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Isoprostanes in wastewater as biomarkers of oxidative stress during **COVID-19** pandemic



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HIGHLIGHTS

G R A P H I C A L A B S T R A C T

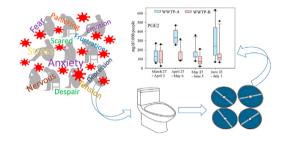
- Four oxidative stress hormones were determined in wastewater
- PGE2 and 5-iPF2α-VI can be a reliable biomarker of community oxidative anxiety.
- Average mass load of isoprostanes ranged from 22.9 to 807 mg/d/1000 people.
- Mass load of isoprostanes increased from the first to second month but decreased in the third month of COVID-pandemic.

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ABSTRACT

Isoprostanes are the potential biomarkers of endogenous human metabolism and proven clinically to provide the quantitative measure of systematic oxidative injury. An ultra-performance liquid chromatography-tandem mass spectrometric analytical method capable of determining four biomarkers of oxidative stress (8-iso-PGF2a, 2,3-dinor-iPF2a-III, PGE2, and 5-iPF2a-VI) in wastewater was developed and validated. Isoprostanes were quantified in the range of 31.1-1270 ng/L in raw wastewater samples in two communities in western Kentucky and Tennessee during the first four months of the COVID-19 pandemic. Consistent detection of PGE2 and 5-iPF2a-VI in wastewater suggested that PGE2 and 5iPF2α-VI can be a reliable biomarker of community oxidative anxiety. The higher 4-month average mass load of isoprostanes [(ranged from 22.9 mg/d/1000 people to 807 mg/d/1000 people] may be attributed to the elevated community level oxidative anxiety owing COVID-19 uncertainties. The average mass loads of PGE2 and 5-iPF2 α -VI in a community were significantly increased (two-tailed p < 0.001) from the first month of COVID-19 pandemic to the second month; however, significantly decreased (twotailed p < 0.001) in the third month. Wastewater-based-epidemiological determination of isoprostanes can be a near-real-time and cost-effective approach of a trend in community depression. This is the first report of the quantification of PGE2 and 5-iPF2α-VI in wastewater and estimation of the community level oxidative anxiety.

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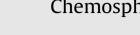
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Mr. Isaac Bowers: collected wastewater samples, performed sample preparation, instrumental analysis, analytical data interpretation, and prepared the first draft of the manuscript. Dr. Bikram Subedi: conceptualized the project, collected wastewater samples, performed sample preparation, instrumental analysis, analytical







data interpretation, prepared and reviewed the manuscript, and supervised overall research project. Both authors reviewed and edited the paper.

1. Introduction

The World Health Organization's global survey among 130 countries reported the devastating impact of the novel coronavirus disease (COVID-19) including bereavement, isolation, loss of income, and fear that trigerred existing mental health conditions and exacerbating the existing ones (WHO, 2020). The psychological impact of COVID-19 has been reported in Italy (Rossi et al., 2020), Spain (Parrado-Gonzalez and Leon-Jariego, 2020), China (Talevi et al., 2020; Zhang and Ma, 2020), and the U.S. (Fitzpatrick et al., 2020). Typically, the elevated level of reactive oxygen species (ROS) during psychological stress catalyzes the oxidation of arachidonic acid and releases isoprostanes, such as the F2isoprostanes. The ROS catalyzed oxidation of cellular constituent can eventually result in oxidative damage to DNA (Morrow and Roberts, 1991; Sivonova et al., 2004; Erie et al., 2003). Rats under stress (electrical shock, forced swimming, water immersion, noise, etc.) secrete higher levels of renal oxidative stress inducers (NO and O₂), a marker (malondialdehyde) that increased oxidative DNA damage (measured by DNA fragmentation in the gastral mucosa in rats) (Bagchi et al., 1999; Gidron et al., 2006; Pedreanez et al., 2011). Individuals' habitual stress (smoking) (Yan et al., 2007), diseases (cardiovascular diseases, select cancers, diabetes) (Barocas et al., 2011; Zhang, 2013; Murai et al., 2000) and depression (high levels of psychological stress) (Blak et al., 2015) were also correlated with elevated levels of isoprostanes. Interestingly, the situational depression such as exam anxiety for students resulted in a significantly higher sensitivity to the lipid peroxidation and DNA damage (Sivonova et al., 2004).

Raw wastewater can be utilized as a pooled urine sample to comprehend the community consumption or exposure to chemicals of concern and the associated community health (Daughton, 2012). There are reports of several F2-Isoprostanes excretion through human urine including 8-iso-prostaglandin F2a (8-iso-PGF2a; 0.25 \pm 0.15 μ g/g of creatinine), 2,3-dinor-8-iso-prostaglandin F₂ α (2,3-dinor-iPF2 α -III, 4.6 \pm 1.6 μ g/g of creatinine), 5-iso Prostaglandin $F_{2\alpha}$ -VI (5-iPF2 α -VI, 7.371 ± 0.021 µg/g of creatinine), and prostaglandin E_2 (PGE2, 178 ± 80 ng/day) (Liang et al., 2003; Practico et al., 1998; Cibbattoni et al., 1979). Excreted isoprostanes are typically flushed down the drain and reach the centralized wastewater treatment plants (WWTP). Three isoprostanes including 8-iso-PGF2a, 2,3-dinor-iPF2a-III, and PGE2 were found to be sufficiently stable under typical sewer conditions with half-lives of 130 h, 110 h, and 110 h, respectively (O'Brien et al., 2019). Moreover, 8-iso-PGF2 α showed no degradation in raw wastewater even at the ambient temperature for 24 h (Ryu et al., 2015). Two studies in Europe reported 8-iso-PGF2a in wastewater (8.7–23.3 ng/L) collected from 12 European cities in typical weeks which was equivalent to a daily mass load of 4.8-14.3 mg/day/1000 people to the WWTPs (Ryu et al., 2016; Ryu et al., 2019). In the U.S. only one study reported 8-iso-PGF2 α in wastewater (~25 ng/L) in a typical week which was equivalent to ~16 mg/day/1000 people (Santos et al., 2015). To the authors' knowledge, there are no reports of 2,3-dinor-iPF2α-III, PGE2, and 5-iPF2α-VI in wastewater.

In this study, an analytical method capable of determining four isoprostanes - 8-iso-PGF2 α , 2,3-dinor-iPF2 α -III (a major metabolite of 8-iso-PGF2 α), PGE2, and 5-iPF2 α -VI (Fig. 1) – in wastewater was developed and validated. Target isoprostanes were monitored in one community in western Kentucky and one community in western Tennessee for 10 consecutive days in typical weeks every month starting from March 27th to July 1st, 2020. The sampling

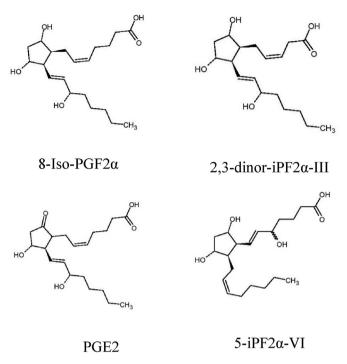


Fig. 1. Structure of target isoprostanes. 8-iso-PGF2 α : 8-iso-prostaglandin F2 α ; 2,3-dinor-iPF2 α -III: 2,3-dinor-8-iso-prostaglandin F₂ α ; PGE2: prostaglandin E₂; 5-iPF2 α -VI (5-iso Prostaglandin F_{2 α}-VI.

period covers the first four months of COVID-19 outbreak in the U.S.; therefore, the quantitation of isoprostanes potentially portrays the community level oxidative anxiety. The cumulative community anxiety after the unprecedented COVID-19 closure of schools, public gatherings, and non-life-sustaining business in March, fears of civil unrest for prolonged period, and the conditional reopening of select businesses and public gatherings in May potentially vary the community-level stress, and hence the biomarkers of oxidative stress. To the authors' knowledge, this is the first study that is reporting quantitative measurement of community-level stress during the COVID-19 outbreak.

2. Materials and methods

2.1. Chemicals and reagents

Standard stock solutions (10 or 100 ppm) of individual isoprostane and their respective deuterated internal standards were purchased at the highest available purity from Cayman Chemical Company (Ann Harbor, Michigan). Target isoprostanes include 8-iso-prostaglandin F2 α (8-iso-PGF2 α), 2,3-dinor-8-iso-prostaglandin F2 α (2,3-dinor-iPF2 α -III); a major metabolite of 8-iso-PGF2 α), prostaglandin E₂ (PGE2), and 5-iPF2 α -VI and the corresponding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding internal standards include 8-iso-PGF2 α -VI and the correshonding interna

2.2. Sample collection and preparation

Two WWTPs from two communities (A and B), one in western Kentucky and one in western Tennessee, were sampled for ten consecutive days every month starting March 27 to July 1st, 2020. Twenty-four-hour composite samples of raw wastewater (one aliquot every 15 min) were collected using a time-proportional autosampler and maintained at 4 °C during the collection period. WWTP-A treats an average of 4.40 million gallons per day (MGD) of primarily domestic sewage whereas WWTP-B treats an average of 1.45 MGD of sewage. All samples collected in polypropylene bottles were transported on ice to the laboratory and stored at -20 °C until further analysis.

The ammoniacal nitrogen (NH₄–N) level in raw wastewater was determined as reported elsewhere (O'Rourke and Subedi, 2020; Croft et al., 2020), and used to estimate the size of the population served by the WWTPs. The concentration of ammoniacal nitrogen (ng/L), wastewater inflow (L/d), and the average per capita daily production of NH₄-N (6900 mg/d) (Been et al., 2014) were utilized to determine near-accurate population. The average populations found were 16187 \pm 5311 for community A and 10941 \pm 3297 for community B. The average population for community A and B determined using ammoniacal nitrogen load were 16.3% lower and 3% higher, respectively, than the census-based population (USCB, 2019). The limited non-human sources of ammoniacal nitrogen could not be quantified and corrected; however, it is least affected by non-human sources compared to other conventional hydrological markers used for the estimation of population (Been et al., 2014).

The samples were equilibrated to room temperature, thoroughly mixed, centrifuged (100 mL) at 4500 rpm for 5 min, and vacuum filtered using 0.45 μ m nylon membrane (47 mm diameter). The filtered samples were spiked with 500 μ L of β -glucuronidase type H-2 enzyme, mixed well, and incubated at 37 °C for 2 h. The incubated samples were then allowed to cool to room temperature, spiked with a mixture of internal standards [250 ng of 8-iso-PGF2a- D_4 , 187.5 ng of 5-iPF2 α -VI- D_{11} , 250 ng of PGE2- D_4], and mixed thoroughly. The spiked samples were extracted using Oasis® hydrophilic-lipophilic balance (HLB) solid-phase extraction cartridges (6 c.c., 200 mg) preconditioned with 3 mL methanol followed by 3 mL of ultrapure water. Samples were extracted under a gentle vacuum at ~1 mL/min, dried under a high vacuum for ~3 min, and eluted with 4 mL of methanol followed by the 3 mL of 5% ammonia in methanol. The combined eluates were concentrated to ~100 μ L using a gentle flow of nitrogen gas at ambient temperature, transferred quantitatively (with multiple wash using methanol) to silanized amber glass LC-vials, adjusted final volume to 500 μ L using methanol, and vortex-mixed. The final concentrate was then syringe filtered through 0.2 μ m nylon filter (MilliporeSigma, Burlington, MA), and injected 1.0 µL for the analysis of target isoprostanes using ultra-performance liquid chromatography (UPLC) tandem mass spectrometry (MS/MS).

2.3. Isotope dilution liquid chromatography-tandem mass spectrometry

The prepared samples were analyzed for target isoprostanes

Table 1

using UPLC (Agilent 1290 Infinity II LC System) coupled with Triple Quadrupole mass spectrometer (Agilent 6460) (Santa Clara, CA). The gradient flow of HPLC grade methanol and 0.1% aqueous solution of formic acid (Table S1) were used to separate analytes using an ACQUITY UPLC® Oligonucleotide BEH C18 column (150 mm × 2.1 mm i.d. × 1.7 µm particle size). Target analyte's peak identification was based on the relative retention time (\pm 0.05 min) to their standard solution, two multiple reaction monitoring (MRM) transitions (Table 1) in the negative mode of ionization, and the ratio of the abundance of quantitative to qualitative ions (\pm 20%). The gas temperature, sheath gas temperature, and capillary voltage were optimized to 350 °C, 300 °C, and 4000 V, respectively.

The isotopic dilution mass spectrometry method was applied where a known quantity of deuterated isotopes of each target analyte (internal standard) is spiked directly into the sample prior to extraction and analytes were quantified based on the relative response factors of isotopic-labeled internal standard and the corresponding analyte. This method allows for accurate quantification by correcting for the loss of analytes during the sample preparation and instrumental analysis process. However, 2,3-dinor-iPF2α-III was quantified using the internal standard for PGE2, due to lack of commercial availability of its deuterated standard. The seven-toten-point calibration curves (10.0 ng/L to 10,000 ng/L) of each target analyte were prepared by plotting the concentrationdependent response factor against the response-dependent concentration factor. The linear or quadratic (PGE2) regression coefficients (r^2) determined using Agilent MassHunter Workstation for the Quantitative Analysis were 0.999 for all analytes.

2.4. Optimization of simultaneous extraction

Raw wastewater samples (n = 3) were centrifuged and filtered as described in section 2.2., spiked with a mixture of target analytes (125 ng each), internal standards, and extracted as described above. One wastewater sample (not spiked with analytes) was included to determine the background concentration of target isoprostanes in wastewater and one ultrapure water sample was included to evaluate the procedural contamination. Moreover, one wastewater sample spiked at 125 ng of each target analyte was extracted using Oasis® MCX mix-mode cation exchange (6 cc, 150 mg) cartridge. All extracts were concentrated and adjusted the final volume to 500 μ L using methanol and analyzed.

2.5. Enzymatic hydrolysis

To evaluate the effectiveness of enzyme hydrolysis (deconjugation) of the potential glucuronidated forms of isoprostanes in wastewater, 500 μ L of β -glucuronidase type H-2 enzyme was added in each of three 100 mL raw wastewater samples and incubated at 37 °C for 1.5 h prior extraction. In addition, three raw wastewater samples not spiked with enzymes were also incubated at 37 °C for 1.5 h (to evaluate the effect of incubation) and three other raw

Target isoprostanes	MRM transitions (collision energy in eV)	Average absolute % recoveries $(n = 3)$	1	CCV % recoveries $(n = 14)$	^a Intraday repeatability (n = 10) at 25 ng level	^a Intraday repeatability (n = 10) at 125 ng level	LOD (ng/L)	LOQ (ng/L)
2,3-dinor- iPF2α-III	325.1 > 237, 136.9 (15 eV)	64 ± 12	128 ± 9.4	106 ± 16	±20.0	±3.8	25.6	85.5
8-iso-PGF2α	353.2 > 193.2, 164.9 (28 eV)	99 ± 30	71.9 ± 1.4	106 ± 4.4	±3.3	±1.4	9.30	31.1
5-iPF2α-VI PGE2	353 > 114.9, 299.2 (30 eV) 352.1 > 271.8, 333.9 (12 eV)	_	110 ± 2.4 117 ± 1.1	116 ± 4.1 107 ± 10	±1.1 ±2.0	$_{\pm 4.8}$ $_{\pm 1.7}$	8.81 13.5	29.5 45.0

^a Standard errors; CCV: continuing calibration verification; LOD: limit of detection; LOQ: limit of quantitation.

wastewater samples not spiked with enzymes were placed at ambient lab conditions (to evaluate the effect of enzyme hydrolysis and incubation). After incubation, all samples were allowed to equilibrate to room temperature, spiked with the internal standard mixture, mixed thoroughly, extracted, and analyzed as described in section 2.2.

3. Results and discussion

3.1. Figure of merit

In a triplicate spiking and recovery experiment (section 2.4), the average absolute recoveries of 8-iso-PGF2 α , 2,3-dinor-iPF2 α -III, PGE2, and 5-iPF2 α -VI were 99 ± 30.4%, 64 ± 11.9%, 58 ± 5.7%, and 126 ± 6.2%, respectively (Table 1). However, MCX cartridge extraction of isoprostanes resulted in the extraction recoveries range from 71.2% (PGE2) to 169% (8-iso-PGF2 α). Due to the unacceptable recovery of 8-iso-PGF2 α , HLB cartridges were considered for further extraction of isoprostanes from raw wastewater.

The intraday repeatability of wastewater extracts spiked at 25 ng and 125 ng levels were determined by analyzing ten consecutive runs. The intraday precision ranged from ± 1.1 (PGE2) to ± 20.0 (8-iso-PGF2 α) at 25 ng level and from ± 1.4 (2,3-dinor-iPF2 α -III) to ± 4.8 (5-iPF2 α -VI at 125 ng level (Table 1).

The limit of detection (LOD) and limit of quantitation (LOQ) were calculated in wastewater using the minimum target isoprostane concentration that results in a signal to noise ratio of 3 and 10, respectively. LOD ranged from 8.8 ng/L (5-iPF2 α -VI) to 25.6 ng/L (2,3-dinor-iPF2 α -III) and LOQ ranged from 29.5 ng/L (5-iPF2 α -VI) to 85.5 ng/L (2,3-dinor-iPF2 α -III) (Table 1). Data points that fell below LOQ were adjusted by substituting them with their respective $\frac{1}{2}$ LOQ values when 70% of the data from that sampling period were above their LOQ.

As part of quality assurance and quality control, a randomly selected sample was spiked (matrix spike, n = 2) with 125 ng of each target isoprostanes and their corresponding internal standards. The average relative recoveries of isoprostanes ranged from 71.9 \pm 1.0% (8-iso-PGF2 α) to 128 \pm 9% (2,3-dinor-iPF2 α -III) (Table 1). A calibration standard mixture was considered as a continuing calibration verification (CCV) solution and analyzed before and after each set of ten samples. The average percent recoveries for all target isoprostanes were within \pm 20% (Table 1).

3.2. Effect of enzyme hydrolysis

PGE2 and 5-iPF2 α -VI were not detected in wastewater samples that were not enzymatically hydrolyzed (and not incubated); however, quantified at the average concentrations of 378 ± 61.2 ng/L and 423 ± 15.1 ng/L, respectively, in enzymatically hydrolyzed samples (Table S2). The average concentration of 8-iso-PGF2 α was increased by 71% (444 ± 110 ng/L to 760 ± 98.1 ng/L) after enzymatic hydrolysis. The higher concentration of isoprostanes after enzymatic hydrolysis was potentially due to the deconjugation of

glucuronidated forms of isoprostanes. Ryu et al. (2016) also reported a 20-fold increase in 8-iso-PGF2 α level in wastewater after enzyme hydrolysis. The glucuronide conjugated forms of 8-isoprostane in human urine samples can vary between 30 and 80% of the total iso-prostane (Holder et al., 2020; Yan et al., 2010). However, 2,3-dinor-iPF2 α -III was not detected even after enzyme hydrolysis. On incubation (without adding enzyme), the average concentration of 8-iso-PGF2 α was 544 ± 28.1 ng/L compared to 444 ± 110 ng/L at ambient conditions. It is important to note that enzyme hydrolysis was performed only at 37 °C for 1.5 h and the optimum deconjugation of the potential glucuronidated forms of isoprostanes may require to further optimize the incubation time.

3.3. Occurrence of isoprostanes in raw wastewater

Sivonova et al. (2004) reported the significantly increased level of oxidative damage to DNA and sensitivity to lipid peroxidation during daily life stress such as examinations for university students in Slovakia. Oxidative anxiety has also been related to tumor, cardiovascular, neurogenerative, and pulmonary sickness. Severe life stress resulted from several traumatic events such as divorce, child abuse, wars, death of loved ones, loss of a job, etc. induce psychotic symptoms including depression, schizophrenia, and bipolar disorders (Schiavone et al., 2013).

In this study, 8-iso-PGF2a, 5-iPF2a-VI, and PGE2 were found at an average from 21.4 to 1270 ng/L in raw wastewater samples from both communities (Table 2). The average level of 8-iso-PGF2 α quantified from community A in this study was ~50 folds higher than that reported by Ryu et al. (2016) in European cities and ~32 folds higher than that reported in Detroit, Michigan (Santos et al., 2015) in typical days. The population served by the WWTP and wastewater inflow can impact the final concentrations of isoprostanes in wastewater (section 3.4); however, wastewater samples in this study were collected during the COVID-19 pandemic that caused many disruptions in people's lives and the economy and leading to the adverse economic effects of the social-health crisis (Parrado-Gonzalez and Leon-Jariego, 2020). The quarterly GDP growth in the U.S. plunged by 4.8%, the largest decrease since the economic recession in 2008 (Casselman, 2020). However, it is also important to note that the level of isoprostanes could not be compared to that prior COVID-19 due to the unavailability of prior samples.

The peroxidation of arachidonic acid in membrane phospholipids forms 5-iPF2 α -VI (5- and 15-series regioisomers) at a significantly higher level than 8-iso-PGF2 α (8- and 12-series regioisomers) (Milne et al., 2011). In community B, 8-iso-PGF2 α was detected only in 15% of samples at an average concentration of 38.7 ng/L; however, 5-iPF2 α -VI was quantified in 93% of the samples at an average concentration of 151 ng/L. The arachidonyl endoperoxide intermediate can isomerize to PGE2 and PGD2. In this study, PGE2 was detected most consistently at average 221 ng/L and 240 ng/L level in communities A and B, respectively. Therefore, determining PGE2 and 5-iPF2 α -VI in raw wastewater can be a

Table 2

Concentration of isoprostanes (ng/L) in raw wastewater. Detection frequencies are provided in parentheses.

	Average concentration $(n = 10)$ at WWTP-A				Average concentration $(n = 10)$ at WWTP-B			
	8-iso-PGF2a	5-iPF2a-VI	PGE2	2,3-dinor- iPF2α-III	8-iso-PGF2a	5-iPF2α-VI	PGE2	2,3-dinor- iPF2α-III
Mar 27-Apr 5 April 27-May 6 May 27- Jun 5 Jun 22-Jul 1	$404 \pm 129 (90\%)$ $734 \pm 114 (100\%)$ $1270 \pm 356 (100\%)$ 815 + 226 (90%)	$147 \pm 49.8 (100\%) 253 \pm 57.6 (100\%) 197 \pm 79.7 (100\%) 251 + 65.8 (100\%) $	$144 \pm 75.1 (100\%) 299 \pm 71.1 (100\%) 201 \pm 101 (100\%) 231 + 122 (100\%) 231 + 120 (100\%) 231 + 12$	nd nd nd nd	$\begin{array}{c} 21.4 \pm 11.0 \; (30\%) \\ 52.4 \pm 10.6 \; (20\%) \\ 31.1 \; (10\%) \\ nd \end{array}$	$110 \pm 57.6 (100\%) 156 \pm 86.0 (100\%) 166 \pm 143 (70\%) 162 \pm 86.8 (100\%) $	$206 \pm 124 (100\%) 252 \pm 74.2 (100\%) 194 \pm 91.0 (70\%) 339 + 130 (100\%)$	nd nd nd nd

nd: non detected.

reliable biomarker of community oxidative anxiety. Future studies are suggested to validate the potential of PGE2 and 5-iPF2 α -VI in raw wastewater collected from diverse communities on typical days. To our knowledge, this is the first report of the quantification of 5-iPF2 α -VI and PGE2 in wastewater. The target metabolite of 8-iso-PGF2 α (2,3-dinor-iPF2 α -III) was not detected in any wastewater samples.

3.4. Mass load of isoprostanes to the WWTPs

The 4-month average mass load of isoprostanes in both communities ranged from 10.1 mg/d/1000 people (8-iso-PGF2 α) to 807 mg/d/1000 people (8-iso-PGF2 α) (Fig. 2). The average mass load of 8-iso-PGF2 α is > 1.4 fold and >50 fold higher in communities A and B, respectively than that reported in typical days in European cities (Ryu et al., 2016; Ryu et al., 2019) and Detroit, Michigan (Santos et al., 2015). Sims et al. (2019) also detected 8-iso-PGF2 β , an isomer of 8-iso-PGF2 α , in raw wastewater collected from a WWTP serving a community (~900,000 people) in southwest of England (Sims et al., 2019). Interestingly, the average mass loads of PGE2 and 5-iPF2α-VI in community A were significantly increased (two-tailed p < 0.001) from the first month of COVID-19 pandemic to the second month; however, significantly decreased (two-tailed p < 0.001) in the third month (Fig. 2). In Kentucky and Tennessee, the state announced executive orders to close schools, public gatherings, and non-life-sustaining business in March after COVID-19 state of emergency declared in the first week of March. In the second week of May, the state announced the conditional reopening of select businesses and public gatherings. The significant drop in the mass load of PGE2 and 5-iPF2a-VI in the third month (wastewater sampled from May 27 to June 5) can have resulted from the partial reopening of businesses. Despite the partial reopening of businesses, surge in COVID-19 cases in late June in the U.S. including Kentucky and Tennessee can have significantly contributed to the elevated level of isoprostanes again, and hence community anxiety.

Moreover, ~25% of college students in China reported mild to severe anxiety and depression associated with the COVID-19 pandemic (Cao et al., 2020). College students comprise a significant proportion (~45%) of both of these communities. The combined COVID-19 anxiety and examination stress in April–May could have contributed to the high level of isoprostanes in wastewater.

4. Implications and summary

Overall, an analytical method capable of determining four biomarkers of oxidative stress (8-iso-PGF2a, 2,3-dinor-iPF2a-III, PGE2, and 5-iPF2 α -VI) in wastewater was developed and validated. The enzyme hydrolysis of wastewater significantly increased the free isoprostanes in wastewater. The higher level of isoprostanes determined in wastewater in this study can have resulted from the potential oxidative anxiety induced by the COVID-19 uncertainties. For the first time, PGE2 and 5-iPF2α-VI were quantified consistently in both communities. We conclude that the determination of PGE2 and in raw wastewater may provide a more reliable biomarker of community oxidative anxiety than 8-iso-PGF2a. The higher mass load of isoprostanes in this study, compared to the reported values, may be attributed to the community level oxidative anxiety owing to COVID-19 uncertainties. It is important to note that the stabilities of target isoprostanes in wastewater were not evaluated in this study as previous studies (Ryu et al., 2015; O'Brien et al., 2019) reported the prominent stability of isoprostanes even at room temperature; however, the stability of target isoprostanes in wastewater would minimize the uncertainties on mass loads of target isoprostanes into the WWTPs. Further research is warranted

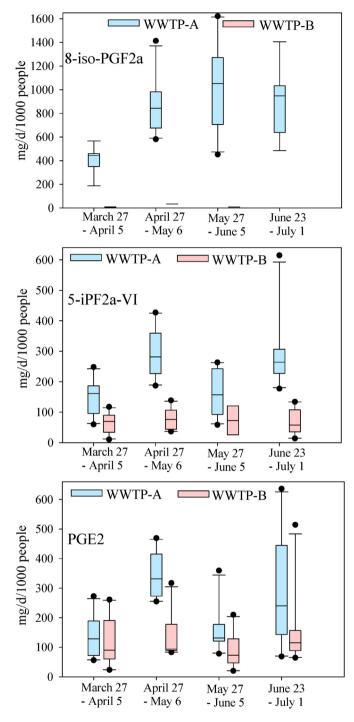


Fig. 2. Mass load of 8-iso-PGF2 α (8-iso-prostaglandin F2 α), 5-iPF2 α -VI (5-iso Prostaglandin F2 α -VI, and PGE2 (prostaglandin E2) in communities A and B. Boxplots showing the median line, interquartile range (25–75 percentiles), whiskers (10 and 90 percentiles), and outliers.

to establish isoprostanes including PGE2 and 5-iPF2 α -VI as the reliable biomarker of community-level oxidative injuries.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

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