BPA-induced changes in the gut microbiome. Overall, BPA exposure produces sex-specific outcomes on the gut microbiome, implying that BPA-exposed males and females may be at different risk for certain diseases.

In support of the disease-driving effect of BPA exposure, studies found that BPA-induced alterations in gut microbiota are often associated with several physiological disorders in the host (Table 1). BPA-induced microbial alterations are found to increase oxidative stress, which is a trigger to inflammation. *In vitro* studies with human colonic cell lines show increased oxidative stress in response to BPA treatment (Qu et al., 2018; Zhao et al., 2019). A study using the *in vitro* Simulator of the Human Intestinal Microbial Ecosystem (SHIME) system to predict the effects of BPA on the human gut also found increased expression of genes related to oxidative stress (Wang et al., 2018b). Furthermore, developmental BPA exposure in mice and rabbits reduced the diversity of gut microbiota composition with a decrease in protective bacteria, such as short-chain fatty acid (SCFA) producers. This is accompanied by chronic intestinal and hepatic inflammation and metabolic disorder (Malaise et al., 2017, 2018; Reddivari et al., 2017). The mechanisms by which BPA executes its actions on the gut microbiome remain unclear. Interestingly, in comparison with the effects of the synthetic estrogen ethinylestradiol (EE), BPA caused many of the

same microbial effects as EE on the gut, such as decreased relative abundance of Firmicutes and Proteobacteria, implying BPA may modify gut microbiota through an estrogenic mechanism (Javurek et al., 2016; Liu et al., 2016).

In addition to the effect on the gut, BPA treatment is known to affect brain development, behavior, and lipid and glucose metabolism (Collins et al., 2012; Guan et al., 2019; Nicholson et al., 2012). As gut microbiota function is known to be important for both brain and metabolic health, it is tempting to speculate that BPA-induced gut microbiome alterations partially mediate the adverse effects of BPA on neurological and metabolic health. Although BPA studies reveal a strong effect of this chemical on altering gut microbiota composition (Catron et al., 2019; Javurek et al., 2016; Wang et al., 2018b), studies that evaluate the contribution of such alterations to BPA-induced adverse neurological and metabolic health outcomes are critically needed.

Because the problematic endocrine disrupting properties of BPA have come to light, alternative bisphenols have increased in prevalence and use (Rochester and Bolden, 2015). Considerably less research has been performed on alternative bisphenols, despite the fact that some may be more developmentally toxic than BPA (Kinch et al., 2015). A study examining the effects of bisphenol S (BPS), bisphenol F (BPF), bisphenol B (BPB), bisphenol AF (BPAF), as well as BPA on the gut microbiome in zebrafish embryos found that the least developmentally toxic bisphenols (BPA, BPS, and BPF) had the most substantial impacts on the microbial community (Catron et al., 2019). In contrast, the most developmentally toxic bisphenols (BPB and BPAF) did not impact microbiota (Catron et al., 2019). The observed developmental toxicity ranking mirrors bisphenol potency in estrogen receptor activation *in vitro*, implying a causal relationship between estrogenicity and microbiota disruption. Future studies should determine the effects of other bisphenol alternatives, such as BPS, BPF, BPB, and BPAF, on the gut microbiome using mouse or pig models for human research and examine if they lead to similar outcomes in mammals. Furthermore, more studies are needed to determine the mechanism by which the changes in the gut microbiome lead to neurological and metabolic health defects.

PHthalATES AND THE GUT MICROBIOME

Phthalates are plasticizers and stabilizers found in vinyl flooring, clothing, detergents, personal care products, children's toys, medical equipment, and plastic packaging film (Babich and Osterhout, 2010; Wittassek et al., 2011). Phthalates are commonly found in the environment as phthalate vapors or dust particles, but they are also found in food and beverages. Because phthalates are noncovalently bound to materials, they can easily leach into the environment. As a result, foodstuff is the primary source of phthalate exposure. Besides ingestion, other routes of exposure to phthalates include inhalation, dermal absorption, and intravenous exposure.

Recent studies in several species indicate that developmental phthalate exposure alters gut microbiota composition and may have significant health consequences (Table 2). Humans exposed to high levels of phthalates at birth (commonly through intravenous infusions) have altered gut microbiota, and these changes have been associated with enhanced immunoglobulin M responses against hepatitis B vaccine vaccination, suggesting that phthalate-induced gut microbial changes in early life may alter immune responses to vaccination (Yang et al., 2019b). Mono-(2-ethylhexyl) phthalate (MEHP) exposure in pubertal mice shifts the gut microbiota composition, and this

Table 1. Bisphenols and the Gut Microbiome

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|----------|--|--|------------------|-----------------------|---|---|--|
| BPA | Developmental | 0, 0.2, 0.6, 1.7, 2.9, 5.7, 11.5, 23.0, or 45.0 μ M in water environment | Zebrafish embryo | Not applicable | \uparrow <i>Chromatiaaceae</i> \downarrow <i>Neisseriaceae</i> (-) <i>Rheinheimera</i> , <i>Pseudomonas</i> , <i>Leptothrix</i> \downarrow <i>Bifidobacterium</i> \uparrow <i>Bacteroidetes</i> | <ul style="list-style-type: none"> High variability between vehicle control groups in each experiment BPA and BPF caused similar microbial community changes BPA acted as an obesogen, induced inflammation, and altered immune responses Supports link between alterations in the gut microbiome and metabolic disease BPA exposure induced defects in fecal antimicrobial activity and impaired protection of the gut BPA treatment altered the gut-associated immune system and systemic immune response BPA-induced colonic and liver inflammation BPA altered the colonic metabolome | Catron et al. (2019) Malaise et al. (2017) Malaise et al. (2018) |
| | 50 μ g/kg bw dams dosed orally from GD 15 to weaning | Mice | Male offspring | Female offspring | Not applicable | | |
| | 50 μ g/kg bw dams dosed orally from GD 15 to weaning | Mice | | | | | |
| | 200 μ g/kg bw dams dosed orally from GD 15 to PND 7 | Rabbits | Male offspring | | \downarrow <i>Bacteroidetes</i> \downarrow <i>Ruminococcaceae</i> \downarrow <i>Oscillospira</i> (dams) | | Reddivari et al. (2017) |
| | 50 mg/kg feed weight (approximately 10mg/kg bw) F0 dams exposed through chow 2 weeks before mating through PND 30 (weaning), F0 males exposed breeding through weaning | Mice | Both; yes | | \uparrow <i>Mogibacteriaceae</i> , <i>Sutterella</i> spp., and <i>Clostridiales</i> in F0 females \uparrow <i>Mollicutes</i> and <i>Prevotellaceae</i> compared in F0 males \uparrow <i>Bifidobacterium</i> , <i>Mogibacteriaceae</i> in F1 females \uparrow <i>Akkermansi</i> , <i>Methanobrevibacter</i> in F1 males \uparrow <i>Verrucomicrobia</i> \downarrow <i>Nitrosipra</i> , <i>OD1</i> , <i>AD3</i> , and <i>Gemmataimonadetes</i> | <ul style="list-style-type: none"> BPA exposure causes some of the same effects on the gut as ethynodiol (EE) exposure and some unique effects compared with EE Clear sex differences in the effects of BPA were evident in both generations BPA-mediated changes in the gut microbiota were inversely related to specific amino acid, lipid, and xenobiotic metabolism/degradation in females BPA-mediated changes in the gut were positively associated with changes in sulfur metabolism, insulin signaling pathway, and steroid hormone biosynthesis in males | Javurek et al. (2016) |
| | 30 μ g/kg BW gavaged from PND 28 to 56 | Nonobese diabetic mice | Female | | | | Xu et al. (2019) |
| Adult | 0, 2, and 20 μ g/L and 2 or 20 μ g/L BPA + 100 μ g/L nano-TiO ₂ in water environment for 3 months | Zebrafish | Both; yes | | \uparrow <i>Actinobacteria</i> at 2 μ g/L \uparrow <i>Lawsonia</i> \downarrow <i>Hyphomicrobium</i> | <ul style="list-style-type: none"> Coexposure to TiO₂ antagonized BPA effects at the low dose and synergized at the high dose Similar microbiome effects of BPA in both sexes Serotonin levels in the gut were decreased in BPA-exposed males | Chen et al. (2018b) |

Table 1. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---|--|---------------------------------|---|---|--|--|-----------------------|
| 12–18 µg/kg feed weight in dog food contaminated with BPA containing, 14 days | Dogs, gonadectomized | Both; yes but small n | ↓ <i>Bacteroides</i> spp., <i>Streptophyta</i> , <i>Erysipelotrichaceae</i> , and <i>Flexispira</i> spp ↑ <i>Bacteroides ovatus</i> , [<i>Prevotella</i> spp.], [<i>Ruminococcus</i> spp.], and <i>Cetobacterium somerae</i> | • Changes in gut microbes related to oxidative stress • Exposure to BPA led to an increase in serum BPA levels | • Some diet by sex interactions observed, but hard to interpret with small sample size • Bacteria known to metabolize bisphenols were suppressed with BPA treatment | • Exposure to BPA led to an increase in serum BPA levels | Koestel et al. (2017) |
| 2000 µg/L in water environment for 5 weeks | Zebrafish | Male, age not specified | BPA altered microbial community; no analysis of statistical significance between groups | Not applicable | • BPA exposure caused similar changes in the gut as EE • Untreated male and female zebrafish have similar gut microbiota | • BPA treatment decreased cell viability, caused oxidative damage due to ROS accumulation, disrupted mitochondrial function, and promoted apoptosis | Liu et al. (2017) |
| In vitro | 0–400 µM for 24 h | HCT116 human colon cancer cells | Not applicable | BPA ↓ microbial community richness at low doses and increased it at high dose; no analysis of statistical significance between groups ↑ <i>Microbacterium</i> and <i>Alcaligenes</i> | • BPA exposure increased genes related to oxidative stress and altered expression of estrogen receptors • BPA is metabolized in the system | • BPA treatment decreased cell viability, caused oxidative damage due to ROS accumulation, disrupted mitochondrial function, and promoted apoptosis | Qu et al. (2018) |
| 25, 250, and 2500 µg/L in nutritional medium for 10 days | In vitro Simulator of the Human Intestinal Microbial Ecosystem | Not applicable | Not applicable | Not applicable | • BPA exposure increased genes related to oxidative stress and altered expression of estrogen receptors • BPA is metabolized in the system | • BPA treatment decreased cell viability, caused oxidative damage due to ROS accumulation, disrupted mitochondrial function, and promoted apoptosis | Wang et al. (2018b) |
| 0–400 µM for 24 h | LS174T human colonic goblet cells | Not applicable | Not applicable | Not applicable | • BPA affected the secretory function of intestinal goblet cells by inducing mitochondrial dysfunction, oxidative stress, and apoptosis | • Exposure to the least developmentally toxic bisphenols impacted microbiota the most (BPS, BPA, BPF), the most developmentally toxic bisphenols did not impact microbiota (BPB, BPAF) | Zhao et al. (2018) |
| Other bisphenols | Development | Zebrafish embryo | Not applicable | ↓ <i>Neisseriaceae</i> (BPS) ↑ <i>Cryomorphaceae</i> (BPS) ↑ <i>Chromatiaceae</i> (BPF) ↓ <i>Neisseriaceae</i> (BPF) | • Similar microbial changes for BPA and BPF exposure, unique profile for BPS exposure • High variability between vehicle control groups in each experiment | • Exposure to the least developmentally toxic bisphenols impacted microbiota the most (BPS, BPA, BPF), the most developmentally toxic bisphenols did not impact microbiota (BPB, BPAF) | Catron et al. (2019) |

Table 2. Phthalates and the Gut Microbiome

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|-------------------------------|---|--|--|---|--|---|----------------------|
| DEHP | Birth | Newborns who did not receive intravenous infusion (control) vs newborns with respiratory distress who were given intravenous infusions | Newborn infants | Both sexes were used; however, sex differences were not assessed as both sexes were combined for analyses | ↓ <i>Rothia</i> spp. and <i>Bifidobacterium longum</i> | • DEHP exposure at medically relevant doses altered the gut microbiota and decreased β -diversity, which may change immune responses in later life | Yang et al. (2019b) |
| DEP, methylparaben, triclosan | From birth to PND 62 and 191 | 0.1735, 0.105, and 0.05 mg/kg/day | Sprague Dawley rats | Female | ↑ Bacteroidetes (<i>Prevotella</i>), Elusimicrobia in PND 62 ↓ Firmicutes (<i>Bacilli</i>) in PND 62 | • Postnatal exposure to environmentally relevant doses of the chemicals from birth to adolescence (PND 62) in rats modified the gut microbiota | Hu et al. (2016) |
| MEHP | 4-week exposure starting at puberty (4 weeks old) | 0 or 0.05 mg/kg MEHP with normal diet or high-fat diet | 4-week-old C57BL/6 mice | Male | ↑ Firmicutes; <i>Intestinimonas</i> , <i>Hollemaniella</i> , <i>Coprobacter</i> , and <i>Parasutterella</i> ↓ <i>Verrucomicrobia</i> ; <i>Akkermansia</i> , <i>Tannerella</i> , and <i>Alloprevotella</i> genus | • Microbiota differences diminished by adulthood (PND 181) • Postnatal exposure to 0.05 mg/kg MEHP altered gut microbiota and induced adipocyte hypertrophy and cholesterol overloading, deposition, and transportation | Wang et al. (2019a) |
| DEHP | Pubertal exposure for 7 and 14 days | 0, 1, or 10 mg/kg/day (oral gavage every other day) 0, 10, or 100 μ M DEHP in cecal contents (in vitro) | 6- to 8-week-old C57BL/6 mice | Female | ↑ <i>Lachnospiridium</i> (in vivo) ↓ <i>Akkermansia</i> , <i>Odoribacter</i> , and <i>Clostridium</i> sensu stricto (in vivo) ↑ <i>Alisipes</i> , <i>Paenibacillus</i> , and <i>Lachnospiridium</i> (in vitro cultured cecal microbiota) | • Decreased fecal α -diversity in control compared with 7-day-DEHP group • Increased production of <i>p</i> -cresol in 7-day-DEHP treatment groups (10 and 100 μ M) compared with control • Decreased butyrate synthesis in 7-day exposure of 100 μ M DEHP | Lei et al. (2019) |
| DMP, DEP, and DBP combined | Not Available | 500 mg/l | Asian carp species silver (Hypophthalmichthys molitrix) and bighead (Hypophthalmichthys Nobilis) | Not applicable | Achromobacter <i>aegifaciens</i> , <i>Pseudomonas japonica</i> , <i>Bacillus subtilis</i> , and <i>Pseudomonas putida</i> isolated and enriched | • Isolated bacteria found in Asian carp eliminated phthalates from experimental systems | Kolber et al. (2019) |

Abbreviations: DBP, dibutyl phthalate; DEHP, di-2-ethylhexyl phthalate; DEP, diethyl phthalate; DMP, dimethyl phthalate; MEHP, mono-(2-ethylhexyl) phthalate.

was accompanied by adipocyte hypertrophy and cholesterol overloading (Wang et al., 2019a), implying a causal relationship between MEHP-induced gut microbiota changes and impaired lipid metabolism. Subchronic exposure to di-2-ethylhexyl phthalate (DEHP) in female mice enhanced the production of *p*-cresol, which inhibits the synthesis of butyrate (Lei et al., 2019). Given that butyrate is a microbial metabolite essential for intestinal homeostasis (Parada Venegas et al., 2019), immune regulation (Correa-Oliveira et al., 2016), and neurological function (Bourassa et al., 2016), exposure to DEHP may lead to intestinal defects, immune dysfunction, and neurobehavioral impairments. The known effects of phthalates on gut microbiota in humans, rodents, and aquatic fishes are summarized in **Table 2**. Future studies should continue to examine whether phthalate exposure alters the gut microbiome via microbiome sequencing. In addition to sequencing studies, future studies should determine which gut microbes degrade phthalates and inoculate a single strain of bacteria into gnotobiotic animals to determine the function of that microbe.

POPs AND THE GUT MICROBIOME

Persistent organic pollutants are organic chemicals that contain hydrophobic and lipophilic properties and can be carried long distances through the air and water. Because these toxic chemicals are resistant to environmental degradation, POPs and their by-products can biomagnify and bioaccumulate in the ecosystem. Few POPs occur naturally in the environment. Instead, most POPs are synthesized for agricultural and industrial use, such as pesticides/insecticides, fast-food packaging, nonstick coatings for cookware, and heat exchange fluids.

Accumulating evidence indicates that POPs are toxic to human health and are linked to reproductive disorders, neurobehavioral impairments, and immune dysfunction (Li et al., 2006; Siwen et al., 2013). With increasing recognition of the roles of the gut microbiota on the immune system and knowing that the gut is the largest immune organ, recent studies have begun to investigate the impact of POP exposure on the gut microbiota and the potential roles it may have on health and disease. Specifically, these studies have investigated the impact of POP exposure on the gut microbiota during the developmental, juvenile, and adult stages in a variety of animals, including mice, fish, and humans (**Table 3**). In this section, we discuss highly studied POPs, including polychlorinated biphenyls (PCBs), polycyclic aromatic hydrocarbons (PAHs), perfluorochemicals (PFCs), and polybrominated diphenyl ethers (PBDEs), polychlorinated dibenz-p-dioxins (PCDDs), and the impact these chemicals have on the gut microbiome.

Polychlorinated Biphenyls

Polychlorinated biphenyls were commercially used in thermal insulation materials, floor finishes, and electrical equipment before being banned in 1979 in the United States. However, PCBs persist in the environment and contaminate the food supply. Therefore, a primary route of exposure is by ingestion of PCBs. Developmental exposure to PCBs in the maternal diet significantly increased gut permeability and immune factors (*Il6*) in the murine ileum compared with control (Rude et al., 2019). Gene expression of inflammatory markers (*Il6*, *IL1 β* , and *IL22*) in the colon was also increased in PCB-exposed mice compared with control (Rude et al., 2019). In addition to these functional changes in gastrointestinal (GI) physiology and immunology, PCB exposure markedly decreased the relative abundance of

Proteobacteria, Bacteroidales family S7-25, and *Alistipes* compared with control (Rude et al., 2019). These findings support that PCB exposure impacts GI physiology, modifies the gut microbiota composition, and alters mucosal immune responses.

In adult male mice, exposure to PCB-126 reduced cecal α -diversity and increased the Firmicutes to Bacteroidetes ratio (F/B ratio) (Petriello et al., 2018). High F/B ratio has been observed in adult, obese populations in mice (Ley et al., 2005) and humans (Ley et al., 2006), implying that PCB-126 exposure may increase the risk of developing obesity. However, associations between the gut microbiome, specifically F/B ratio, and obesity are uncertain as Sze and collaborators revisited and reanalyzed 10 studies and determined that the association between F/B ratio and obesity was relatively weak (Sze and Schloss, 2016). Upon further analysis, the studies included in their analyses had small sample sizes and large interpersonal variation (Sze and Schloss, 2016). PCB exposure also increased intestinal inflammation as demonstrated by increased *hepcidin* (*Hemp*) and *Tnfa* in the colon and *Il-6* and *Il-18* in the jejunum (Petriello et al., 2018). Furthermore, research in fish indicates that PCB-induced intestinal microbiota changes involve aryl hydrocarbon receptor (AhR) signaling (Chen et al., 2018b; Gao et al., 2018b; Ji and Qu, 2019).

Interestingly, a few studies suggest that the effect of PCBs on gut microbiota may be age-dependent. Acute PCB exposure in aging mice (11–13 months) decreased the relative abundance of Proteobacteria phyla (Choi et al., 2013). This is in contrast to the effects of PCB exposure during development exposure through the maternal diet, which increased Proteobacteria (Rude et al., 2019). It is important to note that the exposure methods in the studies were slightly different, which is a limitation. However, it is known that the gut microbiome changes with age (Cresci and Bawden, 2015).

Recently, a study identified that it is possible to attenuate PCB-induced gut microbiota changes in mice through exercise (Choi et al., 2013). However, more studies are needed to address and provide solutions on how to attenuate or prevent PCB-induced gut microbiota changes that are associated with increased gut permeability, intestinal inflammation, and cognitive dysfunction.

Polycyclic Aromatic Hydrocarbons

Polycyclic aromatic hydrocarbons (PAHs) are naturally found in the environment and are generated when burning wood, oil, and gas. Compared with PCBs, not many studies have investigated the impact of PAHs on the gut microbiota. One study used the *in vitro* SHIME system to show that PAH exposure in the colonic microbiome biotransforms PAHs to acquire more estrogenic activity than the parent compounds (Van de Wiele et al., 2005). Although these estrogenic metabolites have similar estrogenicity equivalent to 0.31–2.70 nmol 17 α -ethynodiol (EE2), further examination needs to be conducted on the actual PAH metabolites. A study in Atlantic cod exposed to PAH-contaminated waters showed different microbial communities compared with Atlantic cod residing in uncontaminated waters. PAH-exposed Atlantic cod contains PAH-degrading bacteria such as *Novosphingobium*, *Sphingobium*, and *Sphingomonas*, which are not present in unexposed Atlantic cod (Walter et al., 2019). More studies are needed in models such as mice or rats to determine if these changes also occur in mammals and how we can attenuate the changes in the gut microbiome.

Table 3. Persistent Organic Pollutants and the Gut Microbiome

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|---|--|--|---|---|---|-------------------------|
| PCB mixture | Developmental 0, 0.1, 1, or 6 mg/kg/day starting 2 weeks before gestation and continuing through PND 21 | Congenic wildtype (WT) (75% C57BL/6 and 25% Sv129) | No sex effects were observed, so all the data are combined and represented as equal proportions of males and females | ↑ Proteobacteria phylum, specifically within the Deltaproteobacteria and Betaproteobacteria class | • Increased gut permeability in juvenile double mutant (DM) and WT mice • Genes of some inflammatory markers were significantly altered in treatment groups (regIIIgamma) of the same genotype. | Rude et al. (2019) | |
| PCB-126 | Adult, 7 weeks old | 0 or 1 μmol/kg/day at weeks 2 and 4 | Ldhr ^{-/-} mice | Male | ↑ Diversity in cecum ↑ Firmicutes to Bacteroidetes ratio (F/B ratio) | • Gut microbiota were altered significantly when comparing WT and DM mice. However, PCB exposure also altered gut microbiota in WT mice when compared with control | Petriello et al. (2018) |
| Atrazine, estradiol, Adult (4 months) | PCB126, PCB153 | DMSO or 1.0 μg/l of environmental pollutant mixture | Males and females | ↑ Aeromonas in females Sex-specific xenobiotic responses | • PCB126 exposure altered the gut microbiota and host metabolism • PCB exposure increased Aeromonas in females, but not males. The increase in Aeromonas was also positively correlated with oxidative damage • <i>Histophilus</i> , <i>Mannheimia</i> , and <i>Blastococcus</i> were positively correlated with the integrity of the intestinal epithelial barrier • PCB exposure in males, but not females significantly decreased serotonin levels and tight junction protein 2 compared with control | Chen et al. (2018b) | |
| PCB congeners (PCB153, PCB138, PCB180) | Adults 11–13 months of age | Exercise for 5 days followed by 2 days of oral exposure to PCB mixture (150 μmol/kg) | C57Bl/6j mice | Males | ↓ Proteobacteria (<i>Pseudomonas plegmoglossicida</i> strain CCUGCC 2093, <i>P. plegmoglossicida</i> strain R18, <i>Pseudomonas putida</i> strain SRI156) | • PCB exposure altered gut microbiome, but voluntary exercise attenuated PCB-induced changes in gut microbiome • PCB exposure decreased the abundance of 1133 bacterial taxa and increased the | Choi et al. (2013) |

Table 3. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---|--|--|-----------------------------|---|--|---|----------------------------|
| PAH parent compounds: naphthalene, phenanthrene, pyrene, and benzo(a)pyrene | In vitro | 62.5 nmol | In vitro SHIME | Not applicable | and Firmicutes (<i>Streptococcus infantis</i>) phyla ↑ <i>Candidatus aquaresitis calciphila</i> (fold change 1.9), <i>Staphylococcus epidermidis</i> (fold change 1.7), <i>Bacteroides thetaiotomicron</i> strain 8669 (fold change 1.6), <i>Tropheryma whipplei</i> (fold change 1.5), <i>Corynebacteriaceae</i> , <i>Verrucomicrobiaceae</i> , <i>Uvhophyceae</i> , <i>Porphyrimonadaceae</i> | abundance of 90 taxa compared with control | Van de Wiele et al. (2005) |
| PAH | Information not available | No exposure to oil (northern Norway area) or 0.01 ppm exposure to oil (southern Norway area) | Atlantic cod (Gadus Morhua) | Not applicable | Clean water fish: Fusobacteria and Proteobacteria were the most abundant, followed by Firmicutes and Bacteroidetes. Fish exposed to oil-contaminated waters showed dominance in Firmicutes, followed by Proteobacteria, Bacteroidetes, and Fusobacteria. Vibrionales was the most abundant order identified | Fish guts from a clean area had a lower microbial community richness compared with the fish guts from oil contaminated waters | Walter et al. (2019) |
| PFBS | In utero exposure (eggs exposed to PFBS) | 0, 1.0, 2.9, 9.5 µg/l for the Marine medaka entire lifecycle | Males and females | Greater bioaccumulation of PFBS in males than females at all doses (1.0, 2.9, and 9.5 µg/l) | Intestines of F0-exposed males: ↑ <i>Cetobacterium</i> (2.9 µg/l) compared with control ↓ <i>Vibrio</i> in F1 females after exposure of F0 parents to 9.5 µg/1 PFBS Males had greater inflammatory responses than females Females exposed to PFBS had impaired lipid metabolism | PFBS exposure affected gut microbiota in fertile adults, and this persisted in the offspring • <i>Cetobacterium</i> was significantly and positively correlated with host TJP2 expression, a biomarker of epithelial barrier integrity | Chen et al. (2018a) |
| PFAS; F-53B | Adult, 6 weeks old | 0, 1, 3, or 10 µg/l for 10 weeks | C57Bl/6 mice | | ↓ Abundance of Firmicutes ↓ Verrucomicrobia after F-53B did not cause a sex-dependent effect | • MUC2 protein decreased in response to PFAS, but transcription increased in | Pan et al. (2019) |

Table 3. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|----------|-----------------------------------|--|---|--|---|--|------------------------|
| BDE-47 | In utero and lactational exposure | 0, 0.002, and 0.2 mg/kg BDE-47 from GD 6 to PND 21. Male mice selected to be either on normal diet or HFD for 14 weeks | Primigravida female ICR mice | Only male mice were selected for analysis | ↓ α -Diversity in fecal extracts ↑ In the relative abundances of <i>Parasutterella</i> and <i>Gemella</i> ↓ In the relative abundances of <i>Christensenellaceae</i> , <i>Atopostipes</i> , and <i>Bacillus</i> ↑ In relative abundances of <i>Candidatus saccharimonus</i> , <i>Ruminococcaceae</i> , <i>Staphylococcus</i> , <i>Gemella</i> , <i>Eubacterium</i> , <i>Corynebacterium</i> , <i>Paenibacillus</i> when compared with normal diet + vehicle group | • Exposure to BDE-47 reduced alpha diversity in fecal extracts, altered microbial composition, and impaired metabolic functions, including impaired glucose homeostasis, hepatic steatosis, and injury | Wang et al. (2018a) |
| OBS | BDE-47 or BDE-99 | Adult, 9 weeks old | 0 or 100 μ mol/kg/day for C57Bl/6 mice (conventional and germ-free) | Males | PBDE exposure altered 23 gut microbial taxa | • PBDE exposure in the intestinal microbiome decreased branched-chain and aromatic amino acid metabolites | Scoville et al. (2019) |
| | | Adult, 6 weeks old | 0, 0.1, 1, or 10 μ g/l | ICR mice | Males | • No microbiome data—but the authors reported decreased mucus secretion and gut barrier dysfunction in treatment groups compared with control. | Wang et al. (2019b) |
| HCH | Adult | N/A | Mothers-humans | Males and females were combined for analyses, so sex differences were not determined | • HCH altered microbial composition in human colostrum and the colonization of the infant gut | Tang et al. (2019) | |

Table 3. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|--------------------------|---|-------------------------------|-----------------------|---|---|-------------------------|
| Endosulfan | Adult, 8 weeks old | 0, 0.5, or 3.5 mg/kg/day for 2 weeks | ICR mice | Male | Microbial diversity at the genus level differed between samples (dose-dependent) Not applicable | • Endosulfan exposure-induced liver injury and disrupted amino acid, lipid, and gut microbiota metabolism • TCDD-induced host response was significantly modulated by the presence of SFB in the gut microbiome • TCDD exposure enriched ARG and MGE- harboring members of Enterobacteriaceae in the gut microbiome | Zhang et al. (2017) |
| TCDD | Juvenile, 4 weeks old | 0 or 30 µg/kg/day every 4 days for 28 days | C57Bl/6 Female | | ↑ Segmented filamentous bacteria (SFB). <i>Bacteroides fragilis</i> | Stedfeld et al. (2017b) | |
| TCDD | Juvenile, PND 28, 29 | 0–30 µg/kg/day for 28 and 92 days with the latter having a 30-day recovery period | C57Bl/6 mice Female | | ↑ Enterobacteriaceae | Stedfeld et al. (2017c) | |
| TCDD | Adult, 5–8 weeks old | 0, 0.1, 1.0, and 10 µg/kg/day for 4 days | B6C3F1 mice Female | | Expression of SFB in mouse ileum | • Addition of activated carbon decreased the bioavailability of TCDD in the host and may have influenced the gut microbiome • Dysregulated gut microbiome may have contributed to liver and immune toxicity | Stedfeld et al. (2017a) |
| TCDD | Adult, 6 and 7 weeks old | 0 or 6 µg/kg biweekly for 26 weeks | Male | | ↑ Firmicutes ↓ Bacteroidetes ↑ Lactobacillaceae and Desulfovibrionaceae ↓ Prevotellaceae and Actinobacteria cluster ACK M1 | Lefever et al. (2016) | |
| TCDD, PhIP, HBD, Bl[α]P, delta-methrin, and PAHs | Adults | 0.005, 0.90, 2.60, 5, 21, and 38 ng/ml | Human in vitro cultured feces | Not applicable | Not applicable | • Pollutant disturbance may have promoted inflammation with the release of IL-8 from intestinal epithelial cells | Defoix et al. (2018) |
| TCDF | Adult (8 weeks old) | 0 or 24 µg/kg/day for 5 days | C57Bl/6J mice | Males | ↓ F/B ratio ↑ In <i>Butyrivibrio</i> spp. and <i>Flavobacteria</i> ↓ In <i>Oscillibacter</i> and <i>Clostridia</i> ↓ SFB | Changes in microbiota were associated with marked increases in bile acids, SCFA, altered liver function, increased intestinal inflammation, and inhibited signaling of FXR, a key regulator of fat and glucose metabolism | Zhang et al. (2015a) |

Abbreviations: *n*-HRCD, *n*-hexabromocyclododecane; γ -HBCD, γ -hexabromocyclododecane; BDE-47, 2,2',4,4'-tetrabromodiphenyl ether; DMSO, dimethyl sulfoxide; FXR, farnesoid X receptor; HCH, hexachlorocyclohexane; OBS, sodium *p*-perfluororous nonoxybenzene sulfonate; PAH, polycyclic aromatic hydrocarbons; PBDE, polybrominated diphenyl ether; PCB, polychlorinated biphenyls; PFAS, per- and polyfluoroalkylated substances; PBS, perfluorobutane sulfonate; SCFA, short-chain fatty acid; SHIME, Simulator of the Human Intestinal Microbial Ecosystem; TCDD, 2,3,7,8-tetrachlorodibenzo-*p*-dioxin; TCDF, 2,3,7,8-tetrachloro dibenzofuran.

Perfluorochemicals

Perfluorochemicals are commercially used in nonstick cookware, food packaging, and stain-resistant carpets. Although little is known about the impact of perfluorochemicals on the gut microbiome, a few studies have examined per- and polyfluorooxylated substance (PFAS) and per- and polyfluorobutane substance (PFBS) exposure in utero and during adult stages of life (Table 3). Male fish exposed to PFBS (0, 1.0, 2.9, 9.5 µg/l) had an increased abundance of *Cetobacterium* compared with control male fish. The increase in *Cetobacterium* was positively correlated with altered tight junction protein expression, a marker of epithelial barrier integrity (Chen et al., 2018a). In addition to experiencing alteration of gut microbiota composition and expression of tight junction proteins, males had hyperactive inflammatory responses (Chen et al., 2018a). In contrast to males, PFBS-treated female fish only had slightly increased inflammatory responses, but had severely impaired lipid metabolism as demonstrated by their decreased levels of triglycerides and free fatty acids (Chen et al., 2018a). The alteration in gut microbiota in fertile adults persisted in the offspring. These offspring had increased mortality, especially at higher doses of exposure compared with controls (Chen et al., 2018a). These data suggest that PFC exposure alters the gut microbiome in a sex-dependent manner in fish, and this comes with sex-dependent health outcomes. In addition, the data suggest that gut microbiota alterations and the adverse effects on gut homeostasis, such as increased inflammation, upregulated oxidative stress, and a compromised intestinal epithelial barrier, can be carried over to the next generation.

Unfortunately, there are conflicting results on whether the results observed in fish are translatable to mammals such as mice. PFBS exposure in mouse adipose cells showed that triglyceride levels were increased compared with control (Qi et al., 2018). In contrast, PFC exposure, specifically perfluorooctanesulfonic acid (PFOS) exposure, increased triglyceride levels in mice, whereas PFC exposure, specifically PFBS, decreased triglyceride levels in fish (Wang et al., 2015). Major reasons for these conflicting results include the use of distinct models and species, different PFCs, and research performed outside a living organism versus within a living organism. Another limitation of the study on PFBS exposure is that fish were not exposed to environmentally relevant doses (Chen et al., 2018a). The study used doses in the µg/l range (0, 1, 2.9, and 9.5 µg/l) (Chen et al., 2018a). European adults are exposed to 0.03–3.72 ng/kg/day PFBS (EFSA, 2012; EPA, 2018), and Americans are exposed to roughly 4.2 ng/ml PFBS (Olsen et al., 2017). Although the study in fish did not use environmentally relevant doses, it is still beneficial to study high doses above relevant exposures because these chemicals can bioaccumulate in the body.

Chlorinated polyfluorinated ether sulfonate or F-53B/(6:2) is another perfluorochemical commonly used in electroplating wastewater in China (Siwen et al., 2013). F-53B exposure in adult mice decreased the abundance of Firmicutes and increased the abundance of Verrucomicrobia in both male and female mice (Pan et al., 2019). F-53B exposure also caused sex-specific responses in the gut microbiota. For example, treated male mice had a significant increase in *Akkermansia* compared with control male mice, but treatment did not cause any variation in female mice. Treated female mice had a significant decrease in *Parabacteroides* compared with control female mice, but this did not occur in male mice (Wang et al., 2019c). In addition to sex-specific gut microbial deviations, both male and female mice had significantly decreased mucus production and gene

expression of ion transporters at the highest dose of F-53B exposure (Pan et al., 2019). The potential link between these changes and gut microbiome alterations remain unclear. The different responses of the gut microbiome in males and females to F-53B exposure may explain sex differences in the effects of F-53B exposure on health outcomes.

Polybrominated Diphenyl Ethers

BDE-47 and BDE-99 are 2 PBDEs commonly used as flame retardants in building materials, furniture, airplanes, polyurethane foams, and electronics. Exposure to PBDEs in utero and adulthood perturbs the gut microbiome. Specifically, BDE-47 exposure in utero and during lactation reduced diversity of gut microbiota composition and impaired glucose homeostasis in male mice (Wang et al., 2018a). In adult male mice, BDE-47 and BDE-99 exposure altered 23 gut microbial taxa and amino acid and carbohydrate metabolism (Scoville et al., 2019). Specifically, BDE-treated mice had decreased branched-chain and aromatic amino acid metabolites compared with controls. Although these previous studies provide important information, they both used male mice only. It is important to study the female sex as well and compare that to the male sex because males and females have different hormone profiles, which can affect the impact of environmental chemicals on the entire body, including the gastrointestinal and reproductive tracts. Overall, these studies indicate that exposure to PBDEs can alter gut microbiome and impair metabolism. It would be interesting to know if PBDE-induced gut microbiome alterations mediate some aspects of the metabolic disorders in PBDEs-treated animal models.

Polychlorinated Dibenzo-p-Dioxins

Polychlorinated dibenzo-p-dioxins are naturally formed from volcanic eruptions and forest fires. However, human activities create the most dioxins by manufacturing pesticides and burning organic materials, such as fossil fuels and garbage. Exposure to dioxins can occur through inhalation, ingestion, and dermal absorption. Two dioxins that have been studied are 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) and 2,3,7,8-tetrachlorodibenzofuran (TCDF).

In a series of studies carried out by the same laboratory, conventional, and gnotobiotic female mice were orally dosed with vehicle control or TCDD to assess the effects of TCDD on the gut microbiota (Stedtfeld et al., 2017a,b,c). In the first study, conventional female mice treated with TCDD showed increased Enterobacteriaceae compared with control mice (Stedtfeld et al., 2017c). Gut microbiota alterations due to TCDD treatment were associated with an increase in the reservoir of antibiotic-resistant genes in the gut microbiome. In a second study, gnotobiotic mice of the same sex (female) were colonized with and without segmented filamentous bacteria (SFB) and polysaccharide A (Stedtfeld et al., 2017b). SFB is an immune activator, whereas polysaccharide A is a background immune suppressor. TCDD treatment increased SFB and decreased *Bacteroides fragilis* compared with control. Thus, the TCDD-induced host response was significantly impacted by the presence of SFB in the gut microbiome (Stedtfeld et al., 2017b). The addition of activated carbon decreased the bioavailability of TCDD and obliterated the expression of the SFB 16S rRNA gene (Stedtfeld et al., 2017a). Activated carbon had minimal influences on the murine gut microbiome and affected the relative abundance of *Lactobacillaceae* and 2 other minor groups. TCDD exposure had

drastic effects on the gut microbiome, and activated carbon attenuated the changes in the gut microbiome (Stedtfeld *et al.*, 2017a). Altogether, these studies indicate that TCDD causes alterations in the gut microbiota, which are associated with antibiotic resistance; however, the presence of activated carbon reduces TCDD exposure and possibly antibiotic resistance.

Chronic TCDD exposure in adult male mice also caused alterations in gut microbiota. The dysregulation in the gut microbiome may have contributed to immune toxicities, and this aligns with several studies that examined the effects of TCDD on immune responses. An *in vitro* study cultured human fecal suspensions with TCDD. Following TCDD-gut microbiota interaction, fecal suspensions were processed for its cellular supernatant and incubated with intestinal epithelial cells to measure immune responses (Defoix *et al.*, 2018). TCDF exposure increased IL-8 secretion from intestinal epithelial cells, suggesting that dioxin disturbance may promote inflammation and is immunotoxic (Zhang *et al.*, 2015a). Beyond associative changes in gut microbiome with the immune system, changes in the gut microbiota due to TCDD exposure link to liver toxicity as there were marked increases in bile acids and SCFAs and decreases in farnesoid X receptor signaling (Wang *et al.*, 2019b; Zhang *et al.*, 2017). Ultimately, these studies suggest that dioxin exposure disrupts both gut microbiome and lipid and glucose metabolism. Whether there is a causal relationship between dioxin-induced alterations in gut microbiome and metabolic disorder remains to be studied.

HEAVY METALS AND THE GUT MICROBIOME

Heavy metals are natural, high-density elements found in the Earth's crust and are highly toxic even at low concentrations. Like POPs, heavy metals bioaccumulate in some organisms and further increase the burden of heavy metals. Several studies have examined the impact of heavy metals, such as lead, cadmium, arsenic, and mercury on the gut microbiome (Table 4). In mice, lead exposure reduced diversity of gut microbiota and altered the metabolism of many pathways, including vitamin E, bile acid, and nitrogen metabolism (Gao *et al.*, 2017c).

Exposures to cadmium altered the gut microbiota in frogs (Ya *et al.*, 2019), fish (Chang *et al.*, 2019), and mice (Breton *et al.*, 2013, Zhang *et al.*, 2015b) (Table 4). However, whether there are sex-specific changes remains to be determined because previous *in vivo* studies either did not examine sex differences or used only male or female animals.

Exposure to arsenic and its impact on the gut microbiome have been studied in mice and humans. Arsenic exposure has been shown to alter the gut microbiota in various strains of mice during neonatal, juvenile, and adult stages (Table 4). These alterations were also associated with changes in the metabolism of lipids, vitamin E, bile acids, oxidative stress, and detoxification. Arsenic exposure also presents a problem in humans through ingestion of contaminated waters. An adult human study in a Bangladesh population reported that arsenic exposure altered the gut microbiome and resulted in an overproduction of genus *Citrobacter* (Wu *et al.*, 2019). Depending on the species, *Citrobacter* can result in various health problems such as urinary tract infections, respiratory diseases, inflamed gastrointestinal tract, and sepsis in immunocompromised individuals. Analysis showed a significant association between arsenic exposure in water, *Citrobacter*, and vascular intima-media thickness (IMT). IMT is accessed by measuring the distance of the luminal-intima to the medial-adventitia of the carotid artery by ultrasound. High IMT increases risk of developing

cardiovascular diseases, specifically atherosclerosis. Ultimately, IMT is used as a subclinical marker for atherosclerosis (Simova, 2015). Thus, the data on arsenic exposure suggest that the altered gut microbiome may play an essential role in the development of atherosclerosis, possibly through a microbiota-derived metabolite called trimethylamine N-oxide (TMAO) (Jonsson and Bäckhed, 2017; Lindskog Jonsson *et al.*, 2018). TMAO disrupts bile acid synthesis and metabolism, which is associated with atherosclerosis due to the buildup of cholesterol (Ding *et al.*, 2018). Overall, having a stable microbiome, which includes the presence of *Faecalibacterium*, may reduce arsenic toxicity and arsenic-related diseases (Coryell *et al.*, 2018).

The effects of mercury exposure on the microbiome have been studied in various species including mice, rats, chickens, fish, and humans during the adult stage (Table 4). The forms of mercury included in this literature review are the following: methylmercury (MeHg), monomethylmercury (MMHg), and mercuric chloride (MgCl₂). Exposure to MeHg damaged the GI tract and altered the gut microbiota and its metabolites (Bridges *et al.*, 2018, Lin *et al.*, 2020). However, nutrition may provide some protection from MeHg toxicities. For example, supplementation of selenium reversed some changes in the gut microbiota of MeHg-poisoned rats so that the gut microbiota presented more like the control group (Liu *et al.*, 2019). Protein addition also seems to enhance MMHg metabolism *in vitro* (Guo *et al.*, 2018). Similar to MeHg, MgCl₂ damaged the GI tract and resulted in thickened muscle walls, decreased goblet cells, widened submucosa, and necrotic enterocytes in the cecum of mice and chickens (Ruan *et al.*, 2019; Zhou *et al.*, 2020). In addition to changes in the morphology of the cecum and cecal microbiome, MgCl₂ exposure was associated with altered metabolism of vitamin C, aldarate, xenobiotics, and PAHs (Zhou *et al.*, 2020).

Although a myriad of studies have examined the effects of heavy metals on the gut microbiome, many of these microbiome studies only reported descriptive changes in the makeup of the gut microbiota and used only 1 type of sex or combined both sexes in their analyses (Table 4). In the future, it is important to conduct research beyond characterizing the gut microbiota to evaluate the effects of heavy metals on other aspects of health, such as metabolism and reproduction. Experiments that are designed to investigate the sex-dependent effects and multigenerational or transgenerational effects as a result of heavy metal exposure are especially important.

PESTICIDES AND THE GUT MICROBIOME

Pesticides, which include insecticides, herbicides, and fungicides, are used globally to aid in food production. Although pesticides are generally highly toxic to target species through mechanisms that may not affect humans and most nontarget species, many pesticides are toxic to humans and wildlife through other mechanisms, including endocrine disruption (Mnif *et al.*, 2011). Recent studies on environmentally relevant doses of pesticides have revealed the long-term consequences of the environmental metabolites of legacy pesticides (Liang *et al.*, 2019; Liu *et al.*, 2016) as well as the harms of their replacements, especially on nontarget organ systems including the gut. As scientists are beginning to understand the complex functions of the gut microbiome, the literature on the effects of pesticides on gut microbiota has significantly expanded in the past 3 years (Table 5).

Both pure glyphosate and various glyphosate-based herbicide (GBH) formulations alter the bacterial makeup of the gut microbiome in rodents (Aitbali *et al.*, 2018; Dechartres *et al.*,

Table 4. Heavy Metals and the Gut Microbiome

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|----------|------------------------|---|--------------------------------|--|---|---|-----------------------|
| Lead | Perinatal | 32 ppm until 40 weeks of age in maternal drinking water | Mice | Male and female Pb exposure was highly correlated with increased adult body weight in male, but not female offspring | Bacteroidetes and Firmicutes shifted in response to Pb exposure | • Lead exposure impacted the gut microbiota | Wu et al. (2016) |
| Lead | Juvenile (6 weeks old) | 100 and 500 ppm for 8 weeks in drinking water | Balb/c mice | Female | ↓ Lachnospiraceae ↑ Lactobacillaceae and Erysipelotrichaceae Within Erysipelotrichaceae family, <i>Turicibacter</i> , <i>Blautia</i> , <i>Barnesiella</i> , and <i>Alistipes</i> were higher in treatment group compared with control | • Lead exposure impacted the gut microbiota | Bretton et al. (2013) |
| Lead | Adult, 7 weeks old | 10 ppm for 13 weeks in drinking water | C57BL/6 mice | Female | ↓ Genus diversity in treatment group than control Lead exposure reduced phylogenetic diversity ↓ Clostridiales, Lachnospiraceae, and Ruminococcaceae | • Lead exposure altered the gut microbiome and significantly affected vitamin E, bile acid, nitrogen metabolism, oxidative stress, and detoxification mechanisms | Gao et al. (2017c) |
| Cadmium | Larval stage | 0, 5, 100, and 200 µg/1 from Gosner stage 26 to 38 | <i>Bufo gargarizans</i> larvae | Not specified | ↑ Relative abundance of Proteobacteria ↓ Relative abundance of Bacteroidetes and Firmicutes | • Chronic cadmium exposure caused significant intestinal damage and gut microbiota changes | Ya et al. (2019) |
| Cadmium | Juvenile | 0, 50, and 500 µg Cd/1 in drinking water for 4 weeks | <i>Cyprinus carpio</i> | Not applicable | ↓ Fusobacteria, <i>Cetobacterium</i> , and <i>Akkermansia muciniphila</i> ↑ Firmicutes | • Cd exposure decreased diversity of gut microbiota | Chang et al. (2019) |
| Cadmium | Juvenile (5 weeks old) | 10 ppm in drinking water for 10 weeks | C57BL/6 mice | Male | ↓ Relative abundance of Firmicutes and γ-Proteobacteria Quantity of Firmicutes decreased significantly in the cecum contents after 10 weeks of Cd exposure | • Cd increased liver TG, FFA, and TG levels; mRNA of key genes in de novo FFA synthesis and transport pathways and TG synthesis in mice | Zhang et al. (2015b) |
| Cadmium | Juvenile (6 weeks old) | 20 and 100 ppm for 8 weeks in drinking water | Balb/c mice | Female | ↓ Lachnospiraceae ↑ Lactobacillaceae ↑ Erysipelotrichaceae → <i>Turicibacter</i> , <i>Blautia</i> , <i>Barnesiella</i> , and <i>Alistipes</i> were higher in heavy metal-treated groups than in control | • Cd exposure dysregulated Cd exposure and altered the gut microbiome in mice • Heavy metal exposure altered gut microbiota and may interfere with gut homeostasis | Bretton et al. (2013) |

Table 4. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|------------------|-----------------------------------|---|--|-----------------------|--|--|-----------------------|
| Arsenic | Neonatal (PND 10) | 0.05 mg/kg bw | CD-1 mice | Female | ↓ Bacteroidetes ↑ <i>Eubacterium plexicaudatum</i> , <i>Lachnospirillum</i> , and <i>Mucispirillum schaedleri</i> , Deferribacteres ↑ <i>Shewanella</i> , <i>Rheinheimera</i> , and <i>Bifidobacterium</i> ↓ <i>Herpetosiphonales</i> | • Arsenic exposure resulted in a distinct bacterial population and gut-associated immune status | Gokulan et al. (2018) |
| Arsenic and zinc | Juvenile | Zinc restriction and 0, 50, 500 ppb inorganic arsenic in drinking water for 6 weeks | C57BL/6 mice | Female | | • Zinc deficiency was associated with reductions in the host's and microbiome's ability to detoxify arsenic | Gaulke et al. (2018) |
| Arsenic | Juvenile | 100 ppb for 13 weeks | C57BL/6 mice | Female | ↑ Verrucomicrobia ↓ Firmicutes | • Arsenic exposure altered the gut microbiome and its ability to carry out carbohydrate and pyruvate metabolism | Chi et al. (2017) |
| | | | | | | • Arsenic exposure increased oxidative stress response genes; increased vitamin biosynthesis genes; increased folic acid synthesis; altered LPS-related genes | Guo et al. (2014) |
| | | | | | | • Exposure to arsenicaltered the gut microbiota and the metabolic profiles of the mouse | Li et al. (2019b) |
| Arsenic | Juvenile (5 weeks old) | 3 mg/l As, 5 mg/l Fe + 3 mg/l As, and 5 mg/l Fe | ICR mice | Male | ↑ Firmicutes, Proteobacteria, Acidobacteria, Cyanobacteria ↓ Bacteroidetes and <i>Saccharibacteria</i> (TM7) | • Arsenic exposure was associated with bile acid molecular families | Dheer et al. (2015) |
| Arsenic | Juvenile (5 weeks old) | 50 ppm in drinking water for 2 weeks | C57BL/6 mice | Not specified | ↓ Gut microbiota diversity but not significant | • Arsenic exposure altered the gut microbiome and was linked to changes in amino acid metabolism | |
| Arsenic | Juvenile to adult (6–8 weeks old) | 0, 10, or 250 ppb arsenite As(III) for 2, 5, or 10 weeks | C57BL/6 Tac mice | Male | ↑ Bacteroidetes, <i>Clostridia</i> ↓ Firmicutes, <i>Bacteroides</i> | | |
| Arsenic | Adult, 7 weeks old | 0, 250 ppb, and 1 ppm in drinking water | C57BL/6 mice (with normal or disrupted gut microbiome) | Female | Not applicable | • Arsenic-induced gut alterations altered arsenic biotransformation and increased toxicity | Chi et al. (2019) |
| Arsenic | Adult, 8 weeks old | 0 or 10 ppm for 4 weeks | C57BL/6 mice | Female | ↓ Cyanobacteria, <i>Tenericutes</i> ↑/↓ Firmicutes depending on class | • Arsenic exposure altered the gut microbiota | Lu et al. (2014a) |
| | | | | | | • Metagenomic sequencing revealed that changes in gut microbiota were strongly associated with changes in bile acids, amino acid derivatives, lipids, fatty acids, isoflavones, indole derivatives, glucuronide, and carnitine | |

Table 4. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|----------------------|---------------------------|---|--|--|---|--|-----------------------|
| Arsenic | Adult, 9 and 10 weeks old | 0 or 10 ppm in drinking water for 4 weeks | C57BL/6 mice (free of Helicobacter) | Not specified | Not applicable | <ul style="list-style-type: none"> • Arsenic exposure altered phospholipid, cholesterol, and tryptophan metabolism in the gut • Arsenic upregulated several unsaturated fatty acids • Significantly different gut microbiome phenotypes may have had an impact on arsenic biotransformation | Xue et al. (2019) |
| Arsenic | Adult, 8 weeks old | 0 or 10 ppm in drinking water for 4 weeks | C57BL/6 mice, WT and IL10-/- | Females | ↑ Bacteroidetes ↓ Firmicutes | <ul style="list-style-type: none"> • Nitrogen, carbon, and sulfate metabolism were significantly altered in arsenic-treated males • The gut microbiome is distinct in males and females as well as treated and nontreated mice | Chi et al. (2016) |
| Arsenic | Adult, 8 weeks old | 10 ppm in drinking water for 4 weeks | C57BL/6 mice | Male and female Microbiome of female mice treated with As was more significantly affected than male mice treated with As | Female-treated mice: ↓ Dorea ↑ Akkermansia These changes not observed in males | <ul style="list-style-type: none"> • Stable human gut microbiome transplanted in mice protected against arsenic-induced mortality | Chi et al. (2016) |
| Arsenic | Adult, 7–13 weeks old | 0, 25, or 100 ppm inorganic arsenic in drinking water for 2 weeks | Germ-free C57BL/6 WT and AS3MT-KO mice | Male and female were combined for the analyses; therefore, sex differences were not determined | Presence of <i>Faecalibacterium prausnitzii</i> suggest this microbe provides protection by decreasing arsenic toxicity increasing butyrate synthesis | <ul style="list-style-type: none"> • As3mt and intact gut microbiome (along with the presence of <i>F. prausnitzii</i>) provided protection against arsenic toxicity in mice • Stable human gut microbiome transplanted in mice protected against arsenic-induced mortality | Coryell et al. (2018) |
| Arsenic | 25- to 50-year-old adults | Not applicable | Humans | Male and female adults were included in the study but sex differences were not determined | Did not identify associations between arsenic exposure and the composition of the gut microbiome after correcting for multiple testing | <ul style="list-style-type: none"> • Significant association of genus Citrobacter with IMT | Wu et al. (2019) |
| Methylmercury (MeHg) | Adult, 4 weeks old | 0 or 10 µg/kg bw (oral dosing) | Sprague Dawley rats | Male | ↓ Bacteroidetes and Proteobacteria; Lactobacillaceae, Bacteroidaceae, Streptococcaceae, and Sutterellaceae ↑ Firmicutes; Desulfovibrionaceae, Helicobacteraceae, Peptococcaceae, and Rhodospirillaceae | <ul style="list-style-type: none"> • MeHg exposure damaged intestinal villi and walls in the small and large intestines • MeHg exposure promoted inflammatory responses (increased IL-1β and IL-6; decreased IgE and BDNF) in treatment groups compared with control • MeHg exposure altered the metabolism of nucleotides, carbohydrates, amino acids, and lipids • MeHg altered the gut microbiome and metabolites | Lin et al. (2020) |

Table 4. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---------------|--------------------|---|---|---------------------------------------|---|---|--------------------------|
| Methylmercury | Pregnant mothers | Not applicable | Human (recruited from Greenville, South Carolina), n = 17 | Females | Microbial richness (α - and β -diversity) was not markedly changed for gut microbiota with high and low concentrations of mercury in hair and stool | <ul style="list-style-type: none"> • MeHg concentrations in stool likely do not reflect biotransformation by the gut microbiota • 17 bacterial genera were associated with Hg in stool and hair • Gut microbiome composition differed between high and low MeHg concentrations in blood in stool during early gestation • α-Diversity decreased with MeHg exposure in late gestation compared with early gestation • Several taxa were correlated with MeHg biomarkers, but associations of taxa and MeHg exposure differed depending on whether sample was collected during early or late gestation | Rothenberg et al. (2016) |
| Methylmercury | Pregnant mothers | Not applicable | Humans (recruited from only from South Caroline) | Females | Not applicable | <ul style="list-style-type: none"> • Gut microbiome composition differed between high and low MeHg concentrations in blood in stool during early gestation • α-Diversity decreased with MeHg exposure in late gestation compared with early gestation • Several taxa were correlated with MeHg biomarkers, but associations of taxa and MeHg exposure differed depending on whether sample was collected during early or late gestation | Rothenberg et al. (2019) |
| Methylmercury | Adult | Fathead minnow (FHM); control (0.02 $\mu\text{g/g}$, 0.72 $\mu\text{g/g}$, 5.50 $\mu\text{g/g}$ Hg dry weight for twice a day for 30 days Mice: control (0.02 $\mu\text{g/g}$, 0.43 $\mu\text{g/g}$, 4.39 $\mu\text{g/g}$ Hg dry weight in diet for twice a day for 30 days | FHM CD-1 mice | FHM: males and females Mice: males | <p>Dominant phyla in low MeHg group: Planctomycetes (<i>Pirellula</i> and <i>Planctomyces</i> genera), Fusobacteria (<i>Cetobacterium</i>) ↓ Cetobacterium in high MeHg group compared with control</p> <p>↑ <i>Aeromonas</i>, <i>Acinetobacter</i>, and an unclassified member of the Neisseriaceae family</p> | <ul style="list-style-type: none"> • MeHg exposure increased microbes associated with xeno-biotic metabolism and increased microbes related to immune suppression compared with control • FHM: increased putrescine, L-serine, and glycerol, decreased stearic acid, palmitic acid, oleic acid, and L-glutamine • Mice and FHM exposed to MeHg had decreased fatty acids in backbone of triglycerides. Gene expression of enzymes in were also dysregulated, indicating that fatty acid catabolism was altered | Bridges et al. (2018) |
| Methylmercury | Adult, 8 weeks old | 0 or 4 mg/kg bw MeHg by intragastric gavage for 4 weeks every other day | Sprague Dawley rats | Males | ↓ Cyanobacteria and Bacteroidetes in MeHg treatment groups compared with control; Firmicutes in MeHg + sodium selenite compared with MeHg group ↑ Firmicutes in MeHg treatment groups compared with control; Bacteroidetes in MeHg + sodium selenite compared with MeHg | <ul style="list-style-type: none"> • MeHg exposure altered the gut microbiota compared with control; however, sodium selenite supplementation reversed some of the changes in the gut microbiota, suggesting that selenium may reduce toxicity and improve intestinal health | Liu et al. (2019) |

Table 4. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--------------------------------|-------------------------|--|--|---|--|--|---------------------|
| Methylmercury | Adults, 18–80 years old | Subjects were asked to consume 3 fish meals per week for 2 weeks | Human (recruited from the University of Rochester Medical Center community, $n = 37$) | Males and females Sex differences not specified because males and females were combined for analyses | Not applicable | <ul style="list-style-type: none"> Elimination rates in subjects prescribed antibiotics ($n = 2$) showed significant decreases in MeHg compared with control subjects who did not take antibiotics, suggesting the gut microbiome plays a role in the metabolism of MeHg | Caito et al. (2018) |
| Monomethylmercury | Not applicable | 10 ng/g | In vitro experiment with human fecal samples ($n = 2$) | Not specified | Not applicable | <ul style="list-style-type: none"> Microbes, such as <i>Sutterella parvibrabra</i> and <i>Acidaminococcus intestinalis</i>, play a role in MMHg degradation, but the extent of MMHg metabolism is highly dependent on nutrients such as carbohydrates and proteins in vitro Compared with control, MMHg metabolism is enhanced with the addition of protein | Guo et al. (2018) |
| Mercuric chloride ($HgCl_2$) | Chicks, 1 day old | 0 or 150 ppm in drinking water for 30, 60, and 90 days | Hyline chicken | Male | <p>Day 30 results: ↑ Proteobacteria, Tenericutes</p> <p>Day 60 results: ↑ Tenericutes</p> <p>Day 90 results: ↓ Spirochaetes</p> | <ul style="list-style-type: none"> $HgCl_2$ exposure decreased growth performance and altered cecal and colonic morphology on days 60 and 90 compared with control Species richness increased in Hg groups compared with control after 30 days of exposure but decreased after 90 days of exposure in the cecum Functional profiling of microbial communities showed $HgCl_2$ exposure for 60 days increased genes related to ascorbate and aldarate metabolism, xenobiotic metabolism by cytochrome P450, and PAH degradation $HgCl_2$ exposure for 30 days may affect infectious bacterial diseases No significant functional changes in chickens exposed to $HgCl_2$ for 90 days | Zhou et al. (2020) |
| Mercuric chloride | Adult, 8 weeks old | 0 or 2 mg/kg bw $HgCl_2$ for 90 days | Kunming mice | Female | <p>↑ <i>Butyrimonas</i>, <i>Dehalobacterium</i>, <i>Corprococcus</i>, <i>Oscillospira</i>, and <i>Bilophila</i></p> <p>↓ <i>Sporosarcina</i>, <i>Jeotgalicoccus</i>, <i>Staphylococcus</i>, and <i>Acinetobacter</i></p> | <ul style="list-style-type: none"> $HgCl_2$ exposure decreased body weight, caused histopathological lesions in the cecum, and altered cecal microbiota | Ruan et al. (2019) |

Abbreviations: As, arsenic; Cd, cadmium; IMT, intima-media thickness; MeHg, methylmercury; MgCl₂, mercuric chloride; MMHg, monomethylmercury; PAH, polycyclic aromatic hydrocarbons; Pb, lead; WT, wildtype.

Table 5. Pesticides and the Gut Microbiome

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|-----------------|---|------------------------------|--|---|--|----------------------|
| Aldicarb (I) | Adult | 2 ppm (approximately 0.3 mg/kg bw/day) in drinking water for 13 weeks beginning at 8 weeks of age | C57BL/6 mice | Male | ↓ Christensenellaceae, Coriobacteriaceae, Bacillales, Anaerostipes, Roseburia ↑ Erysipelotrichaceae, Clostridium | <ul style="list-style-type: none"> Exposure to aldicarb increased pathogenicity of gut bacteria—increased genes for virulence, adhesion, and bacteriocins Lipid profile and lipid metabolism altered in feces of treated mice compared with control Aldicarb increased antioxidant gene expression, suggesting increased oxidative stress Aldicarb increased DNA damage response gene expression, suggesting increased DNA damage Brain metabolome/energy metabolism altered in response to aldicarb AMPA did not alter gut microbiome | Gao et al. (2019) |
| Aminomethylphosphonic acid (AMPA; glyphosate metabolite) | Adult | 1.5 or 7.5 mM in sugar syrup or cotreatment with equal amounts of AMPA and glyphosate for 15 days | Honeybee | Female | None | <ul style="list-style-type: none"> Atrazine exposure did not cause significant alterations in gut bacteria in tadpoles and adults Gut microbiome changes are not likely the route through which atrazine affects tolerance to amphibian fungal infections | Knutie et al. (2018) |
| Atrazine (H) | Developmental | 200 µg/l in water environment for 6 days or in ex vivo culture of whole gut for 6 days | Cuban tree frog | Male and female, but sex differences not discussed | <ul style="list-style-type: none"> ↑ <i>Acinetobacter</i> (females) ↓ <i>Streptococcus</i> (females) ↑ <i>Capnocytophaga</i> (males and females) | <ul style="list-style-type: none"> Atrazine exposure decreased body weight and gonadal weight in females Atrazine caused sex-dependent alterations of the microbiome at phyla level Atrazine-induced inflammation and oxidative stress in male intestines | Chen et al. (2018b) |
| Azoxystrobin (F) | Adult | 1.0 µg/l atrazine dose (0.42 µg/l measured) in water environment starting at 4.5 months of age for 7 days | Zebrafish | Both; yes | <ul style="list-style-type: none"> ↓ <i>Proteobacteria</i> (low dose) ↑ <i>Proteobacteria</i> (high dose) | <ul style="list-style-type: none"> Different effects on gut microbiota observed at different doses of azoxystrobin Changes in gut were different from the changes in soil, suggesting gut changes were due to direct ingestion of azoxystrobin | Zhang et al. (2019a) |
| Carbendazim (F) | Adult | 0.1–5 mg/kg azoxystrobin in dry soil environment for 28 days | <i>Enchytraeus crypticus</i> | Not discussed | <ul style="list-style-type: none"> ↓ Bacteroidetes ↑ Firmicutes, Actinobacteria, Proteobacteria | <ul style="list-style-type: none"> Gut alterations occurred after 1 week of treatment, reduced richness and diversity of microbiota Treatment caused inflammation, hepatic lipid metabolism disorder, and liver damage | Jin et al. (2015) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--------------------|-----------------|---|----------------------|---|---|---|-----------------------------|
| | Adult | 0.2–5 mg/kg orally via drinking water for 14 weeks starting at 6 weeks of age | C57BL/6 mice | Male | ↓ Bacteroidetes, Verrucomicrobia ↑ Actinobacteria, Bacteroidetes | • Treatment altered gut microbiota, starting after 1 week of exposure • Treatment altered lipid synthesis and metabolism, causing hyperlipidemia and inflammation | Jin et al. (2018) |
| Chlorothalonil (F) | Adult | Chlorothalonil: 10 µg/l in Honeybees 30% sucrose solution for 6 weeks | Honeybees | Both; however, sex differences were not assessed as both sexes were combined for analyses | ↓ Lactobacillaceae | • Chlorothalonil had the strongest effect on the microbial community of 3 pesticides studied | Kakumanu et al. (2016) |
| Chloryrifos (I) | Developmental | 1 mg/kg bw/day by oral gavage from GD 0 to PND 21 (dams) and PND 21–60 (pups) | Hannover Wistar rats | Male (unclear) | ↑ Bacteroides ↓ Lactobacillus, Bifidobacterium | • Chlorothalonil altered structure and function of gut microbiome including altered gene expression • Similar changes in microbiome observed between SHIME and rats | Joly et al. (2013) |
| | Developmental | 1 or 5 mg/kg bw/day by oral gavage from GD 0 to PND 21 (dams) and PND 21–60 (pups) | Wistar rats | Male | ↓ Lactobacillus ↑ Clostridium, Staphylococcus | • SHIME overestimated effects observed in rats • Exposure significantly decreased body mass and body length vs controls • Exposure delayed maturation of intestinal barriers | Joly Condette et al. (2015) |
| | Developmental | 1 or 3.5 mg/kg gavage to dams from GD 1 until PND 21, gavage to pups from PND 21 to 60; inulin cotreatment daily 25 weeks starting at 4 weeks of age or 20 weeks starting at 9 weeks of age with normal- or high-fat diet | Wistar rats | Male | ↓ Firmicutes, Clostridium coccoides (CPF only) | • Strongest effects of chloryrifos on the gut were observed at weaning • Inulin supplementation alleviated some of the negative effects of prenatal chloryrifos treatment on the microbiome | Reynger et al. (2016a) |
| | Adult | 0.3 mg/kg by gavage daily for 9 weeks starting at 9 weeks of age with normal- or high-fat diet | Wistar rats | Male | Effects varies for all treatment groups | • Chloryrifos treatment had the strongest effect in the newly weaned mice fed a high-fat diet • Diet strongly affected the gut microbiome • Chloryrifos altered gut-brain communication | Liang et al. (2019) |
| | Adult | 0.3 or 3 mg/kg by gavage daily for 9 weeks starting at 9 weeks of age with normal- or high-fat diet | Wistar rats | Male | ↓ Aerococcus, Brenundimonas, Trichococcus (normal diet) ↓ Olsenella, Clostridium sensu stricto 1, Amphibacillus, Enterorhabdus, and Alloprevotella (high-fat diet) | • Low dose with normal diet significantly increased body weight, whereas high-fat diet had no effect on body weight • Chloryrifos-induced diet-specific effects on the microbiome • Insulin levels were decreased under all treatment conditions with bacterial changes related to glucose metabolism | Fang et al. (2018) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---------------|-----------------|--|---|---|--|---|------------------------|
| | Adult | 1 mg/kg by gavage daily for 30 days beginning at 12 weeks of age | Chinese Kunming mice | Male | ↓ Firmicutes, <i>Lactobacillaceae</i> ↑ <i>Bacteroidetes</i> , <i>Bacteroidaceae</i> | <ul style="list-style-type: none"> Treatment altered gut microbial composition Observed gut changes were correlated with changes in metabolites Treatment caused intestinal inflammation and abnormal intestinal permeability | Zhao et al. (2016) |
| | Adult | 30–100 µg/l in water environment for 21 days | Zebrafish | Male | ↓ Proteobacteria | <ul style="list-style-type: none"> Treatment altered expression of oxidative stress genes suggesting increased oxidative stress Treatment caused changes in glucose and lipid metabolism in the liver Treatment increased gut permeability, causing inflammation, and increased lipopolysaccharide levels Treatment caused obesity in normal diet mice Diet and strain did not significantly alter results | Li et al. (2019a) |
| | Adult | 5 mg/kg by gavage daily at 4 weeks of age for 12 weeks with normal- or high-fat diet | C57BL/6 and CD-1 (ICR) mice | Male | ↑ Proteobacteria ↓ <i>Bacteroidetes</i> | <ul style="list-style-type: none"> Treatment increased gut permeability, causing inflammation, and increased lipopolysaccharide levels Treatment caused obesity in normal diet mice Diet and strain did not significantly alter results | Liang et al. (2019) |
| | In vitro | 1 mg per day for 30 days | SHIME | Not applicable | ↑ <i>Enterococcus</i> , <i>Bacteroides</i> ↓ <i>Lactobacillus</i> , <i>Bifidobacterium</i> (no stats) | <ul style="list-style-type: none"> Similar changes in microbiome observed between SHIME and rats | Joly et al. (2013) |
| | In vitro | 3.5 mg/day injection in media into SHIME reactor for 30 days; cells treated with SHIME supernatant for 6 or 24 h; inulin cotreatment | SHIME and Caco-2/TC7 human intestinal cells | Not applicable | ↓ <i>Lactobacillus</i> , <i>Bifidobacterium</i> | <ul style="list-style-type: none"> Inulin cotreatment partially reversed CPF-induced gut alterations and increased SCFA production in the SHIME Cotreatment impacted tight junction gene expression and inhibited proinflammatory signaling in the Caco-2/TC7 intestinal cell line | Réquile et al. (2018) |
| | In vitro | 1 mg/day in oil to SHIME for 30 days | SHIME human <i>in vitro</i> | Not applicable | ↑ <i>Enterobacteria</i> , <i>Bacteroides</i> , <i>Clostridia</i> ↓ <i>Bifidobacterium</i> | <ul style="list-style-type: none"> Treatment altered the microbial community in the SHIME Effects varies between compartments and was strongest for culturable bacteria Changes in SCFA and lactate levels also observed | Reyner et al. (2016a) |
| Coumaphos (I) | Adult | Treated strips following manufacturer's instructions for 6 weeks | Honeybees | Males and females were used; however, sex differences were not assessed because both sexes were combined for analyses | ↑ <i>Burkholderiales</i> , <i>Bifidobacteriales</i> | <ul style="list-style-type: none"> Treatment altered bacterial community No changes in fungal communities were observed due to coumaphos exposure | Kakumanu et al. (2016) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|-----------------|--|---------------------|--|---|---|--------------------------|
| Diazinon (I) | Adult | 650 µg/kg in sugar syrup for 24–34 h of interior worker bees | Honeybees | Female | ↓ <i>Lactobacillus</i> , Bifidobacterium | • Honeybee gut microbiota varied by season • Gut microbiota exhibited similar changes due to different pesticides, suggesting a shared general mechanism | Rouzé et al. (2019) |
| | | 4 mg/l in drinking water for 13 weeks starting at 8 weeks of age | G57BL/6 mice | Male and female guts of males were more disrupted than females | ↓ Lachnospiraceae (both) ↑ Burkholderiales, Bacteroidetes (male) ↓ Firmicutes (male) ↓ Ruminococcaceae, Clostridiaceae, and Erysipelotrichaceae (female) | • Perturbation of the gut may contribute to the neurotoxicity of organophosphate pesticides | Gao et al. (2017b) |
| | Adult | 4 mg/l in drinking water for 13 weeks starting at 8 weeks of age | G57BL/6 mice | Male | N/A | • Treatment altered gut metatranscriptome and quorum sensing system • Treatment-activated stress response pathways and altered energy metabolism of gut microbiota | Gao et al. (2017a) |
| Diethyl phosphate (metabolite of CPF + others) (I) | Adult | 0.08 or 0.13 mg/kg by gavage daily for 20 weeks starting at 8 weeks of age | Wistar rats | Male | ↑ <i>Paraprevotella</i> , Parabacteroides, Alloprevotella, Helicobacter | • Treatment increased pathogenic bacteria levels • Treatment resulted in increased estrogen, decreased triglycerides, and enrichment of SCFA | Yang et al. (2019a) |
| Epoxiconazole (F) | Adult | 4 or 100 mg/kg via diet for 90 days starting at 9 weeks of age | Sprague Dawley rats | Female | ↓ Firmicutes ↑ Bacteroidetes, Proteobacteria ↓ <i>Lactobacillus</i> , Bifidobacterium ↑ <i>Cillitamella apicola</i> , <i>Snodgrassella alvi</i> | • Treatment altered the abundance and composition of gut microbiota | Xu et al. (2014) |
| Fipronil (I) | Adult | 0.25 or 1.0 µg/kg in sugar syrup for 24–34 h of interior worker bees | Honeybees | Female | | • The high dose was lethal • Honeybee gut microbiota varied by season • Gut microbiota exhibited similar changes due to different pesticides, suggesting a shared general mechanism | Rouzé et al. (2019) |
| Glyphosate-based herbicide (GBH) | Developmental | 5 mg/kg/day oral dose via biscuits daily from GD 10 to PND 22 (weaning) | Sprague Dawley rats | Female | ↑ Bacteroidetes ↓ Firmicutes | • Treatment altered maternal behavior; Roundup-treated dams spent significantly more time licking pups compared with control • Glyphosate and Roundup altered the microbiome of the dams at PND 22 differently | Dechartres et al. (2019) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---------------|--|-------------------------------------|--|---|--|---|-----------------------|
| Developmental | 1.75 mg/kg bw/day in drinking water to dams from GD 6 through weaning (PND 28); pups dosed in drinking water following weaning for 6 or 13 weeks | Sprague Dawley rats | Both; however, sex differences were not assessed because both sexes were combined for analyses | ↑ Prevotella, <i>Mucispirillum</i> , <i>Parabacteroides</i> , <i>Veillonella</i> | • No effect of treatment on the dams' microbiome • Different effects observed between Roundup vs glyphosate treatment in the pups | • Microbiome composition changed at different sample times in pups before puberty | Mao et al. (2018) |
| Adult | 250 or 500 mg/kg bw/day dose by gavage beginning at 1 month of age for 1 day, 6 weeks, or 12 weeks | Swiss mice | Male | ↓ <i>Lactobacillus</i> ↓ <i>Bacteroidetes</i> ↓ <i>Corynebacterium</i> ↓ Firmicutes | • Subchronic and chronic exposure increased anxious and depressive behaviors at both doses | • Subchronic and chronic exposure decreased intestinal bacterial count | Aitbali et al. (2018) |
| Adult | 0.00022 to 3.6 g/l in cultured bacteria from caecum of wild turtles | Hawaiian green turtles | Both; however, sex differences were not assessed because both sexes were combined for analyses | ↓ <i>Pantoaea</i> , <i>Proteus</i> , <i>Shigella</i> , <i>Staphylococcus</i> | • Glyphosate significantly decreased bacterial density at all doses | Kittle et al. (2018) | |
| Adult | 0.1 ppb, 400 ppm, or 5000 ppm in tap water for 673 days (50 ng/l, 0.1 g/l, and 2.25 g/l of glyphosate, respectively) | Sprague Dawley rats | Male and female Treated females showed significant differences in gut microbial composition compared with controls and treated males | ↑ <i>Bacteroidetes</i> (females) ↓ <i>Lactobacillaceae</i> (females) | • GBH exposure produced sex-specific gut microbiota composition • Very small study, but unique in treatment length | • Treatment altered levels of select bacterial strains | Lozano et al. (2018) |
| In vitro | Bacterial strains from rat feces were treated with 0.1 ppb, 400 ppm, or 5000 ppm in media for 24 h for up to 5 days | Sprague Dawley rats | Not applicable | ↓ <i>Bifidobacteria</i> , <i>Clostridia</i> , <i>Enterococci</i> (male, highest 2 doses) | • In vivo microbiome disturbances following Roundup treatment may be due to direct bactericidal action | • Highly tolerant strain of <i>Escherichia coli</i> observed in culture | Lozano et al. (2018) |
| In vitro | 0.075–5 mg/ml in media | Many bacterial strains from poultry | Not applicable | Pathogenic bacteria <i>Salmonella</i> and <i>Clostridium</i> were highly resistant to treatment | • Pathogenic bacteria were resistant to glyphosate, whereas beneficial bacteria were moderate to highly susceptible to resistance | • Glyphosate was toxic to the gut microbiome | Shehata et al. (2013) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|-----------------|--|--|--|--|--|--------------------------|
| Glyphosate (H) | Developmental | 0.8, 4, and 20 mg/l/in diet from days 2 to 5 after grafting (larval stage) | Honeybees | Both; however, sex differences were not assessed because both sexes were combined for analyses | ↑ Firmicutes, Clostridia, γ -Proteobacteria, Clostridiales, Lachnospiraceae, Prevotellaceae, and Ruminococcaceae ↓ Bacteroidetes | <ul style="list-style-type: none"> Glyphosate-exposed larvae had lower survival and lower weights compared with controls Glyphosate (20 mg/l) exposure significantly decreased diversity in intestinal bacteria in newly emerged bees Glyphosate had fewer effects on the microbiome than Roundup | Dai et al. (2018) |
| | Developmental | 5 mg/kg/day oral dose via biscuits daily from GD 10 to PND 22 (weaning) | Sprague Dawley rats | Female | ↑ Firmicutes, Butyrivibacillus | | Dechartres et al. (2019) |
| | Developmental | 1.75 mg/kg bw/day in drinking water to dams from GD 6 through weaning (PND 28); pups dosed in drinking water following weaning for 6 or 13 weeks | Sprague Dawley rats | Both; however, sex differences were not assessed because both sexes were combined for analyses | ↑ Prevotella, Mucispirillum, Blautia ↓ Lactobacillus, Aggregatibacter, Streptococcus, Rothia | <ul style="list-style-type: none"> No effect of treatment on the dams' microbiome Different effects observed between Roundup vs glyphosate treatment in pups Microbiome composition changed at different sample times in pups before puberty | Mao et al. (2018) |
| | Adult | 5 or 10 mg/l for 5 days of adult workers; 1 mM for 2 days or 0.1 mM for 5 days of newly hatched workers | Honeybees | Female | ↓ <i>S. alvi</i> , <i>Bifidobacterium</i> , <i>Lactobacillus</i> (5 mg) ↑ <i>G. apicola</i> (5 mg) | <ul style="list-style-type: none"> Exposure decreased total bacteria in adults Stronger effects observed for the 5 mg group than the 10 mg group Exposure interfered with gut colonization in newly hatched bees Glyphosate increased susceptibility of newly hatched bees to bacterial pathogen | Motta et al. (2018) |
| | Adult | 1.5 or 7.5 mM in sugar syrup for 15 days; infection with parasite | Honeybees | Female | ↓ <i>S. alvi</i> ↓ <i>G. apicola</i> ↓ <i>Lactobacillus</i> | <ul style="list-style-type: none"> No synergism observed between glyphosate and <i>Nosema ceranae</i> Glyphosate strongly affected gut microbiota | Blott et al. (2019) |
| In vitro | | 1, 10, 100 μ g/ml in rumen fluid cultured in DAISY incubators for 48 h | Nonlactating Holstein-Friesian cow rumen fluid | Not applicable | ↓ <i>Entodinium</i> spp., <i>Diplodinium</i> spp., <i>Epidinium</i> spp., <i>Ophryoscolex</i> spp., <i>Dasytricha</i> spp. | <ul style="list-style-type: none"> Effects of glyphosate were strongest with fiber-rich diet Glyphosate exposure increased the population of pathogenic species High dose of glyphosate led to increased production of botulinum neurotoxin in cultures incubated with <i>C. botulinum</i> | Ackermann et al. (2015) |
| Hexachlorocyclohexane (HCH, all isomers) (I) | | Colostrum samples collected from eastern China | Human breast milk | Not applicable | ↑ <i>Pseudomonas</i> , <i>Proteus</i> ↓ <i>Enterococcus</i> | <ul style="list-style-type: none"> Different microbial populations were identified in breast milk containing higher levels of HCH isomers More contaminated milk contained more <i>Pseudomonas</i>, bacteria known to break down HCH | Tang et al. (2019) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|------------------------------------|-----------------|--|-------------------------|-----------------------|---|---|-----------------------|
| β -Hexachlorocyclohexane (I) | Adult | 10 mg/kg daily by gavage for 8 weeks | C57BL/6 mice | Male | ↑ Firmicutes, Proteobacteria ↓ Actinobacteria, Verrucomicrobia, Bacteroidetes | • Treatment altered gut microbiota composition • Treatment-activated bile acid synthesis, reduced ileal bile acid readorption, and altered biliary bile acid profiles | Liu et al. (2017) |
| Imazalil (F) | Adult | 100 or 1000 μ g/l for 1,7, or 21 days in aquatic environment at 6 months of age | Zebrafish | Male | ↓ Proteobacteria, Bacteroidetes, Alistipes, Akkermansia ↑ Fusobacteria, Firmicutes | • Exposure increased microbiome diversity in gut • Treatment decreased mucin secretion in the gut, disrupted hepatic metabolism, and altered genes associated with glycolysis and lipid metabolism | Jin et al. (2017) |
| | Adult | 0.1, 0.5, or 2.5 mg/kg bw/day orally via diet for 2, 5, or 15 weeks beginning at 6 weeks of age | C57BL/6 mice | Male | ↓ Bacteroidetes | • Gut changes began by 2 weeks and stabilized by 15 weeks • Changes in cecal bacterial were abnormal and unpredictable • Chronic exposure disturbed intestinal barrier function and decreased mucus secretion | Jin et al. (2018) |
| | Adult | 25, 50, or 100 mg/kg bw/day orally via diet for 4 weeks beginning at 6 weeks of age | ICR mice | Male | ↓ Bacteroidetes, Firmicutes, Actinobacteria | • After 45 days of recovery after treatment finished, the colon only partially recovered • Exposure significantly reduced richness and diversity of cecal and fecal microbiota | Jin et al. (2016) |
| Imidacloprid (I) | Adult | 3.5 μ g/kg in sugar syrup for 24–34 h to interior worker bees | Honeybees | Female | ↓ Lactobacillus, Bifidobacterium | • Treatment-induced colonic inflammation • Honeybee gut microbiota varies by season • Gut microbiota exhibited similar changes due to different pesticides, suggesting a shared general mechanism | Rouzé et al. (2019) |
| | Adult | 500 μ g/l in sugar syrup for 3 days to worker bees | Honeybees | Female | None | • Treatment was lethal to bees, but did not alter the microbiome | Raymann et al. (2018) |
| | Adult | 10–100 μ M in food (length of treatment not given) | Drosophila melanogaster | Female | ↑ Acetobacter, Lactobacillus | • Nonlethal low dose treatment altered gut microbiota • Antibiotic treatment increased fly survival | Daisley et al. (2017) |
| Malathion (I) | Adult | 2 mg/l (approximately 0.6 mg/kg BW/day) in drinking water for 13 weeks beginning at 8 weeks of age | C57BL/6 mice | Male | ↑ Corynebacterium, Clostridium ↓ Planococcaceae, Christensenellaceae, Anaerostipes, Blautia, Roseburia | • Treatment altered expression of genes important for bacterial cell-to-cell communication (quorum sensing) and increased motility and pathogenicity-related genes (related to quorum sensing) • These changes may lead to pathogen invasion | Gao et al. (2018a) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|--|-----------------|---|------------------------|--|---|--|--------------------------|
| Monocrotophos (I) | Adult | 28 µg/kg bw/day in drinking water for 180 days starting at 8 weeks of age followed by fecal transplantation | BALB/c mice | Female | N/A | <ul style="list-style-type: none"> Treatment-induced glucose intolerance Recipients of fecal transplants from treated mice showed glucose intolerance compared with control transplants, suggesting these effects are mediated through gut microbiota | Velmurugan et al. (2017) |
| p,p'-Dichlorodiphenyldichloroethylene (metabolite of DDT, DDE) (I) | Adult | 1 mg/kg bw/day daily by gavage for 8 weeks | C57BL/6 mice | Male | <p>↑ Firmicutes, Proteobacteria ↓ Actinobacteria, Verrucomicrobia, Bacteroidetes</p> | <ul style="list-style-type: none"> Treatment altered gut microbiota composition Treatment-activated bile acid synthesis, reduced ileal bile acid reabsorption, and altered biliary bile acid profiles DDE treatment altered gut microbial composition Pectin treatment altered gut microbiota during cotreatment and recovery periods | Liu et al. (2017) |
| | Adult | 2 mg/kg bw/day by gavage for 8 weeks beginning at 4 weeks of age; cotreatment and recovery with pectin | C57BL/6J mice | Male | <p>↑ Bacteroidetes (cotreatment) ↓ Proteobacteria, Deferribacteres, Cyanobacteria (cotreatment)</p> | <ul style="list-style-type: none"> Pectin treatment altered gut microbiota during cotreatment and recovery Cotreatment reduced the bioaccumulation of DDE in fatty tissues Bioaccumulation in fish liver occurred during exposure Exposure was not lethal but reduced body and liver weight and caused liver damage Exposure significantly altered gut microbiome | Zhan et al. (2019) |
| Pentachlorophenol (I, H) | Juvenile | 1, 50, or 100 µg/l in water environment for 28 days | Goldfish | Both; however, sex differences were not assessed because both sexes were combined for analyses | <p>↑ Bacteroidetes ↓ Firmicutes</p> | <ul style="list-style-type: none"> Bioaccumulation in fish liver occurred during exposure Exposure was not lethal but reduced body and liver weight and caused liver damage | Kan et al. (2015) |
| Permethrin (I) | Juvenile | 34 mg/kg body weight gavage from PND 6 to 21, 4-month follow-up period | Wistar rats | Male | <p>↑ Lactobacillus, Bacteroides, Prevotella, Porphyromonas (PND 21, 51)</p> | <ul style="list-style-type: none"> Treatment altered gut microbial communities and levels of SCFA | Nasuti et al. (2016) |
| | In vitro | 3.2–0.00625 mg/ml in culture media | Many bacterial strains | Not applicable | N/A | <ul style="list-style-type: none"> Blautia producta and <i>Bifidobacterium</i> were most sensitive to permethrin, and <i>E. coli</i> and <i>Pseudomonas aeruginosa</i> were least sensitive Pathogenic bacteria were not very sensitive to treatment | Nasuti et al. (2016) |
| Propamocarb (F) | Adult | 3–300 mg/l in drinking water (0.5–50 mg/kg bw/day) for 28 days starting at 6 weeks of age | ICR mice | Male | <p>↓ Proteobacteria, Bacteroidetes (fecal) ↑↓ Firmicutes, Actinobacteria (fecal) ↓ Firmicutes, Proteobacteria, Actinobacteria, Tenericutes, TM7 (fecal)</p> | <ul style="list-style-type: none"> Changes of gut bacterial populations in feces changed over time during treatment period Different changes observed between fecal and cecal samples Fecal metabolites were altered, indicating metabolic disturbance through or partly through the gut | Wu et al. (2018a) |

Table 5. (continued)

| Chemical | Exposure Window | Dose | Model | Sex & Sex Differences | Effect on Gut Microbiota | Conclusions | References |
|---------------------|-----------------|--|----------------|---|--|---|-------------------------------|
| | Adult | 1–10 mg/l in drinking water for 10 weeks beginning at 6 weeks of age | C57BL/6J mice | Male | ↑ Bacteroidetes ↓ Firmicutes | • Some different changes observed between fecal and cecal microbiota • Fecal metabolites related to energy metabolism were altered | Wu et al. (2018b) |
| | Adult | 100 and 1000 µg/l in water environment for 7 days | Zebrafish | Male | ↑ Proteobacteria, Bacteroidetes, Firmicutes (1000 µg) | • Treatment significantly altered the gut microbiome in the 1000 µg treatment group • Treatment decreased gene expression and altered metabolites related to glycolysis and lipid metabolism | Zhang et al. (2019b) |
| Tau-fluvalinate (I) | Adult | Treated strips following manufacturer's instructions for 6 weeks | Honeybees | Male and female; sex differences not assessed because both sexes were combined for analyses | ↓ Enterobacteriaceae, Caulobacteraceae | • Treatment altered bacterial community • No changes in fungal communities were observed due to pesticide exposure | Kakumanu et al. (2016) |
| Thiamethoxam (I) | Adult | 1.7 µg/kg in sugar syrup for 24–34 h of interior worker bees | Honeybees | Female | ↓ Lactobacillus, Bifidobacterium, Alphaproteobacteria ↑ <i>G. apicola</i> | • Honeybee gut microbiota varied by season • Gut microbiota exhibited similar changes due to different pesticides, suggesting a shared general mechanism | Rouzé et al. (2019) |
| Trichlorfon (I) | Adult | 12 µg/g bw single dose by oral gavage at 8 weeks of age, collected 24 h after dosing | Japanese quail | Male and female Female gut more impacted by treatment than male | ↓ Lactobacillus ↑ Proteobacteria | • Treatment caused different changes in microbial populations in the cecum, large intestine, and feces | Crisol-Martinez et al. (2016) |

Abbreviations: F, fungicide; H, herbicide; I, insecticide; SCFA, short-chain fatty acid; SHIME, Simulator of the Human Intestinal Microbial Ecosystem.

2019; Lozano et al., 2018; Mao et al., 2018) and honeybees (Blot et al., 2019; Dai et al., 2018; Motta et al., 2018). In contrast, amino-methylphosphonic acid (AMPA), a glyphosate metabolite, does not alter the gut microbiome of exposed honeybees (Blot et al., 2019). Collectively, these studies suggest that the parent compounds—not the metabolites—are responsible for the changes in the gut microbiome.

Extensive evidence indicates that the organophosphate chlorpyrifos affects microbial populations in male rodents and fish exposed during development and adulthood (Condette et al., 2015; Li et al., 2019a; Liang et al., 2019; Reygner et al., 2016b; Stunes et al., 2017). Chlorpyrifos exposure additionally causes inflammation and oxidative stress in the gut (Li et al., 2019a; Liang et al., 2019; Zhao et al., 2016). However, information on the effects of chlorpyrifos in females is lacking.

Few studies have examined the developmental effects of pesticide exposures beyond glyphosate and chlorpyrifos on the gut. However, studies in adults across multiple species using a wide range of insecticides, fungicides, and herbicides show alterations of the gut microbiome as well as other common effects, including altered lipid metabolism, inflammation, and oxidative stress (Table 5). The few studies that directly measured sex differences found differential alterations of gut microbiome between male and female animals following exposure to atrazine (Chen et al., 2018b), diazinon (Gao et al., 2017b), GBH (Lozano et al., 2018), and trichlorfon (Crisol-Martinez et al., 2016). However, there is disagreement on whether males or females have more alterations in the gut microbiome.

FUTURE DIRECTIONS

Collectively, the existing data suggest that exposure to environmental chemicals during various stages of life causes alterations in the gut microbiome and is associated with changes in health, including immune dysfunction, altered carbohydrate and lipid metabolism, and neurobehavioral impairments. Moreover, the effects of the environmental chemicals on gut microbiota highly depend on sex and age. A major question raised from the studies reviewed is whether, and to what extent, microbiota mediate the disease-causing effects of the environmental chemicals. The field of gut microbiome toxicology is still relatively new, which is why many studies have only reported changes in the gut microbiota composition and not on the mechanisms by which chemicals interfere with gut microbiota. In the future, it is important to characterize which microbes are present by sequencing the full length of 16S rRNA gene to correctly identify strains and species that are in the gut. A partial 16S rRNA sequencing, commonly on hypervariable regions 3 and 4 (V3 and V4), does not yield enough depth to give information on the species or strains of microbes. While partial 16S rRNA sequencing can give us some information on changes at higher taxonomic levels, this information is not enough because different strains of the same species can have different effects. For example, *Escherichia coli* O157: H7 is a disease-causing strain whereas *E. coli* Nissle 1917 is a probiotic strain (Wassenaar, 2016). In addition to identifying changes of gut bacteria at the species and strain levels, studies should also include metagenomic sequencing to study alterations of other microorganisms in the gut microbiome, which would provide insights into the functional changes in the gut microbiome. On the whole, future studies should examine the causal relationship between chemically induced changes in the gut microbiome and chemically induced adverse health outcomes. As such, research should supplement sequencing studies with

mechanistic studies using conventional and gnotobiotic animal models, such as fish, mice, and pigs, because these models are essential for understanding the impact of environmental chemicals on the gut microbiome and the health consequences resulting from the altered gut microbiota (Nguyen et al., 2015).

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