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Detection of SARS-CoV-2 in raw and treated wastewater in Germany – Suitability for COVID-19 surveillance and potential transmission risks

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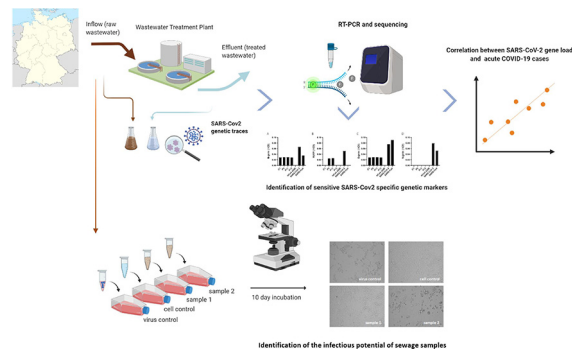
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HIGHLIGHTS

- The first study that reports the detection of SARS-CoV-2 in wastewater in Germany using RT-qPCR.
- The presence of SARS-CoV-2 was confirmed by sequencing, but also the risk of false-positive results has been elucidated.
- In raw wastewater, 3.0 to 20 gene equivalents/mL was found in raw wastewater.
- The replication potential tests were negative for wastewater samples.
- Sanger sequencing was required to differentiate the genetic pattern clearly.

GRAPHICAL ABSTRACT



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ABSTRACT

Wastewater-based monitoring of the spread of the new SARS-CoV-2 virus, also referred to as wastewater-based epidemiology (WBE), has been suggested as a tool to support epidemiology. An extensive sampling campaign, including nine municipal wastewater treatment plants, has been conducted in different cities of the Federal State of North Rhine-Westphalia (Germany) on the same day in April 2020, close to the first peak of the corona crisis. Samples were processed and analysed for a set of SARS-CoV-2-specific genes, as well as pan-genotypic gene sequences also covering other coronavirus types, using reverse transcription-quantitative polymerase chain reaction (RT-qPCR). Additionally, a comprehensive set of chemical reference parameters and bioindicators was analysed to characterize the wastewater quality and composition. Results of the RT-qPCR based gene analysis indicate the presence of SARS-CoV-2 genetic traces in different raw wastewaters. Furthermore, selected samples have been sequenced using Sanger technology to confirm the specificity of the RT-qPCR and the origin of the coronavirus. A comparison of the particle-bound and the dissolved portion of SARS-CoV-2 virus genes shows that

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quantifications must not neglect the solid-phase reservoir. The infectivity of the raw wastewater has also been assessed by viral outgrowth assay with a potential SARS-CoV-2 host cell line in vitro, which were not infected when exposed to the samples. This first evidence suggests that wastewater might be no major route for transmission to humans. Our findings draw attention to the need for further methodological and molecular assay validation for enveloped viruses in wastewater.

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1. Introduction

The current SARS-CoV-2 pandemic has far-reaching global consequences on public health, economic activities, and societies as a whole, which are unprecedented in many respects and cannot be fully assessed yet (WHO, 2020). The number and proportion of persons infected with COVID-19 are mainly determined based on individual testing and laboratory-based bio-molecular diagnostics using, e.g., reverse transcription-quantitative polymerase chain reaction (RT-qPCR). COVID-19 cases are reported by regional health authorities and aggregated on the level of, e.g., the Federal States in Germany, as well as on a national level (RKI, 2020). It is expected that, based on the individual testing that is often triggered by symptoms of test candidates or their respective risk profile, the actual state of infection in a specific region can only be very roughly estimated (Wurtzer et al., 2020). A study from China has also indicated that non-detected cases can make a significant contribution to the development of the infection as a whole (Li et al., 2020). Various researchers have shown that SARS-CoV-2 RNA is detectable in the feces of infected persons (e.g., Amirian, 2020). Furthermore, SARS-CoV-2 RNA can be found in stool samples weeks after the infection is no longer detectable in oral swab samples (Y. Wu et al., 2020). Therefore, it must be systematically assessed whether the virus might, in addition to respiratory droplets, also be transmitted via feces in wastewater (Nemudryi et al., 2020; Tang et al., 2020; Wang et al., 2020; Y. Wu et al., 2020; F. Wu et al., 2020). It could be shown that the duration of viral shedding differed among patients between 14 and 21 days past the onset of the

infection (Wu et al. 2020, Xu et al., 2020). Furthermore, the magnitude of shedding varied from 10^2 to 10^8 RNA copies per gram feces (Lescure et al., 2020; Pan et al., 2020). Wastewater-based epidemiology (WBE) has been suggested as a potentially useful complementary tool to gain insights into the degree of infection spread in a population (Wu et al. 2020, Choi et al., 2018, Rodriguez-Manzano et al., 2010). Recently, various studies detected SARS-CoV-2 RNA in wastewater worldwide (cf. Table 1) like the Netherlands (Medema et al., 2020), France (Wurtzer et al., 2020), USA (Wu et al. 2020), Australia (Ahmed et al., 2020), and Italy (La Rosa et al., 2020). Wurtzer et al. (2020) reported the analysis of SARS-CoV-2 genes in the Greater Paris (France) area and were able to correlate trends in gene occurrence in the wastewater of different wastewater treatment plants (WWTP) with the number of infected individuals. Medema et al. (2020) have shown a good correlation between the number of COVID-19 cases and the gene concentration in the wastewater of different Dutch cities.

However, the studies differed in, e.g., the type of samples, processing procedure, and targeted genes (like N1, N2, N3, and ORF1ab) in RT-qPCR analysis. Only a few studies were complemented by infectivity tests of the genes to determine whether the genetic material was present in intact virus particles or as free nucleic acids. Furthermore, only a few studies comprised phylogenetic analyses to better identify the genetic profile of the obtained material. An overview of recent studies conducted and reported in 2020 is given in Table 1.

Carducci et al. (2020) reviewed the current knowledge about the fate of coronaviruses in different environmental compartments,

Table 1
Overview of selected SARS-CoV2 studies in wastewater conducted in 2020.

Study region	Genetic traces/genes analysed	Results	Reference
Paris (France)	RdRp	Samples in three WWTPs positive between 5th March and 23th April	Wurtzer et al., 2020
Milan and Rome (Italy)	ORF1ab, S	6 out of 12 samples positive in raw wastewater	La Rosa et al., 2020
Netherlands	N1, N2, N3, E	4 out of 7 wastewater samples positive at the beginning of March. 9 out of 9 wastewater samples positive in the middle of March	Medema et al., 2020
Milan (Italy) (WWTP and river)	ORF1ab, N, E	First sampling: 3 out of 4 raw wastewater samples positive 2 out of 2 effluent samples negative 2 out of 2 river samples positive	Rimoldi et al., 2020
Brisbane (Australia)	N and confirmation via Sanger and MiSeq Illumina sequencing	2 out of 9 samples positive in raw wastewater	Ahmed et al., 2020
Massachusetts (USA)	N1, N2, N3,	10 out of 10 raw wastewater samples positive	F. Wu et al., 2020
Israel (different cities and facilities)	E	2 out of 15 samples positive in March 2020 9 out of 11 samples positive in April 2020	Bar Or et al., 2020
Bozeman, Montana (USA)	N1, N2	7 out of 7 positive in March/April 2020	Nemudryi et al., 2020
Istanbul (Turkey)	RdRp	9 out of 9 sludge samples positive	Kocamemi et al., 2020
Valencia (Spain)	N1, N2, N3	35 out of 42 influent samples positive 2 out of 18 secondary treated samples positive 0 out of 12 tertiary effluent samples positive	Randazzo et al., 2020
Yamanashi Prefecture, Japan	N1, N2	None of 5 influent samples positive 1 out of 5 secondary effluent samples positive None out of 3 river water samples positive	Haramoto et al., 2020
Ahmedabad, Gujara, India	ORF1ab, N and S	2 out of 2 influent samples positive 2 out of 2 effluent samples negative	Kumar et al., 2020
Louisiana, USA	N1 and N2	2 out of 15 raw wastewater samples positive Effluent samples all negative	Sherchan et al., 2020
Quito, Ecuador	N1 and N2	3 out of 3 river water samples positive	Guerrero-Latorre et al., 2020

including wastewater. They revealed that only 22 studies were done since 1978 and that more in-depth studies are needed to gain a better understanding of the possible circulation of SARS-CoV-2 in water systems.

In particular, there are still open methodological questions on how to derive correlations between wastewater-based analysis and acute infection cases in the sampled region, as well as concerning the infectivity of the wastewater-borne genetic material of viral origin (Lodder and de Roda Husman, 2020). Additionally, little is known about the potential distribution of the virus in the aquatic environment and the survival of SARS-CoV-2 in water (Naddeo and Liu, 2020). On the one hand, it could be shown that SARS-CoV-2 from treated wastewater was potentially still infectious (Wurtzer et al., 2020). On the other hand, in the study of Rimoldi et al. (2020), the test of viability showed that pathogenicity of the virus in wastewater was negligible. Therefore, they assumed that the risk for public health was not significant.

The present study encompasses a comprehensive analysis of a set of samples from a coordinated campaign from nine wastewater treatment plants (WWTP) in the Federal State of North Rhine-Westphalia, Germany. The liquid and solid phase of the inflow samples were analysed, and a broad spectrum of reference parameters (e.g., bioindicators, nutrients, sum parameters, pharmaceuticals) was measured. Moreover, several SARS-CoV-2 specific genes were compared for their sensitivity and selectivity. A confirmation was conducted via Sanger sequencing. Replication tests were conducted to check the viability of the viral genetic material.

2. Material and methods

2.1. Sewage sampling

Nine municipal WWTP operated by six different water boards were selected for analysis throughout North-Rhine Westphalia (Germany). The plants differed in their design capacity, treatment processes, and connected catchment area characteristics (Table 2). Using installed autosampler devices, the operators of the WWTP collected 24 h flow-dependent composite samples on April 8th, 2020, during dry-weather conditions, either midnight to midnight or between the morning of April 8th, and the morning of April 9th. In addition to raw wastewater (sewage) inflow samples collected after

the sand trap, treated sewage was sampled at selected locations (Table 2). All samples were transported to the laboratory in glass flasks on melting ice on the day of sampling.

For controls, sewage samples collected in earlier dry-weather sampling campaigns prior to the assumed onset of the COVID-19 pandemics in Germany that were kept frozen at $-18\text{ }^{\circ}\text{C}$ until analysis were used. Controls were sampled at WWTP Moers-Gerdt (C1-inflow sampled on July 18th, 2019), and Aachen-Soers (C2-effluent sampled in November 2018, C3-effluent in September 2019, and C4-inflow in January early 2017).

2.2. Sample processing for isolation and quantification of viral RNA

Frozen wastewater samples were thawed at $4\text{ }^{\circ}\text{C}$, and a total of 45 mL were further processed. First, wastewater was centrifuged at $4700\times g$ for 30 min without break, and the clear supernatant was harvested. Purified wastewater was then concentrated using centrifugal ultrafiltration units (Amicon® Ultra-15 Centrifugal Filter Unit, Sigma). Therefore, 15 mL of wastewater was added to the filter unit and centrifuged for 15 min at $3500\times g$, and the concentrated supernatant was harvested. This was repeated twice until 45 mL of wastewater were completely concentrated (final volume of concentrated supernatant approximately 450 μL). For the solid phase of the wastewater sample, the pellet of the 45 mL sample was first washed with deionized water to remove aqueous remains from the sample and centrifuged at $4700\times g$ for 5 min before being resuspended in 150 μL deionized water and centrifuged at $4700\times g$ for 5 min again. A volume of 150 μL of the supernatant was then harvested for further RNA extraction, and the mass was determined (approx. 150 mg in influent and 5–10 mg in effluent). Contaminated equipment (e.g., reaction tubes, tips, filter units) were collected and autoclaved according to the daily cleaning program in the lab.

2.3. RNA extraction

RNA was isolated using the NucleoSpin RNA Virus kit (Macherey Nagel) following the manufacturer's instructions. Briefly, 150 μL supernatant was mixed with lysis buffer supplemented with carrier RNA. After binding on silica membranes, samples were washed several times and eluted in 50 μL RNase-free water. Isolated RNA was stored at $-80\text{ }^{\circ}\text{C}$.

Table 2
Key properties of WWTPs sampled.

Acronym	WWTP	Treatment process	Design capacity ^a PE _{design} [-]	Nominal number of connected residents ^b PE _{nom} [-]	Annual wastewater flow in 2019 ^a [m ³ /a]	Actual wastewater flow on April 8th, 2020 ^b Q _{actual} [m ³ /day]	Sample ID and sampling location on April 8th, 2020
KLEM	Kläranlage Emscher-mündung ^c	AS	2,400,000	2,228,933	347,602,053	1,159,488	P1 Inflow
							P2 Effluent
DU	Duisburg-Kaßlerfeld	AS	450,000	257,262	33,028,604	82,957	P3 Inflow
E	Essen-Süd	AS	135,000	119,226	12,543,902	29,695	P4 Inflow
MO	Moers-Gerdt	AS	250,000	118,018	8,626,863	25,299	P5 Inflow
MG	Mönchenglad-bach-Neuwerk	AS	632,500	494,077	33,382,628	85,131	P6 Inflow
EUS	Euskirchen-Kessenich	AS	132,000	69,496	8,213,543	20,285	P7 Inflow
ERFT	Bergheim-Kenten	AS	120,000	98,898	7,331,398	13,432	P8 Inflow
DN	Düren	AS	310,000	134,723	20,104,329	54,253	P9 Inflow
							P10 Effluent
							P11 Inflow
AC	Aachen-Soers	AS followed by full-scale ozonation and filtration	458,000	205,000	25,057,198	66,052	P12 After tertiary treatment
							P13 Effluent after ozonation and filtration

AS: activated sludge treatment, PE: population equivalent.

^a Data from ELWAS (2020).

^b Data provided by operator.

^c WWTP KLEM treats the complete flow of river Emscher. PE_{nom} refers to the complete upstream catchment area, including PE_{nom} of all upstream WWTPs discharging into river Emscher.

2.4. SARS-CoV-2 specific quantitative RT-qPCR

RNA was analysed by OneStep RT-qPCR using Luna Universal Probe One-Step RT-qPCR Kit (New England Biolabs) or LightCycler® Multiplex RNA Virus Master (Roche) and the CFX96 Real-Time System, with a C1000 Touch Thermal Cycler (Bio-Rad). Primer pairs for E-, N- and M-gene (Toptan et al., 2020) specific PCRs were used in equimolar concentrations (0.4 μ M each per reaction), while RdRP-primer pairs were used according to Corman et al. (2020) with 0.6 μ M and 0.8 μ M for forward and reverse primers, respectively. The sequences of primers and probes are found in Table 3. For probe-based Luna Universal One-Step RT-qPCR kits, 2 μ L of RNA were subjected to reverse transcription performed at 55 °C for 10 min. Initial denaturation was performed for 1 min at 95 °C followed by 45 cycles of denaturation for 10 s and combined annealing and extension for 30 s at 60 °C. For the LightCycler® Multiplex RNA Virus Master kit, 5 μ L of template RNA were used. Reverse transcription was performed at 55 °C for 10 min. Initial denaturation was allowed for 30 s at 95 °C followed by 45 cycles of denaturation for 5 s, extension for 30 s at 60 °C and final cool-down to 40 °C for 30 s. The PCR runs were analysed with Bio-Rad CFX Manager software version 3.1 (Bio-Rad Laboratories). For quantifications, standard curves using plasmid DNA (RdRP) or in vitro transcribed RNA (M-gene) were used as described previously (Toptan et al., 2020). Depending on the RT-qPCR for M-gene or RdRP the detection limit was 200 copies per reaction. If not indicated otherwise, a reaction was considered positive if the cycle threshold (C_T) was below 40 cycles. To calculate from gene equivalents per reaction back to copies per mL, a PCR correction factor (cf) was determined with Luna Universal Probe One-Step RT-qPCR Kit (cf = 0.0037) and LightCycler® Multiplex RNA Virus Master (cf = 0.0015), respectively. To control PCR, viral RNA of SARS-CoV-2 (isolate SARS-CoV-2-FFM1) was used as a positive control (Hoehl et al., 2020; Toptan et al., 2020) and water as a negative control. Furthermore, to verify specificity and sensitivity, RNA of SARS-CoV (isolate SARS-CoV-FFM1) (Drosten et al., 2003) and HCoV-229E from cell culture supernatant was isolated and used in PCR.

2.5. Sanger sequencing analysis

RdRP-PCR products (wastewater sample P12, WTPP Aachen Soers after tertiary treatment; April 8th, 2020, effluent C2 sampled in November 2018 and inflow C4 sampled in January early 2017) were purified using the GenElute™ PCR Clean-Up Kit according to manufacturer's instructions and sequenced from both ends on the Applied Biosystems 3730 DNA Analyzer platform at the BiK-F Laboratory Centre, Frankfurt, Germany. Sequences were identified in terms of the closest homology sequence using BLAST, <https://blast.ncbi.nlm.nih.gov/Blast.cgi> (database Nucleotide collection (nr/nt) using Blastn (optimized for somewhat similar sequences) or Megablast (optimized for highly similar sequences)). Additional alignments were performed using Clustal Omega program for multiple sequence alignment (EMBL-EBI; <https://www.ebi.ac.uk/Tools/msa/clustalo/>). Results of the BLAST search and alignments can be found in Supplementary Fig. S1.

Results of the BLAST search and alignments can be found in Supplementary Fig. S1.

2.6. Replication studies with SARS-CoV-2 positive wastewater

Purified wastewater samples that tested positive for SARS-CoV-2 (P2, P5, P11, P12) were investigated for replication-competent viruses following a recently published procedure (Hoehl et al., 2020). Potential infectious work was performed under biosafety level 3 (BSL-3) conditions in a BSL-3 facility according to the Committee on Biological Agents (ABAS) and the Central Committee for Biological Safety (ZKBS). Briefly, wastewater samples were processed as described in Section 2.2. A human epithelial cell line from colon adenocarcinoma (Caco-2) was grown in cell culture tubes and infected with a 1:10 dilution of purified wastewater in Minimum Essential Media (MEM) supplemented with 1% fetal calf serum, 2% L-glutamine and 1% penicillin/streptomycin. As a control, cells were infected with SARS-CoV-2-containing cell culture supernatant. Cells were cultured for up to 10 days post-infection. Infection with replication-competent virus results in the development of a cytopathic effect, which was measured visually by light microscopy.

2.7. Analysis of chemical and physicochemical reference parameters

All physicochemical and chemical parameters were analysed immediately after sampling following DIN, EN, or ISO standard protocols. If this was not possible, the samples were chemically stabilized and then measured within the approved timeline of the standard protocols. The pH and conductivity were measured using a sensION MM374 Hach Lange instrument (EN ISO 10523 (2012-04), EN 27888 (1993-11)). For the content of the dry residues, a defined volume of wastewater sample was dried at 105 °C for 24 h (DIN 38409-1 (1987-01)), and the residue was weighed by a Sartorius A 120 S balance. Ready-to-use-cell test kits were used to analyse the chemical oxygen demand (COD) with a Macherey & Nagel PF12 and Vario 4 (test kits 985022, 985033 and 98029, ISO 15705 (2003-01)). Total organic carbon (TOC) and total bounded nitrogen (TN_b) were analysed via combustion using a Shimadzu carbon/nitrogen analyser (EN 1484 (1997-08), EN 12260 (2003-12)). For the analysis of ammonia, nitrate, nitrite, total organic nitrogen, ortho-phosphate, and other ions such as chloride and sulfate, a Gallery discrete analyser with photometric and turbidity methods were used after filtration through a 0.45- μ m filter (Thermo Fisher Scientific, ISO 15923-1 (2014-07)).

Two different types of population size markers (exogenous and endogenous) were measured by liquid chromatography coupled with mass spectrometric detection (LC-MS) and photometric detection with a discrete analyser (Choi et al., 2018). LC-MS analysis for the exogenous population size indicators was carried out by using an HP Series 1100 LC coupled to an MS/MS-System (Thermo Finnigan TSQ Quantum, Thermo Fisher Scientific) in positive electrospray ionization mode. Before measurement, 100-mL wastewater samples were spiked with internal standards (Supplementary Table S1) and then enriched and

Table 3

Sequences of primers and probe established and verified for the detection of SARS-CoV-2 (in patient material) by RT-qPCR.

Oligo name	Oligonucleotide sequence (5'-3')	Gene	Reference
E_Sarbeco_F1	ACAGGTACGTTAATAGTAAATAGCGT	E-gene	Corman et al., 2020
E_Sarbeco_R2	ATATTGCAGCAGTACGCACACA		
E_Sarbeco_P1	6-Fam AACTAGCCATCCTTACTGCGCTTCG BBQ1		
RdRP_SARSr-F2	GTGARATGGTCATGTGTGGCGG	RdRP	Corman et al., 2020
RdRP_SARSr-R1	CARATGTTAAASACACTATTAGCATA		
RdRP_SARSr-P2	6-Fam CAGGTGGAACCTCATCAGGAGATGC BBQ1		
M-475-F	TGTGACATCAAGGACCTGCC	M-gene	Toptan et al., 2020
M-574-R	CTGAGTCACTGCTACACGC		
M-507-P	6-Fam TGTTCATCATCAGGACGC BHQ1		
SARS-CoV-2 N-gene F	TGGCCGCAAATTGCACAATT	N-gene	Toptan et al., 2020
SARS-CoV-2 N-gene R	TGTAGGTCAACCAGCTTCCC		
SARS-CoV-2 Probe-N	6-Fam CGCATTGGCATGGAAGTCAC BHQ1		

cleaned on an Oasis HLB cartridge, diluted with methanol, dried under nitrogen flow, and re-dissolved in 1 mL methanol/water (50/50 v/v) for injection. For separation, an RP C18-column with polar endcapping was used (Hypersil GOLD™ aQ C18, Thermo Fisher Scientific). The method was based on a German standard DIN 38407-47: 2017-07. The mobile phase was a gradient containing Millipore water and methanol. 2 mM ammonium acetate and 0.1% acetic acid were added to the methanol and water. The separation was run after injection of 10 µL sample with a linear gradient starting with 20% up to 90% and ending again with 20% methanol. All data were collected in SRM mode by using mass-labeled internal standards (Supplementary Table S1) and Xcalibur 2.0.7. SP1 software for data acquisition and quantification. The LOQs are in the range of 10 to 30 ng/L.

As endogenous parameters, creatinine and urea were quantified after filtration via a 0.45 µm filter by using a Gallery discrete analyser (Thermo Fisher Scientific). Creatinine was measured by photometry at 540 nm as quinonimine-chromogen after enzymatic reactions of creatine with creatinase, sarcosine oxidase, and ascorbate oxidase with a modified clinical method using Thermo Fisher Scientific reagent kit (enzyme. colortest PAP 981896) (Dörner, 2003). The Urea-method is based on a standardized bathwater method. In the first step, the contents of urea and ammonium are detected simultaneously. After enzymatic reaction of urea with urease to ammonium, ammonium reacts with nicotinamide adenine dinucleotide phosphate (NADPH) in a buffered solution at pH 8. The loss of NADPH is directly proportional to the ammonium content (Thermo Fisher Scientific test kit 981820). The urea content is calculated by subtraction of the ammonium content (ISO 15923-1 (2014-07)) from the total content of urea + ammonium.

2.8. Prevalence of reported COVID-19 cases in WWTP catchment area

In Germany, local health authorities report the number of COVID-19 cases confirmed by laboratory diagnosis after aggregation from the community to district level to state and federal authorities. Whenever available, the cumulative prevalence and the cumulative number of COVID-19 patients recovered and deceased were obtained from community-level reports published on the homepages of the responsible local health authorities on April 9th, 2020. If community-level reports were not available, district-level data published by the state health ministry of North Rhine-Westphalia (MAGS) on April 9th, 2020, was used. Since the catchment areas of WWTPs rarely correspond to administrative boundaries, the cases were estimated by calculating weighted sums relative to the residents of each community connected to that sewer network. Acute prevalence was calculated by subtracting reported recovered and deceased patients (Table 4). Nominal incidences (I_{nom}) were

Table 4

Estimated nominal cumulative and acute prevalence of COVID-19 cases on April 8th, 2020, in the catchment areas of the WWTP studied.

WWTP catchment area	Cumulative prevalence of COVID-19		Acute prevalence of COVID-19	
	Number of cases (N_{nom})	Nominal incidence per 100,000 ($I_{nom} = N_{nom} / PE_{nom}$)	Number of cases (N_{nom})	Nominal incidence per 100,000 ($I_{nom} = N_{nom} / PE_{nom}$)
KLEM	1924	86	1037	47
DU	271	105	144	56
E	117	98	62	52
MO	85	72	36	30
MG	554	112	299	60
EUS	110	158	74	107
ERFT	218	220	172	174
DN	158	117	82	61
AC	292	142	128	62

calculated by dividing the estimated number of cases (N_{nom}) in a catchment area by the nominal number of connected residents to the sewer (PE_{nom}). We note that these are estimates with considerable uncertainty since cases might be unevenly distributed between neighbouring districts, and recovered cases are often not based on laboratory diagnosis but the end of ordered quarantine. In addition, the actual number of persons staying in the catchment area on the day of sampling (PE_{actual}) may also differ from the nominal number of connected residents (PE_{nom}). Treating the complete flow of the river Emscher, WWTP KLEM is partly treating water that was treated upstream by other treatment plants. Based on the assumption of poor removal of SARS-CoV-2 in conventional WWTP, reported cases for KLEM in Table 4 are upper estimates covering the complete upstream catchment.

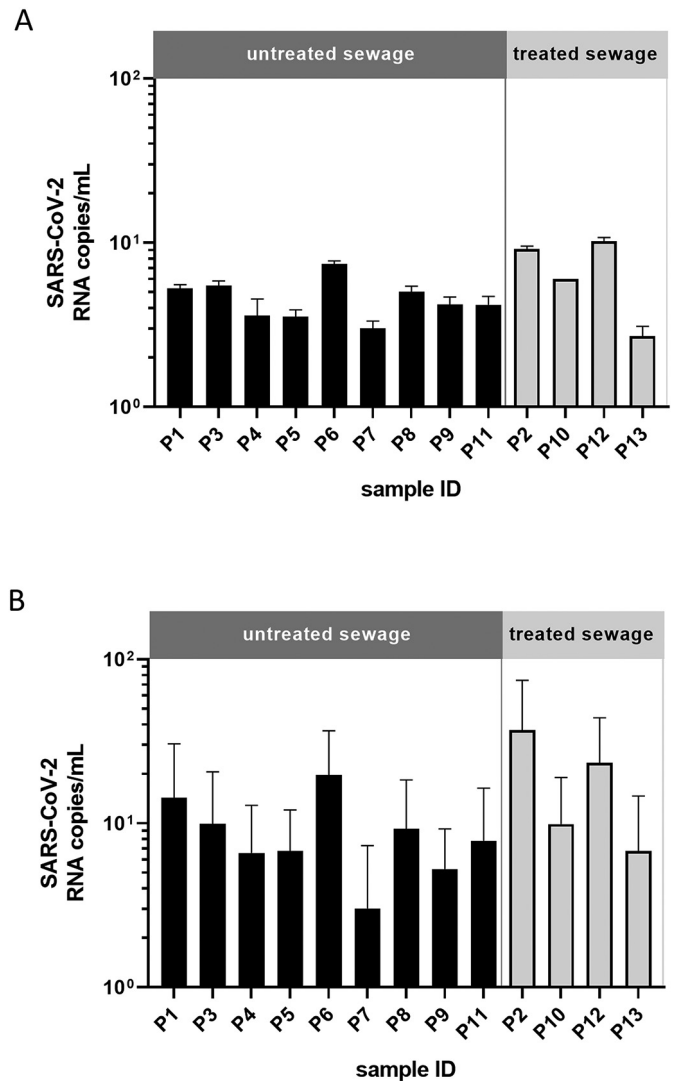


Fig. 1. SARS-CoV-2 specific RNA fragment detected by (A) M-gene and (B) RdRP RT-qPCR in the aqueous phase of untreated and treated municipal sewage. Error bars indicate one standard deviation of duplicate samples. Results of a single PCR measurement are shown for M-gene. For M-gene RT-qPCR, C_T values for the standard ranged between 8.6 (standard 1 = 10^9) and 38.4 (standard 6 = 10^2). Values of tested wastewater above C_T 39 were considered negative for SARS-CoV-2. Standard curve was calculated using Bio-Rad CFX Manager software with $E = 61.4\%$, $R^2 = 0.952$ and slope = -4.813 . Results of two independent PCR measurements of the same samples are shown for RdRp. For RdRP RT-qPCR C_T values for standard range between 18 (standard 1 = 10^8) and 37 (standard 6 = 10^2). Values of tested wastewater above C_T 38 were considered negative for SARS-CoV-2. Standard curve was calculated using Bio-Rad CFX Manager software with $E = 96.5\%$, $R^2 = 0.990$ and slope = -3.408 .

3. Results

3.1. SARS-CoV-2 detection in wastewater: aqueous phase

Wastewater samples from 9 distinct municipal WWTP were collected in April 2020 during the pandemic outbreak of SARS-CoV-2 in North Rhine Westphalia to investigate the presence of virus-specific RNA in untreated and treated wastewater samples by quantitative RT-qPCR. Using well-established quantitative RT-qPCRs for viral M (Fig. 1A) and RdRP genes (Fig. 1B), PCRs provided evidence of SARS-CoV-2-specific RNA traces in the aqueous phase of the tested sewage samples. SARS-CoV-2 RNA copies were found in a range from 3.0 and 20 gene equivalents per mL in the inflow, and 2.7 to 37 gene equivalents in the WWTP effluent (Fig. 1), which is in line with other reports in the Netherlands and Australia (Medema et al., 2020; Ahmed et al., 2020). C_T values for the tested wastewater samples ranged between 32 and 35 for M-gene, and 33 and 37 for RdRP. Interestingly, all wastewater samples showed detectable virus RNA. The higher number of SARS-CoV-2 RNA copies and the higher standard deviation found for the viral RdRP analyses might possibly be attributed to variability between single PCR measurements given by RNA isolation and storage conditions for RNA extraction compared to samples analysed for M-genes. However, the tendencies obtained with both PCR targets were consistent.

3.2. Specificity and sensitivity of different SARS-CoV-2 target genes in RT-qPCR

Given the complex nature of municipal sewage, disturbing factors accumulated during sample processing such as nucleic acids of other human coronaviruses circulating in the human population and potentially concentrated in the wastewater samples could influence analysis results. To verify the specificity and sensitivity of the PCR, we compared different viral genes (N-gene, E-gene, M-gene, and RdRP) and performed RT-qPCR with three different samples of sewage from April 2020, one control sample (C1) and RNA of three different coronavirus isolates (SARS-CoV-2, SARS-CoV, and HCoV-229E).

High specificity for SARS-CoV-2 with good sensitivity was found for the viral RdRP gene compared to the other tested viral genes (Fig. 2). Although no signal was detected for HCoV-229E with the RdRP gene PCRs, unspecific binding of primer and probe to other human coronaviruses like HCoV-NL63, HCoV-OC43, or HCoV-HKU1 cannot be excluded. Previous evaluation of the primers and probe within a different test setting analysing patient material or cell culture-derived SARS-CoV-2 showed minor cross-reactivity with other human coronaviruses (data not shown). Using a more sensitive and specific PCR for RdRP and M-gene may explain the detection of SARS-CoV-2 in all wastewater samples compared to other studies

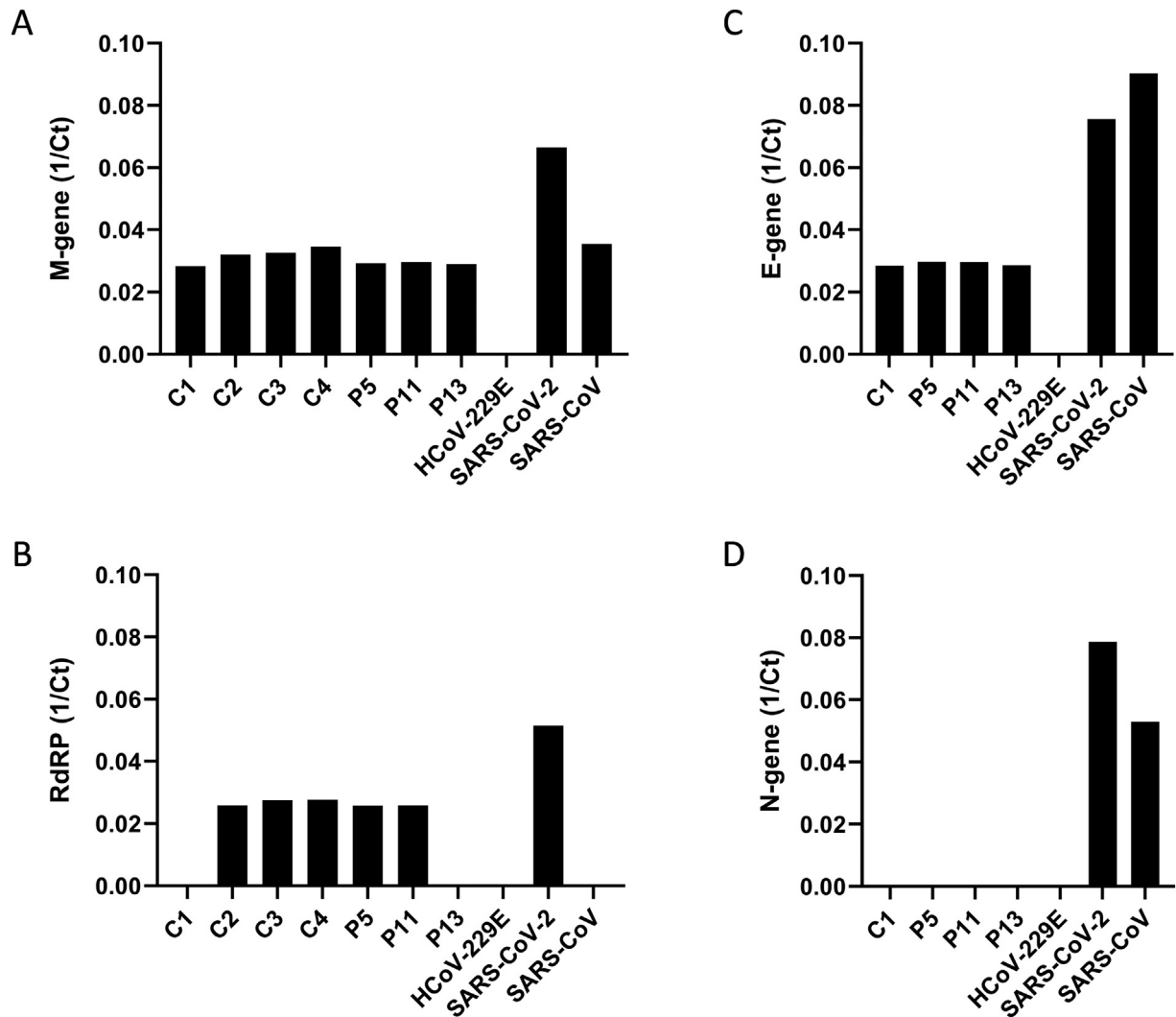


Fig. 2. Specificity and sensitivity of viral (A) M-gene, (B) RdRP-gene, (C) E-gene and (D) N-gene RT-qPCR with respect to control C1 sampled in July 2019 before the known onset of the pandemics in Germany, untreated sewage P5, P11, P13 sampled during the pandemics, SARS-CoV-2 control, and two other human coronaviruses HCoV-229E and SARS-CoV. Results are from wastewater samples measured in duplicates in a single PCR. Standard curve for M-gene PCR was calculated using *Bio-Rad CFX Manager* software with $E = 90.8\%$, $R^2 = 0.996$ and slope = -3.565 . Efficiencies of other primer pairs were not calculated by serial dilution.

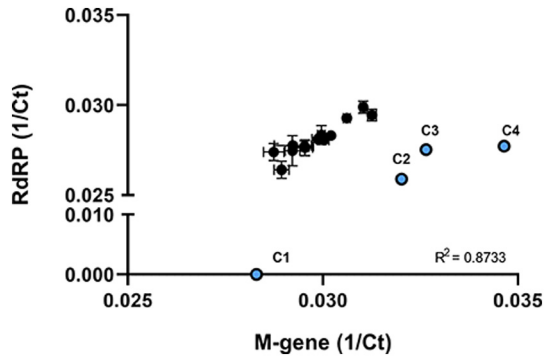


Fig. 3. Correlation of reciprocal C_T values between M-gene and RdRP-gene detection.

where PCR against N-gene was used to detect the presence of viral RNA in the wastewater. Comparing M-gene and RdRP PCR results to N- and E-gene PCR, specificity and sensitivity was clearly increased for RdRP and M-gene. However, taking into account that the M-gene and RdRP PCR may detect other endemic human coronaviruses, wastewater samples collected before the SARS-CoV-2 outbreak in Germany were investigated due to the detection of a positive PCR signal. Therefore, samples from WWTP Aachen-Soers collected in January 2017 (C4), November 2018 (C2), and September 2019 (C3) were processed comparably to the test samples, and RdRP-specific PCR was performed. Indeed, we could detect a minor signal in the PCR with C_T values (2018: C_T 38; 2019: C_T 36 and 2020: C_T 36) close to the detection limit of the PCR (C_T 38) (data not shown). Comparing the amplification curve of SARS-CoV-2 positive control and retrained water controls showed comparable shape, but with 15 C_T difference. A moderate correlation ($R^2 = 0.8733$) of the inverse C_T values of RdRP and M-gene PCR was detectable. Although there were positive results also for retrained controls (C2 to C4), these samples grouped apart from the wastewater samples collected in April 2020 (Fig. 3).

3.3. Sanger sequencing of PCR product

To further analyse retained wastewater samples showing a positive PCR result for M- and RdRP- gene, sequencing of the RdRP-PCR product of C2, C4 and P12 should evaluate if a specific signal was measured (Supplementary Fig. S1). While the sequence of the P12 PCR product showed significant similarity to SARS-CoV-2 sequence (isolate SARS-CoV-2/human/USA/CA-CZB-1525/2020; Accession: MT628199.1) confirming the positive identification of human SARS-CoV2 by means of RT-qPCR (Supplementary Fig. S1A), no match to SARS-CoV-2 was detectable for C2 (Supplementary Fig. S1B). In addition, a positive BLAST result for C4 with SARS-CoV-2 synthetic construct (accession: MT458696.1) (E value 0.002) (Supplementary Fig. S1C) was found. Aligning all three sequenced amplicons (P12, C2, and C4) with the SARS-CoV-2 specific sequence, only the amplicon of P12 showed a sequence identity above 90%. The similarity of C2 and C4 amplicons to the SARS-CoV-2 specific sequence was 50% and 47.3%, respectively (Fig. 4). Therefore, we conclude that the retained samples were negative for SARS-CoV-2.

3.4. SARS-CoV-2 removal in wastewater treatment: role of the solid-phase reservoir

In wastewater treatment, solid residues are largely removed. We thus compared the amount of viral RNA in the aqueous and solid-phase of both inflow and effluent samples. During centrifugation in the processing of the wastewater samples, solid and aqueous phases are separated. To our understanding and as shown in other studies (Kocameci et al., 2020), viral RNA of solid and liquid phase needs to be considered in wastewater surveillance. When comparing the aqueous and solid phase of the samples, we found a one log unit higher SARS-CoV-2 RNA copy number per mL in the solid phase (25 copies/mL) compared to the aqueous phase of the inflow sample (1.8 copies/mL), respectively (Fig. 5A). The difference between the aqueous and solid phase was less evident in the effluent sample, with 8.8 and 13 gene equivalents per mL, respectively (Fig. 5B). When comparing the



Fig. 4. Alignment of sequenced amplicons of P12, C2, and C4. Sequences of wastewater sample P12 and retrained samples C2 and C4 were analysed using Geneious - Bioinformatics Software for Sequence Data Analysis and compared to the SARS-CoV-2 synthetic construct sequence (GenBank: MT458696.1). In green primers binding sites are indicated. Dots in the sequence represent sequence mismatches to the SARS-CoV-2 synthetic construct sequence.

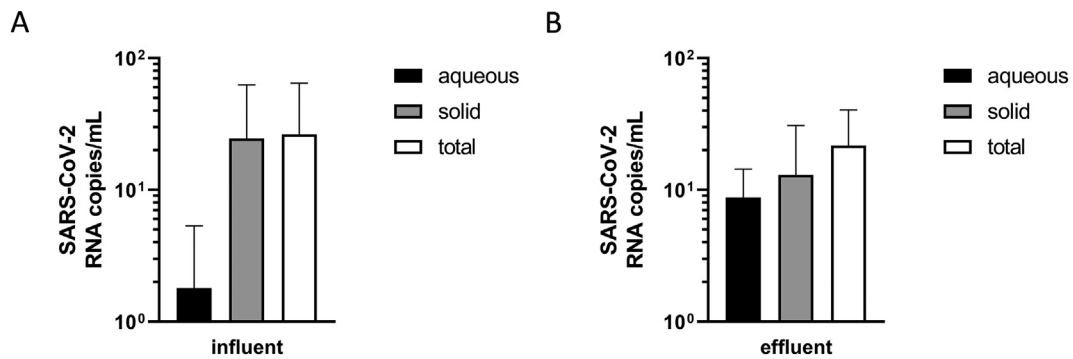


Fig. 5. SARS-CoV-2 removal detected by RdRP RT-qPCR measured by the sum of aqueous and solid-phase of (A) untreated (AC-Inflow) and (B) treated sewage after ozonation. Results are shown as mean + SD of two independent PCR measurements from same sample material. RdRP RT-qPCR C_T values for standard range between 18 (standard 1 = 10^8) and 37 (standard 6 = 10^2). Values of tested wastewater above C_T 38 were considered negative for SARS-CoV-2. Standard curve was calculated using *Bio-Rad CFX Manager software* with $E = 96.5\%$, $R^2 = 0.990$ and slope = -3.408 .

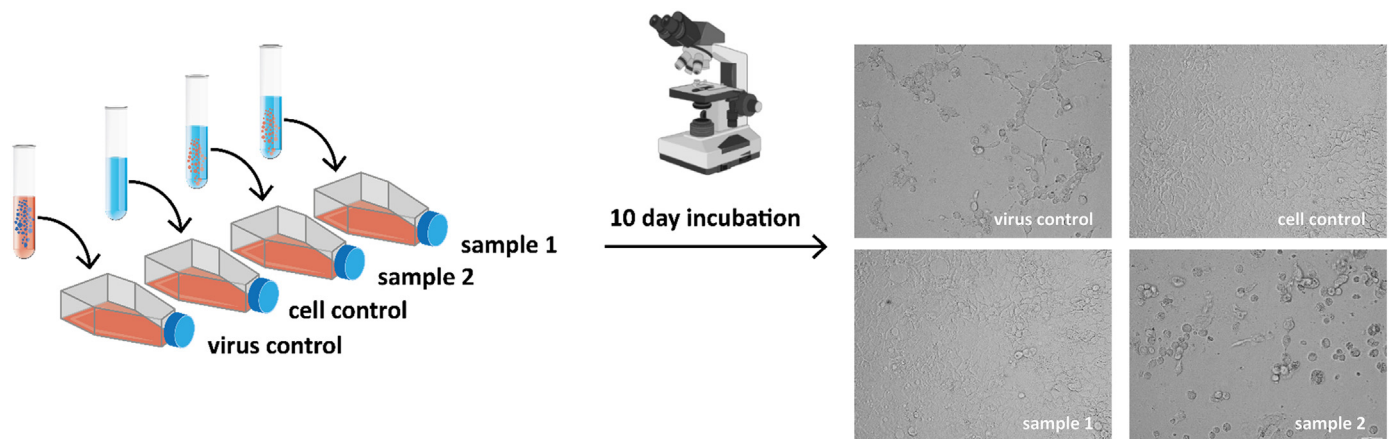


Fig. 6. Viral outgrowth assay setup. Caco-2 cells were inoculated with either wastewater samples or with SARS-CoV-2 containing cell culture supernatant (positive control) or culture medium only (negative control). Cells were investigated daily by light microscopy for the appearance of a cytopathic effect caused by replication-competent SARS-CoV-2 infection for up to 10 days. The experiment was performed in a single replicate for four different wastewater samples (P2, P5, P11, P12) from April 8th 2020.

aqueous phase only, the effluent exhibited a higher gene equivalent concentrations than the influent, which we attribute to repartitioning of gene material from the solid- to the liquid phase during wastewater treatment. In contrast, a difference of total viral copy numbers per mL between untreated and treated wastewater was not detectable (Fig. 5A and B). This might be explained by the sample conditions. While the residence time of the wastewater is about one day at the WWTP, sludge is continuously exchanged with an average sludge age of 10 to 14 days. We assume that solid-phase concentrations are lower-estimates since we did not test whether further extraction steps could mobilize additional SARS-CoV-2 RNA from the solid phase.

3.5. Infectious potential of sewage samples

Although the wastewater in the investigated plants was treated by different processes, viral RNA is not eliminated, as shown above. To test the infectious potential of untreated and treated wastewater, a cell culture model was used to analyse that impact (Fig. 6). Infection of differentiated Caco-2 cells with cell culture grown, replication-competent SARS-CoV-2 (MOI 0.01) led to the induction of a cytopathic effect (CPE after 48–72 h). The CPE is characterized by round and shrinking cells that detach from the culture surface. Observation of CPE after two to three days at a low MOI indicated rapid to moderate replication of the virus. Inoculation of differentiated Caco-2 cells for ten days with purified and concentrated wastewater (P2, P5, P11, and P12) did not result in the production of infectious SARS-CoV-2 particles (data not shown), which suggests that treated sewage appears to be non-infectious even though viral RNA fragments can be detected. Due to the lack of CPE, quantification and verification by RT-qPCR was not performed.

3.6. Reference parameters

The wastewater samples analysed showed typical wastewater characteristics for the different types of WWTPs studied (Brückner et al., 2018; Metcalf and Eddy, 2012; Supplementary Table S2). The measured concentrations of nutrients and physico-chemical parameters are in the same range as found in the last years in the official monitoring by the environmental protection agency of North Rhine-Westphalia (ELWAS, 2020). The samples P2, P10, P12, and P13 are taken in the effluent of the WWTP or the effluent of the clarification. As expected, the substance concentrations in these samples are lower than in the corresponding influent samples P1, P9, and P11. Only the effluent of the WWTP KLEM had a higher content of dry residues in the effluent than in the inflow, which is important for evaluating the partitioning between the liquid and solid phase.

Different bioindicators were analysed in the wastewater to evaluate whether they could serve to estimate the actual number of persons

staying in the catchment area at the day of sampling (PE_{actual}) more accurately than the nominal number of connected residents (PE_{nom}) or the design capacity (PE_{design}). Results are provided in Supplementary Table S3. The load of the bioindicators was found to correlate with the nominal number of connected residents (Fig. 7; Supplementary Table S5).

4. Discussion

4.1. Method validation: comparison to published papers and SOPs

During the SARS-CoV-2 pandemic, testing different sample material (e.g., throat swabs) was a crucial step in the detection of hotspots and limiting transmission. The first SARS-CoV-2 PCR assays were developed against the E-gene and the N-gene to detect the virus in patient samples, and numerous studies analysing wastewater for SARS-CoV-2 surveillance used E-gene or N-gene specific PCRs (Medema et al., 2020; Ahmed et al., 2020; Y. Wu et al., 2020; F. Wu et al., 2020). To a minor proportion, also PCR targeting SARS-CoV-2 RdRP was used (Wurtzer et al., 2020). In this study, wastewater samples were analysed with two different PCRs targeting SARS-CoV-2 RdRP and M-gene. In an initial analysis, a control sample (retained control) and wastewater test samples from April 2020 were tested for the presence of four different SARS-CoV-2 genes (N-, E-, M-gene, and RdRP), and specificity and sensitivity were evaluated based on other coronaviruses, including the

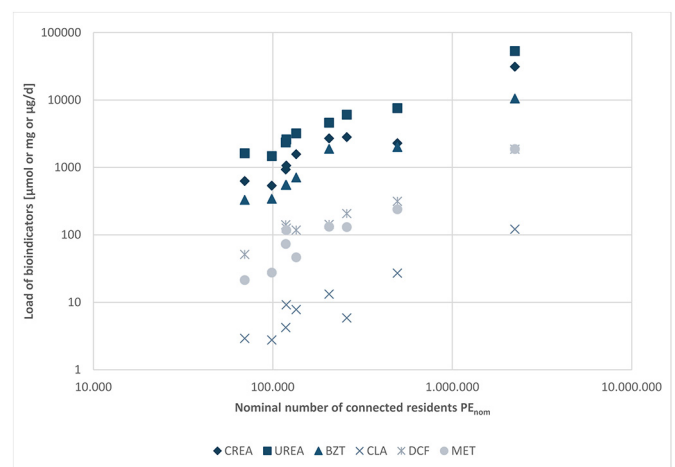


Fig. 7. Bioindicator load for each WWTP: Creatinine [CREA $\mu\text{mol/day}$], urea [Urea mg/day], benzotriazole [BZT $\mu\text{g/day}$], clarithromycin [CLA $\mu\text{g/day}$], diclofenac [DCF $\mu\text{g/day}$], metoprolol [MET $\mu\text{g/day}$]. For regression lines and coefficients of determination, see Supplementary Table S5.

closest relative SARS-CoV. RdRP-PCR was highly specific for SARS-CoV-2 and M-gene PCR being the most sensitive for SARS-CoV-2. These results are in line with recently published data by [Toptan et al. \(2020\)](#), where the efficiency of RT-qPCR primer and probes pairs to detect SARS-CoV-2 genes was compared. At least in this study, N-gene primers were less specific and were excluded due to limited linearity compared to other viral genes. Although other viral genes for SARS-CoV-2 detection in wastewater were used in this study, total gene equivalents per mL are comparable to other published studies ([Ahmed et al., 2020](#); [Y. Wu et al., 2020](#); [F. Wu et al., 2020](#)). To further control the specificity and sensitivity of our PCR, retained wastewater samples from 2017 to 2019, which should be negative for SARS-CoV-2, were analysed. Interestingly, a positive SARS-CoV-2 signal in PCR of retained samples from 2017 and 2018 was observed. One possible explanation could be a cross-contamination during the purification or RNA isolation step. To monitor the isolation process and identify possible cross-contamination, additional controls could be introduced such as isolation from a clean water control. So far unpublished data, analysing retained wastewater samples from Barcelona, Spain, taken in March 2019 showed positive RT-qPCR results for SARS-CoV-2 ([Chavarría-Miró et al., 2020](#)). The authors conclude that the virus was introduced to the Barcelona population due to global travel and remained undetected.

When sequencing the different PCR amplicons obtained using wastewater samples taken during the SARS-CoV-2 pandemic and long before the outbreak (retained samples), distinct results for the test sample (P12) which fits SARS-CoV-2/human/USA/CA-CZB-1525/2020 isolate from NCBI database was found, but not for the retained samples suggesting an unspecific signal in RT-qPCR, respectively. Although RNA samples used for PCR are highly purified, there might have been environmental contaminants affecting PCR. PCR is a

highly sensitive method to detect nucleic acid and is strongly affected by pH, salt concentration, and contamination with extraneous nucleic acids that lead to false-negative or false-positive results, respectively ([Schrader et al., 2012](#); [Paul et al., 1991](#)). BLASTn search showed minor sequence similarity (E values > 0.44) to Pithovirus, a giant DNA virus that infects amoebas ([Legendre et al., 2014](#)), bacteria like *Streptococcus* or animals (e.g., *Xenopus laevis*, *Sparus aurata*). To avoid false-positive results, optimization of PCR conditions can be taken in consideration. Primer pairs used in this study were optimized for RT-qPCR of patient material and cell culture-derived samples. The matrix of wastewater strongly differs from this defined sample material. Furthermore, primers adapted to RT-qPCR results in amplicons of approximately 100 bp, which are complicated to sequence. Establishing qualitative nested PCR could be an option to improve sequencing results. Furthermore, extraneous nucleic acids can be accumulated during the concentration process of the wastewater sample. While single studies may avoid the accumulation of extraneous nucleic acids using filter units with 10 kDa cut-off and therefore also could limit detection of SARS-CoV-2, most studies used filter units with 100 kDa cut-off or other methods for concentrating wastewater samples ([Ahmed et al., 2020](#); [Medema et al., 2020](#); [Nemudryi et al., 2020](#); [La Rosa et al., 2020](#)). Although no false-positive results are described in these analyses, these studies did not use retained wastewater samples before the SARS-CoV-2 pandemic as control. Nevertheless, recently published data by [La Rosa et al. \(2020\)](#) or [Chavarría-Miró et al. \(2020\)](#) could detect SARS-CoV-2 RNA in retained wastewater samples taken before the first outbreak. Especially, wastewater analysis in Barcelona showed SARS-CoV-2 detection already in early 2019. However, results were not verified by an independent method like verifying PCR product

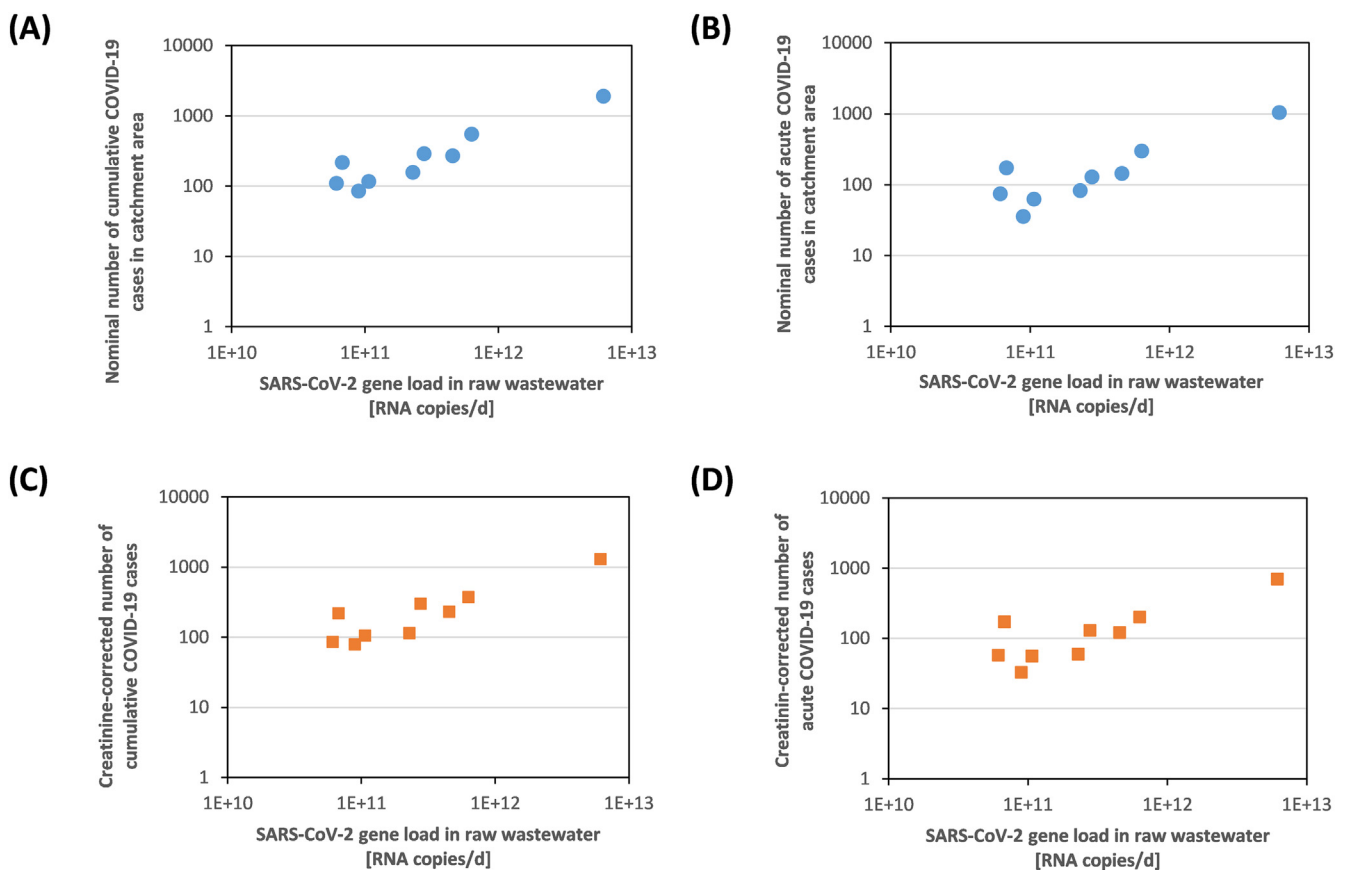


Fig. 8. Correlation between the measured SARS-CoV-2 M-gene copy loads and the nominal numbers of (A) cumulative and (B) acute COVID-19 cases on April 8th, 2020, in the catchment areas of the 9 WWTP studied. In comparison, creatinine-corrected numbers of (C) cumulative and (D) acute COVID-19 cases are displayed. For regressions and coefficients of determination, see Supplementary Table S5.

size using gel electrophoresis, or sequencing (Chavarria-Miró et al., 2020). Taken together, it is important to consider environmental contaminants that might interfere with PCR and may lead to false-positive results. Appropriate controls, such as sequencing, are thus recommended to confirm PCR results.

4.2. Suitability for wastewater surveillance of COVID-19

To establish SARS-CoV-2 wastewater-based epidemiology as a reliable tool for early-warning and surveillance of the current or future COVID-19 outbreaks, the method's detection limit, precision, accuracy, and reliability must meet certain criteria to be useful for public surveillance. For example, early-warning requires a low detection limit to detect the very first cases, while in later phases of the epidemics, high precision and accuracy are needed to survey whether certain thresholds are exceeded, for example, the threshold currently set in Germany of 50 acute cases per 100,000 residents.

On the day of our sampling campaign, we estimate the nominal acute incidence of COVID-19 to range between 30 and 174 cases per 100,000 residents in the nine catchment areas studied (Table 4). Based on surveillance results (Fig. 4), our findings indicate that the RT-qPCR employed is capable of resolving acute incidences of 50 cases per 100,000 residents in dry-weather conditions, although the method has not yet been optimized in terms of sensitivity and precision to detect even lower incidence.

The nine catchment areas studied differ considerably in size, wastewater flow (Table 2), and nominal numbers of COVID-19 cases at the day of sampling ranging from 36 to 1037 acute cases and from 85 to 1924 cumulative cases (Table 4). Inter-comparing these nine catchment areas, we plotted the estimated cumulative and the acute prevalence against the measured SARS-CoV-2 load (Fig. 8), the latter calculated from RT-qPCR measured M-gene copy concentration (Fig. 4) and the actual wastewater flow Q_{actual} on the day of sampling (Table 2). Clearly, the estimated nominal number of COVID-19 cases increases with increasing measured SARS-CoV-2 load, resulting in correlations both for cumulative and acute prevalence (Fig. 8A and B). Scatter in the correlations might be attributed to various reasons, including variations in virus RNA recovery and uncertainty in the estimated prevalence data. In contrast, plotting the incidence against SARS-CoV-2 concentration did not yield a conclusive correlation (not shown), likely because the precision of the qPCR employed was not sufficient to discriminate relatively minor differences in the incidence prevailing in the studied catchment areas at the time of sampling, ranging from 30 to 174 cases per 100,000 residents (less than an order of magnitude, Table 4).

While the number of unreported cases is not known, also the number of persons staying in the catchment area (PE_{actual}) may differ from the nominal number of residents connected to the sewer (PE_{nom}). We evaluated the utilisation of various bioindicators to calculate the incidence in the catchment area by correcting PE_{nom} to PE_{actual} by multiplying PE_{nom} with a bioindicator factor (BIF, Supplementary Table S4). Here we show results for creatinine that humans excrete in the urine (Fig. 8C and D). Although it cannot be evaluated based on the available data, if this correction is superior, it illustrates the potential of bioindicators to account for differences in wastewater composition.

In contrast to our approach, Medema et al. (2020) reported catchment-specific correlations plotting the increase in concentration (not load) of the N gene assays (and cycles of the E gene assays) against the increase in cumulative incidence from 0.1 to 100 cases per 100,000 residents ($3 \log_{10}$ units) over a 7-weeks outbreak of the pandemics. While catchment-specific temporal correlations may be less susceptible to interference with other factors, we consider RNA load-based evaluations more reliable to cope with variations in wastewater flow, for example, stormwater events in combined sewer systems or variations in industrial wastewater production during a lockdown of industry and businesses.

4.3. Potential transmission risks

Our results suggest that detected RNA fragments appear to be non-infectious based on small-volume laboratory studies testing about 1 to 10 gene copies in one reaction. However, given the large capacity and flow rate of WWTPs (Table 2), we calculate that each of the studied WWTP emits 6×10^{10} to 6×10^{12} SARS-CoV-2 gene equivalents per day to the receiving water bodies, some of which serve as crucial water resources for drinking water production, cooling water abstraction, public swimming, irrigation, recreation, and natural habitats. This viral load can be very roughly compared to the potential viral shedding by infected persons in the catchment area, although many uncertainties come with such an appraisal: Rose et al. (2015) report a mean amount of feces excreted of about 129 g/cap/day. Viral gene copies detected in human excreta can vary broadly, probably depending on factors such as the stage of infection, and are given in the range of 10^2 – 10^8 gene copies per g feces (Kitajima et al., 2020). Wölfel et al. (2020) showed the development of gene copy findings on the time course of the infection. With, e.g., 100 infected persons in a sampled catchment, these assumptions would yield a viral load in a range as broad as 10^6 – 10^{12} gene copies per d in the wastewater. The findings, as shown in Fig. 8, are within but on the upper end of this broad range. Few studies have evaluated the fate of coronaviruses and other enveloped viruses in wastewater treatment and surface water (Gundy et al., 2009; Ye et al., 2016). In a recent review by Kitajima et al. (2020), it is stated that currently no applicable dose-response models exist for SARS-CoV-2, which would allow for a better risk assessment. In a recent review by Folorunso et al. (2020), the current knowledge about the fate of SARS-CoV-2 in wastewater treatment systems was summarized, and the few studies undertaken indicate that there is a significant reduction in viral load through the treatment process, but still, gene fragments are detectable in effluents. Although it cannot be excluded with the current testing system of RT-qPCR targeting SARS-CoV-2 genes that virus emitted in large quantities is still infectious, studies analysing SARS-CoV-2 environmental stability support the suggestion that infectivity and replication-competence of released viral particles are very unlikely (Liu et al., 2020; van Doremalen et al., 2020).

5. Conclusions

Based on our findings, we conclude the following:

- SARS-CoV-2 RNA was detected in the inflow of all 9 studied WWTP at concentrations similar to those reported in other studies. Our screening of different target genes points to shortcomings in the selectivity of gene primers used in studies previously published by other authors that might also detect genes non-specific to SARS-CoV-2 viruses.
- Using Sanger sequencing, we confirmed human SARS-CoV-2 for the investigated sample taken during the SARS-CoV-2 pandemic outbreak. In contrast, positive signals in RT-qPCR were not confirmed by sequencing for the retained samples from 2017 and 2018. These results are suggesting an unspecific signal in RT-qPCR, stressing again the urgent need for a confirmation of positive RT-qPCR results by sequencing or other appropriate techniques in order to avoid false-positive results.
- Quantification of SARS-CoV-2 gene concentrations and loads in wastewater needs to consider both aqueous and solid-phase SARS-CoV-2 detection methods and establish robust standard protocols for clean-up and RT-qPCR measurements.
- We observed poor removal of SARS-CoV-2 in all three of the studied conventional activated-sludge WWTP. Full-scale ozonation at one plant seemed to reduce SARS-CoV-2 fragments in the effluent. Membrane-based WWTP planned to be included in future studies.
- We found that the total load of gene equivalents in wastewater correlated with the cumulative and the acute number of COVID-19 cases reported in the respective catchment areas. We consider

load-based correlations superior to gene copy concentration-based approaches. Analysis of suitable bioindicators in the wastewater may improve the assessment of SARS-CoV-2 loads data in catchment areas.

- While our results indicate that RNA fragments are not infectious to Caco-2 cells in cell culture-based assays, further studies are needed to evaluate the risk SARS-CoV-2 may pose in the water cycle.

It is recommended to further develop and implement the concept of wastewater-based epidemiology as a complementary measure to survey the outbreak of the SARS-CoV-2 pandemic and impose catchment-specific measures if necessary.

CRedit authorship contribution statement

Sandra Westhaus: Conceptualization, Methodology, Investigation, Writing - original draft, Data curation, Formal analysis. **Frank-Andreas Weber:** Conceptualization, Methodology, Resources, Writing - original draft, Data curation, Formal analysis. **Sabrina Schiwly:** Writing - original draft, Data curation, Visualization. **Volker Linnemann:** Conceptualization, Methodology, Resources, Writing - original draft, Data curation, Formal analysis. **Markus Brinkmann:** Writing - review & editing, Validation. **Marek Widera:** Methodology, Writing - review & editing. **Carola Greve:** Investigation, Writing - review & editing, Data curation. **Axel Janke:** Resources, Writing - review & editing. **Henner Hollert:** Conceptualization, Resources, Formal analysis, Supervision, Writing - review & editing. **Thomas Wintgens:** Conceptualization, Resources, Formal analysis, Supervision, Writing - review & editing. **Sandra Ciesek:** Resources, Supervision, Writing - review & editing.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.scitotenv.2020.141750>.

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