

Critical window variable selection: estimating the impact of air pollution on very preterm birth

JOSHUA L. WARREN[∗], WENJING KONG

Department of Biostatistics, Yale University, New Haven, CT 06520, USA joshua.warren@yale.edu

THOMAS J. LUBEN

United States Environmental Protection Agency, Durham, NC 27709, USA

HOWARD H. CHANG

Department of Biostatistics and Bioinformatics, Emory University, Atlanta, GA 30322, USA

SUMMARY

Understanding the impact that environmental exposure during different stages of pregnancy has on the risk of adverse birth outcomes is vital for protection of the fetus and for the development of mechanistic explanations of exposure–disease relationships. As a result, statistical models to estimate critical windows of susceptibility have been developed for several different reproductive outcomes and pollutants. However, these current methods fail to adequately address the primary objective of this line of research; how to statistically identify a critical window of susceptibility. In this article, we introduce critical window variable selection (CWVS), a hierarchical Bayesian framework that directly addresses this question while simultaneously providing improved estimation of the risk parameters. Through simulation, we show that CWVS outperforms existing competing techniques in the setting of highly temporally correlated exposures in terms of (i) correctly identifying critical windows and (ii) accurately estimating risk parameters. We apply all competing methods to a case/control analysis of pregnant women in North Carolina, 2005–2008, with respect to the development of very preterm birth and exposure to ambient ozone and particulate matter $\langle 2.5 \mu m \rangle$ in aerodynamic diameter, and identify/estimate the critical windows of susceptibility. The newly developed method is implemented in the R package CWVS.

Keywords: Bayesian variable selection; Gaussian process; Linear model of coregionalization.

1. INTRODUCTION

Adverse birth outcomes have been linked to increases in neonatal morbidity and mortality, long-term health and developmental problems, and appreciable medical costs [\(McCormick](#page-15-0), [1985;](#page-15-0) Butler *[and others](#page-14-0)*, [2007\)](#page-14-0). Recently, there has been considerable interest in examining associations between ambient air pollution exposure during pregnancy and the risks of adverse birth outcomes. Epidemiological work in this area has focused primarily on linking a pregnant woman's exposure to a selected pollutant, averaged over different

[∗]To whom correspondence should be addressed.

pregnancy periods, to various adverse birth outcomes in order to estimate general associations between exposure and risk (Vrijheid *[and others](#page-15-0)*, [2011](#page-15-0); Stieb *[and others](#page-15-0)*, [2012](#page-15-0)). However, due to the temporal aggregation of the exposures over long periods, these studies were unable to investigate the impact of exposure timing on adverse birth outcome development in a more continuous manner across pregnancy.

In its set of strategic goals for 2012–2017, the National Institute of Environmental Health Sciences (NIEHS) included the identification of these more specific periods of vulnerability to environmental exposures, also known as "critical windows of susceptibility," as part of its mission [\(NIEHS](#page-15-0), [2012](#page-15-0)). Identification of critical windows can lead to an improved understanding of possible mechanisms of disease development and provide guidelines for protection of the fetus. As a result, a number of advanced statistical methods have recently been introduced and applied in the reproductive epidemiological setting.

Accounting for correlation in exposures and risk parameters across pregnancy has been handled in a variety of ways in these past studies. Much of the work has relied on a Gaussian process (GP) prior distribution with correlation structures to encourage borrowing of information between parameters across pregnancy per[iods, spatial location, multiple pollutants, multivariate outcomes, and outcome weeks \(](#page-15-0)Warren *and others*, [2012,](#page-15-0) [2013](#page-15-0); Chang *[and others](#page-14-0)*, [2015](#page-14-0); Warren *[and others](#page-15-0)*, [2016](#page-15-0)). Warren *[and others](#page-15-0)* [\(2012\)](#page-15-0) extended this work to the Bayesian semiparametric setting while Wilson *[and others](#page-15-0)* [\(2017\)](#page-15-0) considered alternative smoothing techniques including the use of principal components and splines. Similarly defined distributed lag models [\(Schwartz](#page-15-0), [2000\)](#page-15-0) have been considered to estimate the impact of time-varying exposures using penalized splines (Zanobetti *[and others](#page-16-0)*, [2000\)](#page-16-0), Gaussian processes [\(Heaton and Peng,](#page-15-0) [2012](#page-15-0)), hierarchical modeling (Welty *[and others](#page-15-0)*, [2009\)](#page-15-0), and autoregressive priors [\(Fahrmeir and Knorr-Held,](#page-14-0) [1997\)](#page-14-0).

These methods have successfully uncovered more informative associations between specific exposure timing and adverse health outcomes. However, none of the current methods are designed to directly address the primary question of interest; how to statistically define a critical window of susceptibility. Instead, the models most often focus on stabilized estimation (i.e., smoothness) of the risk parameters across time in order to identify general temporal trends in exposure risk. These methods define a critical window based on whether the calculated confidence/credible interval (CI/CrI) corresponding to a particular time period excludes (critical window) or includes (non-critical window) zero. It is currently unknown how well this definition performs at recovering the true critical window set. In addition, Warren *[and others](#page-15-0)* [\(2012\)](#page-15-0) showed that use of a Gaussian process may lead to over-smoothing during estimation of risk parameters, potentially complicating the identification of critical windows, and that a more flexible model may help to reduce this error. The impact that this potential over-smoothing has on risk parameter estimation and on identifying critical windows is also currently unknown.

In this work, we introduce critical window variable selection (CWVS) as a method for explicitly identifying the true critical windows of susceptibility and avoiding the over-smoothing inherit to many existing methods, allowing for improved estimation of risk parameters. CWVS is an innovative Bayesian variable selection method that introduces temporal smoothness during estimation of the risk parameters as well as in the variable selection process. This is accomplished via a flexible cross-covariance model based on the linear model of coregionalization (LMC) for the inclusion probabilities and risk parameters. Bayesian v[ariable](#page-14-0) [selection](#page-14-0) [under](#page-14-0) [collinearity](#page-14-0) [of](#page-14-0) [the](#page-14-0) [predictors](#page-14-0) [has](#page-14-0) [been](#page-14-0) [studied](#page-14-0) [previously](#page-14-0) [by](#page-14-0) Ghosh and Ghattas [\(2015](#page-14-0)) who recommended that under complex correlation structures, information regarding grouping patterns of covariates should be incorporated into the prior distribution. We take this approach when developing CWVS.

We investigate the properties and performance of CWVS in comparison to existing methods through simulation, and use CWVS to identify and estimate critical windows of susceptibility with respect to prenatal exposure to ambient ozone and particulate matter less than 2.5 μ m in aerodynamic diameter (PM2.5), and very preterm birth (VPTB) in North Carolina (NC), 2005–2008. Both pollutants have been

792 J. L. WARREN AND OTHERS

linked with adverse birth outcomes in past studies (Stieb *[and others](#page-15-0)*, [2012\)](#page-15-0). Preterm birth, a birth occurring before 37 completed gestational weeks, is one of the leading causes of mortality in children under 5 years [of age and is often associated with significant developmental issues for survivors \(](#page-14-0)Blencowe *and others*, [2012\)](#page-14-0). Children born even earlier are at a far greater risk of morbidity and/or mortality. A number of epidemiological studies have investigated the association between prenatal exposure to ambient air pollution and preterm birth development but fewer have focused on children born well before 37 weeks as we do in this study [\(Landgren](#page-15-0), [1996](#page-15-0); Brauer *[and others](#page-14-0)*, [2008](#page-14-0); Wu *[and others](#page-15-0)*, [2009\)](#page-15-0).

2. DATA

We analyze birth records data from NC in 2005–2008. The dataset includes all live births occurring across the entire state during the selected years.We limit the analyses to singleton births, without any documented congenital anomalies, and with a gestational age of at least 27 weeks (based on the clinical estimate of gestational age) (Chang *[and others](#page-14-0)*, [2015](#page-14-0)). In addition, the mother must be between 15 and 50 years of age, inclusively. The primary study outcome, VPTB, is defined as a birth occurring before 32 completed weeks of gestation (Blencowe *[and others](#page-14-0)*, [2012\)](#page-14-0). Due to the low prevalence of VPTB in our study population (3672/454,058; 0.81%) and the computational complexities of working with large datasets within the Bayesian setting, we opt for a case/control analysis of the data by retaining each of the eligible VPTB cases in the data and randomly selecting four controls (children born at \geq 32 weeks) for each. This study was approved by the Institutional Review Board at the University of North Carolina at Chapel Hill.

We utilize the Environmental Protection Agency (EPA) fused downscaler estimates (https://www.epa.gov/hesc/rsig-related-downloadable-data-files) to link pollution exposures with each of the women across the relevant calendar periods of pregnancy. The statistical model that produced these estimates has been previously described (Berrocal *[and others](#page-14-0)*, [2010\)](#page-14-0). Briefly, in its most general form, the model links observed pollution values from EPA monitoring sites with deterministic chemistry/meteorology model output from the Community Multiscale Air Quality Modeling System (CMAQ) through a spatially and temporally varying regression coefficients model. The full spatial and temporal coverage of the CMAQ output allows for prediction of air pollution concentrations in locations and on days without recorded measurements from a fixed-site monitor, while the varying coefficients allow the form of the bias in the CMAQ output (with respect to the monitoring data) to change across space and time. Daily ozone [8 h maximum; parts per billion (ppb)] and $PM_{2,5}$ (24 h average; micrograms per cubic meter (μ g/m³)) point estimates are provided at the year 2000 Census tract centroids for 2004–2008 (based on data downloaded in June–July, 2017). We link each woman's residence at delivery (latitude/longitude) with the closest Census tract centroid (possibly outside of NC) to create estimated weekly average pollution exposures across the entire pregnancy. [Figure S1 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online displays the full set of Census tract centroids in NC and relevant surrounding areas (4054 locations in total). Individuals with missing covariates were removed from the final analysis. [Table S2 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online displays study population information by case/control status.

3. CRITICAL WINDOW VARIABLE SELECTION

We introduce CWVS, a general framework for modeling a binary outcome as a function of time-varying predictors with the goals of (i) identifying an important, and temporally proximal, subset of those predictors and (ii) making proper inference on the parameters corresponding to that subset. In the context of reproductive epidemiology, these informative predictors/parameters can be thought of as critical windows of susceptibility where higher levels of environmental exposures lead to increased risk of an adverse birth outcome.

We model the observed outcome as $Y_i|\beta, \alpha \stackrel{\text{ind}}{\sim} \text{Bernoulli}\{p_i(\beta, \alpha)\}, i = 1, ..., n$, where Y_i is the binary outcome associated with subject *i*; $p_i(\beta, \alpha)$ is the probability that subject *i*'s response results in the adverse outcome and is a function of unknown parameters β and α ; and *n* is the number of subjects. We model the probability that an outcome occurs as a function of subject-specific covariates, some of which are static across time and others that are time-varying and represent the main exposures of interest. We introduce a logistic regression framework such that

$$
\ln\left\{\frac{p_i(\boldsymbol{\beta},\boldsymbol{\alpha})}{1-p_i(\boldsymbol{\beta},\boldsymbol{\alpha})}\right\} = \mathbf{x}_i^{\mathrm{T}}\boldsymbol{\beta} + \sum_{i=1}^m z_i \left\{c_i\left(t\right)\right\} \alpha\left(t\right),\tag{3.1}
$$

where \mathbf{x}_i is the vector (length *p*) of static covariates/confounders specific to subject *i*, including the intercept; *β* is the accompanying vector of unknown regression parameters; *m* represents the number of exposure time periods that are considered; z_i { c_i (t)} is the average exposure at subject *i*'s spatial location occurring during time period *t* (e.g., week of pregnancy) that covers calendar interval *ci* (*t*) (e.g., 7 day calendar date range of pregnancy week *t* for subject *i*); and α (*t*) is the unknown regression parameter that describes the association between an exposure occurring during time period *t* and the risk of outcome development.

Our interest is in performing temporally smoothed variable selection in order to create a method for identifying a time period of great importance while maintaining the stability and smoothness of risk parameter estimation that results through use of commonly applied penalized regression techniques. In order to achieve both of these aims, we decompose $\alpha(t)$ as $\alpha(t) = \theta(t) \gamma(t)$ and introduce separate models for θ (*t*) (effect magnitude component) and γ (*t*) (variable selection component) that each account for correlation across time and potential cross-covariance.

The real-valued parameter that describes the association between exposure during period *t* and occurrence of the binary outcome is denoted by $\theta(t)$. We assume *a priori* that $\theta(t)$ and $\theta(t')$ are potentially more similar when $|t - t'|$ is small, and model this relationship using a GP with time series correlation structure. The binary random variable that describes the overall importance of exposure during period *t* on development of the outcome and is denoted by $\gamma(t)$ and is modeled such that $\gamma(t)$ | $\pi(t) \stackrel{\text{ind}}{\sim}$ Bernoulli { $\pi(t)$ }, Φ^{-1} { $\pi(t)$ } = $\eta(t)$. The probability that exposure period *t* is included in the regression model is given as π (*t*) and is defined as a function of an underlying set of real-valued latent parameters, $\eta(t)$, through use of a probit link function (Φ^{-1}) is the inverse cumulative distribution function of the standard normal distribution). Similar to θ (*t*), we model the vector of η (*t*) parameters using a separate GP with time series correlation structure under the assumption that important (and conversely unimportant) predictors are temporally clustered. This GP allows for temporal smoothness and stability in the identification of important time-varying predictors.

In order to flexibly model temporal correlation separately for the θ (*t*) and η (*t*) processes, as well as the cross-covariance between the two processes, we utilize the LMC [\(Wackernagel](#page-15-0), [2013](#page-15-0)) such that

$$
\left[\begin{array}{c} \theta(t) \\ \eta(t) \end{array}\right] = A \left[\begin{array}{c} \delta_1(t) \\ \delta_2(t) \end{array}\right], A = \left[\begin{array}{cc} A_{11} & 0 \\ A_{21} & A_{22} \end{array}\right],
$$

where $A_{jj} > 0$ for $j = 1, 2$ and $A_{21} \in \mathbb{R}$. This constructive definition leads to

$$
\theta(t) = A_{11}\delta_1(t), \ \eta(t) = A_{21}\delta_1(t) + A_{22}\delta_2(t),
$$

where we define $\delta_j = \{\delta_j(1), \ldots, \delta_j(m)\}^{\text{T}} |\phi_j \stackrel{\text{ind}}{\sim} \text{MVN} {\{\mathbf{0}_m, \Sigma(\phi_j)\}\}$, $j = 1, 2$, where MVN (μ, Ω) is the multivariate normal distribution with mean vector μ and variance/covariance matrix Ω ; $\mathbf{0}_m$ is a column vector of length *m* where each entry is equal to 0; and $\Sigma \left(\phi_j \right)_{t,t'} = \exp \left\{ -\phi_j |t-t'| \right\}$ with $\phi_j > 0$. We note that any general 1D isotropic correlation function can be used, and we select the exponential version to be consistent with existing critical window methods. Correlation between θ (*t*) and η (*t*) is induced by the shared latent $\delta_1(t)$ parameter.

In order to complete the model specification, we assign prior distribution to each of the model parameters. We opt for weakly informative priors in order to place more emphasis on the data rather than our prior beliefs during posterior estimation. The vector of covariate/confounder regression parameters is modeled as $\beta \sim MVN$ $(0_p, \sigma_\beta^2 I_p)$ where σ_β^2 is fixed at a large value and I_p is the *p* by *p* identity matrix. The LMC correlation parameters are specified as $\phi_j \stackrel{\text{iid}}{\sim} \text{Gamma}(\alpha_\phi, \beta_\phi)$, $j=1,2,$ where α_ϕ and β_ϕ are fixed at small values; while the LMC variance/covariance parameters are modeled as $\ln(A_{jj}) \sim \text{Normal}(0, \sigma_A^2), j = 1, 2$; and $A_{21} \sim \text{Normal}(0, \sigma_A^2)$, where σ_A^2 is fixed at a large value. Full details on the Markov chain Monte Carlo (MCMC) posterior sampling algorithm are given in [Section S1 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online.

3.1. *Induced covariance structure*

Use of the LMC leads to a flexible variance–covariance structure for the set of latent parameters that define α (*t*), the main risk parameters. Understanding the induced covariance structure is key in understanding how our model balances temporal smoothness in parameter estimation with abrupt changes in risk modeled through the variable selection components. The LMC allows for separate levels of temporal smoothness in parameter estimation for θ (*t*) and η (*t*) while simultaneously modeling of the cross-covariance between both sets of parameters.

The covariance between two of the continuous components of α (*t*) on exposure periods *t* and *t'* is given as Covariance { θ (*t*), θ (*t'*) $|A_{11}, \phi_1\rangle = A_{11}^2 \exp{-\phi_1|t-t'|}$ where A_{11}^2 represents the variance of the process and ϕ_1 describes the level of temporal smoothness in the risk parameters. Smaller values of ϕ_1 indicate more temporal similarity in parameter values across time. As $|t - t'|$ increases, the level of correlation between the parameters decreases.

The covariance between two parameters that ultimately define the variable selection inclusion probabilities for α (*t*) on exposure periods *t* and *t'* is given as Covariance $\{\eta$ (*t*), η (*t'*) $|A_{21}, A_{22}, \phi_1, \phi_2\}$ $= A_{21}^2 \exp \{-\phi_1 |t-t'| + A_{22}^2 \exp \{-\phi_2 |t-t'| \}.$ Compared to the previously defined covariance, this expression includes an additional term that is similar in structure. Allowing for separate smoothness parameters increases the flexibility of the model and allows for the possibility that the two processes differ greatly in terms of temporal smoothness. We note that if we define $\phi_1 = \phi_2 \equiv \phi$ then the LMC collapses to a more commonly used separable cross-covariance framework. Finally, the cross-covariance between the two sets of parameters on exposure periods *t* and *t'* is defined as Covariance { θ (*t*), η (*t'*) $|A_{11}, A_{21}, \phi_1$ } = $A_{11}A_{21}$ exp { $-\phi_1|t-t'$ } where $A_{21} \in \mathbb{R}$ determines if the correlation is positive or negative.

4. SIMULATION STUDY

We design a simulation study to determine the most appropriate definition of a critical window using CWVS and to explore various properties of CWVS in comparison with existing approaches. Specifically, we are interested in each method's ability to (i) correctly identify the true set of critical windows and (ii) to properly estimate the parameters associated with these critical windows with respect to mean squared error (MSE) and CrI coverage.

We begin by describing the process of generating a single simulated dataset for analysis in the study. Our main priority is to simulate data that closely resemble data from our area of application (see Section 5) so that the simulation study findings can provide relevant insights into the use of our model within that

Setting	Critical week (s)	α (<i>t</i>) on critical week(s)		
Early $1(E1)$		λ		
Middle $1 (M1)$	14	λ		
Early $4(E4)$	$1 - 4$	$\lambda/4$		
Middle $4 (M4)$	$13 - 16$	$\lambda/4$		
Early $7(E7)$	$1 - 7$	$\lambda/7$		
Middle $7 (M7)$	$11 - 17$	$\lambda/7$		

Table 1. *Simulation study data generating settings for* α (*t*) *where* $\lambda = 0.64$

setting. The underlying data generating model is given as

$$
Y_i|p_i \stackrel{\text{ind}}{\sim} \text{Bernoulli}(p_i), i = 1,\ldots,n; \text{ logit}(p_i) = \beta_0 + \sum_{i=1}^{27} z_i \{c_i(t)\} \alpha(t).
$$

We choose the sample size of the simulated dataset to exactly match the NC VPTB analysis sample size (*n* = 18 360) and similarly, β_0 is set at −1.39 so that ≈20% of the simulated responses result in the outcome. The pollution exposures for a particular woman in the dataset across the first 27 weeks (*m* = 27 weeks, chosen to match the NC VPTB application) of pregnancy are randomly sampled without replacement directly from the full cohort of pregnant women ozone exposures in NC (454 048 women in total) in order to obtain realistic exposure correlation and magnitudes across pregnancy. An entire time series of exposure connected to an actual individual is selected and assigned to a simulated person/response. Finally, we explore a number of different options for the pollution risk parameters, α (*t*).

In each α (*t*) setting, we ensure that the magnitude of simulated risk is realistic by using results obtained from fitting CWVS to the NC VPTB dataset while focusing on ozone exposure. We define total exposure risk across pregnancy by summing over the posterior means of $\alpha(t)$ | $\gamma(t) = 1$ for all *t*, where the posterior marginal inclusion probability was at least 0.50 (i.e., $P\{y(t) = 1 | Y\} \ge 0.50$). We refer to this summation as

$$
\lambda = \sum_{t=1}^{27} \mathbb{E} \{ \alpha \left(t \right) | \gamma \left(t \right) = 1, Y \} I \left[P \{ \gamma \left(t \right) = 1 | Y \} \geq 0.50 \right],
$$

where *I* (.) represents the indicator function, and note that $\lambda = 0.64$ in the NC VPTB ozone application. Therefore, if we assume there is a short critical window only covering one pregnancy week, all of this risk would be placed in that particular week. More generally, if we assume a critical window set spanning *k* pregnancy weeks, the risk would be divided evenly across those *k* parameters such that α (*t*) = λ/k for all *t* included in the critical window set. We formally define the critical window set by setting α (*t*) = 0 for the unimportant time periods. In total, we choose six settings for α (*t*) as shown in Table 1. Figure S2 of the supplementary material available at *Biostatistics* online displays each of the different settings on the odds ratio scale. "Early" and "Middle" refer to the location of the critical window within the duration of pregnancy. These settings allow us to explore how each method performs when the true critical window length varies and when the timing of importance shifts from the tail to the middle of pregnancy.

Once the exposures are selected for each simulated individual in the dataset and we have chosen the $\alpha(t)$ setting, we calculate the probability of the outcome for each individual and simulate the binary response variables using these probabilities. For each of the six different settings, we simulated 50 independent datasets (simulating different exposures each time).

4.1. *Defining a critical time period*

We consider three different options for defining a critical time period using CWVS and investigate the most appropriate version. First, we explore the use of the median probability model (Barbieri *[and others](#page-14-0)*, [2004\)](#page-14-0), where we define the critical window set to include all time periods *t* such that $P\{\gamma(t) = 1 | Y\} \geq 0.50$. Next, we focus attention on the continuous component of α (*t*) and define a time period *t* as critical if the 95% CrI of α (*t*) | γ (*t*) = 1 excludes zero (in either direction). Finally, we combine both ideas such that time period *t* is in the critical window set if its marginal posterior inclusion probability is ≥ 0.50 and the 95% CrI for $\alpha(t)$ | $\gamma(t) = 1$ excludes zero.

4.2. *Competing methods*

Along with fitting CWVS to the simulated data and identifying the most appropriate definition of a critical time period, we also explore a number of competing methods to determine the benefits of our newly developed framework. Each method uses the statistical model in (3.1), but differs in the prior distribution introduced for the α (*t*) parameters.

- Gaussian process (GP): The original method used by Warren *[and others](#page-15-0)* [\(2012\)](#page-15-0) to identify and estimate critical windows of susceptibility, where $\alpha | \sigma_{\alpha}^2, \phi \sim \text{MVN} \{ \mathbf{0}_{27}, \sigma_{\alpha}^2 \Sigma (\phi) \}$ and $\Sigma (\phi)_{t,t'} =$ $\exp \{-\phi |t - t'|\}.$
- Stochastic search variable selection (SSVS): A Bayesian variable selection technique originally introduced by [George and McCulloch](#page-14-0) [\(1993\)](#page-14-0), where $\alpha|\gamma \sim \text{MVN }(\mathbf{0}_{27}, D)$; *D* is a diagonal matrix with $D_t = \tau^2[\{1 - \gamma(t)\} + c^2 \gamma(t)]$; and $\gamma(t) \stackrel{\text{iid}}{\sim}$ Bernoulli (π) for $t = 1, ..., 27$.
- CWVS Separable: A simplified form of CWVS where we no longer include the LMC specification and instead model the joint process using a separable covariance matrix such that

$$
\left[\begin{array}{c} \theta \\ \eta \end{array}\right]|\Omega,\phi\sim \text{MVN}\left\{0_{27},\Omega\otimes\Sigma\left(\phi\right)\right\},\
$$

where $\Sigma(\phi)$ matches the matrix described in GP, Ω is an unstructured two-by-two cross-covariance matrix, and ⊗ represents the Kronecker product.

For each of the competing methods, we select weakly informative prior distributions to complete the model specifications. Across all models, $\beta_0 \sim N(0, 1000)$. For GP, $\sigma_\alpha^2 \sim$ Inverse Gamma (0.10, 0.10) and $\phi \sim$ Gamma (1.00, 1.00). In SSVS, $\pi \sim$ Uniform (0, 1) while *c* and τ are selected based on guidance from [George and McCulloch](#page-14-0) [\(1993\)](#page-14-0) and output from GP fitted to each simulated dataset. Specifically, *c* is fixed at 100 and σ_{β}/τ is fixed at 10, where σ_{β} represents the median of the posterior standard deviations for the weekly effect estimates from GP. We then solve for τ and note that this value changes across simulated datasets based on the results from GP. In CWVS Separable, $Ω⁻¹ ~$ Wishart(*I*₂, 3)[, allowing for marginally uniform prior correlations between the parameters \(](#page-14-0)Gelman *and others*, [2014](#page-14-0)). For CWVS, we follow the priors specified in the NC VPTB analysis (see Section 5 for full details).

Each of the competing methods were fit to each of the simulated datasets. All methods were run for 15 000 MCMC iterations, with the first 5000 discarded as a burn-in period. For CWVS, CWVS Separable, and SSVS, we explore the most appropriate definition of a critical time period based on each definition's ability to correctly identify the true critical window set. GP does not include a variable selection component, so we define a critical time period as an individual α (*t*) parameter whose 95% quantile-based CrI fails to include zero (in either direction) as is standard in previous work. We also calculate the posterior mean and 95% CrI for each α (*t*) parameter from every model. For the variable selection methods, we only consider the posterior samples where $\gamma(t) = 1$ when calculating these quantities, resulting in inference for $\alpha(t)|\gamma(t)=1$.

Using the output from each fitted model, we estimate the average MSE for the true critical window set for a particular method and setting as $\frac{1}{50|B_1|} \sum_{j=1}^{50} \sum_{t \in B_1} {\alpha(t) - \hat{\alpha}_j(t)}^2$, $B_1 = {t : \alpha(t) \neq 0}$, where $\hat{\alpha}$ _{*j*} (*t*) is the posterior mean estimated from simulated dataset *j* and time period *t* for a selected method and $|B_1|$ represents the number of time periods included in the true set of critical windows. We also evaluate two other definitions of *B* where B_2 represents the set of time periods *t* such that α (*t*) = 0 but α (*t* + 1) \neq 0 or α (*t* − 1) \neq 0. This represents the time period(s) that immediately surround the true critical window set. We define B_3 as the set of time periods that are not in B_1 or B_2 . These reflect the unimportant periods of pregnancy, far removed from the true critical window time periods. These three versions of *B* allow us to compare the performances of the methods across every possible option of risk estimation behavior during the pregnancy; significant periods, bordering significant periods, and more general non-significant periods. Next, we estimate the average empirical coverage of the 95% CrIs for the different forms of *B* for a particular method such as $\frac{1}{50|B|} \sum_{j=1}^{50} \sum_{t \in B} I \{ \alpha(t) \in \text{CrI}_j(t) \}$, where $\text{CrI}_j(t)$ is the 95% CrI for α (*t*) calculated from simulated dataset *j* and time period *t* for a selected method. Finally, we calculate the proportion of times that each method correctly identifies the true critical window set exactly (no more or less periods identified) based on each method's definition of a critical window such as $\frac{1}{50} \sum_{j=1}^{50} I\left\{ \left(\sum_{t=1}^{27} I\left[\text{CW}_j(t) = I\left\{ \alpha(t) \neq 0 \right\} \right] \right) = 27 \right\}$, where CW_j (*t*) is the indicator for whether time period *t* is identified as a critical window based on simulated dataset *j* for a selected method. This metric is also used to determine the most appropriate definition of a critical time period among the competing definitions. As a sensitivity analysis, we also explore the ability of each method to correctly identify the true critical window set without penalty for classifying non-critical time periods as important.

The results for determining the most appropriate critical window definition for each method are displayed in Table [2.](#page-8-0) For simulation settings E1 and M1, the CrI definition performs poorly across all variable selection methods. When comparing the median probability model with the combined definition, it is clear that the combined definition leads to more accurate identification of the true critical window set for CWVS, particularly for settings with more than one critical time period. SSVS and CWVS Separable are relatively robust to this choice. GP struggles to identify the true critical window set in comparison to CWVS for the majority of simulation settings. The ability of each method to correctly identify the true critical window set declines as the number of critical periods increases. This may be due to the decreasing risk magnitude at each critical period when the total number of non-zero time periods increases, resulting in reduced statistical power (see Table [1\)](#page-5-0). For the sensitivity analysis results in Table S3 of the supplementary material available at *Biostatistics* online, the median probability model for CWVS more often identifies the critical window set across all simulation settings than the other definitions when there is no penalty for false positives. However, the results in Table [2](#page-8-0) suggest that the median probability model often has very low specificity. GP and CWVS perform similarly when the combined definition is considered and no penalty is enforced. Based on these findings, from this point forward we use the combined definition of a critical time period when applying the variable selection techniques to simulated and real data.

Table 2. *Simulation study results to determine the most appropriate critical time period definition. The proportion of times that each definition correctly identifies the exact true set of critical time periods is presented across all simulation settings. Standard errors are presented in Table S4 of the supplementary material available at Biostatistics online. Median probability: P* { γ (*t*) = 1|*Y*} ≥ 0.50*; Credible interval: 95% credible interval of* α (*t*) $|\gamma(t)| = 1$ *excludes zero*

Method	E1	M1	E4	M4	E7	M7
Median probability						
SSVS	0.98	1.00	0.30	0.14	0.00	0.00
CWVS Sep.	0.96	0.98	0.68	0.42	0.10	0.00
CWVS LMC	0.90	0.96	0.38	0.14	0.26	0.02
Credible interval						
SSVS	0.50	0.30	0.36	0.30	0.00	0.00
CWVS Sep.	0.48	0.44	0.50	0.52	0.04	0.06
CWVS LMC	0.36	0.32	0.80	0.60	0.44	0.16
Median probability and credible interval						
SSVS	0.98	1.00	0.30	0.14	0.00	0.00
CWVS Sep.	0.96	0.98	0.66	0.42	0.04	0.02
CWVS LMC	0.90	0.96	0.80	0.60	0.44	0.16
GP	0.40	0.38	0.62	0.58	0.34	0.20

The average MSE and empirical coverage results across all definitions of set *B* are displayed in Table [3.](#page-9-0) In [Figures S3–S14 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, we display boxplots of the posterior means and CrI lengths across the 50 simulated datasets for each method and simulation setting. With respect to average MSE, CWVS consistently performs well across the different simulation settings and set *B* definitions while GP struggles when the number of critical periods is small. SSVS and CWVS Separable perform worse when the number of critical periods is large. In terms of average empirical coverage, CWVS more often performs near the advertised coverage of 95% than GP (as do the other variable selection methods), though in some settings the estimated coverage is below or above 95%. GP fails to consistently cover the true effect size for the B_1 and B_2 sets of weeks.

These results suggest that GP may be over-smoothing when the length of the critical window is short, leading to difficulties in properly identifying and estimating significant risk parameters in the true critical window set and those bordering the true set. CWVS on the other hand has strong performance in all categories across all data generating settings due to the combination of GP smoothness and variable selection discreteness. In [Section S2 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, we present additional simulation studies to compare the performance of GP and CWVS when the true risk parameters are smoothly varying across pregnancy and their magnitudes are inflated. The results suggest similar performances of the methods when the risk parameter magnitudes are larger, regardless of their level of smoothness.

5. VERY PRETERM BIRTH IN NORTH CAROLINA, 2005–2008

We apply CWVS and each competing method from Section 4.2 to analyze critical windows of susceptibility for the VPTB outcome with respect to ozone (8 h maximum) and $PM_{2.5}$ (24 h average) weekly average exposures across pregnancy. In (3.1), we select $m = 27$ in order to only consider the impact of the first 27

Metric	Method	E1	M1	E4	M4	E7	M7
Critical weeks (B_1)							
MSE*1000	GP	12.57	16.76	1.23	2.72	0.68	0.78
	SSVS	1.43	1.28	3.66	3.62	5.05	4.91
	CWVS Sep.	1.55	1.47	1.90	1.71	1.89	1.40
	CWVS LMC	1.79	2.23	1.17	1.92	0.76	0.68
Coverage	GP	0.66	0.60	0.90	0.85	0.96	0.95
	SSVS	0.92	1.00	0.93	0.94	0.90	0.91
	CWVS Sep.	0.90	1.00	0.96	0.99	0.97	0.99
	CWVS LMC	0.88	0.98	0.98	0.96	0.99	0.99
Bordering critical weeks (B_2)							
MSE*1000	GP	8.34	3.67	2.78	2.75	1.25	1.53
	SSVS	2.40	2.24	3.14	4.13	4.02	4.32
	CWVS Sep.	2.95	2.49	2.62	3.29	2.16	2.74
	CWVS LMC	2.96	2.51	2.21	2.48	1.31	1.67
Coverage	GP	0.72	0.93	0.76	0.82	0.88	0.87
	SSVS	0.96	0.97	0.96	0.91	0.94	0.92
	CWVS Sep.	0.92	0.95	0.96	0.94	0.90	0.88
	CWVS LMC	0.88	0.94	0.92	0.89	0.90	0.90
	Non-critical, non-bordering critical weeks (B_3)						
MSE*1000	GP	2.25	2.71	0.55	0.61	0.45	0.50
	SSVS	1.64	2.04	2.32	2.36	2.92	2.98
	CWVS Sep.	1.86	2.25	1.23	1.47	0.96	1.27
	CWVS LMC	1.68	2.13	0.69	0.77	0.51	0.68
Coverage	GP	0.98	0.97	0.99	0.99	0.99	1.00
	SSVS	0.97	0.96	0.96	0.96	0.95	0.95
	CWVS Sep.	0.98	0.96	0.98	0.98	0.99	0.98
	CWVS LMC	0.94	0.95	1.00	0.99	0.99	0.99

Table 3. *Average mean squared error (MSE) and empirical coverage simulation study results across all simulation settings and sets B. Standard errors are presented in [Table S5 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at Biostatistics online*

weeks of pregnancy on VPTB risk. Extending past this point would lead to women giving birth and leaving the risk set, potentially biasing the results; an issue explored in depth by Chang *[and others](#page-14-0)* [\(2015](#page-14-0)). Use of the case–control study design means that we are unable to estimate VPTB prevalence in the population. However, the estimated odds ratios are still appropriate measures of risk [\(Seaman and Richardson,](#page-15-0) [2004](#page-15-0)).

The covariates in the **x***ⁱ* vector are available from the birth certificate and include maternal information: smoking status, diabetes status, age category ([15,25), [25–30), [30,35), [35,50]), race/ethnicity (White non-Hispanic, Black non-Hispanic, Hispanic, Other), education level (<high school, high school, and >high school), and marital status (married, unmarried); spatiotemporal information: an indicator for year of delivery to capture long-term trends in VPTB risk, an indicator for month of delivery to capture seasonal variation in VPTB risk, and the latitude and longitude of maternal residence at delivery (and their interaction) to account for spatial variability in VPTB risk across the state; and miscellaneous information on each birth: the trimester that prenatal care began (first, second, or third) and sex of the child. See [Table S2 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online for full information.

800 J. L. WARREN AND OTHERS

Continuous covariates were standardized to have a mean of 0 and standard deviation of 1 without loss of generality in order to stabilize model fitting. For the pollution data, we standardized exposure on each pregnancy week by subtracting off the median and dividing by the interquartile range (IQR), allowing for an IOR increase/decrease interpretation of the α (*t*) parameters. Figures S15 and S16 of the supplementary material available at *Biostatistics* online display the IQRs for each pollutant across each week of pregnancy in order to facilitate interpretation of the parameters. In general, the variability in IQRs across pregnancy week was low, with values ranging from 20.54–21.59 ppb and 5.52–5.74 μ g/m³ for ozone and PM_{2.5}, respectively.

From Section 3, we select $\sigma_{\beta}^2 = 1000$, $\alpha_{\phi} = \beta_{\phi} = 1$, and $\sigma_A^2 = 16$ to complete the model specification. Each of the methods are fit for 100 000 MCMC iterations, discarding the first 10 000 as a burn-in period. We thin the remaining 90 000 posterior samples resulting in 10 000 posterior samples with which to make inference. Convergence was assessed through visual inspection of individual parameter trace plots as well as calculation of the Geweke convergence diagnostic (Geweke *[and others](#page-14-0)*, [1991\)](#page-14-0). Neither tool suggested signs of non-convergence for any of the models. See [Table S6 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online for more details on the Geweke results.

In order to compare the results with traditional epidemiological analyses, we also fit logistic regression models that included the same set of covariates/confounders as CWVS along with average first and second trimester exposures to each pollutant (fit separately to each pollutant for comparison purposes). For CWVS, we summarize the posterior distribution of $\exp \left\{iQR_{tri}\sum_{t=1}^{13} \frac{\alpha(t)}{iQR_{t}}\right\}$ $\}$ for each pollutant where IQR_{tril} is the IQR of the first trimester average exposure and IQR_t is the IQR of the exposure during week *t* of pregnancy. This represents the impact of experiencing an IQR_{tri} increase in first trimester average exposure under the assumption that this increase is spread evenly across the first 13 weeks of pregnancy. We similarly summarize the posterior distribution for the second trimester exposure as well (weeks 14– 26). Results between the two models are expected to be comparable when viewed on the same exposure time scale.

5.1. *Results*

[Table S7 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online displays the covariate results (\mathbf{x}_i) from the CWVS model fit using the ozone exposure data. The results using the PM_{2.5} data were nearly identical and are not shown as a result. Maternal information including smoking during pregnancy, having less vs. greater than a high school education, having no previous live births, advanced age, being Black (non-Hispanic), and unmarried were found to be associated with increased risk of VPTB development. Having a male child and beginning prenatal care in the first trimester were also shown to be associated with elevated risk. The average correlation between weekly ozone exposures separated by 1 week is 0.79, dropping to 0.64 at 5 weeks, and 0.25 at 10 weeks. For $PM_{2.5}$, the comparable correlations are 0.50, 0.36, and 0.11.

Figures [1](#page-11-0) and [2](#page-12-0) display the graphical output from each of the competing models fit to the ozone exposure dataset, while similar results are shown in [Figures S17 and S18 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online for PM2.5. In Figure [1](#page-11-0) and [Figure S17 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, we present the posterior means and 95% CrIs of exp { $\alpha(t)$ } $\gamma(t) = 1$ (odds ratio scale) for CWVS, CWVS Separable, and SSVS. For GP, we present the same posterior summaries for exp {α (*t*)}. Results shown as dashed lines indicate that the pregnancy week is included in the critical window set (as defined in Section 4.3 for each method). In Figure [2](#page-12-0) and [Figure S18 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, we present the marginal posterior inclusion probabilities, $P(\gamma(t) = 1|Y)$, for the models that include a variable selection component (CWVS, CWVS Separable, and SSVS).

Fig. 1. Posterior mean and 95% credible interval results from the very preterm birth and ozone (8 h maximum) exposure analysis in North Carolina, 2005–2008. Results based on an IQR increase in weekly exposure. Weeks identified as part of the critical window set are shown as dashed lines. These definitions depend partly on the posterior inclusion probabilities in Figure 2 for the variable selection methods.

For the ozone results in Figure 1, CWVS (and CWVS Separable) identifies pregnancy weeks 4, 5, 7, 9, 10 as a the critical window set.We also observe that while CWVS and CWVS Separable agree with respect to critical window identification, CWVS has shorter CrIs. SSVS has relatively long CrIs and produces less stable results as it ignores temporal correlation during parameter estimation. CWVS and GP result in CrIs of similar length though GP identifies fewer critical windows than CWVS. This is potentially due to the over-smoothing of GP as is evident when observing the parameter estimates from weeks 4–6. Weeks 4 and 5 are nearly significant, but due to the behavior of week 6, they are pulled towards the null in the GP results. However, CWVS does not penalize weeks 4 and 5 as harshly and identifies these weeks as relevant. A similar phenomenon may be occurring in the opposite direction during week 11 in both models. In Figure [2,](#page-12-0) the smoothness of the posterior inclusion probabilities across pregnancy due to the modeling of temporal correlation in the η (*t*) parameters is evident. Similar patterns are seen across the three variable selection methods, with a clearer indication of the important weeks shown using CWVS. The 95% CrI for A_{21} is positive and excludes zero, indicating cross-correlation between θ (*t*) and η (*t*) for CWVS.

For the PM2.5 results in [Figures S17 and S18 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, CWVS (and CWVS Separable) suggest no critical windows of susceptibility while GP identifies two protective periods. The first 2 weeks of pregnancy have 95% CrIs that do not include zero but posterior inclusion probabilities below 0.50 for CWVS Separable. As a result, these 2 weeks are not identified as part of the critical window set. The smoothness in parameter estimation is once again evident for CWVS from these figures. The posterior distribution of A_{21} is centered around zero, suggesting no cross-correlation.

Fig. 2. Posterior inclusion probability results from the very preterm birth and ozone (8 h maximum) exposure analysis in North Carolina, 2005–2008.

For the traditional epidemiological analysis, the odds of VPTB increased 47.47% for an IQR increase of 18.03 ppb in average first trimester ozone exposure (odds ratio 1.47; 95% CI 1.31, 1.66), while no statistically significant association was seen in the second trimester. CWVS estimated the comparable first trimester odds ratio as 1.43 (95% CrI 1.25, 1.63) with a similar null response in the second trimester. For PM_{2.5} in the second trimester, the odds decreased 9.40% for an IQR increase of 3.70 μ g/m³ (odds ratio 0.91; 95% CI 0.84, 0.98) using the traditional model, while no statistically significant association was seen in the first trimester. CWVS estimated the comparable second trimester odds ratio as 0.94 (95% CrI: 0.87, 1.00), suggesting no impact of exposure, with a similar null result observed in the first trimester.

In the [supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online, we present a sensitivity analysis to the choice of prior distributions for the temporal smoothness parameters (ϕ_i) and an investigation into the need for spatial random effects in the modeling framework. We find that the primary results are generally robust to the choice of prior distribution for ϕ _{*j*} (see [Figures S19 and S20 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online). By analyzing the Pearson residuals obtained from the traditional epidemiological models using a spatial semivariogram analysis, we conclude that there is no meaningful residual spatial correlation in either the ozone or $PM_{2.5}$ results (see [Figures S21 and S22 of the supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) available at *Biostatistics* online).

6. DISCUSSION

Through simulation, we have shown that CWVS has the ability to more accurately identify critical windows of susceptibility and improve risk parameter estimation in a number of different data generating settings compared to existing methods. By incorporating a model for temporal correlation into the variable selection process, CWVS maintains smoothness in parameter estimation as well as the posterior inclusion probabilities. Use of the LMC allows for flexibility in the degree of smoothness for the different processes as well as a non-separable cross-covariance structure. Unlike SSVS, which struggles due to the high correlation between predictors that leads to high correlation between risk parameters, and GP, which can often over-smooth during parameter estimation, CWVS represents a balance between both modeling strategies. When using CWVS and other variable selection methods that account for adjustment uncertainty to estimate causal effects, it is important to consider a weighted average effect across probable models instead of relying on inference from a single selected model. The confounders/covariates in CWVS (**x***i*) do not include a variable selection component. Therefore, all models being averaged across include the same set of confounders/covariates and this may help to stabilize estimation of the environmental exposure risk paramete[rs that are typically smaller in magnitude. Please see](#page-16-0) Wang *[and others](#page-15-0)* [\(2012](#page-15-0)) and Zigler and Dominici [\(2014](#page-16-0)) for further discussion.

When used to examine VPTB risk due to ambient air pollution exposure in NC, 2005–2008, CWVS identified a different critical window set than the standard GP model. However, both methods suggested that elevated exposure to ozone during selected weeks in the first trimester was associated with increased risk of VPTB. The identified critical window set from CWVS suggested the possibility of multiple critical periods (weeks 4, 5, 7, 9, 10). It is possible that the true critical window set also includes weeks 6 and 8 and that these weeks were omitted due different sources of error leading to exposure misclassification or due to power issues because of differing risk magnitudes across the weeks. However, the simulation study results suggest that CWVS has improved estimation and critical window set identification properties than the competing methods. For $PM_{2.5}$, CWVS suggested no relevant weeks of interest while GP identified a protective effect in the first 2 weeks of pregnancy. CWVS was also shown to produce results with shorter CrIs in comparison to CWVS Separable and SSVS. Standard logistic regression analyses that included the same set of covariates/confounders confirmed these findings.

A strength of this work is use of the dataset of pollution exposure estimates from the EPA fused downscaler statistical model that has improved spatial/temporal coverage compared to observed monitoring data. However, use of the residence at delivery to link exposures results in exposure misclassification as some women move between date of conception and delivery date. A summary of recent work in this area suggests that 9–32% of pregnant women move at some point during pregnancy, while the distances actually traveled tend to be short; typically less than 10 kilometers [\(Bell and Belanger,](#page-14-0) [2012\)](#page-14-0). Additionally, Warren *[and others](#page-15-0)* [\(2017\)](#page-15-0) found that critical window estimation was remarkably robust to exposure misclassification due to this movement in the population of pregnant women. Another limitation is the use of outdoor ambient air pollution concentration as a surrogate for individual exposure.

Future work in this area should focus on relaxing the linearity assumption often made in this framework. Exposure at a given time period is almost always assumed to be linearly associated with risk, with different prior structures specified for these regression parameters. However, non-linear dose–response relationships have been observed in previous studies and accounting for these non-linear associations may uncover previously unobserved information regarding exposure timing and risk. Extensions to the modeling of multiple pollutants will also be needed given the importance of understanding the impact of pollution mixtures on human health. Additionally, alternative smoothing and variable selection methods within the functional data analysis setting may be useful for identifying critical windows. Finally, future work and uses of CWVS should consider the level of correlation between the temporal predictors. Our simulation study results were based on weekly ozone exposures in NC that may be more or less correlated across time than alternative exposures and/or different averaging times (e.g., daily, monthly).

804 J. L. WARREN AND OTHERS

SUPPLEMENTARY MATERIAL

[Supplementary material](https://academic.oup.com/biostatistics/article-lookup/doi/10.1093/biostatistics/kxz006#supplementary-data) is available at http://biostatistics.oxfordjournals.org. The R package CWVS and worked example are available at https://github.com/warrenjl/CWVS.

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