


## Article

# Highly Efficient Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub> Z-Scheme Photocatalyst for Its Enhanced Photocatalytic Performance in Degradation of Rhodamine B and Phenol

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**Abstract:** Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub> heterojunctions, with different g-C<sub>3</sub>N<sub>4</sub> dosages, were synthesized using an in situ deposition method, and the photocatalytic performance of g-C<sub>3</sub>N<sub>4</sub>/Ag<sub>3</sub>PO<sub>4</sub> heterojunctions was studied under simulated sunlight conditions. The results revealed that Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub> exhibited excellent photocatalytic degradation activity for rhodamine B (Rh B) and phenol under the same light conditions. When the dosage of g-C<sub>3</sub>N<sub>4</sub> was 30%, the degradation rate of Rh B at 9 min and phenol at 30 min was found to be 99.4% and 97.3%, respectively. After five cycles of the degradation experiment for Rh B, g-C<sub>3</sub>N<sub>4</sub>/Ag<sub>3</sub>PO<sub>4</sub> still demonstrated stable photodegradation characteristics. The significant improvement in the photocatalytic activity and stability of g-C<sub>3</sub>N<sub>4</sub>/Ag<sub>3</sub>PO<sub>4</sub> was attributed to the rapid charge separation between g-C<sub>3</sub>N<sub>4</sub> and Ag<sub>3</sub>PO<sub>4</sub> during the Z-scheme charge transfer and recombination process.

**Keywords:** Ag<sub>3</sub>PO<sub>4</sub>; g-C<sub>3</sub>N<sub>4</sub>; semiconductor photocatalyst; Z-scheme mechanism

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## 1. Introduction

With the rapid development of industry, environmental pollution caused by industrial wastewater is becoming increasingly serious. Photocatalysis is an effective technology to degrade pollutants in water, which has been widely researched [1,2]. However, one-component semiconductor photocatalysts always face various defects, such as low visible-light availability and easy recombination of photogenerated charges. It has been proven that the construction of semiconductor heterostructures is an effective route to improve photocatalytic efficiency [3,4]. In recent years, an all-solid Z-scheme semiconductor composite photocatalyst has been applied in photocatalysis [5–9]. When Z-scheme photocatalysts are excited, h<sup>+</sup> from the valence band (VB) at a higher energy level can combine with e<sup>−</sup> from the conduction band (CB) at a lower energy level, while e<sup>−</sup> with a stronger reducing ability in CB at a higher energy level and h<sup>+</sup> with a stronger oxidation ability in lower VB at a lower energy level can participate in the reduction and oxidation processes during photocatalytic degradation, respectively. This method is conducive to obtain high charge separation efficiency and strong redox ability simultaneously, thus improving the photocatalytic efficiency [8,9].

In recent years, Z-scheme Ag<sub>3</sub>PO<sub>4</sub>-based photocatalysts with a high photocatalytic activity have been designed and applied in wastewater treatment and environmental control [10–13], including Ag<sub>3</sub>PO<sub>4</sub>/MoS<sub>2</sub> [14], Bi<sub>2</sub>MoO<sub>6</sub>/Ag<sub>3</sub>PO<sub>4</sub> [15], Ag<sub>3</sub>PO<sub>4</sub>/Bi<sub>2</sub>WO<sub>6</sub> [16], Ag<sub>3</sub>PO<sub>4</sub>RGO/BiMoO<sub>4</sub> [17], AgPO<sub>4</sub>/Ag/WO<sub>3-x</sub> [18], and Ag<sub>3</sub>PO<sub>4</sub>/Pd/LaPO<sub>4</sub> [19]. Lamellar g-C<sub>3</sub>N<sub>4</sub> nanosheets possess high surface area, suitable band gap (2.7 eV), low cost, and good

thermal and chemical stability, which has attracted extensive attention in the field of photocatalysis [20–23]. When  $g\text{-C}_3\text{N}_4$  is combined with  $\text{Ag}_3\text{PO}_4$ , the resultant  $g\text{-C}_3\text{N}_4/\text{Ag}_3\text{PO}_4$  photocatalyst is expected to show significantly enhanced photocatalytic activity.

Among the many types of pollutants, dyes and dangerous compounds are two main pollutants in industrial wastewater. Rh B and phenol are the typical substances of the two pollutants, respectively. Rh B is very harmful to human health. It can cause redness of skin and viscera, mild congestion of cerebral vascular, rupture of myocardial fiber, and other symptoms. Phenol has a strong corrosive effect on skin and mucous membrane, inhibiting the central nervous system and damaging the function of liver and kidney, etc. In addition, phenol is more difficult to degrade than other pollutants in water. Thus, they were chosen as the degradation object in photocatalytic experiments.

In this paper, we synthesized the  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$  Z-scheme heterojunction photocatalyst using the in situ deposition method and evaluated the photocatalytic activity by the degradation experiment for Rh B and phenol. The influence of  $g\text{-C}_3\text{N}_4$  and  $\text{Ag}_3\text{PO}_4$  on photocatalytic activity was studied in detail and the probable photocatalytic mechanism of  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$  was proposed.

## 2. Experimental Section

### 2.1. Sample Preparation

**Preparation of  $g\text{-C}_3\text{N}_4$ :** A typical calcination method was used to prepare  $g\text{-C}_3\text{N}_4$ . Briefly, 10 g urea powder was placed in an alumina crucible with a lid. The crucible was heated in air at a heating rate of  $2\text{ }^\circ\text{C}\cdot\text{min}^{-1}$  to  $550\text{ }^\circ\text{C}$  and, then held at this temperature for 2 h to obtain  $g\text{-C}_3\text{N}_4$ . Subsequently, the bulk  $g\text{-C}_3\text{N}_4$  was thermally exfoliated into  $g\text{-C}_3\text{N}_4$  nanosheets by calcination at  $600\text{ }^\circ\text{C}$  for 2 h in air. The light yellow product was collected and ground using an agate mortar for subsequent use.

**Synthesis of  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ :** Fifty milligrams of  $g\text{-C}_3\text{N}_4$  nanosheets were dispersed in 80 mL of deionized water by ultrasonication. Silver ammonia solution ( $0.1\text{ g}\cdot\text{L}^{-1}$ ) was dropped into the aqueous dispersion of  $g\text{-C}_3\text{N}_4$  nanosheets and, then magnetically stirred for 1 h to fully adsorb  $\text{Ag}(\text{NH}_3)_2^+$  ions on the surface of  $g\text{-C}_3\text{N}_4$  nanosheets. Then, the  $\text{KH}_2\text{PO}_4$  solution ( $0.1\text{ g}\cdot\text{L}^{-1}$ ) was dropped into the above mixture under magnetic agitation and the mixture continued to be stirred for 1 h. The final product was collected by centrifugation, washed with deionized water and ethanol thrice, and dried at  $70\text{ }^\circ\text{C}$  for 1 h. Finally, the product was collected and ground with an agate mortar for further use. According to the theoretical dosage of  $g\text{-C}_3\text{N}_4$ , the as-prepared samples were named  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ -10 wt%,  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ -20 wt%,  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ -30 wt%, and  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ -40 wt%. The actual dosage of  $g\text{-C}_3\text{N}_4$  detected by EDS were 9.2 wt%, 16.3 wt%, 27.7 wt%, and 41.8 wt%, respectively. In addition, the simple physical mixture of  $\text{Ag}_3\text{PO}_4$  and 30 wt%  $g\text{-C}_3\text{N}_4$  was named the  $\text{Ag}_3\text{PO}_4/g\text{-C}_3\text{N}_4$ -30% mixture.

### 2.2. Sample Characterization

The crystal structure was analyzed by a Bruker D8 X-ray diffractometer (XRD, Bruker, Germany), equipped with a  $\text{Cu K}\alpha$  irradiation light source ( $\lambda = 0.154\text{ nm}$ ). The microstructure was observed using a Tecnai G2 F20 transmission electron microscopy (TEM, FEI, Hillsboro, OR, USA). Room-temperature transient photoluminescence (PL) spectra were recorded using an FLS1000 spectrometer (EI, UK). UV-vis diffuse reflectance spectra (UV-Vis, Hitachi, Tokyo, Japan) were measured by using a UH4150 UV-Vis near-infrared spectrophotometer. The photocurrent response was measured using a CHI 760E electrochemical workstation (Chenhua, Shanghai, China).

### 2.3. Photocatalytic Activity Test

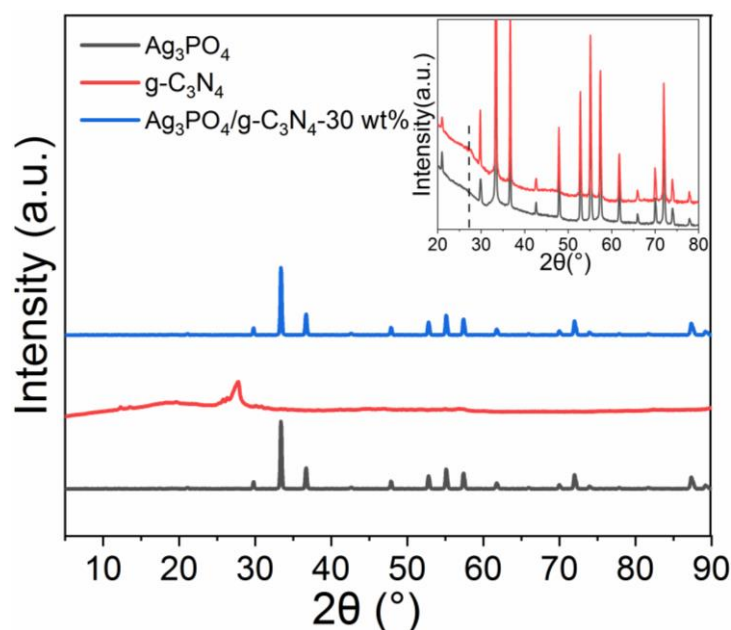
The photocatalytic activity was evaluated by the pollutant degradation experiments at room temperature. A Polifilet xenon lamp (300 W) with a 320-nm filter was used as the light source. The spectra of the xenon lamp are shown in Figure S1 and detailed experimental devices are shown in Figure S2. The reaction solution consisted of 50 mL

of rhodamine B (Rh B,  $5 \text{ mg}\cdot\text{L}^{-1}$ ) or 50 mL of phenol ( $10 \text{ mg}\cdot\text{L}^{-1}$ ), and the photocatalyst was  $0.03 \text{ g Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , or  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ . The photocatalyst was weighed and added to the reaction solution, and the reaction solution was continuously stirred in the dark for 30 min to achieve an adsorption–desorption balance between the photocatalytic material and pollutant. Subsequently, the solution was irradiated by a full-wavelength Xenon lamp, and the absorbance of the supernatant was measured at certain intervals. In the cyclic experiments, the photocatalyst was separated from the reaction system after each degradation experiment, washed with ethanol and deionized water, and re-dispersed in the newly-prepared reaction solution to repeat the degradation experiment.

### 3. Results and Discussion

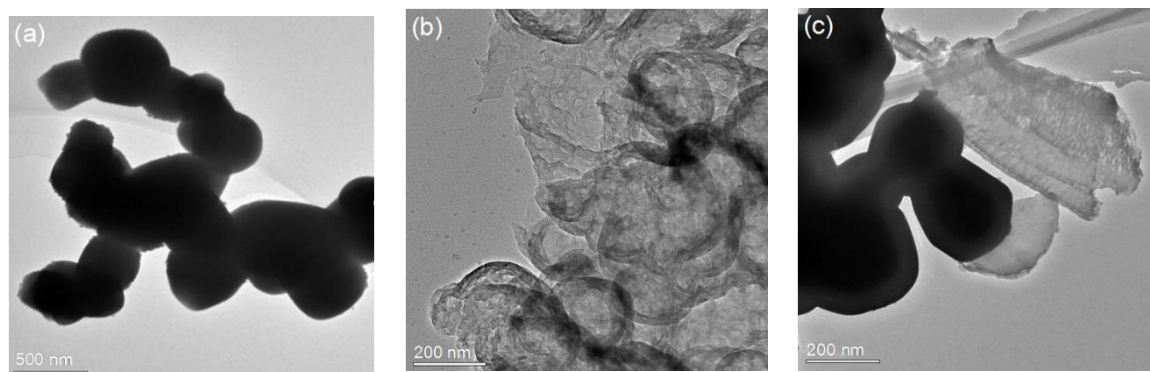
#### 3.1. Structural Analysis and Microstructure

Figure 1 shows the XRD patterns of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$  and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%. As shown in Figure 1, a strong peak appeared in the diffraction pattern of  $\text{g-C}_3\text{N}_4$  at  $2\theta = 26.5^\circ$ , corresponding to the (002) planes of  $\text{g-C}_3\text{N}_4$  (JCPDS card no. 87-1526), which is the characteristic interlayer stacking peak of  $\text{g-C}_3\text{N}_4$  [24]. The  $\text{Ag}_3\text{PO}_4$  and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% exhibited similar XRD patterns and all strong diffraction peaks corresponded to the cubic  $\text{Ag}_3\text{PO}_4$  phase (JCPDS card no. 06-0505). The inset provided the refined XRD patterns of  $\text{Ag}_3\text{PO}_4$  and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%. Compared with  $\text{Ag}_3\text{PO}_4$ , the XRD pattern of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  showed the characteristic peaks of  $\text{g-C}_3\text{N}_4$ ; however, the peak intensities were far weaker than that of  $\text{Ag}_3\text{PO}_4$ . This may be attributed to the inferior crystallinity and lower content of well-exfoliated  $\text{g-C}_3\text{N}_4$ .



**Figure 1.** XRD patterns of as-prepared  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ .

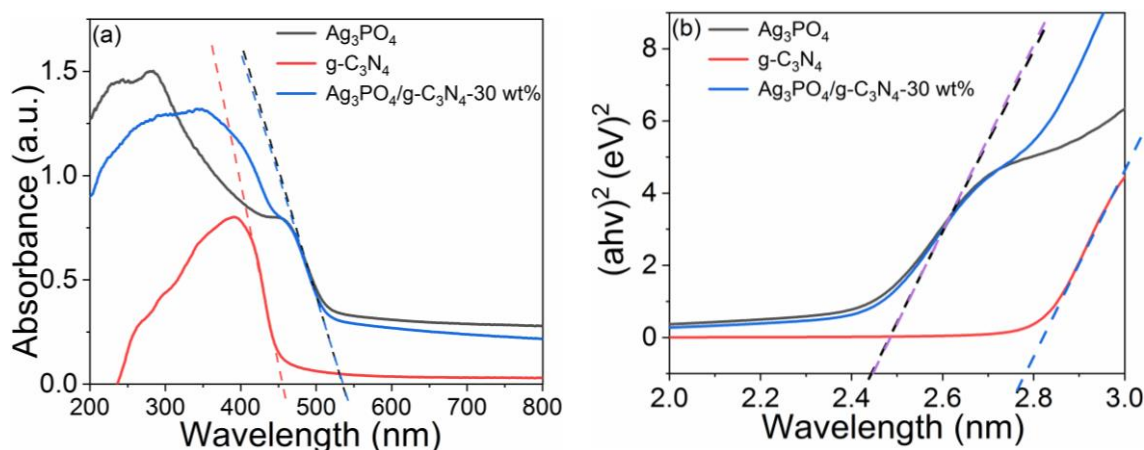
Figure 2 shows TEM images of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  photocatalysts. Figure 2a illustrates that  $\text{Ag}_3\text{PO}_4$  consisted of approximately cubic particles with a size of 200–300 nm. As shown in Figure 2b,  $\text{g-C}_3\text{N}_4$  presented thin wrinkled nanosheets. After thermal exfoliation, the specific surface area of  $\text{g-C}_3\text{N}_4$  increased significantly, due to morphological changes. Figure 2c shows that the small-sized  $\text{Ag}_3\text{PO}_4$  particles were attached to the surface of  $\text{g-C}_3\text{N}_4$ , forming a stable composite.



**Figure 2.** TEM images of (a)  $\text{Ag}_3\text{PO}_4$ , (b)  $\text{g-C}_3\text{N}_4$ , and (c)  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ .

### 3.2. Optical Properties

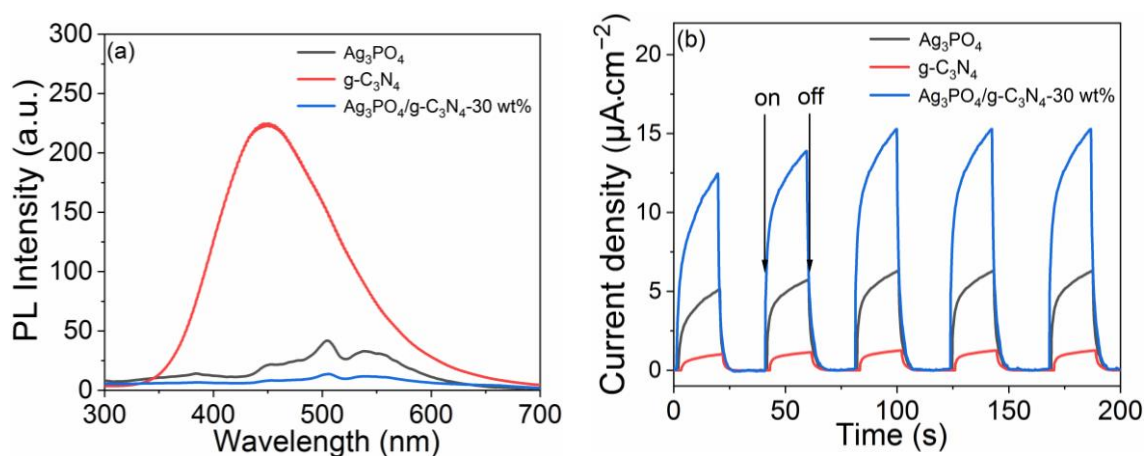
Figure 3 shows the UV-vis diffuse reflectance spectra of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% photocatalysts. As shown in Figure 3a, the absorption cutoff edges of  $\text{Ag}_3\text{PO}_4$  and  $\text{g-C}_3\text{N}_4$  were located at about 460 and 530 nm, respectively. Compared with  $\text{Ag}_3\text{PO}_4$ , the absorption edge of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% was basically unchanged. Based on the UV-vis absorption data, the bandgap width of the photocatalysts was calculated and results are shown in Figure 3b. The calculated bandgap width of  $\text{g-C}_3\text{N}_4$  was about 2.78 eV, whereas the bandgap of  $\text{Ag}_3\text{PO}_4$  and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% decreased to 2.45 eV.



**Figure 3.** (a) UV-vis diffuse reflectance spectra, (b) estimated bandgap of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%.

By testing the photoelectrochemical properties of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% photocatalysts, the separation and transfer efficiency of photogenerated electron-hole pairs were studied and results are shown in Figure 4. Figure 4a presents the photoluminescence (PL) spectra of the as-synthesized photocatalysts. The PL emission peak of  $\text{g-C}_3\text{N}_4$  was located at 460 nm, showing the highest PL intensity and indicating that the photogenerated charge of  $\text{g-C}_3\text{N}_4$  exhibited high recombination efficiency. The PL emission peak of  $\text{Ag}_3\text{PO}_4$  was located at 460 nm, showing a far lower PL intensity than  $\text{g-C}_3\text{N}_4$ . When  $\text{Ag}_3\text{PO}_4$  was combined with  $\text{g-C}_3\text{N}_4$ , the location of the PL emission peak of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% was basically the same as  $\text{Ag}_3\text{PO}_4$ , but the PL peak intensity of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% was significantly lower than  $\text{Ag}_3\text{PO}_4$ . Among  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  exhibited the lowest PL peak intensity, which corresponded to the lowest recombination efficiency for photogenerated charges. As can be observed in Figure 4b, all photocatalyst electrodes exhibited rapid response when irradiated by a Xenon lamp (full wavelength). The  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% showed the highest photocurrent response of about  $16.35 \mu\text{A}\cdot\text{cm}^{-2}$ , which was 2.79 times higher

than  $\text{Ag}_3\text{PO}_4$  ( $5.87 \mu\text{A}\cdot\text{cm}^{-2}$ ) and 21.8 times higher than  $\text{g-C}_3\text{N}_4$  ( $0.75 \mu\text{A}\cdot\text{cm}^{-2}$ ). These results indicate that the combination of  $\text{Ag}_3\text{PO}_4$  and  $\text{g-C}_3\text{N}_4$  reduced the recombination efficiency of photogenerated electrons and holes, and accelerated the charges transfer, which is beneficial for photocatalysis.



**Figure 4.** (a) Photoluminescence spectra and (b) transient photocurrent response curves of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%.

### 3.3. Photocatalytic Activity

Furthermore, using Rh B and phenol as target pollutants, we simulated the photocatalytic reaction under sunlight irradiation using Xenon lamp (full wavelength) irradiation, and evaluated the photocatalytic activity, as shown in Figure 5. Figure 5a shows the photocatalytic activity of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  with different amounts of  $\text{g-C}_3\text{N}_4$ . After irradiation by the Xenon lamp for 9 min, the photocatalytic degradation rate of RhB by  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ ,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -10 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -20 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%, and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -40 wt% was found to be 71.1%, 22.2%, 79.8%, 95.5%, 99.4%, and 89.9%, respectively. With the increase of  $\text{g-C}_3\text{N}_4$  content, the photocatalytic activity of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  initially increased, followed by a decrease. The optimal photocatalytic activity was achieved for  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%. The first-order kinetic model [25,26] was used to calculate the corresponding reaction rate constants ( $k$ ), and the results are shown in Figure 5c. The observed reaction rate constant of  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ ,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -10 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -20 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%, and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -40 wt% was found to be 0.1033, 0.0209, 0.1333, 0.2591, 0.4227, and 0.1911  $\text{min}^{-1}$ , respectively. The  $k$  value of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% ( $0.4227 \text{ min}^{-1}$ ) was the highest, which was  $\approx 4.09$  and 20.24 times higher than  $\text{Ag}_3\text{PO}_4$  and  $\text{g-C}_3\text{N}_4$ , respectively.

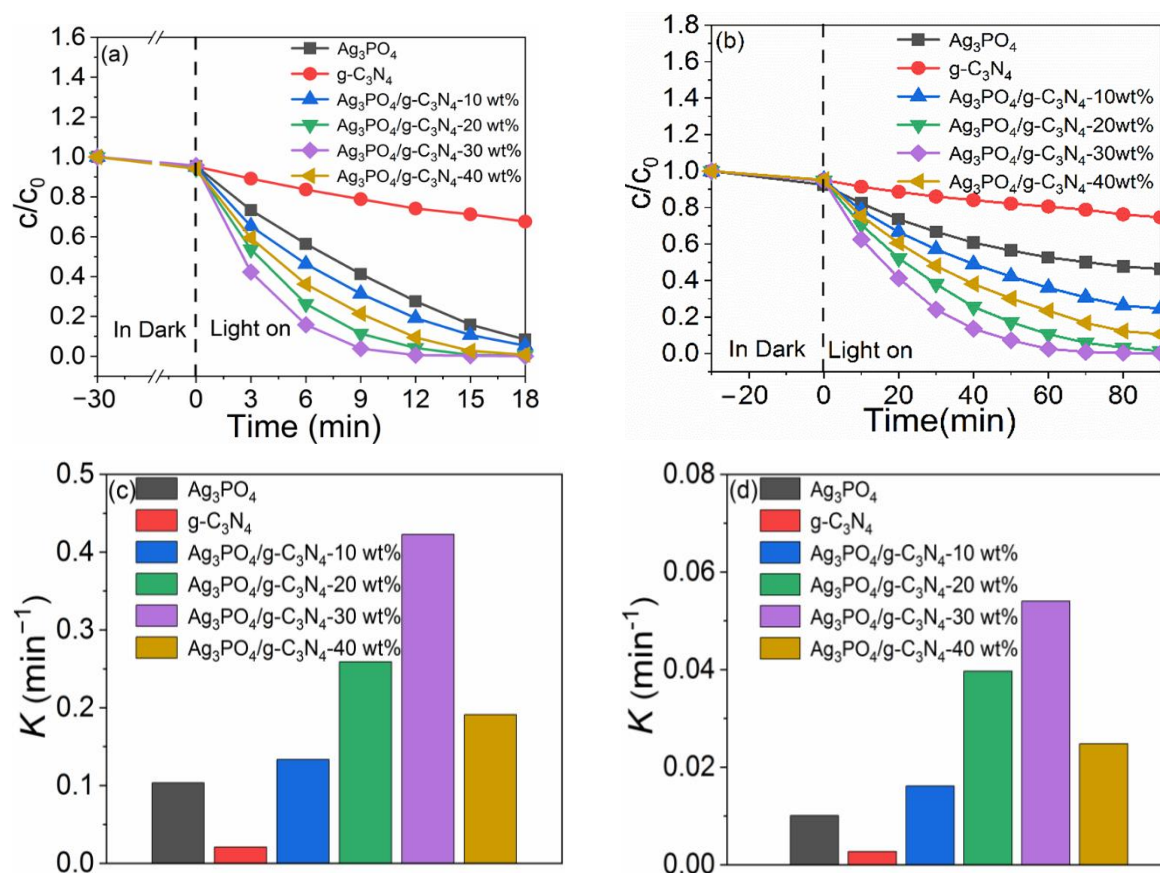
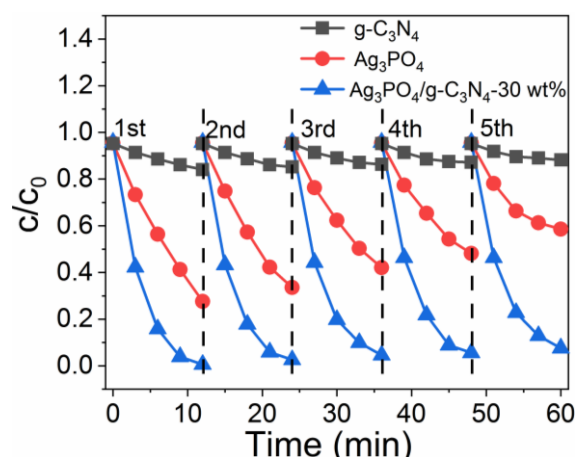


Figure 5. (a,b) Photocatalytic curves, (c,d) rate constants in the degradation of Rh B and phenol, with different  $\text{g-C}_3\text{N}_4$  content.

In order to further verify the superior photocatalytic activity of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ , the photocatalytic degradation experiment for phenol was also carried out and the results are shown in Figure 5b. Under Xenon lamp irradiation for 30 min, the degradation rate of phenol by  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ ,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -10 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -20 wt%,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt%, and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -40 wt% was found to be 43.0%, 15.8%, 63.9%, 90.9%, 99.6%, and 77.5%, respectively. Figure 5d shows that the  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% exhibits the highest rate constant  $k$  ( $0.0540 \text{ min}^{-1}$ ), which was  $\approx 5.35$  and 20.00 times higher than  $\text{Ag}_3\text{PO}_4$  ( $0.01009 \text{ min}^{-1}$ ) and  $\text{g-C}_3\text{N}_4$  ( $0.0027 \text{ min}^{-1}$ ), respectively. Hence,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  showed obvious advantages for the degradation of pollutants.

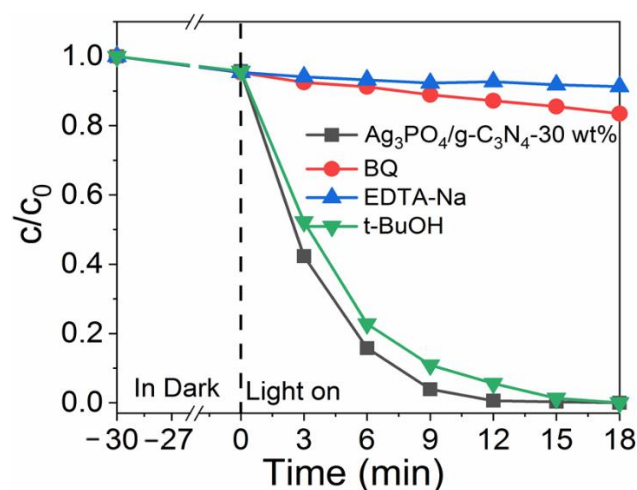
Figure 6 presents the cyclic stability of Rh B degradation by  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% photocatalysts. Under Xenon lamp irradiation, the loss rate of Rh B degradation by  $\text{Ag}_3\text{PO}_4$ ,  $\text{g-C}_3\text{N}_4$ , and  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ -30 wt% during the fifth cycle, compared with the initial degradation, was 32.5%, 11.5%, and 7.3%, respectively. The presence of  $\text{g-C}_3\text{N}_4$  significantly reduced the loss rate for Rh B and phenol degradation. Hence,  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  showed excellent photocatalytic stability.



**Figure 6.** Recycling runs results of Ag<sub>3</sub>PO<sub>4</sub>, g-C<sub>3</sub>N<sub>4</sub>, and Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub>-30 wt% in degradation of Rh B.

### 3.4. Photocatalysis Species

In order to identify the active species during the photocatalytic process, free radical capture experiments were carried out using Rh B as a target pollutant. EDTA-2Na, p-benzoquinone (BZQ), and tert-butanol were introduced during the photocatalytic process as h<sup>+</sup>, ·O<sub>2</sub><sup>-</sup>, and OH<sup>-</sup> inhibitors, respectively, and the results are shown in Figure 7. The introduction of tert-butanol during the photocatalytic process of Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub>-30 wt% rendered no influence on the photodegradation efficiency of Rh B, whereas EDTA-2Na and BZQ both significantly reduced the degradation efficiency of Rh B with a degradation rate of 4.4% and 12.4%, respectively. These results indicate that h<sup>+</sup> and O<sub>2</sub><sup>·-</sup> are the main active species in Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub>-30 wt%.



**Figure 7.** Photocatalytic activities of Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub>-30wt% for the degradation of Rh B in the presence of different scavengers.

### 3.5. Energy Band Structure and Photocatalytic Mechanism

Figure 8 presents the Z-scheme charge transfer pathway of the Ag<sub>3</sub>PO<sub>4</sub>/g-C<sub>3</sub>N<sub>4</sub> composite photocatalyst for the degradation of organic pollutants. The bandgap of g-C<sub>3</sub>N<sub>4</sub> was 2.7 eV with the VB potential of ~1.4 eV and CB potential of ~-1.3 eV [27,28]. The potential of e<sup>-</sup> on the CB of g-C<sub>3</sub>N<sub>4</sub> was -1.3 eV, which can reduce the molecular oxygen O<sub>2</sub> to ·O<sub>2</sub><sup>-</sup> because the potential of O<sub>2</sub>/·O<sub>2</sub><sup>-</sup> was -0.44 eV vs. NHE. Therefore, O<sub>2</sub><sup>·-</sup> was the main active substance during the photocatalytic process by g-C<sub>3</sub>N<sub>4</sub>. The bandgap of Ag<sub>3</sub>PO<sub>4</sub> was 2.45 eV with a VB potential of ~2.9 eV and CB potential of ~0.45 eV [29].

The generated electrons ( $e^-$ ) in the CB of  $\text{Ag}_3\text{PO}_4$  are insufficient to reduce  $\text{O}_2$  into  $\text{O}^{2-}$ . Therefore, holes ( $h^+$ ) play a major role during the photocatalytic degradation of organic matter by  $\text{Ag}_3\text{PO}_4$ .

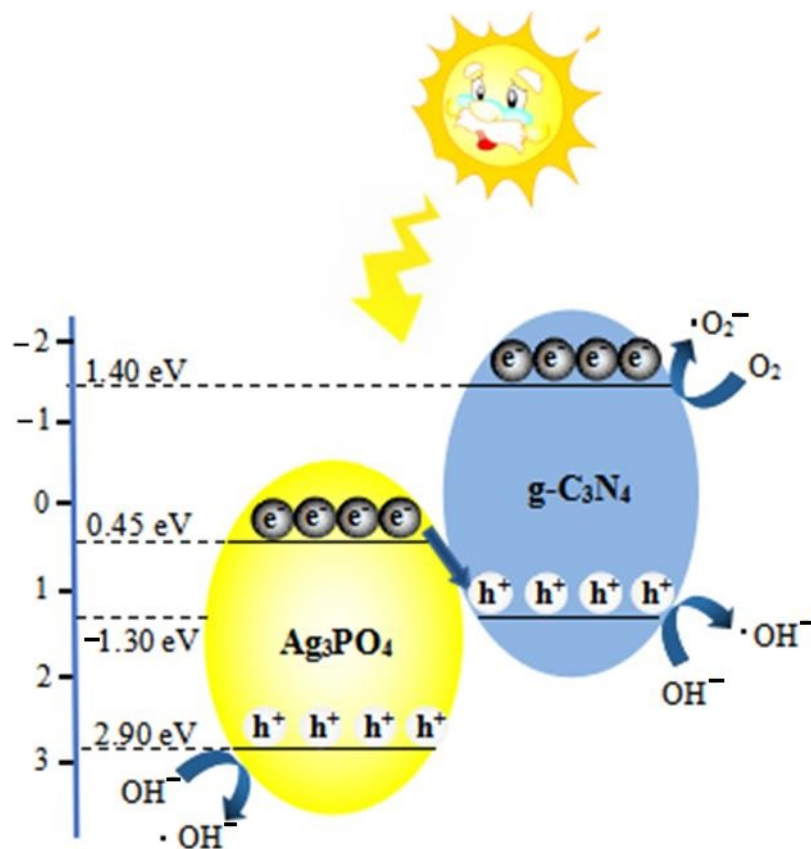


Figure 8. Energy band structure and Z-scheme photocatalytic mechanism of  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$ .

Based on the energy band analysis, it can be inferred that the photogenerated  $e^-$  in the CB of  $\text{Ag}_3\text{PO}_4$  can combine with  $h^+$  in the VB of  $\text{g-C}_3\text{N}_4$  due to the formation of a heterojunction interface between  $\text{Ag}_3\text{PO}_4$  particles and  $\text{g-C}_3\text{N}_4$  nanosheets, resulting in the accumulation of  $e^-$  in the CB of  $\text{g-C}_3\text{N}_4$  and  $h^+$  in VB of  $\text{Ag}_3\text{PO}_4$ . The  $h^+$  in the VB of  $\text{Ag}_3\text{PO}_4$  can directly react with pollutants, whereas the electrons in CB of  $\text{g-C}_3\text{N}_4$  can reduce  $\text{O}_2$  into  $\text{O}^{2-}$ , which reacts with pollutants. The Z-scheme charge transfer mechanism promotes the separation of electron-hole pairs, slows down the photocorrosion of  $\text{Ag}^+$ , and improves photocatalyst activity and stability.

#### 4. Conclusions

In summary, the Z-scheme heterojunction  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  photocatalyst was synthesized using an in situ deposition method and exhibited excellent photocatalytic degradation activity for Rh B and phenol under Xenon lamp irradiation. The observed rate constant ( $k$ ) for the degradation of Rh B by  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  was found to be  $0.4227 \text{ min}^{-1}$ , which was 4.09 and 20.24 times higher than pure  $\text{Ag}_3\text{PO}_4$  and  $\text{g-C}_3\text{N}_4$ , respectively. Moreover, the  $k$  value for the degradation of phenol by  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  was  $0.0540 \text{ min}^{-1}$ , which was 5.35 and 20.00 times higher than pure  $\text{Ag}_3\text{PO}_4$  and  $\text{g-C}_3\text{N}_4$ , respectively. Overall, the formation of the Z-scheme heterojunction hindered the recombination of photogenerated electrons and holes, and accelerated the electron transfer, thus improving the activity and stability of photocatalysts.

**Supplementary Materials:** Figure S1: The spectra of xenon lamp, Figure S2: The picture of experimental setup.



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**Data Availability Statement:** The data presented in this study are available on request from the corresponding author. The data are not publicly available due to the ongoing follow-up studies.

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**Sample Availability:** Samples of the  $\text{Ag}_3\text{PO}_4/\text{g-C}_3\text{N}_4$  photocatalyst are available from the authors. However, it may be necessary to pay properly for the synthesis and mailing of samples.

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