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**Abstract:** Vehicular pollution has become a major problem in urban areas due to the exponential increase in the number of automobiles. Typical exhaust emissions, which include nitrogen oxides (NO<sub>x</sub>), hydrocarbons (HC), carbon monoxide (CO), soot, and particulate matter (PM), doubtless have important negative effects on the environment and human health, including cardiovascular effects such as cardiac arrhythmias and heart attacks, and respiratory effects such as asthma attacks and bronchitis. The mitigation measures comprise either the use of clean alternative fuels or the use of innovative technologies. Several existing emission control technologies have proven effective at controlling emissions individually, such as selective catalytic reduction (SCR) and lean NO<sub>x</sub> trap (LNT) to reduce NO<sub>x</sub> and diesel particulate filter (DPF) specifically for PM abatement. These after-treatment devices are the most profitable means to reduce exhaust emissions to acceptable limits (EURO VI norms) with very little or no impact on the engine performances. Additionally, the relative lack of physical space in which to install emissions-control equipment is a key challenge for cars, especially those of small size. For this reason, to reduce both volume and cost of the after-treatment devices integrated catalytic systems (e.g., a sort of a "single brick") have been proposed, reducing both NO<sub>x</sub> and PM simultaneously. This review will summarize the currently reported materials for the simultaneous removal of  $NO_x$  and soot, with particular attention to their nature, properties, and performances.

Keywords: soot; NO<sub>x</sub>; simultaneous removal; emission control; oxidation catalysts

#### 1. Introduction

In the theoretical or clean diesel combustion only  $CO_2$  and  $H_2O$  are produced as exhaust, being carbon oxidized to  $CO_2$  and hydrogen to  $H_2O$ . However, the oxidation process in actual diesel engines is very far from being an ideal process, so in the exhausts many byproducts and pollutants, both gaseous and solid, are present. Depending on many factors (air–fuel ratio, air–fuel concentration, ignition timing, turbulence in the combustion chamber, combustion form, combustion temperature, etc.) different exhaust compositions are obtained and a number of harmful products are generated; the most significant harmful products are CO, unburnt hydrocarbons (HC),  $NO_x$ , and particulate matter (PM). To the vehicular pollution point of view, great importance related to  $NO_x$ , CO, HC, and smoke (particles matters, PM, or soot), while  $CO_2$  emissions are mainly related to its greenhouse potential in the atmosphere.

To control and regulate the vehicular emissions, two regulatory commissions have been established, i.e., the Environmental Protection Agency (EPA) and the European Parliament (EURO) imposing the most stringent emission regulations; indeed, in Europe Euro VI protocols in force since 2015 have reduced emissions levels of  $NO_x$  by 87% and of PM by 96%. Many efforts have been made to develop adequate technologies to respect the imposed limits. Those are classified as primary or secondary techniques.

Diesel engines use highly compressed hot air to ignite the fuel. Air, mainly composed of oxygen and nitrogen, is initially drawn into the combustion chamber. Then, it is compressed, and the fuel is injected directly into this compressed air at about the top of the compression stroke in the combustion chamber. This generates high temperatures, which are sufficient for the diesel fuel to ignite spontaneously when it is injected into the cylinder. The high temperatures in the cylinders cause the nitrogen to react with oxygen and generate  $NO_x$  emissions, whose amount is a function of the maximum temperature in the cylinder, oxygen concentrations, and residence time. Most of the emitted  $NO_x$  is formed early in the combustion process, when the piston is still near the top of its stroke. This is when the flame temperature is the highest. So, it is an established factor that  $NO_x$  formation is highly dependent on temperature and reducing peak temperatures during combustion obviously reduce  $NO_x$ .

To reduce the NO<sub>x</sub> emissions from vehicles, there are two main approaches, i.e., first minimizing the amount of NO<sub>x</sub> created, and second removing NO<sub>x</sub> from the exhaust. The first task is mainly accomplished by manipulating engine operating characteristics (the so-called primary technology), for example by lowering the intake temperature, reducing power output, retarding the injector timing, reducing the coolant temperature, and/or reducing the combustion temperature. However, the problem of controlling NO<sub>x</sub> in diesel exhaust is really complicated, and for this reason diesel vehicles require different approaches. Electronically controlled fuel injection, engine modification, increasing injection timing, water spray in the combustion chamber, improvement of fuel properties, use of fuel additives, etc., are some of the primary methods to reduce NO<sub>x</sub> emissions. Between the other, cooled exhaust gas recirculation (EGR) is well known and is the method that most engine manufacturers are currently using. EGR system recycles a portion of the exhaust gas into the combustion chamber, mixing with fresh air and thus reducing the oxygen content and increasing the water vapor content of the combustion mixture; the result is a reduction of the peak combustion temperature and as consequence of NO<sub>x</sub> formation. Unfortunately, the decreases in combustion temperature influences also the reaction rates, causing the emission of other pollutant species like CO, CO<sub>2</sub>, HC, and soot.

Reducing  $NO_x$  by manipulating engine operation generally reduces fuel efficiency. Besides, the mere manipulation of engine operation is not sufficient to meet the current stringent emission limits. As a result, after-treatment systems also need to be implemented, that remove the produced pollutant (mainly,  $NO_x$  and/or soot) from the exhausts.

The reduction of NO<sub>x</sub> emissions is currently performed by selective catalytic reduction (SCR) [1] and lean NO<sub>x</sub> traps (LNTs) [2] catalytic after-treatments. In the SCR system an aqueous urea solution (AdBlue<sup>®</sup>) is injected into the exhaust post combustion; the hydrolysis of urea generates ammonia, which reacts with the NO<sub>x</sub> giving nitrogen and water. This catalytic reaction is accomplished by using zeolite-based catalysts, doped with Fe and/or Cu. With regard to the LNT system, NO<sub>x</sub> are stored onto the catalyst surface during a lean phase (i.e., in the excess of oxygen), forming nitrites/nitrates ad-species; these are reduced to nitrogen during a subsequent rich phase (i.e., in the presence of a reductant like H<sub>2</sub>, CO, and HC) of a few seconds. In this case, the catalyst is a PGM-based system, i.e., a catalyst containing platinum-group metals (PGM) like Pt, Pd, and Rh, doped with alkaline and/or alkaline-earth metal oxides as storage components. The detailed mechanism of these systems will be addressed in Section 3.1.

On the other hand, to meet EURO V regulations concerning soot all new vehicles must be equipped with a diesel particulate filter (DPF) that capture soot and other dangerous particles, preventing them releasing in the atmosphere. DPF are usually wall-flow monoliths in cordierite ( $2MgO-2Al_2O_3-5SiO_2$ ) or silicon carbide (SiC) with a honeycomb structure, where the channels are alternatively blocked at the end (Figure 1). A DPF can remove around 85% of the particulates from the exhaust.

If the filters are overloaded, the particles can cause obstruction of the flow of gas, which is manifested by an increase of the particulate filter pressure drop and as a consequence a decrease in the engine efficiency. The filter can operate until it is clogging; thus, it may be regenerated before the problem occurs. During regeneration, the soot is oxidized into gaseous products. Thermal regeneration of DPF, also called active regeneration, requires a temperature of 550–600 °C, so that the controlled oxidation of the particulate with  $O_2$  occurs [3]. Active regeneration systems inject raw diesel fuel into the exhaust stream (post combustion) or into the diesel oxidation catalyst (DOC) to achieve appropriate regeneration temperatures in the DPF. Obviously, this process requires an extra fuel consumption leading to a fuel penalty; moreover, excessive heating can damage filter itself and the other downstream catalytic after-treatment devices.

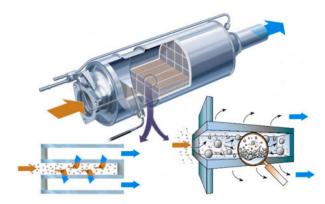


Figure 1. Functioning of a typical diesel particulate filter (DPF).

The development of catalyzed DPFs (CDPFs), i.e., DPFs coated with a catalytic layer, aiming at lowering the soot oxidation temperature (passive regeneration) represents a viable solution to allow energy savings and prevent filter overheating during the regeneration phase. Accordingly, low temperature activity, on one hand, and high thermal stability, on the other hand, represent the main key issues for the development of suitable catalytic materials. Another option is to use the fuel-borne catalysts (FBC), i.e., active fuel additives constituted by metallic compounds (such as transition metals like cerium, iron, or copper compounds) that decrease the soot combustion temperature by improving the catalyst–soot contact [4,5]. However, several drawbacks such as the fuel penalty, possible FBC volatilization, formation of deposits on the DPF, and the high investment costs prevented the broad commercialization of FBCs.

An alternative commercially available technology is the continuously regenerating trap (CRT<sup>®</sup>) system, proposed by Johnson Matthey [6], which uses an oxidation diesel catalyst (DOC) upstream the DPF to generate NO<sub>2</sub> (i.e., a stronger oxidant than O<sub>2</sub> [7]) and consequently to decrease the soot combustion temperature in the downstream filter. Indeed, NO is oxidized to NO<sub>2</sub> in the DOC, then the so-formed NO<sub>2</sub> reacts with trapped soot, forming CO<sub>2</sub> and NO, which in turn is reoxidized to NO<sub>2</sub>. The low NO<sub>x</sub> removal efficiency, due to NO<sub>2</sub> slip, represents the main problem of this technology [8].

Johnson Matthey further refined the CRT system by essentially combining CRT and CDPF in the so-called CCRT system (catalyzed continuously regenerating trap) where filter itself is coated with a catalyst, which improves the operating window for the filter regeneration.

Recently, to ensure the respect of the latest vehicular emission regulations, to reduce both packaging volume and cost of after-treatment systems, combined technologies have been also proposed to significantly remove PM and NO<sub>x</sub> simultaneously. Accordingly, the use of combined DPF-LNT/SCR systems has been proposed in different configurations to take advantage of potential synergisms among the various devices, for example, to advantageously use concentration and/or temperature gradients generated by the catalytic system. Indeed, DPFs are fitted close to the engine where the exhaust is hottest so that passive regeneration is more likely to work; however, combining with LNT and/or SCR, the latter will be closer to the engine thus enjoying the advantage of higher temperatures for the NO<sub>x</sub> removal.

Many types of catalysts have been investigated for the simultaneous control of NO<sub>x</sub> and soot, i.e., PGM-based systems (platinum-group metals like Pt, Pd, Rh, and Ir), and also a wide variety of

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materilas like spinel-type oxides, hydrotalcites, rare earth metal oxides, mixed transient metal oxides, and perovskites. The high raw material cost of PGM catalysts has become a significant issue, so that the development of non-PGM catalysts is one of the most promising challenges.

In this context, the overall aim of this work is to give an overview on catalytic materials for the combined  $NO_x$  and soot removal. However, it is important to note that the concept of simultaneous removal of soot and  $NO_x$  is almost confused in the literature. First of all, the  $NO_x$  reduction expression is used in a double sense: NO<sub>x</sub> reduction as NO<sub>x</sub> concentration decrease, and reduction as formation of N-products with a lower oxidation state with respect to NO. Some authors refer to catalysts able to control simultaneously soot and NO<sub>x</sub> because they are able to oxidize NO to NO<sub>2</sub>, which in turn is active in soot oxidation. In other cases, the simultaneous control is related to the catalysts ability in soot oxidation and NO<sub>x</sub> storage without any deep study on the reduction mechanism of stored NO<sub>x</sub>. Finally, the simultaneous removal of  $NO_x$  and soot strictly speaking refers to oxidation of soot while NO<sub>x</sub> are reduced to nitrogen. With this in mind, first of all we consider PGM-free catalysts that have been proposed for the simultaneous removal of soot and NO<sub>x</sub>, i.e., hydrotalcites differently doped and perovskite-like catalysts. Then, PGM-based catalysts have been considered; particular attention was paid to the DPNR (Diesel Particulate NOx Reduction) system patented by Toyota for the simultaneous removal of NO<sub>x</sub> and soot. In addition to the traditional DPNR system based on traditional Pt-based NO<sub>x</sub> storage-reduction catalyst, new PGM-free formulations have been deeply analyzed. Finally, with the aim of analyzing the combined technologies proposed to reduce simultaneously these pollutants and realize compact systems, SCR/or LNT/DPF systems have been considered and analyzed.

## 2. PGM-Free Catalysts for the Simultaneous Removal of Soot and NO<sub>x</sub>

## 2.1. Hydrotalcites and Mixed Metal Oxides Catalysts

Hydrotalcites (HTLCs) are part of the big family of anionic clays; they have  $Mg^{2+}$ ,  $Al^{3+}$ , and  $CO_3^{2-}$  ions as its constituents. These layered double hydroxides (LDHs) may be represented by the general formula  $[M_{1-x}^{II} M_x^{III}(OH)_2]_x + (A^{n-})_{x/n} \cdot yH_2O$ , where  $M^{2+} (M^{2+} = Mg)$ ,  $M^{3+} (M^{3+} = AI)$  are di- and trivalent metal cations,  $A^{n-}$  is an interlayer anion (e.g.,  $CO_3^{2-}$ ) and x represents the molar fraction of  $M^{3+}$  per total metal. These materials can exhibit different physical and chemical properties depending on the nature of  $M^{2+}$  and/or  $M^{3+}$ , on their molar ratios and on the interlayer anions. For these reasons, they are usually applied in catalysis [9], mainly for catalytic oxidation applications [10]. Since fresh hydrotalcites have high water content, which often makes them inactive for certain reactions, they are activated by thermal decomposition. The calcination temperature ranges from 500 to 800 °C, and the resultant mixed metal oxides after calcinations offer improved catalytic performances [11,12], exhibiting large surface area, basic properties, high metal dispersions, and low propensity to sintering; their redox properties could be further improved by doping with transition metal ions, i.e., Co, Cu, Ni, or Zn as  $M^{2+}$  cations, or Cr, Ga, Mn, or Fe as  $M^{3+}$  cations [11,13]; finally, the presence of alkali metals in the catalyst formulation enhances the catalytic soot combustion of hydrotalcite-derived materials [14–18].

LDHs represent suitable candidates for the simultaneous removal of NO<sub>x</sub> and soot, as summarized in Table 1. Being their catalytic activity strongly related to the nature of metals, their amount and calcination temperature, different formulations have been proposed and between them copper-, cobalt-, or nickel-based oxides have demonstrated high activity in oxidation reactions [19], like soot combustion. Moreover, it is well known that the addition of alkali oxides, and in particular K can further improve their activities favoring the reaction through the formation of low melting point compounds [20], or of eutectics with other catalyst components, thus improving the surface mobility of the active species [12,13,21] and hence favoring the soot–catalyst contact, which has been claimed as a key factor in the soot oxidation process.

Hydrotalcites Families	Catalysts	<b>Preparation Method</b>	<b>Reaction Conditions</b>	Ref.
	A series of K-supported MgAl with different amount of K doping (2, 5, 8 and 15 wt.% of K)	Co-precipitation; K addition by impregnation; calc. 600 °C	Cat/soot = 9/1; 9.97% $O_2$ + He or 1050 ppm NO + He; total flow 100 mL/min; heating rate 5 or 10 °C/min	[22–24]
Binary MgAlO <sub>x</sub> LDH	CuAlO <sub>x</sub> catalyst Co-precipitation; calc. 800 °C		Cat/soot = 9/1; 0.25% NO + 5.0% O <sub>2</sub> + He; total flow 20 cm <sup>3</sup> /min; heating rate 1.6 $^{\circ}$ C/min	[25]
	CoAlO <sub>x</sub> catalyst	Co-precipitation; calc. 500 °C e 800 °C	Cat/soot = 20/1; 0.25 vol.% NO + 5 vol.% O <sub>2</sub> + He, total flow 80 cm <sup>3</sup> /min, heating rate 1.4 °C/min	[26]
Ternary MgAlO <sub>x</sub> LDH	A series of Mn-doped MgAl with different amount of Mn doping (from 0.5 to 3)	Co-precipitation; calc. 800 °C	Cat/soot = 20/1; 750 ppm NO + 10 vol.% $O_2$ + $N_2$ , total flow 240 cm <sup>3</sup> /min, heating rate 10 °C/min	[27]
	A series of $Cu_xMg_{3-x}Al$ with different Cu contents (0.5, 1.0, 1.5, 2.0, 2.5, 3.0)	Co-precipitation; calc. 600 °C–700 °C–800 °C	Cat/soot = 20/1; 0.25 vol.% NO + 5 vol.% O <sub>2</sub> + He; total flow 20 cm <sup>3</sup> /min, heating rate of 1.6 °C/min.	[28]
	Co <sub>2.5</sub> Mg <sub>0.5</sub> Al	Co-precipitation; calc. 500 °C, 600 °C, 700 °C or 800 °C	Cat/soot = 20/1; 400 ppm NO + 10 vol.% O <sub>2</sub> + N <sub>2</sub> ; total flow 400 mL/min; heating rate of 10 °C/min.	[12,13]
Quaternary MgAlO <sub>x</sub> LDH	A series of K-doped Co <sub>2.5</sub> MgAl with different amount of K doping (1.5, 4.5, 7.5, 10)	Co-precipitation; K addition by impregnation; calc. 600 °C	Cat/soot = 20/1; 400 ppm NO + 10% $O_2$ + $N_2$ ; total flow 20 mL/min	[13]
	A series of K-doped $Mn_{1.5}MgAl$ with different amount of K doping (x = 1.5, 4.5, 7.5, 10, 15, 20)	Co-precipitation; K addition by impregnation; calc. 800 °C	Cat/soot = 20/1; 750 ppm NO + 10 vol.% O <sub>2</sub> + N <sub>2</sub> , total flow 240 cm <sup>3</sup> /min, heating rate 10 °C/min	[29]

Table 1. Literature review of hydrotalcite-based catalysts for the simultaneous removal of soot and $NO_x$	•
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MgAlLDH is one of the most common precursors for synthesizing binary LDH-derived catalysts [9]. Zhang et al. [22–24] studied the performances in soot combustion of MgAl LDH derived mixed oxides eventually doped with K. Using a NO<sub>x</sub>/O<sub>2</sub> mixture, the authors demonstrate that the ignition temperature (T<sub>i</sub>) and the temperature for 50% soot conversion (T<sub>50</sub>) decreased with the increase in the K loading, reaching the optimum of K below 8 wt % of the supporting amount. Moreover, also the selectivity to CO<sub>2</sub> results slightly increased by the K presence. However, the total NO<sub>x</sub> reduction efficiency of the K/MgAlO<sub>x</sub> catalysts is still not high enough; indeed, the best NO<sub>x</sub> removal efficiency is no higher than 8%, which is not practically viable. The key reason is because MgAlO<sub>x</sub> mixed oxides itself have relatively poor reductive activity.

Thus, in order to increase the total NO<sub>x</sub> removal efficiency of these LDH-derived catalysts, a noble metal or a metal with redox properties similar to the noble metal is generally demanded. For these reasons, a catalytic system based on CuO has been considered by Wang et al. [25]. As expected, the ignition temperature ( $T_i$ ) is dependent on the soot–catalyst contact, being near 260 °C in tight contact conditions and 314 °C in loose contact; moreover, the maximum conversion of NO to N<sub>2</sub> decreased from 40.4% to 29.2%, definitely higher than that on K/MgAlO<sub>x</sub> catalysts. Wang et al. [26] have examined the catalytic property of Co–Al mixed oxides derived from hydrotalcites as a new active catalyst for the simultaneous NO<sub>x</sub>-soot removal reaction. Additionally, in this case, the catalytic activity was related to the redox properties of the catalyst affected by the Co content and calcination temperature. Indeed, by increasing the calcination temperature from 500 to 800 °C, both the activity of soot oxidation and the selectivity to  $N_2$  formation increased due to the enhancement of redox properties. The active species might come from  $Co_3O_4$ ; indeed, when the Co spinel form is present, its reduction takes place in the same temperature window of soot oxidation suggesting that a redox-type mechanism acts for soot oxidation in the presence of  $O_2/NO_x$  and that the catalyst is redox active, i.e., easily reducible and reoxidizable by gaseous oxidants. The occurrence of simultaneous NO<sub>x</sub>-soot removal reactions was confirmed by the presence of  $CO_2$ ,  $N_2$ , and  $N_2O$  between the reaction products: the oxidation of soot by either NO<sub>x</sub> or O<sub>2</sub> giving CO<sub>2</sub>, and the reduction of NO<sub>x</sub> by soot to N<sub>2</sub> and N<sub>2</sub>O. For the best formulation (i.e., with the Co/Al ratio of 5 and calcinations temperature of 800 °C) the ignition temperature of soot oxidation was 290 °C; however, the NOx selective conversion to N2 remains too low, which is lower than 4%.

Ternary LDH-derived catalysts have also been proposed in this context. Li et al. [27] studied the hydrotalcite-based  $Mn_xMg_{3-x}AlO$  (Table 2) and found that the soot combustion activity follows the order:  $Mn_{1.5} > Mn_{1.0} > Mn_{0.5} = Mn_{2.0} > Mn_{2.5} = Mn_{3.0} > Mn$ -free, while the sample  $Mn_{1.0}$  shows the best performance in the simultaneous soot– $NO_x$  removal. They conclude that the  $Mn^{4+}$  ions are the most active species for soot combustion and simultaneous soot -  $NO_x$  removal.

Sample	NO <sub>x</sub> Uptake (µmol/g <sub>cat</sub> )	NO <sub>x</sub> Reduction Percentage (%)
Mn-free	373 (100–272 °C)	7.2 (278–700 °C)
Mn0.5	657 (100–404 °C)	20.4 (327–614 °C)
Mn1.0	502 (100–386 °C)	24.0 (295–554 °C)
Mn1.5	271 (100–336 °C)	12.6 (308–700 °C)
Mn2.0	233 (100–355 °C)	6.9 (253–460 °C)
Mn2.5	85 (100–202 °C)	6.5 (263–618 °C)
Mn3.0	108 (100–213 °C)	10.4 (212–648 °C)

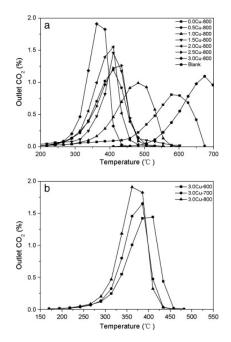
**Table 2.**  $NO_x$  storage capacities and  $NO_x$  reduction percentages of the  $Mn_xMg_{3-x}AlO$  catalysts. Adapted with permission from [27]. Copyright (2010) American Chemical Society.

Later on, Li and their coworker [29] proposed a series of K-promoted hydrotalcite-derived Mn<sub>1.5</sub>Mg<sub>1.5</sub>AlO catalysts; the catalyst exhibits both an efficient soot oxidation and NO<sub>x</sub> storage. Moreover, the authors suggest different pathway for the soot oxidation depending to the loading

of potassium; in particular, when K is less than 10 wt %, the reaction follows the oxygen spillover mechanism, while for higher K content the direct participation of the surface nitrates in soot oxidation is claimed, hence suggesting the involvement of a redox mechanism occurring between nitrates and soot particles. Between all the surface species formed during the storage, DRIFT spectroscopy results revealed that potassium monodentate nitrate are the most reactive with soot and their formation is ruled by K loading. Moreover, a new phase was identified, i.e., K<sub>2</sub>Mn<sub>4</sub>O<sub>8</sub> that proves to be highly active for soot combustion and NO<sub>x</sub> reduction by soot.

A series of Mn-doped MgAlO<sub>x</sub> was also prepared by Cui et al. [30]. The catalysts exhibited high NO<sub>x</sub> storage capacity at low temperatures (150–300 °C), due to its greater surface area, improved reducibility and higher surface  $Mn^{3+}$  content. During the lean-rich cycling tests, the average NO<sub>x</sub> removal rate can reach above 70% after Mn doping. However, the presence of soot has a slight detrimental effect on the NO<sub>x</sub> uptake, that decreases from 426 µmol/g at 200 °C to 406 µmol/g. This could be related to the exothermic combustion of soot destabilized the stored species and the soot combustion produced CO<sub>2</sub> would compete with NO<sub>x</sub> for storage sites on the catalyst. Unfortunately, the authors do not investigate the activity of Mn-doped MgAlO<sub>x</sub> directly in soot combustion.

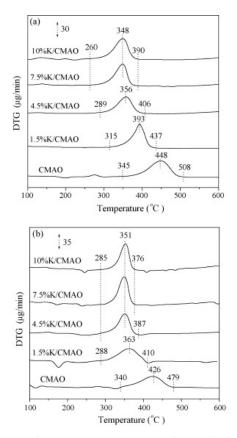
Between the ternary LDH-derived catalysts, the effect of Co and Cu in the formulation has been also studied. Indeed, Cu- and Co-MgAl LDH-derived catalysts have demonstrated to be very active for the simultaneous catalytic removal of soot and NO<sub>x</sub> due to their redox and acid–base properties. As a matter of fact, in the review of Yang et al. [9] some examples of CuMgAlO<sub>x</sub> catalyst active in the simultaneous removal of soot and NO<sub>x</sub> are reported. It has been verified that the catalytic activity strongly depends on the calcination temperature, being the optimal fixed at 800 °C [28,31]. Additionally, the results suggested that the addition of Cu significantly increased the activity of catalysts; among the tested catalysts, Cu(3)MgAlO<sub>x</sub> sample calcined at 800 °C shows the best activity with  $T_i = 260$  °C and total amount of N<sub>2</sub> formed during the TPR run near  $6.0 \times 10^{-5}$  mol, as evident in Figure 2.



**Figure 2.** Catalytic performances for soot oxidation over CuMg/Al catalysts during NO<sub>x</sub> soot reactions. Gas composition: NO 0.25 vol.%, O<sub>2</sub> 5 vol.%, balance He; flow rate: 20 cm<sup>3</sup>/min Reproduced with permission from [28]. Copyright © 2012, Elsevier, Inc. (a) outlet CO<sub>2</sub> concentration over CuMg/Al catalysts calcined at 800 °C at different Cu loading; (b) outlet CO<sub>2</sub> concentration over 3.0 CuMg/Al catalysts calcined at different temperature.

The effect of calcination temperature has been reported by Li et al. [12] for the  $Co_{2.5}Mg_{0.5}Al$  catalyst and 4.5% K-promoted sample, founding that simultaneous catalytic removal of soot and NO<sub>x</sub> can be

achieved over these catalysts in the temperature range of 300–700 °C. The catalyst 4.5% K/Co<sub>2.5</sub>Mg<sub>0.5</sub>Al calcined at 600 °C shows the best performance, not only for soot combustion but also for simultaneous soot–NO<sub>x</sub> removal; in this case, the soot ignition temperature is near 330 °C and the NO<sub>x</sub> reduction percentage results as high as 32%. This is attributed to its high surface K/Co atomic ratio and to the strong interaction between K and Co. To further study the performance of K-promoted Co<sub>2.5</sub>Mg<sub>0.5</sub>Al catalysts, the effects of K loading has been evaluated [13]. The results showed that the soot combustion activity is higher when K is present and it increases with K loading. Indeed, as it is reported in Figure 3, the onset temperature for soot oxidation shifts towards lower values upon increasing K loading.



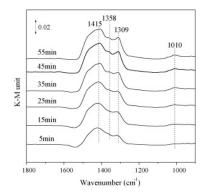
**Figure 3.** DTG profiles of soot combustion on x% K/Co-MgAlO catalysts (x = 0, 1.5, 4.5, 7.5 and 10) in different atmosphere: (a) air, (b) mixture of 400 ppm NO + 10% O<sub>2</sub> balanced by N<sub>2</sub>. Adapted with permission from [13]. Copyright © 2009, Elsevier, Inc.

The data reported in the Figure 3 clearly demonstrate that the soot oxidation activity of Co-MgAlO hydrotalcite is enhanced by the presence of K both in the air and in the presence of NO. Additionally, Zhang et al. [22] come to the same conclusions studying K/MgAlO catalysts. The authors attributed this positive effect of potassium to its strong interactions with Mg(Al) species; this weakens the bonds in CoAl<sub>2</sub>O<sub>4</sub> spinel and Mg(Al)–O, facilitating the mobility of bulk lattice oxygen species. Li et al. [13] performed also NO<sub>x</sub> storage experiments at 350 °C, i.e., the temperature at which the maximum conversion of soot is observed in the same conditions. The results showed that when Co-MgAlO catalysts are promoted with K, the NO<sub>x</sub> storage capacity is increased (Table 3); moreover, the NO<sub>x</sub> storage capacity increases with potassium loading. The NO adsorption is facilitated by activation of gaseous oxygen and NO itself due to the presence of electron-donating K species.

DRIFT studies reported in Figure 4 demonstrated that on the 1.5% K/Co-MgAlO catalyst the main adsorbed species are monodentate nitrates on K sites that have a stronger basicity than Mg sites. When potassium loading increases, monodentate nitrates transform into ionic species.

Sample	NO <sub>x</sub> Uptake (mg/g <sub>cat</sub> )
Co-MgAlO	24.10
1.5% K/Co-MgAlO	31.75
4.5% K/Co-MgAlO	51.88
7.5% K/Co-MgAlO	56.03
10% K/Co-MgAlO	61.24

**Table 3.** NO<sub>x</sub> storage capacities over Co-MgAlO and the K-promoted catalysts. Adapted with permission from [13]. Copyright © 2009, Elsevier, Inc.



**Figure 4.** The in situ DRIFTS spectra of NO<sub>x</sub> sorption on 1.5% K/Co-MgAlO catalyst. Adapted with permission from [13]. Copyright © 2009, Elsevier, Inc.

In summary, in this work the authors assumed a mechanism for NO<sub>x</sub> adsorption over K/Co-MgAlO catalysts similar to that proposed for traditional K-based lean NO<sub>x</sub> trap (LNT) catalysts like Pt/K/Al<sub>2</sub>O<sub>3</sub> catalysts [32–34]. In fact, in these traditional catalytic systems Pt is the active site for NO oxidation, while K<sub>2</sub>O is assumed to be the main potassium species where the adsorption occurs; along the same lines, in K/Co-MgAlO catalysts the Co sites are thought to be responsible for the activation of NO molecules. It is important to note that although the NO<sub>x</sub> storage capacity of K/Co-MgAlO catalysts is undoubted, nothing is reported about the reduction of stored NO<sub>x</sub>.

In the review by Yang et al. [9] the authors concluded that hydrotalcite-based catalysts have demonstrated to be active in the simultaneous removal of soot and NO<sub>x</sub>; indeed, in the K–M–O system new phase reactive oxygen species are present, which easily react with soot. These species, formed due to the strong interaction between K and metals on the support of catalysts, oxidized NO to NO<sub>2</sub>, being the first stored in the form of nitrates over K species while the former reacts with soot. Definitely, even if there is still room to improve both soot oxidation activity and the NO<sub>x</sub> storage capacity of the LDH derived catalysts, in the future they could represent a viable solution for the vehicle emissions control systems.

#### 2.2. Perovskite-Based Catalysts

Today, perovskites are considered as a viable alternative choice to PGM-based catalysts in the oxidation of particulate matter due to their ease of synthesis, high thermal stability and low cost compared to PGMs. In addition, these materials could be modified and doped with a wide range of elements in order to adapt their properties in relation to their specific applications [35–37].

The general chemical formula for perovskite compounds is ABO<sub>3</sub>, where "A" and "B" are two cations of very different sizes, and O is the anion that bonds to both. In general, "A" is rare earths (e.g., La, Ce, and Pr), alkali and alkaline earths (e.g., Cs, Sr, Ba, and Ca) larger than "B" transition metals (e.g., Co, Fe, Cu, Ni, Mn, Cr, and Al). A large number of metal ions having a different valence can replace both A and B ions, thus modulating the catalytic properties of these materials. Note that the catalytic properties of perovskite-type oxides basically depend on the nature of A and B ions; the A site ions are

catalytically inactive, while catalytic activity is generally determined by the B cation. These materials have the capacity to form reactive oxygen species with high mobility, becoming a good option to replace noble metals for soot removal applications, like catalyzed DPF. Furthermore, alkali oxides like K or Li are often added to perovskites to improve soot combustion processes. Owing to their redox properties coupled to a high oxygen mobility, perovskites were used for soot oxidation also in the presence of NO, with particular attention to the NO oxidation reaction (being NO<sub>2</sub> more reactive than NO in soot oxidation) and to the selectivity of the C + NO<sub>2</sub> reaction. These studies are the basis for the development of perovskites catalysts for the simultaneous removal of soot and NO<sub>x</sub> from automotive exhausts. In particular, the incorporation of dopants in the parent perovskite structure was indicated to improve both de-NO<sub>x</sub> and de-soot catalytic activity. The application of perovskites catalysts in soot and NO<sub>x</sub> control is summarized in Table 4.

Perovskites Families	Chemical Composition	Catalysts	Preparation Method	Reaction Conditions	Ref.
Manganites	A <sub>1-x</sub> R <sub>x</sub> MnO <sub>3</sub>	$La_{1-x}K_{x}MnO_{3}$ (x = 0.1, 0.2, 0.25, 0.3)	Co-precipitation; calc. 850–950 °C	Cat/soot = 20/1; NO (0.5%) + O <sub>2</sub> (5%) + He; total flow 20 cm <sup>3</sup> /min; heating rate 1 °C/min	[38]
	La <sub>1-x</sub> K <sub>x</sub> MnO <sub>3</sub>	Citrate-combustion route; calc. 750 °C	$\label{eq:cat/soot} \begin{split} & \text{Cat/soot} = 10/1; \ 10\% \ \text{O}_2 + 0.5\% \\ & \text{NO} + \text{He}; \ \text{total flow 100 mL/min}; \\ & \text{heating rate 1 }^\circ\text{C/min} \end{split}$	[39]	
		La <sub>1-x</sub> K <sub>x</sub> MnO <sub>3</sub>	Complex combustion method; calc. 800 °C	Cat/soot = 10/1; NO 1000 ppm + O <sub>2</sub> 0.5 vol% + He. total flow 70 mL/min; heating rate 5 °C/min.	[40]
		La <sub>0.7</sub> Ag <sub>0.3</sub> MnO <sub>3</sub>	Solid-state method; calc. 600 °C	Cat/soot = 20/1; 10% NO-Ar mixture 30 mL/min, oxygen 10 mL/min, air 60 mL/min; heating rate 10 °C/min	[41]
Cobaltites	$A_{1-x}R_xCoO_3$	La <sub>1-x</sub> K <sub>x</sub> CoO <sub>3</sub>	Citric acid-ligated method; calc. 800 °C	Cat/soot = $5/1$ ; 5% O <sub>2</sub> + 2000 ppm NO + He; total flow 50 cm <sup>3</sup> /min; heating rate 2 °C/min	[42]
		La <sub>1-x</sub> K <sub>x</sub> CoO <sub>3</sub>	Complex combustion method; calc. 800 °C	$\begin{array}{l} \mbox{Cat/soot} = 10/1; \mbox{NO} \ 1000 \ \mbox{ppm} + \\ O_2 \ 0.5 \ \mbox{vol}\% + \mbox{He. total flow 70} \\ \mbox{mL/min; heating rate 5 °C/min.} \end{array}$	[40]
Ferrites	$A_{1-x}R_xFeO_3$	La <sub>1-x</sub> K <sub>x</sub> FeO <sub>3</sub>	complex combustion method; calc. 800 °C	Cat/soot = 10/1; NO 1000 ppm + O <sub>2</sub> 0.5 vol% + He. total flow 70 mL/min; heating rate 5 °C/min.	[40]
Chromites	$A_{1-x}R_xCrO_3$	La <sub>1-x</sub> K <sub>x</sub> CrO <sub>3</sub>	citrate-combustion route; calc. 700 °C	$\label{eq:cat/soot} \begin{split} & \text{Cat/soot} = 10/1; \ 10\% \ \text{O}_2 + 0.5\% \\ & \text{NO} + \text{He}; \ \text{total flow 100 mL/min;} \\ & \text{heating rate 1 }^{\circ}\text{C/min} \end{split}$	[39]
Titanites	A <sub>1-x</sub> R <sub>x</sub> TiO <sub>3</sub>	Sr <sub>0.8</sub> K <sub>0.2</sub> TiO <sub>3</sub>	sol–gel method; calc. 850 °C	$\begin{array}{l} Cat/soot = 4/1; 500 \text{ ppm NOx} + O_2 \\ 5\% + N_2; \text{ heating rate } 10 \ ^\circ\text{C/min.} \end{array}$	[43]

Table 4. Literature review of perovskite-based catalysts for the simultaneous removal of soot and NO<sub>x</sub>.

A large number of investigations focused on mixed oxides with perovskite and spinel structures (i.e., see the review by Hernández-Giménez et al. [44] and references therein) aiming at developing efficient and cheap soot oxidation catalysts. Many catalytic formulations have been reported to promote soot combustion, including noble metals, alkaline metals and alkaline earth metals, transition metals that can accomplish redox cycles (e.g., V, Mn, Co, Cu, Fe, etc.), and internal transition metals (i.e., rare earth metals). In general, alkali metals are the most intensively studied dopants [45–48] even if, in some cases, potassium-containing catalysts are reported to suffer from low stability at high temperatures [49].

Mn-based materials are a good choice for oxidative applications. Teraoka et al. [38,45] first and Wang [50] later on, systematically investigated perovskite or spinel oxides for the simultaneous catalytic removal of soot and NO<sub>x</sub> in oxygen-containing model exhausts, focusing the attention on the promoting effect of potassium. Indeed, doping with K can effectively reduce the ignition temperature of soot and improve the selectivity for reducing NO<sub>x</sub> to N<sub>2</sub>. Both founded that nanosized La<sub>1-x</sub>K<sub>x</sub>MnO<sub>3</sub> (x = 0.2, 0.25) exhibited a very high catalyzing activity under a loose contact conditions. The study of Teraoka [38] revealed that the La–K–Mn–O oxides are good candidates for the simultaneous NO<sub>x</sub>–soot removal reaction, as demonstrated by the formation of  $CO_2$  due to the oxidation of the soot and the reduction of NO into N<sub>2</sub> (and N<sub>2</sub>O) observed in the same temperature range. The ignition temperature for soot oxidation decreases by increasing the K loading down to 208 °C when x = 0.4, while the 30% of reduced NO was converted to N<sub>2</sub>O. It is worth to note that the study has been performed under tight conditions between catalysts and soot particle. However, under practical conditions the contact between the catalysts on the surface of filter and PM particle is loose.

La–K–Mn–O systems have been considered also by Peron and Glisenti [39] in a very recent work, which focused on LaMnO<sub>3</sub> compared with LaCrO<sub>3</sub> as potential starting points for the substitution of noble metal catalysts. The perovskite formulation has been also modified by doping with K, and the resulting materials, i.e., La<sub>0.8</sub>K<sub>0.2</sub>CrO<sub>3</sub> and La<sub>0.8</sub>K<sub>0.2</sub>MnO<sub>3</sub> have been tested in soot oxidation by O<sub>2</sub> and NO. Potassium has improved activity for both Cr- and Mn-containing perovskites, but Mn-containing perovskites results more active than chromites (T<sub>max</sub> 314 °C vs. 337 °C).

The remarkable improvement in the soot oxidation activity also under the loose contact condition reported for chromites (LaCrO<sub>3</sub>) with the respect to manganite (LaMnO<sub>3</sub>), suggests that also other perovskite families, like cobaltites (LaCoO<sub>3</sub>), ferrites (LaFeO<sub>3</sub>), and titanites (SrTiO<sub>3</sub>) where La (or Sr) is partially substituted with K, could be proposed in this field. This is confirmed by Kureti et al. [51] reporting that iron-containing materials are promising catalysts for simultaneous NO<sub>x</sub>–soot removal. On the other hand, Fino et al. [52] investigated the effect of Cr/Fe presence, preparing a series of Cr- and Fe-substituted perovskite samples (LaMnO<sub>3</sub>, LaFeO<sub>3</sub>, LaCrO<sub>3</sub>, LaCrO<sub>3</sub>, and La<sub>0.9</sub>K<sub>0.1</sub>Cr<sub>0.9</sub>O<sub>3- $\delta$ </sub>) and testing them in the soot oxidation. The comparative analysis of such catalysts showed that chromite class are more active than LaMnO<sub>3</sub> and LaFeO<sub>3</sub> catalysts, having La<sub>0.9</sub>K<sub>0.1</sub>Cr<sub>0.9</sub>O<sub>3- $\delta$ </sub> the lowest CO<sub>2</sub> peak temperature (455 °C).

Cobaltites-based catalysts have been considered by Wang et al. [42] studying the catalytic performance of the La–K–Co–O perovskite oxide catalyst. The partial substitution of La<sup>3+</sup> at A-site by alkali metal K<sup>+</sup> enhanced the catalytic activity for the oxidation of soot particle and reduction of NO<sub>x</sub>. The results demonstrated that the La<sub>0.70</sub>K<sub>0.30</sub>CoO<sub>3</sub> sample is a good candidate catalyst for the simultaneous removal of the soot particle and NO<sub>x</sub>; indeed, the combustion temperatures for soot particles are in the range from 289 to 461 °C, the selectivity of CO<sub>2</sub> is very high near 98% and the conversion of NO to N<sub>2</sub> is 34.6% under loose contact conditions. The authors proposed at least three possible mechanisms to explain the enhanced catalytic activity after K-doping: (i) the formation of high valance ion (Co<sup>4+</sup>) at B sites of perovskite, which had better catalytic oxidation activity than Co<sup>3+</sup>; (ii), the formation of oxygen vacancy, whose presence favors the NO adsorption; and (iii) the formation of oxide catalysts with nanometric size and thus the good contact between the catalysts and soot.

Recently, Dhal et al. [40] discuss the properties of nanometric  $La_{1-x}K_xMnO_3$ ,  $La_{1-x}K_xCoO_3$ ,  $La_{1-x}K_xFeO_3$ , and  $La_{1-x}Na_xMnO_3$  perovskite-type oxide catalysts in the simultaneous removal of NO<sub>x</sub> and soot. The reported results demonstrate that all of the catalysts were active in soot combustion, showing the manganite-based catalyst the lower ignition temperature (see Table 5). The removal of NO<sub>x</sub> has been investigated in presence and absence of soot under cycling conditions, i.e., alternating lean-rich phases according to the DPNR concept (see below), and the NO<sub>x</sub> conversion is reported in Table 5.

**Table 5.** Ignition temperature ( $T_i$ ) of soot combustion and NO<sub>x</sub> conversion and of La-based catalysts.Adapted with permission from [40]. Copyright © 2017, Elsevier, Inc.

Catalyst	T <sub>i</sub>	NO <sub>x</sub> Conversion
LaCoO <sub>3</sub>	300 °C	45% @400 °C
La <sub>1-x</sub> K <sub>x</sub> CoO <sub>3</sub>	200 °C	50% @305 °C
La <sub>1-x</sub> K <sub>x</sub> MnO <sub>3</sub>	150 °C	53% @300 °C
La <sub>1-x</sub> K <sub>x</sub> FeO <sub>3</sub>	250 °C	48% @320 °C

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The authors founded that when K is not present in the formulation, the activity order for soot oxidation follows  $LaMnO_3 > LaFeO_3 > LaCoO_3$ . On the other hand, the addition of K makes  $La_{1-x}K_xCoO_3$  more active than  $La_{1-x}K_xFeO_3$ , while  $La_{1-x}K_xMnO_3$  remains the most active. They proposed a soot oxidation mechanism operating through different routes depending on the temperature: in the low temperature region, NO is initially oxidized to NO<sub>2</sub> over catalysts, then NO<sub>2</sub> oxidizes the soot particles; at high temperature besides the soot oxidation by NO<sub>2</sub>, the active oxygen on the surface of the catalysts can directly oxidize soot particles involving superficial oxide complexes intermediates. The N<sub>2</sub> formation occurs through the reaction between the soot particles and adsorbed nitrate species acting as oxidant species.

Meanwhile, some studies have reported that the partial substitution of lanthanum by strontium in LaCoO<sub>3</sub> and LaMnO<sub>3</sub> perovskites improved NO-to-NO<sub>2</sub> oxidation activity beyond that of the model Pt-based catalyst formulation [53,54]; furthermore, other studies have demonstrated that strontium substitution of lanthanum could modulate the NO<sub>x</sub>-to-N<sub>2</sub> efficiency of perovskite-based catalysts [55,56]. Indeed, it has been found that both La<sub>1-x</sub>Sr<sub>x</sub>CoO<sub>3</sub> and La<sub>1-x</sub>Sr<sub>x</sub>MnO<sub>3</sub> perovskites compared favorably with a commercial platinum-based catalyst in DOC (diesel oxidation catalyst) and LNT (lean NO<sub>x</sub> trap) systems [57], with improved activity in soot oxidation [58,59].

Additionally, the dual substitution of Ag and K in place of La in LaMnO<sub>3</sub> has been demonstrated giving improved activity in the simultaneous soot–NO<sub>x</sub> reaction [41]. As a matter of fact, soot combustion is largely accelerated, with the temperature for maximal soot conversion (T<sub>m</sub>) lowered by at least 50 °C; moreover, the silver substitution at A site of perovskite increases the NO<sub>x</sub> reduction efficiency (Table 6). Indeed, the metallic Ag can efficiently adsorb NO and O<sub>2</sub>, and oxidize NO to NO<sub>2</sub> and, given that NO<sub>2</sub> is a better oxidizer for soot than NO or O<sub>2</sub>, it enhances the activity for both soot oxidation and NO<sub>x</sub> reduction. From the data reported in the work, it appears evident that the co-presence of Ag and K at the A-site of LaMnO<sub>3</sub> catalyst improved the NO conversion from 45% to 64%, better than in the only Ag-substituted catalyst (53.28%), even the temperature for maximum NO conversion was higher for K-substituted catalyst than the only Ag-substituted catalyst.

	Soot Oxidation		NO Reduction	
Catalyst	Т <sub>і</sub> (°С)	Τ <sub>m</sub> (°C)	T <sub>max</sub> -NO (°C)	Conversion (%)
LaMnO <sub>3</sub>	205	380	386	45.3
La <sub>0.65</sub> Ag <sub>0.35</sub> MnO <sub>3</sub>	152	300	301	53.2
La <sub>0.65</sub> Ag <sub>0.1</sub> K <sub>0.25</sub> MnO <sub>3</sub>	135	355	368	64.4

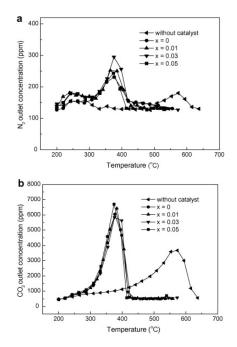
**Table 6.** Different catalysts' activity. Reprinted with permission from [41]. Copyright © 2018, Bulletin of Chemical Reaction Engineering and Catalysis.

Another example is reported by Li and coworkers [60], who prepared a series of nanometric Fe-substituted La0<sub>.9</sub>K<sub>0.1</sub>Co<sub>1-x</sub>Fe<sub>x</sub>O<sub>3- $\delta$ </sub> (x = 0, 0.05, 0.1, 0.2, and 0.3) perovskite catalysts for diesel soot oxidation, NO<sub>x</sub> storage, and simultaneous NO<sub>x</sub>-soot removal. The reported results showed that the formulation with x = 0.1 is the most active in the removal of both soot and NO<sub>x</sub>; in fact, the maximal combustion rate temperature is reached at 362 °C, the storage capacity is 213 µmol/g and the NO<sub>x</sub> reduction by soot is 12.5%. The reason of such higher activity is related to the high oxidation capacity of Fe, so NO is efficiently oxidized to NO<sub>2</sub> that is adsorbed on the catalyst surface more efficiently than NO. However, also in this work the reduction of stored NO<sub>x</sub> is reported in term of NO<sub>x</sub> released under the programming temperature.

Considering the titanites family, the activity and stability of two potassium-perovskite catalysts (K/SrTiO<sub>3</sub> and Sr<sub>0.8</sub>K<sub>0.2</sub>TiO<sub>3</sub>) and a potassium-copper perovskite catalyst (K–Cu/SrTiO<sub>3</sub>) for soot combustion in the NO<sub>x</sub>/O<sub>2</sub> gas mixture has been analyzed by López-Suárez and coworker [43]. Even the authors do not report any data on NO conversion, all the catalysts result in being active in the soot combustion in NO<sub>x</sub>/O<sub>2</sub> mixtures because of a decrease in the onset temperature for soot combustion and an increase significantly in the soot combustion rate. However, all the potassium-containing

catalysts suffer deactivation; the most significant decrease in catalyst activity takes place between the first and the second TPR (temperature programmed reduction) cycle, being the deactivation in further cycles much less important; indeed, the  $T_{50\%}$  (temperature required to achieve the 50% soot conversion) values increase by more than 150 °C from the first to the second cycle (i.e., from near 450 °C to ca. 600 °C).

Other more complex formulations have been proposed in the literature. Zhao et al. [61] studied  $La_{1-x}Ce_xNiO_3$  ( $0 \le x \le 0.05$ ) perovskite catalysts for simultaneous removal of nitrogen oxides and diesel soot. Indeed, the results reported in Figure 5 indicate that N<sub>2</sub> and CO<sub>2</sub> are produced almost at the same temperature range thus evidencing the occurrence of the simultaneous removal of NO<sub>x</sub> and soot. In particular, compared to the non-catalytic combustion of soot under the same reaction conditions, the ignition temperature decreases from 450 to 300 °C. The partial replacement of Ce for La increases the concentration of Ni<sup>2+</sup> thus promoting the catalytic activities. The redox properties of the Ce<sup>3+</sup>/Ce<sup>4+</sup> couple and the capacity of cerium oxide to exchange oxygen with the gas phase are also behind the good catalytic performance of ceria-based materials as soot combustion catalysts [62].



**Figure 5.** The outlet N<sub>2</sub> and CO<sub>2</sub> concentration profiles during temperature programmed reaction over  $La_{1-x}Ce_xNiO_3$  catalysts ( $0 \le x \le 0.05$ ). Substitution degrees (x) are indicated within the figure. (a) Outlet N<sub>2</sub> concentration profiles and (b) outlet CO<sub>2</sub> concentration profiles. Reprinted with permission from [61]. Copyright © 2009, Elsevier, Inc.

Very recently, simultaneous  $NO_x$  storage and soot oxidation performances of doped barium cerate perovskite materials have been reported by Maffei et al. [63]. The  $NO_x$  storage capacity of the co-doped (Zr and Co) barium cerate was found comparable to traditional Pt-based lean  $NO_x$  traps, i.e., 228 µmol/g at 380 °C and in the presence of water; moreover, the lowest  $T_{max}$  for soot oxidation resulted near 436 °C. However, in this case the rich step typical for DPNR systems has not been analyzed.

It is undoubted that the soot– $NO_x$ -catalyst system implies a very complex interplay and that the chemistry of the involved reactions is also complex. For these reasons, how occurs the simultaneous removal of  $NO_x$  and soot under the condition of rich oxygen over perovskite-type catalysts remains almost unclear and several mechanisms have been proposed. A scheme of the complex pathway of the reactions involved in the simultaneous removal of soot and  $NO_x$  has been proposed by Liu et al. [64] in the case of  $La_{2-x}Rb_xCuO_{4-\lambda}$  perovskite-like oxide catalysts.

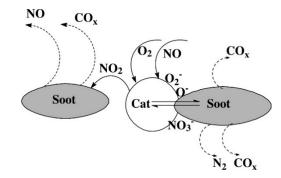
This scheme (Figure 6) supposes the formation of  $O_2^-$  and  $O^-$  species by dissociative adsorption of  $O_2$  on the catalyst surface. When the contact between catalyst and soot is guaranteed, they react to

each other forming CO and CO<sub>2</sub>, otherwise the catalyst cannot promote the soot oxidation reaction. Nevertheless, NO can be oxidized to NO<sub>2</sub> according to reaction (1), and since this species is more reactive towards soot than NO and/or O<sub>2</sub>, the reaction between NO<sub>2</sub> and soot occurs also under loose contact conditions according to global stoichiometry of reactions (2) and (3):

$$2 \text{ NO} + \text{O}_2 \rightarrow 2 \text{ NO}_2 \tag{1}$$

$$C + 2 \operatorname{NO}_2 \to \operatorname{CO}_2 + 2 \operatorname{NO}$$
 (2)

$$C + NO_2 \rightarrow CO + NO$$
 (3)



**Figure 6.** The scheme of the reaction mechanism for the simultaneous removal of soot and NO<sub>x</sub> over  $La_{2x}Rb_xCuO_{4\lambda}$  perovskite-like oxide catalysts. Reprinted with permission from [64]. Copyright © 2008, American Chemical Society.

Additionally, on the catalyst surface nitrate species could be formed, according to reactions (4) and (5), which in turns could be reduced by soot giving nitrogen reactions (6) and (7):

$$NO + O_2^- \to NO_3^- \tag{4}$$

$$NO_2 + O^- \to NO_3^- \tag{5}$$

$$2 \text{ NO}_3^- + 2 \text{ C} \to \text{N}_2 + 2 \text{ CO}_2 + 2 \text{ O}^-$$
(6)

$$2 \text{ NO}_3^- + 4 \text{ C} \to \text{N}_2 + 4 \text{ CO} + 2 \text{ O}^-$$
(7)

Despite the extensive research efforts, the activities of the perovskite catalysts remain typically inferior to those of the precious metal-based counterparts, being the role of noble metal mainly to oxidize NO to NO<sub>2</sub>, which subsequently oxidizes soot to CO and CO<sub>2</sub>. Therefore, NO<sub>2</sub> is used as an intermediate to facilitate an indirect contact between the catalyst and soot. For this reason, a small amount of noble metal like Pt or Pd has been incorporated into the B site of the perovskite structure. It has been reported that the catalytic activity is remarkably improved; in particular, Pd results more effective than Pt for NO<sub>x</sub> reduction at lower temperatures and the Pd–K interaction to promote the reduction of pre-adsorbed NO<sub>x</sub> [65,66].

To conclude, it is worth to note that the concept of "simultaneous removal of soot and  $NO_x$ " in most cases is intended as oxidation of soot and simultaneous reduction of  $NO_x$ , i.e., the oxidation of soot occurs in the presence of  $NO_x$ , which in turn are reduced to  $N_2$ . This is very different from the meaning given by the DPNR Toyota concept, as detailed in the following. Indeed, in the literature rarely the perovskites behaviors are analyzed in the complete lean-rich cycles typical of DPNR catalysts. The main problem is that the perovskite-like structure is extremely damaged under a reducing atmosphere, but it seems to be recovered by calcining at 400 °C under lean conditions. Furthermore, the effect of  $H_2O$  or  $CO_2$  normally present in the exhausts often is not considered [35].

Noble metal-based catalysts have been extensively studied in the catalytic soot combustion in order to enhance the intrinsic redox ability of catalysts. Although many methods are employed to synthesize supported noble metal nanoparticles, including impregnation, ion exchange, liquid-phase chemical reduction, and co-precipitation, there is much interest in novel methods to synthesize improved catalysts. Indeed, it is well-known that the high surface-to-volume ratio of the particles is crucial for the activity of the metal nanoparticle. So, the ability to synthesize stable and precisely engineered nanoparticles in order to optimize their catalytic activity is highly desired. The atomic layer deposition (ALD) technique can offer several benefits when compared to conventional nanoparticle synthesis methods, for example extreme film thickness uniformity, precise thickness control, excellent step coverage, and high reproducibility. The thickness of the films can be easily controlled by controlling the number of deposition cycles. For these reasons, ALD can be used to deposit catalytic coatings on high surface area porous powder supports or on geometrically complex structures [67–69] such as particulate filters in diesel engine exhaust systems.

Metal-doped hydrotalcites have been proposed due to the ability of LDHs to improve the dispersion of precious metals [66]. Indeed, as already discussed above, K-supported MgAlO (K/MgAlO) systems showed improved NO<sub>x</sub> adsorption at high temperatures [18]; in addition the optimization with noble metals (i.e., mainly Pt and Pd) improved the catalytic performance, particularly at low temperatures [70]. Moreover, it has been found that there is an interaction between the hydrotalcite-like derivative structure doped with Pd and the matrix oxide [71], so that Pd promotes the reduction of the composite oxide and increase active oxygen species in the surface, as well. For these reasons, the soot–NO<sub>x</sub> simultaneous removal performance of the hydrotalcite-like catalyst was improved by combining K and Pd, which improve the catalytic soot combustion activity and the NO<sub>x</sub> storage/reduction efficiency, respectively.

Among the others PGM-based catalytic systems, the best known is the already mentioned DPNR system patented by Toyota group in the early 2000s [72–75]. This system applies a NO<sub>x</sub> storage and reduction catalyst (like LNT catalyst) uniformly coated on the wall surface and in the fine pores of a highly porous filter substrate. The direct injection diesel engine configuration includes a newly-developed fuel injector installed on the upper stream of catalyst for adding fuel to the exhaust system, a common-rail fuel injection system capable of carrying out high-pressure, high-precision fuel injection control and an electrically controlled EGR system (Figure 7), in order to make it possible to drive with a rich air-fuel ratio, and to achieve precious catalyst temperature control.

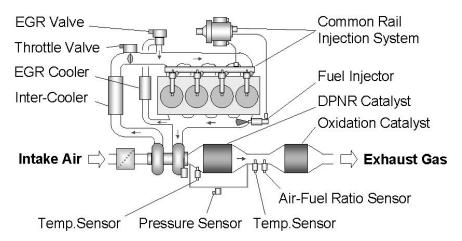
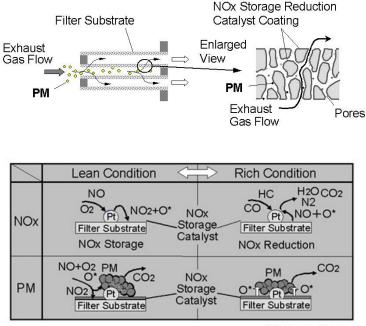


Figure 7. Engine system configuration with the diesel particulate–NO<sub>x</sub> reduction (DPNR) system.

 $NO_x$  are adsorbed on the DPNR catalyst during the lean phase and are reduced during the subsequent rich phase. Particulate matter emitted from the engine is trapped by the filter substrate and then, starting from 300 °C, it is continuously oxidized by the oxygen and  $NO_x$  contained in the

exhaust gas and by an active form of oxygen produced on the catalyst by repeated switching between a lean air-fuel ratio and rich air-fuel ratio (Figure 8). Both pore structure of the filter substrate and the catalyst are specifically optimized to improve both soot trapping and oxidation efficiency.



O\*:Active Oxigen

**Figure 8.** DPNR catalyst cross-section and NO<sub>x</sub> and particulate matter (PM) treatment mechanism. Reprinted with permission from [76]. Copyright © 2003, Springer Nature.

The DPNR catalyst is formally a NSR (NO<sub>x</sub> storage reduction) catalyst used in the lean NO<sub>x</sub> trap (LNT) system. It is constituted by three key components, i.e., a high surface area support (e.g.,  $\gamma$ -alumina), a noble metal (Pt), and an alkaline or alkaline-earth metal oxide; the catalyst presents a high NO<sub>x</sub> storage capacity combined with a high soot oxidation capability.

In this section a survey on the existing literature on the DPNR concept is presented based on both the traditional Pt-based systems and the new Pt-free ones.

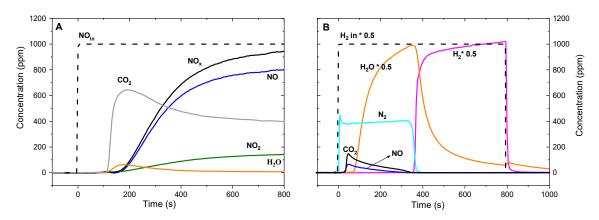
### 3.1. DPNR Catalysts for the Simultaneous Removal of $NO_x$ and Soot

In the last decade, the DPNR technology catalysts have attracted increasing attention due to the well performances of LNT catalysts also in the soot combustion. Among the others, Castoldi et al. [77], extensively studied the DPNR strategy based on a traditional LNT formulations, i.e., Pt-Ba-based catalysts supported on alumina. As a matter of fact, the Pt-Ba/Al<sub>2</sub>O<sub>3</sub> catalyst/soot mixture (9:1 w/w) has been tested at constant temperature under cycling conditions, i.e., alternating rectangular step feeds of NO/O<sub>2</sub> (lean condition) with rectangular step feeds of H<sub>2</sub> (rich condition).

A typical example of lean-rich sequence is reported in Figure 9. During the storage phase (Figure 9A),  $NO_x$  are efficiently stored over the Pt-Ba/Al<sub>2</sub>O<sub>3</sub> in the presence of soot with formation of surface nitrates, according to the global reaction (8):

$$2 \operatorname{BaO} + 4 \operatorname{NO} + 3 \operatorname{O}_2 \to 2 \operatorname{Ba(NO_3)}_2 \tag{8}$$

In the meantime,  $CO_2$  evolution is also observed and ascribed to soot oxidation. It is worth noting that  $CO_2$  is observed late (near 150 s) compared to the beginning of the storage phase due to its adsorption over barium sites with formation of BaCO<sub>3</sub>, and not correlated to a delayed combustion. In fact, when the same lean phase was performed in the presence of  $CO_2$ ,  $CO_2$  formation was immediately observed after NO addition to the reactor.



**Figure 9.** Lean and rich sequence (**A**, **B** respectively) obtained at 350 °C in the case of a Pt–Ba/Al<sub>2</sub>O<sub>3</sub>/soot mixture. Adapted with permission from [77]. Copyright © 2006, Elsevier, Inc.

The promoting effect of NO on soot combustion over Pt-based catalysts was ascribed to the ability of Pt to catalyze the oxidation of NO into  $NO_2$  (i.e., a stronger oxidant than oxygen), which favors the soot combustion according to the following overall reactions, already proposed in the literature:

$$NO + \frac{1}{2}O_2 \rightarrow NO_2$$
$$2 NO_2 + C \rightarrow CO_2 + 2 NO_2$$

Indeed, the amount of NO<sub>2</sub> detected results lower than that measured without soot at the same temperature, suggesting its consumption due to the reaction with soot.

It is noteworthy that the NO originated from reaction (2) can be oxidized again to  $NO_2$  according to reaction (1), resulting in the so-called NO recycle, which is expected to speed up the soot combustion [78]. Moreover, to explain the positive effect of  $NO_2$  formation on soot combustion a cooperative  $NO_2/O_2$  oxidation mechanism was also invoked. On this regard, several C-O<sub>2</sub>-NO<sub>2</sub>- reaction mechanisms were proposed. Indeed, some authors suggest the occurrence of a reaction between O<sub>2</sub> and the surface intermediates species arising from the interaction between soot and  $NO_2$  [79,80]; others report the existence of two parallel routes: (i) direct soot oxidation by  $NO_2$  and (ii) soot combustion involving cooperatively both  $O_2$  and  $NO_2$ , strongly catalyzed by Pt [81–84]. The latter favors the decomposition of surface oxygen species originated by  $O_2$  and reactions that involve species adsorbed onto the catalyst surface.

Furthermore, it has been suggested that another contribute to soot combustion is given by stored nitrates. In line with this hypothesis Castoldi et al. compared the behavior of  $Pt-Ba/Al_2O_3$  and  $Pt/Al_2O_3$  catalysts. They found that the Ba-free sample efficiently oxidizes soot but cannot store  $NO_x$ . On the other side the  $Pt-Ba/Al_2O_3$  sample exhibited a similar soot oxidation capacity despite the occurrence of  $NO_x$  storage on the catalyst surface (i.e., resulting in lower gas phase  $NO_x$  concentration than the Ba-free sample) suggesting specific oxidizing properties of the surface  $NO_x$  species. Accordingly, as it will be detailed later, the following global reaction could be hypothesized:

$$C + 2 NO_3^- \rightarrow CO_3^{2=} + 2 NO + \frac{1}{2} O_2$$
 (9)

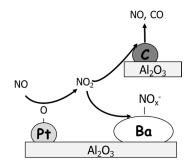
During the reduction of stored  $NO_X$  with  $H_2$  (Figure 9B),  $N_2$  production was observed (along with minor amounts of NO) accordingly to the overall stoichiometry:

$$Ba(NO_3)_2 + 5H_2 \rightarrow BaO + N_2 + 5H_2O \tag{10}$$

The water formed during reduction displaced carbonates previously formed and this accounts for the evolution of CO<sub>2</sub> during the reduction. Moreover, this CO<sub>2</sub> could also derive from limited soot

combustion during the rich phase. In fact, according to the literature the  $NO_x$  reduction involves at first the release of  $NO_x$ , which then are reduced to nitrogen over Pt [2]. Accordingly, the released  $NO_x$  may oxidize soot leading to the formation of  $CO_2$ . The oxidation of soot during rich pulses is also supported by Sullivan [85] and by Toyota researchers that, by means of ESR (electron spin resonance) experiments, correlated the higher soot oxidation rate observed in the presence of rich pulses to the formation of so-called activated oxygen species (e.g., superoxide species) generated under rich conditions [72].

The investigation has been extended by Lietti and coworkers considering the effect of temperature, the presence of  $CO_2/H_2O$ , and the NO inlet concentration on both  $NO_x$  storage reduction activity and soot oxidation capacity of model LNT catalyst [86–88]. As pointed out by the authors, the presence of soot strongly influences the storage behavior of NO<sub>x</sub> over Pt-Ba/Al<sub>2</sub>O<sub>3</sub>, decreasing its storage capacity in the range 250–350 °C regardless of the NO inlet concentration used in the experiment (i.e., 500 ppm and 1000 ppm). However, considering sequential lean-rich cycles, it appeared that the residual soot loading directly affects the performance of  $Pt-Ba/Al_2O_3$ . In fact, both the NO<sub>x</sub> breakthrough and the amounts of NO<sub>x</sub> stored at the steady state were found to progressively increase during the lean-rich sequence, i.e., upon decreasing the soot loading as consequence of soot combustion. Similar trends were also reported by Cortés-Reyes et al. [89] who observed lower  $NO_x$  adsorption rate in the presence of soot than in the absence (from  $9.5 \times 10^{-3}$  to  $4.5 \times 10^{-3}$  mg min<sup>-1</sup>), while once the catalyst becomes regenerated the adsorption rate of  $NO_x$  backs to its original values. This negative effect of soot on the  $NO_x$  storage capacity was explained by a less availability of  $NO_2$  for storage due to its involvement in the soot oxidation, i.e., Ba and soot compete in the reaction with NO<sub>2</sub> [85,90]. On the basis of the so-called "nitrate" route for the storage of  $NO_x$  (i.e., NO oxidation to  $NO_2$  and subsequent  $NO_2$ adsorption in the form of nitrates via a disproportion reaction [91]), soot was blamed to offer another path for the use of  $NO_2$  rather than the  $NO_x$  storage process, i.e., being  $NO_2$  involved in the soot oxidation instead of surface  $NO_x$  formation (Figure 10). Accordingly, in the presence of soot lower amounts of NO<sub>2</sub> were measured during the lean phase than that observed in the absence of soot.

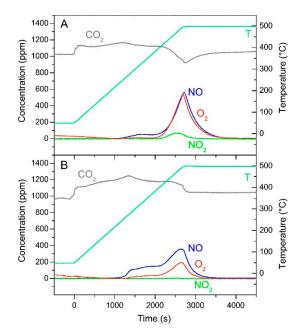


**Figure 10.** Proposed mechanism by which particulate matter decreases the  $NO_x$  adsorption capacity of a  $NO_x$  trap.

On the other side, the direct interaction between surface nitrates and soot particles via a surface reaction was reported to positively affect the soot combustion process. In particular, the direct participation of the surface  $NO_x$  in soot combustion was suggested to occur without the preliminary thermal decomposition of stored species. In fact, specific studies on the thermal stability of stored  $NO_x$  species [86–88] confirmed that in the absence of soot the thermal decomposition of stored species occurs near the adsorption temperature (in this case 350 °C, Figure 11A) and according to the global stoichiometry of reactions (11) and (12). Note that an uptake of  $CO_2$  was observed due to the formation of barium carbonates.

$$Ba(NO_3)_2 \rightarrow BaO + 2 NO + 3/2 O_2 \tag{11}$$

$$Ba(NO_3)_2 \rightarrow BaO + 2 NO_2 + 1/2 O_2$$
 (12)



**Figure 11.** TPD (Temperature Programmed Decomposition) run after NO<sub>x</sub> adsorption at 350 °C over (**A**) Pt-Ba/Al<sub>2</sub>O<sub>3</sub> catalyst and (**B**) Pt-Ba/Al<sub>2</sub>O<sub>3</sub>/soot mixture. Reprinted with permission from [87]. Copyright © 2011, Elsevier, Inc.

On the other hand, in the presence of soot (Figure 11B) the decomposition is observed at a lower temperature; moreover, also the products distribution (i.e., amount of NO, O<sub>2</sub>, and NO<sub>2</sub>) is different, and the registered concentrations obey to the global stoichiometry of reaction between soot and nitrates adsorbed species (reaction (9)), already suggested.

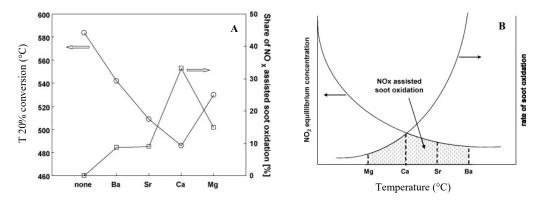
The above results on one side confirmed the destabilizing effect of soot on the stored  $NO_x$  and on the other supported the active role of the surface  $NO_x$  in soot combustion, without the necessity of their prior thermal decomposition. As a matter of fact, the stored nitrates were found to oxidize soot at temperatures lower than those of their decomposition without soot. The direct reaction between stored  $NO_x$  and soot was explained invoking the mobility of surface nitrates in analogy with the reduction mechanism proposed for LNT catalysts [92,93]. Since in the literature it has been suggested that nitrates adsorbed on Ba are mobile in the presence of reducing centers (i.e., reduced Pt sites kept in the reduced state by the reductant), similarly the presence of soot (a reductant) was supposed to be the driving force for the mobility of the nitrates, which may ultimately oxidize soot.

With this in mind,  $NO_x$  surface species act as an oxidant toward soot, which itself function as a reducing center; this redox mechanism represents a pathway parallel to the traditional  $NO_2$ -soot oxidation occurring in the presence of gas-phase  $NO_2$  during the DPNR operations.

The occurrence of a direct surface reaction between carbon particles and NO<sub>x</sub> stored species was also proposed by Tschamber and coworkers [94–97] to explain the decrease in the NO<sub>x</sub> storage activity when soot is present. The proximity between the storage sites, Pt sites, and their contact with soot is a key factor in the NO<sub>x</sub> storage behavior of the catalysts in the presence of soot according to the two possible pathways for NO<sub>x</sub> storage from NO/O<sub>2</sub> mixtures proposed by the literature [91], i.e., the "nitrate" and "nitrite" route, respectively. The "nitrite" route involves the stepwise oxidation of NO resulting to the formation of surface nitrites, and a cooperative interaction between Pt and Ba nearby sites is suggested as crucial for this route, which hence implies the existence of a strong Pt–Ba interaction. Nitrites are then gradually oxidized into nitrates, which prevail at catalyst saturation. Assuming that the proximity of Pt decreases the reducing character of soot via formation of C(O) oxygen surface complexes and consequently the redox interaction occurring between soot and nitrates, the authors suggested that nitrate species formed close to Pt (via "nitrite" route) are less destabilized by soot. Besides, the nitrate species formed far from Pt (via "nitrate route") were indicated to suffer

much more from contact with soot according to the above proposed surface reaction. As a matter of fact, TEM analysis revealed the occurrence of structure modifications following carbon combustion, i.e., platinum sintering and Ba agglomeration. These structural modifications were suggested to decrease the proximity between the platinum and storage sites, resulting in a decrease in the  $NO_x$  storage capacity though the "nitrite route". Additionally, the effect of water was investigated and a non-cumulative effect of carbon and  $H_2O$  on the  $NO_x$  storage capacity was pointed out resulting from the competition between the destabilization of the weakly bonded surface nitrate, by carbon, and the enhancement of bulk nitrates formation, by water.

The importance of the interaction between surface nitrates and soot and the beneficial effect on the soot oxidation process is also documented by Sullivan et al. [98] suggesting that on the Na/Al<sub>2</sub>O<sub>3</sub> catalyst NO adsorbs in the form of nitrites/nitrates, which can further decompose, i.e., releasing NO<sub>2</sub>. Shuang et al. [99] indicated the NO<sub>2</sub> derived from nitrates decomposition on Pt-Mg/Al<sub>2</sub>O<sub>3</sub> catalysts as beneficial for the soot oxidation activity. Krishna and Makkee [100] claimed the involvement of surface nitrates in soot oxidation via the release of NO<sub>2</sub> in the gas phase studying soot oxidation with the NO/O<sub>2</sub> mixture over Pt-K/Al<sub>2</sub>O<sub>3</sub> and Pt-Ba/Al<sub>2</sub>O<sub>3</sub> LNT catalysts. In another work, Kustov and Makkee [101] analyzed the impact of stored nitrates on soot combustion over Al<sub>2</sub>O<sub>3</sub> supported alkali-earth oxides (i.e., Ba, Sr, Ca, and Mg), founding that stored nitrates promote to the soot oxidation by decreasing the temperature of soot combustion up to almost 100 °C, as it clearly appears from Figure 12A where the activity of given catalysts is reported as a function of the temperature of 20% soot conversion.



**Figure 12.** (**A**) Influence of alkali-earth cations on T20% conversion and share of NO<sub>x</sub> assisted soot oxidation and (**B**) schematic explanation of NO<sub>x</sub> assisted soot oxidation for different alkali-earth metals. Reprinted with permission from [101]. Copyright © 2009, Elsevier, Inc.

Moreover, it was found that the temperature of the nitrates decomposition and the temperature of soot oxidation by  $NO_2$  are crucial for an efficient soot oxidation (see Figure 12B). Indeed, the efficiency of  $NO_x$  utilization depends on both these factors, being the soot oxidation by  $NO_2$  limited by the kinetics of the  $NO_2$ -C reaction and by the low thermodynamic stability of  $NO_2$  at high temperatures.

Sanchez and coworkers showed the direct participation of the surface nitrates in soot oxidation, without the need of their preliminary thermal decomposition, over K-containing lanthanum supported catalysts [102,103]. Other authors support the same thesis over Cs-loaded  $MnO_x$ –CeO<sub>2</sub> catalysts [104], and over BaAl<sub>2</sub>O<sub>4</sub> catalysts [105,106].

As already discussed above, the interaction between soot and nitrates (or, in general,  $NO_x$  stored species) influences the oxidation of soot and it is extremely efficient when K is present in the catalyst formulation. This beneficial effect was ascribed to the formation of K-based low melting point/volatile compounds, which can favor the surface mobility of the active species and consequently the soot–catalyst contact, that is essential for soot oxidation [107,108]. Moreover, a synergistic effect occurring between K and Pt has been reported to increase the mobility of actives species formed over the potassium thus enhancing the soot oxidation activity [81,100]. Many authors investigated the performances of K-based LNT catalysts in the soot combustion simultaneously to the  $NO_x$  storage.

Among others, Matarrese et al. [33,34,109,110] investigated the effect of the substitution of Ba by K in LNT catalysts. They founded a significantly higher activity for soot combustion for the Pt-K/Al<sub>2</sub>O<sub>3</sub> catalyst compared to that of Pt-Ba/Al<sub>2</sub>O<sub>3</sub> sample, while the catalysts showed a similar de-NO<sub>x</sub> activity. The high soot oxidation activity of Pt-K/Al<sub>2</sub>O<sub>3</sub> was explained taking into account the high mobility of K-nitrates, likely enhancing the soot–catalyst contact and hence boost in the soot combustion, according to the literature already reported. This was confirmed also by dedicated TPD (temperature programmed desorption) and TPO (temperature programmed oxidation) experiments [111] which pointed out the lower thermal stability of the NO<sub>x</sub> species stored over K and the higher reactivity of K-nitrates towards soot then Ba-ones. TPO results performed in the presence of oxygen by M. Cortés-Reyes et al. [82] suggests the following reactivity order in terms of onset temperature for soot oxidation.

$$Pt-K/Al_2O_3 > K/Al_2O_3 > Pt/Al_2O_3 = Pt-Ba/Al_2O_3 = Ba/Al_2O_3 = Al_2O_3.$$

These results were in agreement with those of Krishna and Makkee [100] who reported that both K/Al<sub>2</sub>O<sub>3</sub> and Pt–K/Al<sub>2</sub>O<sub>3</sub> show a superior soot oxidation capacity than Pt/Al<sub>2</sub>O<sub>3</sub> and Pt-Ba/Al<sub>2</sub>O<sub>3</sub>.

The main drawback of the K-based catalyst was the partial deactivation upon ageing with repeated  $NO_x$  storage reduction cycles in the presence of soot, showing a decreased  $NO_x$  storage capacity and also a lower soot oxidation activity. This was correlated to a reduced availability of the K actives species, due to partial loss of K and/or to participation of the K actives species in interactions with alumina and/or Pt [34].

#### 3.2. Pt-Free DPNR Systems

Well-known PGM catalysts are expensive and because of PGM low abundance, they may undergo increasing prices upon increasing demand [112,113]. For this reason, in the last years great efforts have been made to develop low-cost PGM-free catalysts, also for DPNR applications.

Castoldi et al. [114] proposed silver catalysts supported on Al<sub>2</sub>O<sub>3</sub>, CeO<sub>2</sub>, and ZrO<sub>2</sub> and containing Ba or Sr as storage components. The results showed all the investigated catalysts in  $NO/O_2$  promote the soot combustion at low temperatures (i.e., ca. 250 °C). Such a good oxidation activity was likely related to the ability of metallic silver to form suboxide species and/or superoxide O<sub>2</sub><sup>-</sup> ions, which are expected to assist carbon oxidation by  $O_2$ . Besides, all the catalysts were active in the oxidation of NO to NO<sub>2</sub> so that the presence of silver was indicated to promote also the NO<sub>2</sub>-assisted soot oxidation. All the Ag-containing catalysts were able to remove  $NO_x$  both in the absence and in the presence of soot, being able to adsorb them in lean conditions and subsequently reduce in rich ones. A lower NO<sub>x</sub> storage capacity was observed for the Sr-based samples if compared to the Ba-based samples, which was ascribed to the different basicity of strontium and barium. The reactivity of a model LNT Pt-Ba/Al<sub>2</sub>O<sub>3</sub> system was also considered for comparison purposes. The results showed that on one side, without soot, the storage capacity of the Ag-systems outperforms that of a traditional Pt–Ba/Al<sub>2</sub>O<sub>3</sub> LNT catalyst; on the other side, soot negatively affects the storage capacity of Ag-based catalysts more than model LNT catalysts. Besides the Ag-systems, particularly those supported on ceria, were found much more active in the simultaneous soot combustion than the Pt-based catalyst. This was correlated to the generation and participation of oxygen active species from silver and/or ceria. In conclusion, Ag-based catalysts were proven as a promising alternative to Pt-based catalysts for the simultaneous removal of soot and NO<sub>x</sub> even if their NO<sub>x</sub> reduction performances should be further improved, being low the selectivity to nitrogen.

Matarrese et al. [115] investigated a second class of Pt-free catalysts based on ruthenium supported on different supports ( $Ce_{0.8}Zr_{0.2}O_2$ ,  $ZrO_2$ , and  $Al_2O_3$ ) with Ba or K as NO<sub>x</sub> storage materials. According to TPO experiments, all the investigated systems were found more active in soot combustion than traditional Pt-based LNT materials. The presence of well dispersed ruthenium nanoparticles was suggested to promote the dissociative adsorption of oxygen leading to the formation of active oxygen species, which eventually can be transferred to the carbon promoting its oxidation. In particular, K-containing catalysts exhibited very low onset temperature in the presence of NO/O<sub>2</sub> (i.e., in the range 220–235 °C) pointing out a synergistic interaction between Ru and K. In addition, all the formulations were able to accomplish the simultaneous removal of soot and NO<sub>x</sub>, under isothermal cycling conditions, especially, NO<sub>x</sub> storage performances similar to conventional Pt-based catalysts were reported. Beside, K catalysts exhibited higher de-NOx and de-soot activity than Ba catalysts. However, also the NO<sub>x</sub> reduction efficiency of the Ru-containing catalysts requires further improvements.

Castoldi et al. compared Ag-, Ru-, and Pt-based catalysts supported on  $Al_2O_3$  and containing Ba as  $NO_x$  storage material [116]. The results confirmed the superior soot oxidation activity for Ag-Ba/Al<sub>2</sub>O<sub>3</sub> and even more for Ru-Ba/Al<sub>2</sub>O<sub>3</sub> than model Pt-Ba/Al<sub>2</sub>O<sub>3</sub> LNT catalysts. Besides their  $NO_x$  storage capacity was comparable to that of traditional Pt LNT catalysts. However, also in this case, the  $N_2$  selectivity during reduction was rather low, particularly in the case of Ag-Ba/Al<sub>2</sub>O<sub>3</sub> for which NO was the main reduction product (i.e.,  $N_2$  selectivity near 30%). Of note, as opposed to model Pt-based catalysts, the  $NO_x$  storage capacity of Ru-Ba/Al<sub>2</sub>O<sub>3</sub> was not negatively affected by soot; moreover, in the case of the Ru-Ba/Al<sub>2</sub>O<sub>3</sub> catalyst very similar NO/NO<sub>2</sub> ratios were measured with and without soot. This was correlated to the lower involvement of  $NO_2$  in the soot oxidation (in line with the direct involvement of active oxygen species formed by metallic Ru) or to the higher oxidation efficiency of NO to  $NO_2$ .

Since ceria-based oxides are recognized among the most promising materials for soot combustion [62,117], ceria/zirconia (CZ)-based catalysts, doped with Pt, Au, Ru, or Fe and containing K, have been considered for both soot oxidation and simultaneous removal of NO<sub>x</sub> by Matarrese et al. [118]. The results pointed out that all the CZ formulations and in particular the Ru-containing catalysts are able to decrease the soot ignition temperature in the presence of only oxygen at temperatures below 300 °C. Moreover, when operating under isothermal cycling conditions the Ru-based catalysts were found much more active than Pt-K/Al<sub>2</sub>O<sub>3</sub> in both soot combustion and NO<sub>x</sub> storage capacity. The high NO<sub>x</sub> storage activity was explained basing on FT-IR experiments. In fact, the initial formation of nitrites, which evolve fast to nitrates, was observed on both Ru and Pt-systems. Moreover, the contribution of also bulk nitrates and mono-nitrosyl species on Ru was found for the Ru system. However, also in this case, the Ru-based catalysts showed a poor NO<sub>x</sub> reduction activity if compared to model LNT catalysts.

Of note, it should be remembered that Ru-based catalysts have often been accused of low stability due to the possible loss of active phase (i.e., with consequent catalyst deactivation) via volatilization of Ru oxides. In particular, several recent studies [119,120] reported that at high temperatures (i.e., higher than 700 °C) RuO<sub>2</sub> is oxidized into volatile RuO<sub>4</sub>. Notably, focusing on this crucial aspect, Villani et al. [121] investigated the stability of Ru/Zeolite based catalysts and found 900 °C as the upper limit temperature for practical applications of ruthenium catalysts, while Matarrese et al. [118] performed repeated TPO soot oxidation cycles reporting a quite stable and reproducible behavior.

Bueno-Lopez et al. [122] investigated Cu/Ce<sub>0.8</sub>M<sub>0.2</sub>O<sub> $\delta$ </sub> catalysts (M = Zr, La, Ce, Pr, or Nd) for the simultaneous removal of soot and NO<sub>x</sub> and, also in this case, the presence of soot affected the NO<sub>x</sub> storage activity depending on the different nature of the acid/basic character of the doping metal, resulting in being detrimental for catalysts with very basic supports (e.g., doped with La). This was explained with the possible competition for the adsorption sites between the CO<sub>2</sub> emitted during soot combustion and NO<sub>x</sub>, in agreement with the DRIFT results. The behavior of the Pr-based catalyst (i.e., the best formulation for both NO<sub>x</sub> adsorption and soot oxidation) was further investigated at 400 °C under lean-rich cyclic conditions. N<sub>2</sub> was reported as the main reduction product under rich condition. Besides, soot oxidation was reported during H<sub>2</sub> pulses, which was tentatively explained invoking (i) with the destabilization of the stored nitrates upon H<sub>2</sub> admission leading to the formation of NO<sub>2</sub> that may oxidize soot and/or (ii) with the localized increase in temperature due to the oxidation of H<sub>2</sub>.

More recently copper/ceria-based catalysts were investigated for soot oxidation by Giménez-Mañogil et al. [123] who pointed out the synergistic interaction between copper and cerium, as responsible for the catalytic performances. In particular, copper in close contact with ceria was

indicated to improve the catalysts reducibility, which was found to play a key role in their catalytic behavior towards NO oxidation to NO<sub>2</sub> and soot combustion processes.

### 4. SCR/DPF Combined Technologies for the Simultaneous Removal of Soot and NO<sub>x</sub>

Among the combined technologies proposed to reduce simultaneously soot and NO<sub>x</sub> and realize more compact systems, an important role is played by SCR catalysts coated on a particulate filter, i.e., the so-called SCRF<sup>®</sup> systems (SCRF<sup>®</sup> is a registered trademark of Johnson Matthey Public Limited Company. All rights reserved) [124,125].

There are some difficulties in combining de-NO<sub>x</sub> and de-soot functionalities in terms of technology related to the combination of the de-NO<sub>x</sub> and PM abatement functionalities, mainly to the interactions between the SCR and soot chemistries that can result in a SCRF performance lower than that of the individual devices. Additionally, the impact on the mass-transfer characteristics by the presence of soot can affect in a negative way the de-NO<sub>x</sub> efficiency. In order to minimize the impacts of these factors, the catalytic material and the coating process are very important. Moreover, since it is desirable to introduce the highest possible quantity of catalyst in the pores of the filter, the porosity of the latter must be carefully considered due to the limits established by the maximum pressure loss a filter component can have. In addition, it is necessary to prevent thermal damage of the SCR-coating and the DPF-monolith.

A very crucial aspect to consider is the analysis of the performance of the SCRF<sup>®</sup> in comparison to that of the individual SCR and DPF devices. In particular, the close interplays between the SCR and DPF functions for the general case of soot-loaded devices deserve attention [126–128]. For example, it should be considered the involvement of NO<sub>2</sub> in both SCR reactions and passive soot oxidation but also the right compromise between de-NO<sub>x</sub> performance, filtration efficiency, and pressure drop behavior when selecting the optimal washcoat amount. Moreover, the influence of soot presence on NO<sub>x</sub> conversion must be taken into considerations. In fact, the reaction of NO<sub>2</sub> with soot forming NO affects the in situ NO<sub>2</sub>/NO<sub>x</sub> ratio, which as it is known directly impacts on the de-NO<sub>x</sub> performance. Therefore, a promoting or a detrimental effect takes place depending on the operating conditions [126]. Finally, also the thermal stability of SCR catalyst formulations needs attention bearing in mind the severe thermal conditions occurring within filters [129].

In 2012, Kröcher and a coworker [130] investigated the SCR reaction in the presence of  $NH_3$  over soot (i.e., both model and real diesel soot), in the temperature range between 200 and 350 °C. They observed SCR activity on diesel soot, which could be exploited to improve future diesel exhaust after treatment systems, where both DPF and SCR systems have to be used in series to reach the desired emission limits. The authors proposed a mechanism for  $NO_x$  reduction over diesel soot where the first reaction step is the disproportionation of  $NO_2$ , which is followed by the formation of ammonium nitrates and nitrites.

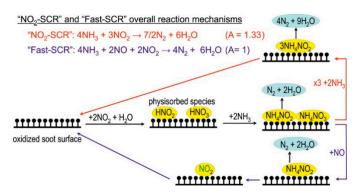
$$2 \operatorname{NO}_2 + \operatorname{H}_2 O \to \operatorname{HONO} + \operatorname{HNO}_3 \tag{13}$$

HONO and  $HNO_3$  are assumed to remain physisorbed on the surface, where they form ammonium nitrate ( $NH_4NO_3$ ) and ammonium nitrite ( $NH_4NO_2$ ) in the presence of  $NH_3$ :

$$2 \text{ NH}_3 + \text{HONO} + \text{HNO}_3 \rightarrow \text{NH}_4\text{NO}_3 + \text{NH}_4\text{NO}_2$$
(14)

The nitrites decompose directly into  $N_2$  and  $H_2O$  (indeed  $NH_4NO_2$  is unstable at temperatures above 60 °C), whereas the nitrates have to be reduced to nitrites either by NO or in its absence by  $NH_3$ , which also leads to the formation of  $N_2$  and  $H_2O$ . The SCR chemistry on the soot surface is summarized schematically in Figure 13.

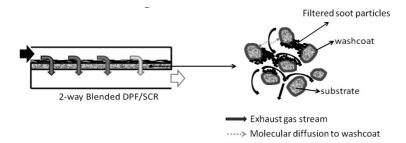




**Figure 13.** Suggested mechanism for the selective catalytic reduction of  $NO_x$  on diesel soot. The dark points symbolize acidic carbon surface species supporting  $NH_3$  adsorption. Reprinted with permission from [130]. Copyright © 2012, American Chemical Society.

Regarding the SCR component, the use of Cu- or Fe-based ion-exchanged zeolite for SCRF<sup>®</sup> applications [131–136] requires specific attention, i.e., in achieving the high NO<sub>x</sub> conversion over a wide temperature range, avoiding NH<sub>3</sub>-slip together with high efficiency for soot filtering and removal, and avoiding thermal damage to both SCR coating and SCRF<sup>®</sup>-monolith [137]. However, high NO<sub>x</sub> conversion and high soot-filtration efficiency cannot often be accomplished simultaneously. Indeed, as reported by different authors [138–140], a soot-loaded SCRF<sup>®</sup>, operated in the cake filtration regime, showed less NO conversion than a soot-free SCRF<sup>®</sup>.

It is worth noting that the combination of DPF and SCR functionalities into a single device may result in problems, which are absent in separate devices; indeed, the operation of the 2-way DPF/SCR device (Figure 14) might be different from that of conventional DPF and SCR, and the fuel penalty during filter regeneration must be taking into account.



**Figure 14.** Schematic diagram of blended 2-way DPF/selective catalytic reduction (SCR). Reprinted with permission from [141]. Copyright © 2012, American Chemical Society.

Park et al. [141] considered a Cu–zeolite catalyst coated inside the DPF substrate material. The results showed a soot ignition temperature near 300 °C, i.e., considerably lower than thermal soot oxidation so that copper zeolite was indicated as a promising catalytic material not only for SCR reactions but also for soot oxidation. In fact, the soot deposited in the form of deep-bed filtration contacts the catalytic material directly and it starts to oxidize first upon increasing the temperature. On the other hand, the authors founded that the de-NO<sub>x</sub> reactions are influenced by the deposited soot. First, the de-NO<sub>x</sub> reaction is hindered by the deposited soot due to the resistance for mass transfer from the exhaust gas stream to the catalytic sites; second, significant amounts of NO<sub>2</sub> in the soot cake layer are consumed by the soot oxidation reactions. In the absence of soot, the de-NO<sub>x</sub> performance becomes poor when the NO<sub>2</sub>/NO<sub>x</sub> ratio exceeds 0.5. However, when soot is fully loaded in the 2-way device and at temperatures suitable for the NO<sub>2</sub>/NO<sub>x</sub> ratio. It is clear from the obtained results that to completely describe the phenomena occurring in a DPF/SCR system, it is necessary to develop and implement new kinetic models taking into account both soot filtration and NO<sub>x</sub> reduction by SCR.

For this purpose, SCR kinetics (see Table 7) was incorporated in the model describing the transport and reaction phenomena inside a wall-flow type substrate, while for soot filtration and oxidation, mathematical formulations based on a soot cake layer filtration model and both NO<sub>2</sub>- and O<sub>2</sub>-based soot oxidation reactions are incorporated [142,143].

Description	Reaction	Reaction Rate	
NH <sub>3</sub> adsorption	$\rm NH_3 + S \rightarrow \rm NH_3\text{-}S$	$R_1 = A_1 C_{NH3} (1 - \theta)$	
NH <sub>3</sub> desorption	$\rm NH_3\text{-}S \rightarrow \rm NH_3 + S$	$R_2 = A_2 expig(rac{E_{a,2}}{RT}ig) heta$	
NH <sub>3</sub> oxidation	$2 \text{ NH}_3\text{-}\text{S} + 3/2 \text{ O}_2 \rightarrow \text{N}_2 + 3 \text{ H}_2\text{O} + 2 \text{ S}$	$R_3 = A_3 exp \Big( rac{E_{a,3}}{RT} \Big) C_{O_2}  heta$	
		$R_4 = A_4 exp \left(\frac{E_{a,4}}{RT}\right) C_{NO} C_{O_2}^{1/2} - k_{4,b} C_{NO_2}$	
NH <sub>3</sub> oxidation	$NO + 1/2 O_2 \rightarrow NO_2$	where $k_{4,b} = A_4 expigg(rac{ar{k}_{a,4}}{RT}igg) C_{NO}  heta$	
		$K_{eq} = exp\left(-rac{\Delta S}{R} ight)exp\left(-rac{\Delta H}{RT} ight)$	
Standard NH <sub>3</sub> SCR	$4 \text{ NH}_3\text{-}\text{S} + 4 \text{ NO} + \text{O}_2 \rightarrow 4 \text{ N}_2 + 6 \text{ H}_2\text{O} + 4 \text{ S}$	$R_5 = A_5 expig(rac{E_{a,5}}{RT}ig) C_{NO}  heta$	
Rapid NH <sub>3</sub> SCR	$2 \text{ NH}_3\text{-}\text{S} + \text{NO} + \text{NO}_2 \rightarrow 2 \text{ N}_2 + 3 \text{ H}_2\text{O} + 2 \text{ S}$	$R_6 = A_6 expig(rac{E_{a,6}}{RT}ig) C_{NO} C_{NO_2}  heta$	
NH <sub>3</sub> SCR with NO <sub>2</sub>	$4 \text{ NH}_3\text{-}S + 3 \text{ NO}_2 \rightarrow 3.5 \text{ N}_2 + 6 \text{ H}_2\text{O} + 4 \text{ S}$	$R_7 = A_7 expigg(rac{E_{a,7}}{RT}igg) C_{NO_2}  heta$	
N <sub>2</sub> O formation by SCR	$2 \text{ NH}_3\text{-}\text{S} + 2 \text{ NO}_2 \rightarrow \text{N}_2\text{O} + \text{N}_2 + 3 \text{ H}_2\text{O} + 2 \text{ S}$	$R_8 = A_8 expig(rac{E_{a,eta}}{RT}ig) C_{NO_2}  heta$	

Table 7. SCR reaction and reaction rate used by Park et al. [141,143].

Watling et al. [127] reported the development, validation, and application of a one-dimensional model for an SCRF<sup>®</sup>. The model described in that paper was developed by combining kinetics for either a Cu-zeolite or an Fe-zeolite SCR catalyst, originally developed for a flow-through monolith, with a physical model for a coated DPF. It has been demonstrated that this model is capable of predicting NO<sub>x</sub> conversion and NH<sub>3</sub> slip from an SCRF<sup>®</sup> system in a real diesel exhaust. Since the design and control of the SCRF<sup>®</sup> strongly depends on the interaction between NO<sub>x</sub> reduction and soot oxidation reactions taking place in close vicinity, this model has been also applied to investigate the interaction between SCR and DPF functionality. The presence of soot on the SCRF<sup>®</sup> is predicted to have no significant impact on NO<sub>x</sub> conversion, while SCR activity (NO<sub>x</sub> reduction) is predicted to significantly retard the rate of soot removal by oxidation with NO<sub>2</sub>. Indeed, NO<sub>x</sub> reduction by SCR occurs much more rapidly than the soot–NO<sub>2</sub> reaction.

Furthermore, Tronconi et al. [144] studied the interactions between soot and Cu/zeolite powder, with particular attention to the chemistries of NH<sub>3</sub>-SCR and of soot combustion at the lab-scale level. All the experimental tests have been performed over physical mixtures consisting of powdered Cu-SCRF and soot, thus emphasizing the interactions between the two reacting systems. The addition of NH<sub>3</sub> was found to greatly reduce the low-temperature combustion of soot by NO<sub>2</sub>, i.e., the actual oxidizing agent of soot at low temperatures. In fact, given that NO<sub>2</sub> is converted in the NH<sub>3</sub>-SCR reactions (i.e., fast SCR and NO<sub>2</sub> SCR reactions), a substantial decrease in the passive soot regeneration by NO<sub>2</sub> can be expected in SCRF systems. On the other hand, the impact of soot on the SCR reactions (i.e., NH<sub>3</sub> oxidation, standard SCR and fast SCR activities) was marginal. Besides the SCR activity in excess of NO<sub>2</sub> was promoted because of the low-temperature interaction between NO<sub>2</sub> and soot, which leads to a more favorable NO<sub>2</sub>/NO ratio, closer to the optimal 1/1 molar ratio.

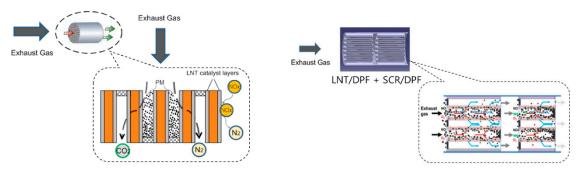
Similar results were found for instance by Schrade et al. [128] and Mihai et al. [145], who observed only a slight decrease of the NO<sub>x</sub> conversion (up to 5%) for the standard-SCR reaction in the presence of engine soot on the SCRF (loaded on an engine test bench). Cavataio et al. [140] also reported a decrease in NO<sub>x</sub> conversion in the standard- and fast-SCR for a soot-loaded SCRF, but on a much larger extent (up to 20%, using model-soot from a soot generator). In both case, these results were explained by blocking of the catalytically active sites.

Along the same lines, Lopez et al. [146] studied vanadia-SCR catalyst coating combined with a wall flow particulate filter. The results confirmed that significant  $NO_x$  and PM reduction could be

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obtained over transient cycles and at steady state conditions, being more than 70% NO<sub>x</sub> conversion over a degreened vanadia-SCR/DPF. Passive filter regeneration was also investigated, obtaining both a good passive regeneration and a good particle number (PN) filtration.

Very recently, Martinovic et al. [147, 148] investigate the integration of soot oxidation and NO<sub>x</sub> SCR by a two-component selective catalytic system and the interaction between them. The SCR catalysts were either Fe- or Cu-ZSM-5, while as the soot oxidation catalyst, CeO<sub>2</sub>-PrO<sub>2</sub> (namely CP) was impregnated with K (called KCP). The authors found that physically mixing the commercial SCR catalyst with the soot oxidation one, it is possible to significantly decrease the onset temperature for soot combustion and simultaneously increase the NO<sub>x</sub> conversion; indeed, NO is oxidized to NO<sub>2</sub>, which participates in the fast SCR reaction. Moreover, in this physical mixture, the soot was oxidized mainly by O<sub>2</sub>, since the contribution of NO<sub>2</sub> was limited because it reacted in the SCR reaction (kinetically much faster). Interestingly, the authors conclude that their results have been obtained at the laboratory scale with the main aim of providing a detailed study on the interaction between a soot oxidation catalyst and a SCR catalyst. Thus, on a real monolith not only chemical interactions, but also fluid-dynamics and pressure drop, catalyst loading and distribution in the monolith, contact between soot and catalyst, contact length with the filtered soot cake represent key parameters to be taken into account. Finally, also coupled LNT-SCR/DPF systems have been also considered given that LNT catalysts can be significantly improved through the addition of a downstream SCR catalyst [149,150]. Moreover, to solve some problems typical of each of these after-treatment units, Kang et al. [125], introduce a LNT/DPF+SCR/DPF hybrid system (Figure 15). In a previous work, Choi and Lee [151] investigated the LNT/CDPF catalyst system for simultaneous removal of NO<sub>x</sub> and soot (Figure 15A), where the LNT/DPF approach closely resembles the DPNR concept. LNT(2Pt20Ba)/CDPF coated with cordierite substrate showed the highest de-NO<sub>x</sub> performance among other LNT catalysts. Moreover, the addition of 5% Co improved both the NO<sub>x</sub> conversion performance and also the PM oxidation rate that resulted higher if compared to bare DPF, which was not coated with the LNT catalyst. Subsequently, Kang et al. [125] introduced a hybrid system of LNT/DPF+SCR/DPF (Figure 15B). The results showed that the  $NO_x$  conversion of the hybrid system was 40% compared to 25% of the LNT/DPF system. Moreover, the PM oxidation activity in the hybrid system was higher than in all the other configurations (i.e., hybrid system > LNT/DPF > bare DPF > SCR/DPF). The conclusions of their work are very interesting, being the de-NO<sub>x</sub> and de-PM activities of the hybrid system superior to that of the single LNT/DPF system. Indeed, NO<sub>2</sub> and NH<sub>3</sub> forming in LNT/DPF under a rich air-to-fuel ratio are used as a reductant for the SCR/DPF catalyst of the hybrid system LNT/DPF+SCR/DPF. In addition, the SCR/DPF increased the  $NO_x$  conversion through HC-SCR [125].



**Figure 15.** Fundamental principles of emission reduction in after-treatment systems: (**A**) lean NO<sub>x</sub> trap (LNT)/DPF system and (**B**) LNT/DPF + SCR/DPF hybrid system. Adapted with permission from [125]. Copyright © 2018, Elsevier, Inc.

#### 5. Conclusions

The need to decrease  $NO_x$  and soot emissions is still critical and it is an unsolved challenge as indicated by the continuous increase of stringent regulation limits for both  $NO_x$  and diesel soot. Indeed, the EURO VI levels are limited to 0.08 g/km for  $NO_x$  and 0.005 g/km for soot. For these reasons, research efforts are strongly focusing on effective techniques to meet the EURO VI standards in general. To do this, the automotive industry has been forced to propose after-treatment solutions often based on several catalytic converters and in addition resulting in rather high-pressure drops, sophisticated control technologies, high cost, considerable weight, and space consumption, as well. Therefore, the simultaneous removal of soot and NO<sub>x</sub> in a single catalyzed device represents a viable solution, also in view of the advantages that can be obtained in terms of both investment costs and pressure drop reduction.

This review focused on papers dealing with catalytic materials proposed and tested in the simultaneous catalytic removal of NO<sub>x</sub> and soot, most of them based on laboratory studies. One of them is represented by hydrotalcite-derived mixed metal oxides. At the laboratory scale, hydrotalcite-based catalysts including potassium and cobalt in their formulations gave promising results in both soot combustion and NO<sub>x</sub> storage in the same temperature range (200–400 °C). With regard to NO<sub>x</sub> storage activity, K-promoted catalysts showed high trapping efficiency due to the formation of several kinds of N-containing surface species. Additionally, perovskite catalysts have been found as efficient catalysts for simultaneous abatement of diesel soot and nitrogen oxides resulting in performance close to those of PGM catalysts. However, in a short time it is more likely the simultaneous use of perovskites and noble metals rather than the total replacement of noble metals by perovskite by exploiting their specific advantages. Moreover, for both two categories of catalysts mainly the ability to store NO<sub>x</sub> is considered and discussed.

Model Pt-based LNT catalysts are able to simultaneously remove NO<sub>x</sub> and soot when operated under cycling conditions, i.e., alternating lean-rich phases according to the DPNR strategy. For these catalytic systems the simultaneous removal is really effective. Indeed, these systems are able to store NO<sub>x</sub> under lean conditions and subsequently reduce them under rich ones even in the presence of soot while soot combustion simultaneously occurs. The soot oxidation is more efficient during the storage phase than during the rich one. Pt promotes soot combustion by catalyzing the oxidation of NO into NO<sub>2</sub> by O<sub>2</sub>, the most effective oxidation agent being NO<sub>2</sub>. Additionally, the nature of the NO<sub>x</sub> storage component (i.e., alkaline or alkaline earth) directly affects soot oxidation. In particular K-based system show higher performances in the soot oxidation if compared to Ba-based because of the high mobility of the active K surface species (i.e., low melting point/volatile compounds, eutectics with other catalyst components, etc.), which can improve the contact between catalyst and soot and consequently the soot oxidation activity. Besides the nature of the Pt-O-alkaline/alkaline earth interaction determines the temperature range in which the combustion process is effective. However, K-based systems are blamed for low thermal stability, which is associated to several technological problems, particularly the loss of active phase with consequent catalysts deactivation.

The current trend is looking for novel Pt-free catalytic formulations aiming at lowering the cost of the DPNR technology. In particular, ceria-based catalysts doped with Ag, Ru, or Cu should be considered as a promising alternative to Pt-based catalysts for the simultaneous removal of soot and  $NO_x$  as a result of higher soot oxidation activity and lower detrimental effect of soot on the amounts of stored  $NO_x$ , as well. However, in most cases their reactivity in the reduction of the stored  $NO_x$  should be further improved towards  $N_2$  selectivity.

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Conflicts of Interest: The authors declare no conflicts of interest.

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