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# Mosquito microevolution drives Plasmodium falciparum dynamics

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#### Author contributions

#### **Competing interests**

The authors declare no competing interests.

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#### Abstract

Malaria, a major cause of child mortality in Africa, is engendered by *Plasmodium* parasites that are transmitted by anopheline mosquitoes. Fitness of *Plasmodium* parasites is closely linked to the ecology and evolution of its anopheline vector. However, whether the genetic structure of vector populations impacts malaria transmission remains unknown. Here, we describe a partitioning of the African malaria vectors into generalists and specialists that evolve along ecological boundaries. We next identify the contribution of mosquito species to *Plasmodium* abundance using Granger causality tests for time-series data collected over two rainy seasons in Mali. We find that mosquito microevolution, defined by changes in the genetic structure of a population over short ecological time-scales, drives *Plasmodium* dynamics in nature, whereas vector abundance, infection prevalence, temperature and rain have low predictive values. Our study demonstrates the power of time-series approaches in vector biology and highlights the importance of focusing local vector control strategies on mosquito species that drive malaria dynamics.

The *Anopheles gambiae* sensu lato (s.l.) species complex inhabits ecosystems ranging from the dry Saharan edge and savanna woodlands, to humid rainforests and saline environments. Population genetics and genomics studies have uncovered signatures of adaptations along ecological boundaries, suggesting an ecotype structure of the morphologically indistinguishable species complex<sup>1–5</sup>. Accumulating evidence indicates the importance of spatiotemporal structuring of vector populations for malaria transmission<sup>6</sup>, but the contribution of vector microevolution remains unknown. This question is particularly important in the light of climate-induced restructuring of vector populations worldwide. Currently used linear correlational approaches to mapping species biogeography<sup>7,8</sup> and differences in resistance to *Plasmodium* are mostly based on analyses of samples collected within a limited window of time or in experimental infections<sup>9,10</sup>. However, such approaches often generate contradictory results as they are not well-suited for studying dynamics of natural complexities<sup>8,11</sup>.

Data-driven approaches such as Granger-causality (G-causality) analysis harness information contained in natural variance and offer powerful tools to decipher complex interactions that cannot be examined in the laboratory. G-causality tests are implemented by time-lagged regression based on the predictive notion of causality whereby a cause precedes and predicts the effect<sup>12</sup>. In contrast to classical epidemiological models, this quantitative approach makes minimal assumptions about the underlying biological mechanisms. Instead, it gathers the unique information contained in the dynamics (time series) of the causal variable to test how well a generated interaction network explains observed responses. Here we exploited G-causality to test whether genetic structure of a mosquito population impacts *Plasmodium* abundance.

We first simplified the genetic characterization of mosquito ecotypes across sub-Saharan Africa. The *Anopheles* species complex is classified into species by inversions of chromosome 2 or microsatellite markers on chromosomes X and 3L<sup>13–15</sup>. The microsatellite marker on chromosome 3L encompasses a highly polymorphic key mosquito immune

factor, the thioester-containing protein 1 (TEP1)<sup>16</sup>. TEP1 is a complement-like protein that mediates mosquito resistance to diverse pathogens, including bacteria<sup>17</sup>, fungi<sup>18</sup> and *Plasmodium<sup>19</sup>*. As *TEP1* polymorphism has been implicated in the adaptive radiation of the species complex<sup>10</sup>, we harnessed its exceptional variability to increase the genotyping resolution beyond species classification (Supplementary Fig. 1). *TEP1* variation can be resolved into four allelic forms (or haplotypes) named \**S1*, \**S2*, \**R1* and \**R2*<sup>10,20</sup>. Initially, \**S1* and \**R1* alleles have been shown to segregate in *Plasmodium*-susceptible and -resistant mosquito strains, respectively<sup>21</sup>. Structural comparison and biochemical studies in vitro of the most divergent R1 and S1 forms pointed to the substitutions within the highly variable thioester domain (TED) as major contributors to the phenotypic divergence between the alleles<sup>21,22</sup> (Fig.1a). To compare the robustness of *TED* as a marker for *TEP1* genotyping, we sequenced full-length *TEP1* and *TED* from laboratory and field mosquitoes and compared their amino acid sequences. Congruent topologies of the cladograms confirmed that *TED* contains all the information necessary for identification of *TEP1* alleles (Fig.1b and Supplementary Tables 1–3).

In a large-scale geographical census, we sampled and genotyped a total of 1,556 A. gambiae s.l. adults and larvae from west to east sub-Saharan Africa (Fig. 2a). The sampling sites spanned Sahelian and savanna ecosystems (Mali, Burkina Faso and Kenya), and rain-forest (Cameroon and Kenya) (Supplementary Table 4). At all sites, A. gambiae sensu stricto (s.s.) and Anopheles coluzzii were the main vector species, except for Kenya, where a temporary decline in A. gambiae s.s. tilted mosquito populations towards Anopheles arabiensis and Anopheles merus<sup>23</sup>. TEP1\*<sup>S</sup> and TEP1\*<sup>R</sup> alleles were identified in all species. To validate our genotyping results, we sequenced a representative set of alleles from each species and country. Marked differences were observed in the geographic and species-specific distribution of individual alleles (Fig. 2b). In addition to the known alleles, we found a previously unknown allele \*R3 which was closely related to \*R2 with four private small nucleotide polymorphisms in the post-β-hairpin loop (Supplementary Fig. 2). Our survey showed broad distribution of TEP1\*S1 and TEP1\*R2 across Africa, whereas other alleles were more geographically restricted. At the molecular level, the two most common alleles either lacked clear geographic structuring (\*S1) or were highly conserved (\*R2) (Fig. 2b,c and Supplementary Fig. 1). By contrast, rare alleles were species- and/or populationspecific. Indeed, \*R1 was only found in A. coluzzii in Mali and Burkina Faso; \*S2 was enriched in A. coluzzii in Cameroon; and \*R3 was restricted to the salt-water-breeding A. merus in coastal Kenya (Fig. 2c). The observed restricted allelic distribution may be caused by non-random mating resulting from local population structure or selection acting on the rare alleles. To detect signatures of non-random mating at a population level, we tested for deviations from Hardy-Weinberg equilibrium. Although significant non-random mating was detected locally for all alleles, no global pattern of inbreeding was found across species or sampling sites, providing further evidence for local selection acting at the TEP1 locus (Supplementary Fig. 3).

As expected from species-specific allele distribution, *TEP1* genotypes were highly structured according to species. \**S1/S1* and \**R2/S1* were enriched in both *A. gambiae* s.s. and *A. arabiensis;* \**R1/R1,* \**S1/S1,* \**S1/S2* and \**R2/S2* were prominent in *A. coluzzii*,and \**R2/S1,* \**R3/R3,* \**R3/S1* and \**S1/S1* were found in *A. merus* (Fig.3a). Overall, the observed

patterns were reminiscent of niche partitioning between generalist and specialist species. In line with our observations, previous studies have demonstrated that A. gambiae s.s. and A. arabiensis occupy a wider variety of ecosystems than the more geographically restricted A. *coluzzii* and *A. merus*<sup>24</sup>, have higher larval competitiveness<sup>25</sup>, faster larval developmental rates<sup>26</sup> and thrive mainly during the rainy season<sup>6</sup>. Unlike specialists, generalist species are believed to be selected for dispersion rather than local competitiveness<sup>27</sup>, which prevents selection for locally adapted phenotypes. We hypothesized that ecological differences within the species complex drive local selection patterns of specialist TEP1 genotypes. To test this hypothesis, we performed principal component analysis of *TEP1* genotype frequencies and observed a clear structuring of A. coluzzii into bioclimate zones according to TEP1 genotype (Fig.3b). \*R1/R1 homozygotes were exclusively identified in dry biomes of the Sahel zone, whereas \*R2/S1 and \*S1/S2 were mostly associated with savanna and rainforest. By contrast, no clear bioclimate structuring was observed in our samples for the generalists A. gambiae s.s. and A. arabiensis. The strongest ecotype partitioning in A. coluzzii was observed by bioclimate for \*R1 and by geography for \*S2 alleles, which have demonstrated roles in *Plasmodium* resistance<sup>20</sup> and male fertility<sup>28</sup>, respectively. Whereas a potential trade-off between these important life-history traits was suggested for heteroallelic populations<sup>28</sup>, our results showed that these alleles occur in allopatric mosquito species. On the basis of these and previous results<sup>3,29</sup>, we propose that the observed biogeographic clines in TEP1 genotype may link TEP1 evolution to local adaptation (Fig.3c): however, other scenarios such as genetic drift or founder effects cannot be ruled out.

Most of the sites examined in this study featured two or more Anopheles species, with the exception of the site 2 in Burkina Faso, which was populated exclusively by Anopheles coluzzii (Fig. 2). How these sympatric vector species contribute to Plasmodium transmission, that is, whether they show similar vectorial capacity, remains unknown. Differences in vectorial capacity between A. gambiae s.s. and A. coluzzii have been examined in laboratory conditions in Mali, Burkina Faso and Cameroon with quite contradictory conclusions<sup>7,8,11</sup>. To address this question in natural settings, we revisited the exceptional Nankilabougou site in Mali, where Plasmodium-resistant A. coluzzii ecotypes (*TEP1*<sup>\*R1</sup>) and susceptible A. gambiae (*TEP1*<sup>S1</sup> and *TEP1*<sup>\*R2</sup>) breed in sympatry (Fig. 2c, site 1). Time-series collections of adults were replicated during two rainy seasons in 2014 and 2015. Mosquito samples were genotyped for species and TEP1 as described above, whereas *P. falciparum* exposure was assayed in whole mosquitoes by *PfCox1* -based PCR (Supplementary Table 2). Whereas the number of infected mosquitoes varied significantly during both rainy seasons, the prevalence of *Plasmodium* in both species was around 10% (Supplementary Fig. 4 and Supplementary Table 7). These results suggested that the two species were equally attracted to humans and had similar exposure to Plasmodium. Species genotyping confirmed the expected TEP1 allelic distribution according to mosquito species (Supplementary Fig. 5). In line with our initial characterization, A. coluzzii was predominantly TEP1 \*R1 and A. gambiae s.s. was represented by TEP1 \*S1 and  $TEP1 * R^2$ . Species abundances followed similar dynamics during the two rainy seasons, with consistently higher numbers of A. coluzzii than A. gambiae s.s. (Fig. 4a). To test whether vector microevolution impacted Plasmodium abundance, we used G-causality, a statistical hypothesis test for determining whether one time series can forecast another. If past (or

lagged) values of a time series  $Y_1$  contain information that helps predict another signal  $Y_2$  (beyond the information contained in past values of  $Y_2$  alone), then  $Y_1$  is said to 'Gcause'  $Y_2$ . Here, we tested whether *Plasmodium* abundance was G-caused by the following mosquito predictors: total mosquito abundance, abundance of A. coluzzii, abundance of A. gambiae s.s. and species frequency. We also gauged whether the environmental factors temperature and rain G-cause Plasmodium or mosquito dynamics. Every time series was modelled by a q-order vector autoregression (VAR(q)) (Methods, equations (3) and (4)) for each season separately, and a Wald test was used to determine which model had a significant contribution (Methods and Supplementary Fig. 6). The best models of both seasons (Supplementary Table 5) were subsequently summarized into one co-integrated G-causality network by Fisher's method (Fig. 4b). Strikingly, species frequency was the only predictor that showed a significant impact on *Plasmodium* abundance in both seasons (adjusted P = 0.01) (Fig. 4b). Reciprocal G-causality tests revealed no impact of the number of *Plasmodium*-infected mosquitoes on mosquito abundance or mosquito species frequencies, indicating that *Plasmodium* exerts low or no selective pressure on mosquito microevolution at this sampling site. Surprisingly, the temperature and rain time series failed to predict *Plasmodium* abundance and species frequencies (Supplementary Table 7), suggesting that linear mathematical approaches and the time-series data collected in our study did not have sufficient power to resolve the impact of these environmental factors on mosquito populations.

As G-causality tests do not define the sign of causal interactions (positive or negative), we explored how an increase in a causal variable affected a cumulative impulse response in the predicted variable. We simulated a theoretical increase of 10% in the frequency of *A. coluzzii* (the change often observed in species frequency during a rainy season) and examined in silico its effect on *Plasmodium* abundance. For both seasons, an increase in *A. coluzzii* frequency had a significant negative impact on *P. falciparum* abundance with a time lag of 21 d, indicating that a marked drop in the number of infected mosquitoes can be expected after one mosquito generation (Fig. 4d). Quantification of the impact of mosquito population structure on *P. falciparum* abundance in natural conditions has revealed significant interspecies differences in driving *Plasmodium* dynamics.

By exploiting the information contained in genetic fluctuations of natural populations, we demonstrated that mosquito microevolution drives *Plasmodium* dynamics. Although further studies of diverse ecosystems and larger datasets are needed for broader generalization of these results, our study contributes important evidence that the structure of vector populations is a crucial variable for prediction of malaria transmission. Nonlinear approaches and longer time series may identify other factors that impact *Plasmodium* abundance. Our genotyping results are consistent with the previously reported extensive genotypic complexity in *Anopheles gambiae* s.l.<sup>1–5</sup>, which is likely to be generated by differential evolutionary forces that promote alternative life-history strategies within the species complex. We found that generalist *TEP1* genotypes occupy diverse environments across Africa, whereas specialized ecotypes inhabit mostly particular niches. Together with the previous reports on geo-distribution of *A. coluzzii TEP1\*R1* (refs. <sup>3,29</sup>), our data suggest that this species thrives in areas with a single short-term annual rainy season and a long dry season, regulated by opposing effects of Harmattan and monsoon in West Africa. Therefore,

we propose that local selection by the environment is a more likely driver of evolution of this vector competence gene than *Plasmodium* exposure. Further studies are needed to formally examine the role of *TEP1* polymorphism on *Plasmodium* dynamics.

In conclusion, this report offers an advanced conceptual framework for studying natural dynamic vector–parasite interactions with potential to predict the consequences of changing environment on vector distribution and to target vector species relevant for malaria transmission.

## Methods

#### Sample collection

**Genotype mapping**—*Anopheles gambiae* s.l. samples (n = 1,538) were collected in a cross-sectional approach in Mali, Burkina Faso, Cameroon and Kenya (Supplementary Table 4) in September and October 2009 and throughout 2012. Adult samples were collected indoors (Kenya, Mali), from mating swarms (Burkina Faso) or in the laboratory from field-collected larvae (Cameroon). Larvae were collected by spoon dipping.

**Time-series collections**—Collection was performed in 2014 (n = 1,091; 28 time points) and replicated in 2015 (n = 4,795; 130 time points) in Nankilabougou, Mali, where *A. gambiae* s.s. (*TEP1*\*<sup>*S1/S1*</sup>, *TEP1*\*<sup>*R2/S1*</sup>, *TEP1*\*<sup>*R2/R2*</sup>) and *A. coluzzii* (TEP1 \*<sup>*R1/R1*</sup>) breed in sympatry. Note that bednet or ACT interventions were not performed at the site during this study. Adults were collected by CDC light traps twice a week (2014) or nightly (2015) throughout the rainy seasons. All samples were processed immediately after collection and stored in 70% ethanol before genotyping. Temperature data were collected using a data logger placed at one of the trapping sites. Rain data were obtained from the Bancumana meteorological station located 6 km from Nankilabougou.

#### Genotyping, sequencing and Plasmodium detection

PCR amplifications were performed on DNA extracted from whole larvae and adult legs (to prevent sperm contamination in mated females) using the DNeasy kit (QIAGEN). All samples were genotyped for species and *TEP1* by previously described PCR-based methods (Supplementary Tables 1 and 2). Genotypes were identified by digestion with BamHI, HindIII and BseNI (Fermentas, according to manufacturer's instructions) according to Supplementary Table 3. All *TEP1\*R1*-like samples were sequenced to confirm *\*R1* identification. For large-scale sequence comparisons of alleles, amplicons of the *TEP1-TED* region (n = 103) were cloned for sequencing and compared to published sequences from the G3, L3-5 and 4Arr laboratory strains. Full-length genomic *TEP1* sequences (n = 25) were obtained by PCR of overlapping fragments generated from homozygous individuals and sequenced.

For the time-series collections, DNA was extracted separately from mosquito legs and carcasses using the Nucleomag VET Kit (Marchery Nagel). The DNA isolated from legs was used for genotyping as described above. The carcass DNA served as a template to identify infected mosquitoes by PCR amplification of the *P. falciparum Cox1* gene (Supplementary Table 2).

#### Data analysis

**Sequence comparison**—Sliding-window analysis of nucleotide diversity was performed by DnaSP<sup>30</sup>. Sequence comparisons were conducted with the MEGA6 Kimura (K80) model<sup>31</sup>, with invariable sites and gamma rate of evolution based on Jmodeltest<sup>32,33</sup> for the optimal nucleotide substitution model.

**Population genetics**—Deviations from Hardy–Weinberg equilibrium (Supplementary Table 4) in *TEP1* genotypes were examined by  $\chi^2$ -test in R using the HWE. chisq function from the genetics package<sup>34</sup>. This function evaluates how well the observed genotype frequencies agree with the expected genotype frequencies predicted by allele frequencies.

The fixation index  $F_{ISik}$  (ref. <sup>35</sup>; Supplementary Fig. 3) for every subpopulation *i*, and every allele *k*, were calculated by allele frequencies *p* as follows:

$$F_{\text{IS}ik} = \frac{P_{ik} - p_{ik}^2}{p_{ik}(1 - p_{ik})} \tag{1}$$

where the difference of observed homozygotes *P* and expected homozygotes  $p^2$ , is divided by the expected proportion of the allele in heterozygotes. The significance of the result can then be tested by  $\chi^2$ -test:

$$\chi^2 = N F_{\text{IS}ik}^2 \tag{2}$$

where N is the sample size.

#### Time-series analysis

We formulated vector autoregression models (VAR) of two variables, where the predicted variable  $y_1(t)$  represents *Plasmodium* abundance and  $y_2(t)$  represents one of the four considered mosquito predictors: total mosquito abundance, abundance of *A. coluzzii*, abundance of *A. gambiae* s.s. or species frequency (that is, a total of four models were built). Additionally, we also tested whether such environmental predictors as temperature and rain impact *Plasmodium* and mosquito dynamics. In the VAR models, the time series of one variable is trained on time-lagged values t - q of both variables, that is, for every time-point t, the value  $y_x(t)$  of one variable is predicted by a historical time-lagged autoregressive term  $y_x(t-q)$ , and historical values  $y_y(t-q)$  of the other variable. The VAR models are given by:

$$y_{1}(t) = c_{1} + \sum_{j=1}^{q} A_{1,1}^{j} y_{1}(t-j) + \sum_{j=1}^{q} A_{1,2}^{j} y_{2}(t-j)$$
(3)

$$y_2(t) = c_2 + \sum_{j=1}^{q} A_{2,1}^j y_1(t-j) + \sum_{j=1}^{q} A_{2,2}^j y_2(t-j)$$
<sup>(4)</sup>

where *q* is the maximum number of lagged observations included in the model (that is, the order), and the matrix *A* contains the time-invariant coefficients (representing the contribution of each time-lagged observation to the predicted value  $y_1(t)$  and  $y_2(t)$ ). The

time-invariant coefficients  $A_{2,1}^{j}$  for both variables and intercept constants *c* were found by linear regression. This time-series analysis was carried out in R<sup>36</sup> using the vars package<sup>37</sup>.

Because the collection frequency of adult mosquitoes differed between 2014 and 2015 (twice a week and daily, respectively), we performed an in silico downsampling of the 2015 dataset. To this end, we summed the values of the time series over non-overlapping time windows of length k = 3-4 d. This approach effectively increased the sample sizes for every time point, thereby increasing measurement accuracy. For every time window k, we repeated this sampling procedure starting at different time points, generating k-1 different time series. Every of these generated datasets was modelled with VAR(q), and the accuracy in predicting both variables was then assessed by calculating the average goodness of fit over all q, that

is,  $\frac{1}{q} \sum_{i=1}^{q} R_i^2$ . The best VAR model was observed for k = 4 and for the datasets starting either

at the earliest or at second time point (Supplementary Table 8). Given that the time series starting at the earliest time point was longer (thus containing more information), it was selected for further modelling of the 2015 season.

The model order q of VAR models have a strong impact on the conclusions that can be drawn from them. To select an appropriate time lag q, we next devised an out-of-sample validation approach, where the models that were trained on the data from one season were used to predict the measurements of another season. The model with the highest out-of-sample accuracy as determined by  $R^2$  was chosen for subsequent G-causality testing (Supplementary Table 9).

#### Granger-causality testing

To test whether mosquito dynamics and environmental factors Granger-cause *Plasmodium* abundance, the dynamics of *P. falciparum* abundance were modelled separately for each sampling season and for each mosquito and environmental variable (Supplementary Fig. 6). To infer Granger causality for both directions of prediction (for example, whether mosquito abundance causes *Plasmodium* abundance or *Plasmodium* causes mosquito abundance), the contribution of the slope coefficients  $y_2(t)$  to equation (3) and  $y_1(t)$  to equation (4), were confirmed by the Wald test:

$$H_{0,1}: A_{1,2}^1 = A_{1,2}^2 = \dots = A_{1,2}^q = 0, H_{1,1}: y_2 \text{Granger} - \text{causes} y_1$$
 (5)

$$H_{0,2}: A_{2,1}^1 = A_{2,1}^2 = \dots = A_{2,1}^q = 0, H_{1,2}: y_1 \text{Granger} - \text{causes} y_2$$
 (6)

Granger non-causality is rejected at low P values of the Wald test and the alternative hypothesis of a causal relationship is accepted. It is important to note that whereas a significant P value implies causality, the absence of significance does not prove the absence of causality.

Granger causality tests require models that are serially uncorrelated and dynamically stable. Therefore, all models where tested for correlated residuals using a portmanteau test using the serial test function of the vars package. To test for stability, the eigenvalues of the model's characteristic polynomial were confirmed to be below 1.

The asymptotic assumption of the Wald test is violated when non-stationary time series are modelled. One approach to make a time series stationary is to difference it. However, to avoid potential bias of the Granger test caused by data transformation, we applied a method developed by Toda and Yamamoto<sup>38</sup> for Granger causality testing in non-stationary time series. The order of integration (*m*) that is required to make the time series stationary was identified by the ndiffs function of the forecast package which implements the Kwiatkowski-Phillips–Schmidt–Shin, augmented Dickey–Fuller and Phillips–Perron tests. Of note, the VAR(*q*+*m*) model was only applied for hypothesis testing, and not for the described model analysis and selection.

To evaluate which interactions can be generalized beyond a single rainy season, a consensus causality network was postulated by *P*-value integration. To this end, *P* values of the Granger causality test for the same hypothesis (for example, species frequency causes *Plasmodium* abundance) but from different seasons were integrated by Fisher's method. This method is commonly used in meta-analysis and yields a  $\chi^2$ -value, with which a cumulative *P* value can be derived. Because the additional replication is appreciated, the resulting *P* value can be lower than the input *P* values. As a result of the out-of-sample validation step, the hypothesis tests cannot be seen as entirely independent replicates of the hypothesis test; therefore, we introduced a false discovery rate correction (*n* = 2). Finally, a Bonferroni correction for multiple hypothesis testing (*n* = 6) with the same dataset was applied.

#### Impulse-response analysis

Having established that mosquito species frequency G-caused *Plasmodium* abundance, we tested for the sign of this interaction using a cumulative impulse response. Generally, an impulse response is the reaction of the dynamic system to an external change, which describes the reaction of the system as a function of time. To estimate the response of *Plasmodium* abundance to a theoretical increase in *A. coluzzii* frequency, cumulative impulse-response functions were generated with the irf function of the vars package with 1,000 bootstrap replicates at a 95% confidence interval. For non-zero impulse-response functions, the sign of the tested interaction is the direction of the impulse response. For example, in a negative interaction, an increase in the predictor will cause a decline in the predicted variable.

#### **Reporting Summary**

Further information on research design is available in the Nature Research Reporting Summary linked to this article.

## **Supplementary Material**

Refer to Web version on PubMed Central for supplementary material.

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# Data Availability

The full-length *TEP1* and *TEP1–TED* sequences are available at the NCBI GenBank under accession numbers MF098568 to MF098592 and MF035727 to MF035924, respectively. The data that support the findings of this study are available from the corresponding author upon request.

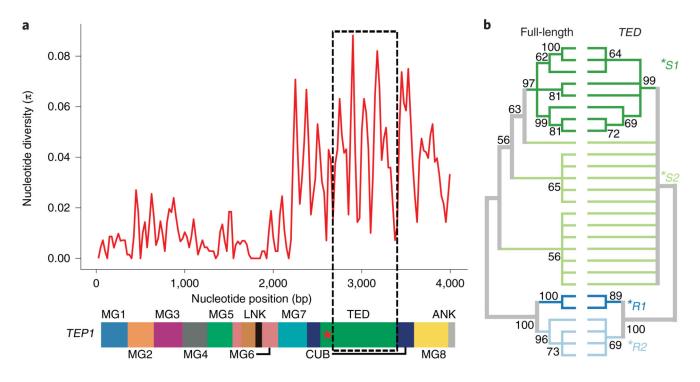
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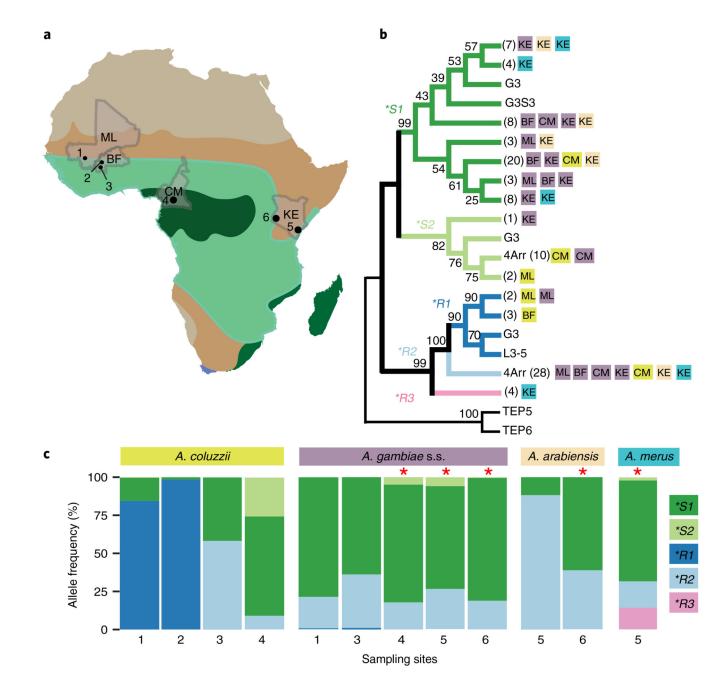
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## Fig. 1. TEP1 diversity and classification.

**a**, Nucleotide diversity over the *TEP1* coding sequence plotted above a schematic representation of the protein domain organization. The dashed box indicates the region chosen for genotyping. **b**, Congruency of *TEP1* allelic segregation based on neighbourjoining comparison of full-length (left) versus *TED* only (right) amino acid sequences, illustrated by bootstrapped (1,000 iterations) phylogenetic trees (50% cut-off). MG, macroglobulin domain; LNK, linker region; CUB, complement C1r/C1s; ANK, 2-disulfide anchor motif.

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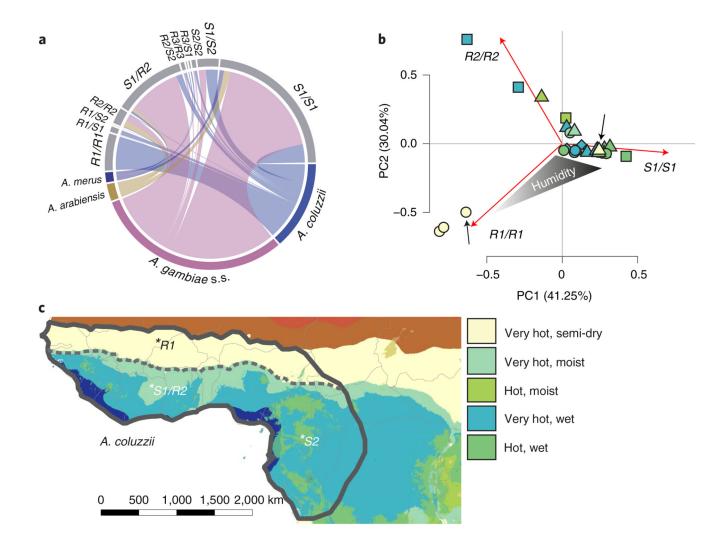


#### Fig. 2. TEP1 population genetics.

**a**, Location of sampling sites across African bioclimate zones (Supplementary Table 4) indicated by colour. Collected *Anopheles* species and sample sizes for each site are shown in brackets. Mali (ML): (1) (*A. coluzzii, n* = 116; *A. gambiae* s.s., *n* = 150); Burkina Faso (BF): (2) (*A. coluzzii, n* = 100) and (3) (*A. coluzzii, n* = 12; *A. gambiae* s.s., *n* = 47); Cameroon (CM): (4) (*A. coluzzii, n* = 210; *A. gambiae* s.s., *n* = 643) Kenya (KE): (5) (*A. gambiae* s.s., *n* = 17; *A. merus, n* = 49; *A. arabiensis, n* = 17) and (6) (*A. gambiae* s.s., *n* = 122, *A. arabiensis, n* = 67). **b**, Neighbour-joining comparison of TED amino acid sequences (*n* = 103) identified in the sampling sites by country (as in **a**) and by species (colour as in **c**),

and compared to the laboratory strains G3, 4Arr and L3-5. Corresponding TED sequences of the closely related TEP5 and TEP6 (*A. gambiae* pest strain) were used as an outgroup for TEP1. Bootstrap values (1,000 iterations) are shown at the nodes of the plotted phylogenetic tree. Numbers of samples with identical sequences are shown in brackets. **c**, Distribution of *TEP1* allelic frequencies per species across sampling sites 1–6. Significant deviations from Hardy–Weinberg equilibrium (calculated by  $\chi^2$ -test and indicated by the red asterisks) were observed in *A. gambiae* s.s. from sites 4, 5 and 6 (*P*=3.0×10<sup>-4</sup>, *P*=5.0×10<sup>-4</sup>, *P*=1.1×10<sup>-5</sup>, respectively); *A. arabiensis* from site 6 (*P*=1.5×10<sup>-5</sup>); and *A. merus* from site 5 (*P*=9.110<sup>-12</sup>). For details see Supplementary Table 4.

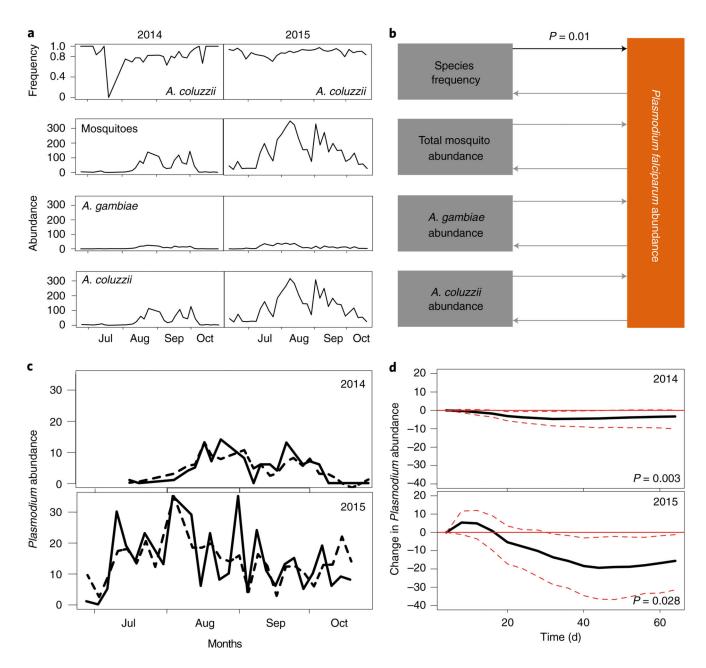
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#### Fig. 3. Genetic and ecotype structures of Anopheles gambiae s.l.

**a**, *TEP1* genotype composition of the species (sample sizes are indicated by the length of the species segment). *A. gambiae* s.s. (magenta) are predominantly represented by \**S1/S1* and \**R2/S1; A. coluzzii* feature \**R1/R1,* \**S1/S1* and \**S1/S2.* **b**, Principal component analysis of genotype frequencies in species and sampling locations (n = 15). Species are coded by shapes (*A. gambiae* s.s.; triangle; *A. coluzzii*, circle; *A. arabiensis*, square; *A. merus*, diamond), colours represent bioclimate zones<sup>39</sup>(as in colour legend, Supplementary Table 4). Arrows indicate the Mali sampling site where *A. gambiae* s.s. and *A. coluzzii* recotypes (delineated by grey lines) according to *TEP1* genotypes over bioclimate zones<sup>39</sup> (as in **b**). *A. coluzzii* ecotype distribution was deduced from this and previous studies<sup>3,29</sup>.

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#### Fig. 4. Granger causality analysis.

**a**, Time-series collections for rainy seasons in 2014 and 2015. *A. coluzzii* frequency, abundance of total mosquitoes, *A. gambiae* and *A. coluzzii*, and *Plasmodium* abundance (**c**, solid lines) were used to generate VAR models for each season separately. The 2014 season was sampled every 3-4 d (n = 1,091 mosquitoes; 28 time points), whereas the 2015 season was modelled as a cumulative summary of 4 d (n = 4,795 mosquitoes; 33 time points). **b**, Schematic representation of G-causality predictions for both seasons, arrows indicating the direction of prediction (non-significant G-causality in grey, significant in black). *P* values were computed using Wald-test statistics for every season, and subsequently integrated by Fisher's method and Bonferroni correction, generating a cumulative *P* value for both seasons

(black). **c**, Comparison of the observed (solid line) *Plasmodium* abundance in the time-series data and the predicted (dashed line) *Plasmodium* abundance by VAR model based on species frequency for two seasons. **d**, Theoretical cumulative impulse-response test. The sign of significant Granger causal interactions was determined for each season separately by estimating in silico the response to an impulse in one of the variables. A simulated 10% increase in *A. coluzzii* frequency in the mosquito population caused a significant drop in *Plasmodium* abundance in both models with a time lag of 21 d. The solid black line depicts the mean out of 1,000 bootstrap replicates, and the red dashed lines show the 95% confidence bands. Significant G-causality for every season (computed using Wald-test statistics) is given by the individual *P* values. For a detailed summary of the analysis, see Supplementary Table 4.