

## ORIGINAL ARTICLE

# Cigarette smoking enhances the metabolic activation of the polycyclic aromatic hydrocarbon phenanthrene in humans

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## Abstract

Although it is well established that human cytochrome P450 1 family enzymes are induced by cigarette smoking through activation of the Ah receptor, it is not known whether this leads to increased metabolic activation or detoxification of carcinogenic polycyclic aromatic hydrocarbons (PAH), which are present in cigarette smoke and the general environment. We gave oral doses of deuterated phenanthrene ( $[D_{10}]Phe$ ), a non-carcinogenic surrogate of carcinogenic PAH such as benzo[a]pyrene, to smokers ( $N = 170$ , 1 or 10  $\mu g$  doses) and non-smokers ( $N = 57$ , 1  $\mu g$  dose). Bioactivation products (dihydrodiol and tetraol) and detoxification products (phenols) of  $[D_{10}]Phe$  were determined in 6-h urine to obtain a comprehensive metabolic profile. Cigarette smoking increased the bioactivation of  $[D_{10}]Phe$  and decreased its detoxification resulting in significantly different metabolic patterns between smokers and non-smokers ( $P < 0.01$ ), consistent with increased cancer risk in smokers. The Phe bioactivation ratios ( $[D_{10}]PheT/total [D_9]OHPhe$ ) were significantly higher (2.3 ( $P < 0.01$ ) to 4.8 ( $P < 0.001$ ) fold) in smokers than non-smokers. With solid human *in vivo* evidence, our results for the first time demonstrate that cigarette smoking enhances the metabolic activation of Phe, structurally representative of carcinogenic PAH, in humans, strongly supporting their causal role in cancers caused by smoking. The results suggest potential new methods for identifying smokers who could be at particularly high risk for cancer.

## Introduction

More than 50 years ago, Welch et al. investigated the metabolism of benzo[a]pyrene (BaP) to 3-hydroxyBaP in placenta from smokers and non-smokers, demonstrating for the first time that cigarette smoking induced the human metabolism of BaP, a prototypic highly carcinogenic polycyclic aromatic hydrocarbon (PAH), now considered “carcinogenic to humans” by the International Agency for Research on Cancer (1,2). Decades of

subsequent research extended and refined this initial finding as summarized in [Supplementary data](#), available at Carcinogenesis Online, but it is still not known whether the induction of human PAH metabolism by cigarette smoking results in increased PAH metabolic activation or detoxification. This question, addressed in the study reported here, is clearly critical to our understanding of mechanisms of lung cancer etiology by cigarette smoking,

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## Abbreviations

3-OHBaP	3-hydroxyBaP
CYP1	cytochrome P450 1
BaP	benzo[a]pyrene
PAH	polycyclic aromatic hydrocarbons
PBR	Phe bioactivation ratios
Phe	phenanthrene

and to the potential causative role in human lung cancer of the multiple carcinogenic PAH in cigarette smoke.

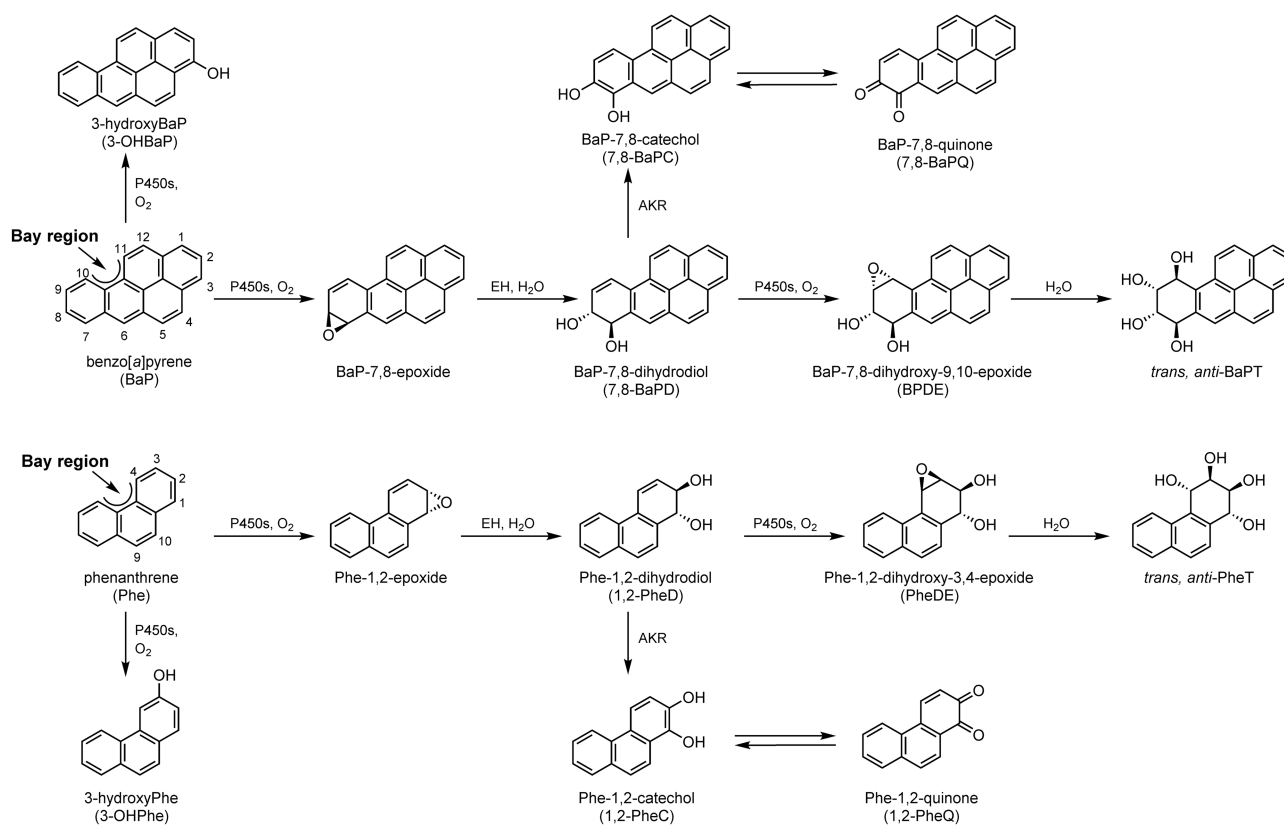
The metabolism of PAH is catalyzed by cytochromes P450 1A1, 1A2 and 1B1 (CYP1A1, CYP1A2 and CYP1B1), which are induced by cigarette smoking via activation of the Ah receptor and are involved in both the metabolic activation and detoxification of BaP and other PAH, as illustrated in Figure 1 for BaP and phenanthrene (Phe). A major pathway of metabolic activation proceeds through BaP-7,8-epoxide to the proximate carcinogen 7,8-BaPD, then to the ultimate carcinogen BPDE, which reacts with DNA to produce adducts that are critical in carcinogenesis by BaP because they cause mutations in the TP53 tumor suppressor gene and other critical genes (2,3). Redox cycling between BaP-7,8-catechol and BaP-7,8-quinone has also been associated with metabolic activation, whereas 3-hydroxyBaP (and other OHBaPs) are products of detoxification (4).

Extensive studies in laboratory animals have investigated the role of Cyp1a1, Cyp1a2 and Cyp1b1 in the metabolism of BaP in mice using genetic engineering tools (5–7). The results of these studies are complex and indicate that experimental design is a key factor and that these enzymes may have a more important

role in carcinogen detoxification than activation in laboratory animals. For example, after reviewing studies involving targeted gene disruption methodologies, Reed et al. concluded that in vitro studies using hepatic enzyme systems showed that BaP was metabolically activated by cytochrome P450s, while studies in vivo indicated that hepatic P450 enzymes played a more pivotal role in BaP detoxification rather than in its activation (5).

Phe (Figure 1) is the simplest PAH with a bay region, a feature closely associated with carcinogenicity, but Phe, with only 3 aromatic rings, is noncarcinogenic (2). The metabolism of Phe closely resembles that of BaP, proceeding through 1,2-PheD and a bay region diol epoxide PheDE, which reacts with H<sub>2</sub>O to produce *trans,anti*-PheT (8,9). Phe is also metabolized to phenols (1-, 2-, 3-, 4-, 9-OHPhe) and quinones (1,2-, 3,4-, 9,10-PheQ) as illustrated (10,11).

In this study, we used a unique carcinogen phenotyping approach, in which smokers and non-smokers took a single oral dose of [D<sub>10</sub>]Phe. The use of [D<sub>10</sub>]Phe allowed us to profile critical metabolic pathways in smokers and non-smokers, without any interference from exposure to environmental Phe or its metabolites which are ubiquitous in the diet and polluted air among others. We have previously shown that orally administered [D<sub>10</sub>]Phe has a similar metabolic profile with respect to formation of *trans,anti*-PheT as [D<sub>10</sub>]Phe administered in a cigarette (8). Thus, we were able to determine the effect of cigarette smoking on the metabolic activation versus detoxification of [D<sub>10</sub>]Phe, as measured by the relative amounts of urinary [D<sub>10</sub>] *trans,anti*-PheT, [D<sub>10</sub>]1,2-PheD and total [D<sub>9</sub>]OHPhe in 170 smokers and 57 non-smokers.



**Figure 1.** Metabolism pathways of BaP and Phe to epoxides (BaP-7,8-epoxide and Phe-1,2-epoxide), dihydrodiols (7,8-BaPD and 1,2-PheD), diol epoxides (BPDE and PheDE), tetraols (*trans,anti*-BaPT and *trans,anti*-PheT), phenols (3-OHBaP and 3-OHPhe), catechols (7,8-BaPC and 1,2-PheC) and quinones (7,8-BaPQ and 1,2-PheQ). The major pathway of metabolic activation proceeds through the epoxides to the diols to the diol epoxides and is measured as 1,2-PheD and *trans,anti*-PheT while detoxification is measured as 3-OHPhe.

## Materials and methods

### Chemicals

[D<sub>10</sub>]Phe (98%) was purchased from Cambridge Isotope Laboratories, Inc. (Andover, MA). [D<sub>10</sub>]Phe for oral dosing was repurified in the University of Minnesota Molecular and Cellular Therapeutics GMP facility by normal-phase high-performance liquid chromatography followed by recrystallization from ethanol. It had >99% chemical purity and 98% isotopic purity. [D<sub>10</sub>]Phe oral dose solutions were prepared by dissolving 1 mg/4 mg of purified [D<sub>10</sub>]Phe (employing a balance that will accurately weigh 200 µg and an anti-static gun) in 1000 ml/400 ml of 100% ethanol and delivering 1-ml aliquots of the solution into amber dosing bottles at the Fairview Compounding Pharmacy (Minneapolis, MN), such that each bottle contained 1 µg/10 µg of [D<sub>10</sub>]Phe. All oral doses were stored at 4 °C before administration. On the day of dosing, 4 ml of drinking water was added to the amber dosing bottle, which was swirled to achieve a composition of 20% ethanol (this was necessary because storage in 20% ethanolic/80% H<sub>2</sub>O resulted in decreased concentrations of [D<sub>10</sub>]Phe).

### Clinical study design

The study was approved by the U.S. Food and Drug Administration, Health Canada, and the University of Minnesota and Mayo Clinic Institutional Review Boards. Healthy volunteers including current cigarette smokers smoking at least 10 cigarettes per day for at least 1 year and non-smokers (age 18–65 years) were invited into the clinic for an orientation visit and informed consent was obtained. Eligible subjects must have been in good physical health (no history of aerodigestive cancer; no unstable medical condition such as liver, renal or cardiac dysfunction) and stable, good mental health (not currently or within the past 6 months experiencing unstable or untreated psychiatric diagnosis including substance abuse disorder as determined by the licensed medical professional). Subjects using tobacco or nicotine-containing products other than cigarettes greater than twice weekly, who were taking other investigational agents or medications that alter relevant metabolic enzymes, and/or women who were pregnant or nursing were also excluded.

One hundred thirty-five current smokers (65 females) received 10 µg (53.2 nmol) of [D<sub>10</sub>]Phe orally and another 35 current smokers (14 females) and 57 non-smokers (29 females) received 1 µg (5.32 nmol) of [D<sub>10</sub>]Phe orally once after fasting a minimum of 4 h to empty the stomach. We did not have regulatory permission to give 10 µg [D<sub>10</sub>]Phe to non-smokers. The smoking status of each subject was determined by questionnaire and confirmed by measurement of urinary cotinine. Those subjects who had ≥30 ng/ml of urinary cotinine were identified as current smokers (12). Urine collections were obtained from 0 to 6 h after dosing with [D<sub>10</sub>]Phe. Participants were encouraged to drink water during the collection period. The volume of the 6-h urine collection was measured, and aliquots (15 ml) of the urine samples were frozen at –20 °C until analysis. The collection period was based on our previous study demonstrating an excellent correlation between 6 and 48 h urinary excretion of [D<sub>10</sub>]PheT ( $r = 0.95$ ,  $P < 0.001$ ) (8). Thus analysis of 6-h urine is an accurate reflection of [D<sub>10</sub>]Phe metabolism in humans.

### Determination of urinary metabolites

The analyses of [D<sub>10</sub>]PheT, [D<sub>10</sub>]PheD and [D<sub>9</sub>]OHPhe were performed essentially as described previously (8,13,14). Briefly, for the analysis of [D<sub>10</sub>]PheT, a 100 µL aliquot of urine containing 100 fmol [<sup>13</sup>C<sub>6</sub>]PheT as internal standard was incubated with β-glucuronidase (1050 units) and arylsulfatase (8400 units) overnight with shaking at 37 °C. After solid-phase extraction and silylation, the samples were analyzed by gas chromatography-negative chemical ionization-tandem mass spectrometry (GC-NCI-MS/MS). For the analysis of [D<sub>10</sub>]1,2-PheD, a 300 µL (10 µg dose group)/1 ml (1 µg dose group) aliquot of urine containing 5 pmol [<sup>13</sup>C<sub>6</sub>]1,2-PheD as internal standard was incubated with β-glucuronidase (1000 units) and arylsulfatase (8000 units) overnight with shaking at 37 °C. After solid-phase extraction and silylation, the samples were analyzed by GC-NCI-MS/MS. For the analysis of [D<sub>9</sub>]OHPhe ([D<sub>9</sub>]1-, 2-, 3-, 4- and 9-OHPhe), a 3-ml aliquot of urine containing 1 ng [<sup>13</sup>C<sub>6</sub>]3-OHPhe as internal standard was incubated with β-glucuronidase (2000 units) and arylsulfatase (16000 units) overnight with shaking at 37 °C. After solid-phase extraction and silylation, the samples were analyzed by GC-electron impact (EI)-MS/MS with a TSQ Quantum

instrument (Thermo Scientific, San Jose, CA). [D<sub>9</sub>]Phe ortho-quinones ([D<sub>9</sub>]PheQ) including [D<sub>9</sub>]1,2-PheQ, [D<sub>9</sub>]3,4-PheQ and [D<sub>9</sub>]9,10-PheQ were also determined in the urine of 50 subjects (16 1-µg dose group non-smokers, 9 1-µg dose group smokers and 25 10-µg dose group smokers) by LC-ESI-MS/MS using the method described previously (10). Typical GC/LC-MS/MS chromatograms of [D<sub>9</sub>]Phe metabolites and corresponding <sup>13</sup>C-labeled internal standards are shown in Figure S1 (see Supplementary data, available at Carcinogenesis Online).

### Statistical analysis

For continuous variables, the Wilcoxon rank sum test was used since they were not normally distributed. To compare the pie charts, a multivariate test for compositional data was used (15,16) because the percentages in each pie chart are linearly correlated (i.e. the sum of the three percentages is 100%) and can be regarded as compositional data. A value of  $P < 0.05$  was considered statistically significant. The data were analyzed by SPSS Version 26.0 for Windows (IBM Inc., Armonk, NY) and R package Compositional Version 3.9 under R Version 4.0.0.

## Results

Smokers received doses of either 1 or 10 µg of [D<sub>10</sub>]Phe, while non-smokers received only 1 µg doses of [D<sub>10</sub>]Phe, due to regulatory limitations. To determine the possible effect of the amount of an oral dose of [D<sub>10</sub>]Phe on the metabolic pattern, five smokers took two different oral doses of [D<sub>10</sub>]Phe (1 or 10 µg) 1-year apart. The percentages of [D<sub>10</sub>]1,2-PheD ( $P = 0.55$ ), [D<sub>10</sub>]PheT ( $P = 0.31$ ), total [D<sub>9</sub>]OHPhe ( $P = 0.15$ ) as well as the compositional test ( $P = 0.35$ ) were not significantly different between the two doses in these 5 subjects, confirming that these different oral doses of [D<sub>10</sub>]Phe did not affect the metabolic pattern in the same person. Thus, comparison of the metabolic patterns of [D<sub>10</sub>]Phe between the 1 µg dose in non-smokers and the 10 µg dose in smokers was valid.

Concentrations of urinary [D<sub>10</sub>]Phe metabolites (nmol/6-h urine) are summarized in Table 1. All the [D<sub>10</sub>]Phe metabolites in the 10 µg dose group were significantly higher than in the 1 µg dose group ( $P < 0.01$ ) as they received different oral doses of [D<sub>10</sub>]Phe. Thus, the mean values of total [D<sub>10</sub>]Phe metabolites ( $\Sigma$ ([D<sub>10</sub>]1,2-PheD, [D<sub>10</sub>]PheT and total [D<sub>9</sub>]OHPhe)) in the 10 µg dose group (9.94 nmol/6-h urine) were 12.7 and 14.8 times higher than that in the 1 µg dose non-smokers (0.78 nmol/6-h urine,  $P < 0.001$ ) and smokers (0.67 nmol/6-h urine,  $P < 0.001$ ). Although there were no significant differences between the 1 µg dose non-smokers and smokers ( $P = 0.472$ ) in total [D<sub>10</sub>]Phe metabolites, the concentration of total [D<sub>9</sub>]OHPhe in the 1 µg dose smokers (0.18 nmol/6-h urine) was significantly lower than that in non-smokers (0.28 nmol/6-h urine,  $P < 0.01$ ), indicating cigarette smoke affected the metabolic pattern, as further shown below. The concentrations of total [D<sub>9</sub>]PheQ were  $0.0044 \pm 0.0079$  (mean  $\pm$  SD) nmol/6-h urine and  $0.072 \pm 0.11$  nmol/6-h urine in the 1 µg dose subjects ( $N = 25$ ) and 10 µg dose subjects ( $N = 25$ ), respectively, accounting for <1% of total deuterated Phe metabolites. The average 6-h conversion rate (6-h total [D<sub>10</sub>]Phe metabolites (nmol)/oral dose of [D<sub>10</sub>]Phe (nmol)) were 14.7 % in the 1 µg dose non-smokers, 12.7% in the 1 µg dose smokers and 18.7% in the 10 µg dose smokers. Our earlier study demonstrated that levels of [D<sub>10</sub>]Phe metabolites at 6 h were about half the amount excreted in 48 h (8).

The percentages of each metabolite of the total [D<sub>10</sub>]Phe metabolites quantified are presented in Table 2 and Figure 2. The composition of the metabolite pattern was significantly different in smokers and non-smokers for [D<sub>10</sub>]Phe metabolites ( $P < 0.01$ ). This was most clearly evident when comparing smokers (10 µg dose,  $N = 135$ ) and non-smokers (1 µg dose,  $N = 57$ ) in Figure 2. Considering the individual metabolites, the percentages of the

**Table 1.** Concentrations of [D<sub>10</sub>]Phe metabolites in 6-h urine (nmol/6-h urine)

		[D <sub>10</sub> ]Phe metabolites		
		[D <sub>10</sub> ]1,2-PheD	[D <sub>10</sub> ]PheT	Total [D <sub>9</sub> ]OHPhe
Non-smokers (1 µg dose, N = 57)	Mean (range)	0.39 (ND ~1.61)	0.11 (0.02–0.25)	0.28 (0.05–1.55)
	SD	0.40	0.06	0.23
	CV (%)	102	52.0	80.8
	Median (25 <sup>th</sup> , 75 <sup>th</sup> )	0.23 (0.12, 0.44)	0.11 (0.07, 0.14)	0.24 (0.15, 0.34)
Smokers (1 µg dose, N = 35)	Mean (range)	0.36 (ND ~1.29)	0.14 (0.003–0.29)	<b>0.18<sup>a</sup></b> (ND ~0.68)
	SD	0.33	0.07	0.14
	CV (%)	90.2	55.2	77.5
	Median (25 <sup>th</sup> , 75 <sup>th</sup> )	0.23 (0.14, 0.56)	0.14 (0.07, 0.17)	0.16 (0.07, 0.24)
Smokers (10 µg dose, N = 135)	Mean (range)	<b>6.55<sup>a</sup></b> (ND ~28.63)	<b>2.26<sup>a</sup></b> (ND ~7.75)	<b>1.13<sup>a</sup></b> (ND ~6.41)
	SD	4.37	1.67	0.75
	CV (%)	66.7	73.9	66.7
	Median (25 <sup>th</sup> , 75 <sup>th</sup> )	5.49 (3.29, 9.33)	1.82 (0.88, 3.38)	1.01 (0.65, 1.47)

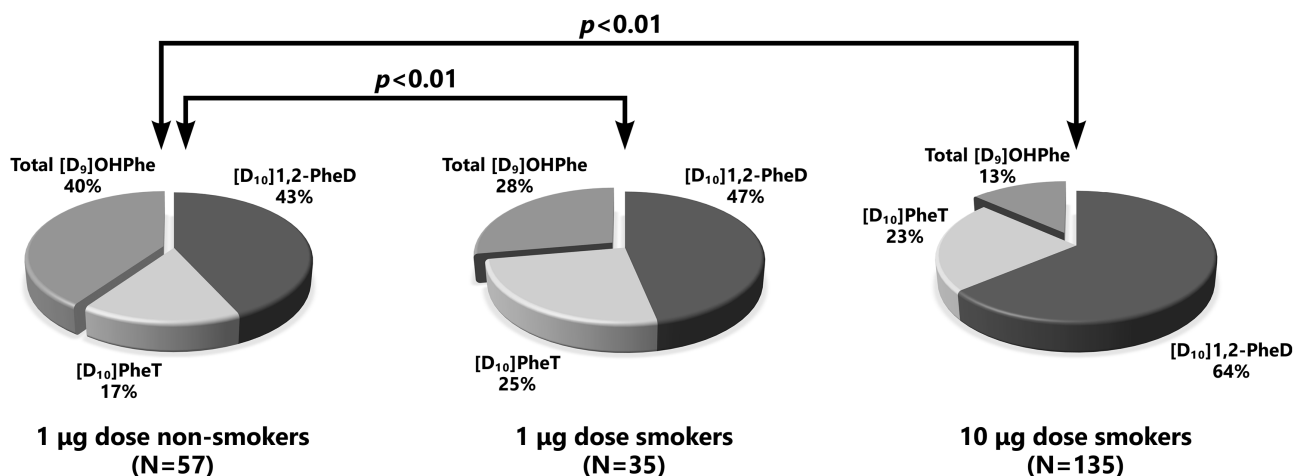
<sup>a</sup>Bolded values,  $P < 0.01$  compared with non-smokers, Wilcoxon rank-sum test. CV, coefficient of variation.

**Table 2.** Percentage of each metabolite of the total [D<sub>10</sub>]Phe metabolites

	Non-smokers (1 µg dose, N = 57)	Smokers (1 µg dose, N = 35)		Smokers (10 µg dose, N = 135)	
	Mean % (SD)	Mean % (SD)	P-value <sup>a</sup>	Mean % (SD)	P-value <sup>a</sup>
[D <sub>10</sub> ]1,2-PheD	43.15 (23.35)	46.74 (19.73)	0.35	63.89 (14.95)	<0.01
[D <sub>10</sub> ]PheT	16.52 (10.18)	25.48 (16.61)	<b>0.02</b>	22.76 (13.46)	<0.01
Total [D <sub>9</sub> ]OHPhe	40.32 (20.46)	27.78 (13.89)	<0.01	13.35 (9.58)	<0.01
Compositional <sup>b</sup>		P-value < 0.01		P-value < 0.01	

<sup>a</sup>Bolded values, statistically significant compared to non-smokers, using the Wilcoxon rank-sum test for each percentage.

<sup>b</sup>The test for compositional data was used, as in the pie charts of Figure 2.

**Figure 2.** Pie charts representing the pattern of metabolites of [D<sub>10</sub>]Phe in smokers and non-smokers.

bioactivation end products [D<sub>10</sub>]PheT were significantly higher in smokers ( $P = 0.02$  in 1 µg dose smokers and  $P < 0.01$  in 10 µg dose smokers) than non-smokers ( $P < 0.01$ ). As for [D<sub>10</sub>]1,2-PheD, an intermediate metabolite in the bioactivation pathway, its percentage was higher in all the smokers than non-smokers while the difference was statistically significant in the 10 µg dose smokers versus 1 µg dose non-smokers ( $P < 0.01$ ). The percentage of total [D<sub>9</sub>]OHPhe, representing detoxification, was lower in all current smokers than non-smokers ( $P < 0.01$ ). Males and females had similar non-smoker versus smoker effects on the metabolic pattern (see [Supplementary data Table 1](#), available at

Carcinogenesis Online). The Phe bioactivation ratios (PBR): [D<sub>10</sub>]PheT/total [D<sub>9</sub>]OHPhe were significantly different in smokers and non-smokers as shown in [Table 3](#) and [Figure 3](#).

Non-deuterated Phe metabolites were also simultaneously detected. For each of the non-deuterated Phe metabolites, the mean urinary concentrations in smokers were 1.61- to 3.56-fold higher than those in non-smokers (all  $P < 0.01$  except for 1,2-PheD). The mean values of total non-deuterated Phe metabolites ( $\Sigma$ (1,2-PheD, PheT and total OHPhe)) in the 1 µg dose group of smokers (5.36 nmol/6-h urine,  $P < 0.01$ ) and 10 µg dose group of smokers (5.98 nmol/6-h urine,  $P < 0.001$ ) were also significantly



Table 3. PBR in smokers and non-smokers

	Non-smokers (1 µg dose [D <sub>10</sub> ]Phe)	Smokers (1 µg dose [D <sub>10</sub> ]Phe)	Smokers (10 µg dose [D <sub>10</sub> ]Phe)
PBR ([D <sub>10</sub> ]PheT/[D <sub>9</sub> ]3-OHPhe)	2.05	3.56 <sup>a</sup>	3.92 <sup>c</sup>
PBR ([D <sub>10</sub> ]PheT/total [D <sub>9</sub> ]OHPhe)	0.40	0.91 <sup>b</sup>	1.91 <sup>c</sup>

<sup>a</sup>P < 0.05.<sup>b</sup>P < 0.01.<sup>c</sup>P < 0.001 compared with non-smokers. Mann–Whitney U-test was used since the data were not normally distributed.

higher than that in non-smokers (2.89 nmol/6-h urine). This result indicates that Phe metabolites are reliable and effective biomarkers of human exposure to PAH in cigarette smoke, generally consistent with previous studies (9,14,17).

## Discussion

In this study, we used a novel approach featuring [D<sub>10</sub>]Phe as an indicator of human PAH metabolism. The use of [D<sub>10</sub>]Phe eliminates possible complications from environmental exposure to metabolites of Phe, a ubiquitous PAH. In our previous study, we demonstrated that the pharmacokinetics of [D<sub>10</sub>]Phe administered to smokers in a cigarette were essentially identical to [D<sub>10</sub>]Phe given orally to the same subjects (8). Thus, oral administration of [D<sub>10</sub>]Phe, as in the present study, reflects exposure to PAH in cigarette smoke, even though the oral dose is likely metabolized in the liver by CYP1A2 while the smoking dose is metabolized in the lung by CYP1A1 and CYP1B1. All three enzymes are induced by cigarette smoking (18–20).

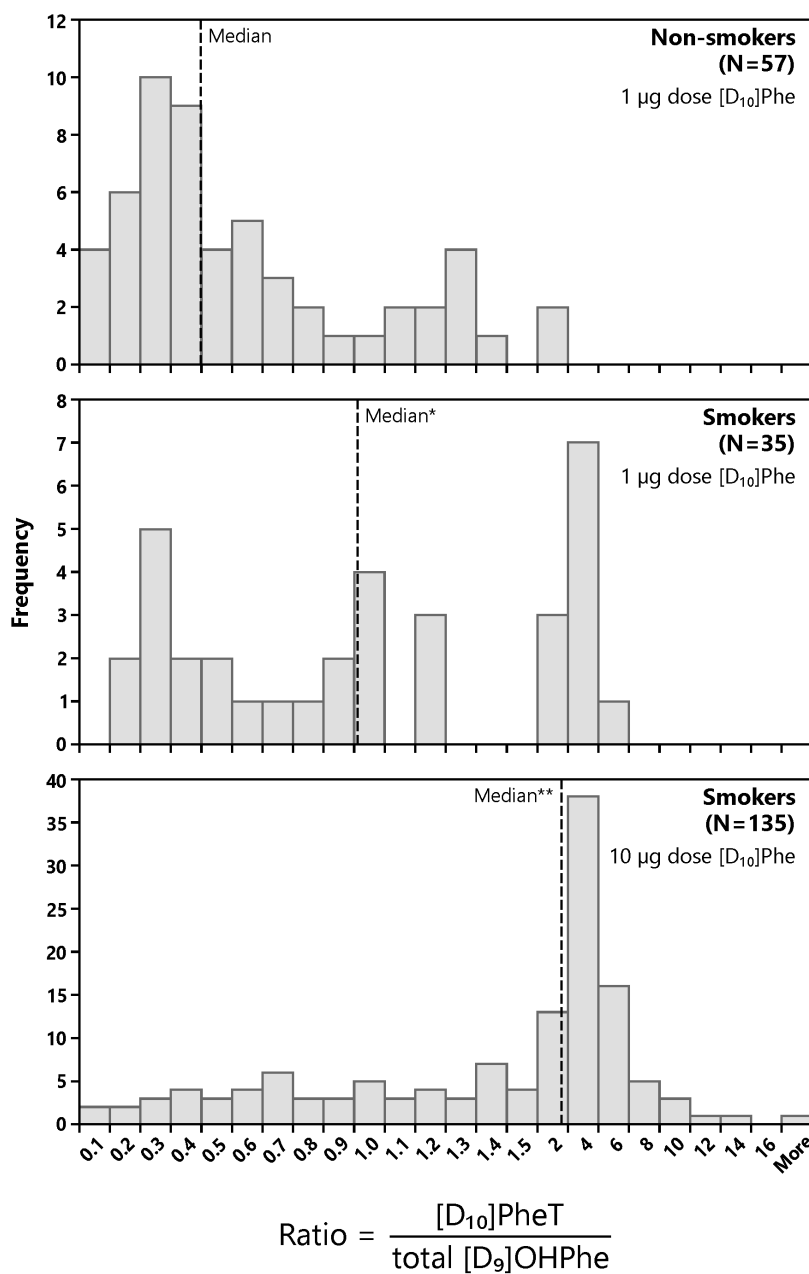
The results of this unique study demonstrate for the first time in humans that the induction of CYP1A1, 1A2 and 1B1 by cigarette smoking via the Ah receptor, a phenomenon that has been known and investigated for more than 50 years, causes an increase in the metabolic activation of Phe, as clearly shown in the metabolic profile of [D<sub>10</sub>]Phe (Figure 2). The structural similarity of Phe and BaP, a highly carcinogenic PAH considered “carcinogenic to humans” by the IARC, and the close relationship of their metabolic pathways as shown in our previous studies, support the conclusion that cigarette smoking increases the metabolic activation of carcinogenic PAH in humans, resulting in increased cancer risk, as discussed further below. While the increased exposure to PAH in smokers compared with non-smokers has been repeatedly demonstrated (9,10,14,17), the increased metabolic activation of these PAH in smokers arguably increases cancer risk and supports the critical role of these potent carcinogens as causes of lung cancer and other PAH-related cancers in smokers.

We developed methods to panoramically characterize [D<sub>10</sub>]Phe metabolites in human urine (8,10,13,14). The use of Phe metabolites has inherent advantages: it has a bay region—a feature closely associated with carcinogenicity, its metabolism is very similar to that of BaP while its metabolite concentrations in human urine are ~ 10 000 times higher than those of BaP, making it much more practical for large studies. A unique superiority of the [D<sub>10</sub>]Phe metabolite phenotyping approach is that it integrates all human enzymatic and genetic factors that impact carcinogenic PAH metabolism without any interference from environmental exposure by giving a precise dose to the subjects. Non-deuterated Phe metabolites detected in human urine can be influenced by extraneous environmental factors such as dietary intake. Hydroxyphenanthrenes (mainly 2-OHPhe and 3-OHPhe) and some other metabolites like epoxides, orthoquinones and glucuronide conjugates have been identified from cow/goat milk, marine products (fish and crab) and egg yolk

(21–23). Consumption of these animal products could lower the credibility of results of non-deuterated Phe in human *in vivo* metabolism analysis.

PAH such as BaP requires metabolic activation to exert their carcinogenic effects. Thus, cancer risk may be related to the metabolic imbalance between bioactivation and detoxification pathways. Differences in individual susceptibility to the adverse effects of carcinogenic PAHs may, in part, be due to differences in the metabolic enzymes, especially CYP enzymes. The CYP1 enzymes—CYP1A1, CYP1A2 and CYP1B1—are responsible for both metabolically activating and detoxifying BaP (24–27). This raises the question: is the induction of CYP1 enzymes more important in bioactivation or detoxification of carcinogenic PAH in humans? Previously, animal research and human *in vitro* studies produced inconsistent results. The Ah receptor-mediated induction of CYP1A1 was found to be related to both the enhancement and suppression of BaP-DNA adduct formation in mouse/rat organs (6,28–33) and human cells (34–39). The biggest limitation of these studies is that they cannot precisely simulate the complex situation *in vivo* in humans. Thus, these previous paradoxical results can hardly be used in extrapolation to describe and predict the authentic role of CYP1 enzymes in PAH carcinogenesis and cancer susceptibility in humans due to the lack of direct evidence from human *in vivo* research. Although the relationship between CYP1 enzyme induction and increased cancer risk has been proposed (26,40,41), no previous studies have precisely examined the consequences of induction of CYP1A1, 1A2 and 1B1 with respect to bioactivation versus detoxification of carcinogenic PAH in humans. From the metabolic perspective, the relationship between induction of CYP1 enzymes and human cancer risk was still unclear.

In this study, we characterized urinary [D<sub>10</sub>]Phe metabolite profiles in cigarette smokers and non-smokers. Cigarette smoke is a strong inducer of CYP1 enzymes. It has been reported that the CYP1A1 enzyme activity is induced by cigarette smoke up to 100-fold in human lung (19,42,43). The inducing effects of smoking on CYP1A2 and CYP1B1 activity in humans have also been confirmed (20,44). We found that bioactivation was the dominant pathway induced in the metabolism of Phe in smokers. PheT, the bioactivation end product of Phe, was significantly and positively correlated with BaP-tetraol in our previous studies (45) and was significantly related to increased lung cancer risk in humans (46). In this study, smokers not only had a significantly higher proportion of activation products ([D<sub>10</sub>]1,2-PheD and [D<sub>10</sub>]PheT, P < 0.01), but also a significantly lower proportion of detoxification products (total [D<sub>9</sub>]OHPhe, P < 0.01) with a significant difference in the whole distribution pattern compared with non-smokers (P < 0.01; Figure 2). This is the first solid human *in vivo* evidence to demonstrate that induction of CYP1 enzymes by cigarette smoke contribute significantly to the metabolic activation of PAH. As critical drivers in the bioactivation pathway, variations of CYP1 enzyme activity must play an important role in individual differences in carcinogenic PAH metabolism and cancer susceptibility.



**Figure 3.** Histograms of the Phe bioactivation ratios (PBR: ratios of  $[\text{D}_{10}]\text{PheT}$  to total  $[\text{D}_9]\text{OHPhe}$ ) in smokers and non-smokers. \* $P < 0.01$ ; \*\* $P < 0.001$  compared with non-smokers. Mann-Whitney  $U$ -test was used since the data were not normally distributed.

These variations can be summarized in the PBR, the ratio of the metabolic activation end product  $[\text{D}_{10}]\text{PheT}$  to the detoxification end products total  $[\text{D}_9]\text{OHPhe}$ . These two end products succinctly summarize the comprehensive effects of P450s on bioactivation and detoxification pathways. The PBR could be characteristic of a given individual's bioactivation capability of carcinogenic PAH. We observed a large interindividual variation among the subjects in their PBR: the PBR varied by as much as 23-fold among 1  $\mu\text{g}$  dose non-smokers, 32-fold among 1  $\mu\text{g}$  dose current smokers and 271-fold among 10  $\mu\text{g}$  dose current smokers. This approach allows us to directly and accurately assess *in vivo* interindividual variation in PAH metabolism in humans. The PBR also represents the relative dominance of one pathway over the other and may be modifiable by the induction of CYP1 enzymes. In this study, PBR were 2.28- ( $P < 0.01$ ) to 4.78

( $P < 0.001$ )-fold higher in smokers than that in non-smokers (Table 3 and Figure 3). Our previous studies of non-deuterated Phe metabolites have shown that this ratio is induced by cigarette smoking, consistent with the present results. However, as noted above, those results might have been compromised by environmental or dietary exposure to Phe metabolites (21–23). We propose that the PBR can be used to identify cigarette smokers and ex-smokers who are at high risk for lung cancer and other tobacco-related cancers. The PBR can be readily determined by analysis of a urine sample 6 h after a dose of 1  $\mu\text{g}$   $[\text{D}_{10}]\text{Phe}$  as in this study. This could be particularly important for ex-smokers who remain at high risk for lung cancer for years after quitting. A high PBR would signal immediate participation in a lung cancer screening program to detect cancer at a relatively early stage while it is still amenable to treatment.

The formation of *ortho*-quinones and reactive oxygen species has been proposed as an important bioactivation pathway of carcinogenic PAH such as BaP (47,48). BaP-7,8-dione can not only generate reactive oxygen species resulting in oxidative DNA damage but can also react with DNA to form stable B[a]P-7,8-dione-DNA adducts (47,49,50). The results obtained upon analysis of the urine of 50 subjects in this study indicate that total [D<sub>8</sub>]PheQs make up only a small percentage (<1%) of [D<sub>10</sub>]Phe metabolites. In contrast, our previous study found that unlabeled total PheQs can reach 14.5% of unlabeled Phe metabolites in humans (10). These results indicate that most of the PAH *ortho*-quinones in human urine originate from external exposures in the environment but not internal metabolism of parent PAH. Thus, [D<sub>8</sub>]PheQs are not further discussed in the human bioactivation studies described here.

One limitation of the study was that, while we confirmed current cigarette smoking or non-smoking status by analysis of urinary cotinine, we could not positively identify former smokers and had to rely on questionnaire data. It is possible that some subjects identified as non-smokers were actually former smokers, and that the induction of cytochrome P450s may have persisted into the time frame of the current study. This would tend to decrease our observed effects by potentially increasing the PBR in non-smokers. The persistence of induction of CYP enzymes by smoking is unclear and requires further study.

## Conclusion

The results presented in this study demonstrate that cigarette smoking increases the metabolic activation of Phe, a structurally representative PAH, via the well-established induction of CYP1 enzymes. With solid human *in vivo* evidence, this study for the first time revealed that cigarette smoking raises a dual risk of carcinogenic PAH in humans: both higher exposure and higher metabolic activation. The results of this study significantly improve our understanding of PAH metabolism in cigarette smokers and are applicable to investigations of lung cancer susceptibility and prevention, potentially through the simple measurement of the urinary PBR.

## Abbreviations

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