



Article Impact of Environmental Indicators on the COVID-19 Pandemic in Delhi, India

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Abstract: Currently, there is a massive debate on whether meteorological and air quality parameters play a crucial role in the transmission of COVID-19 across the globe. With this background, this study aims to evaluate the impact of air pollutants (PM2.5, PM10, CO, NO, NO2, and O3) and meteorological parameters (temperature, humidity, wind speed, and rainfall) on the spread and mortality due to the COVID-19 outbreak in Delhi from 14 Mar 2020 to 3 May 2021. The Spearman's rank correlation method employed on secondary data shows a significant correlation between the COVID-19 incidences and the PM_{2.5}, PM₁₀, CO, NO, NO₂, and O₃ concentrations. Amongst the four meteorological parameters, temperature is strongly correlated with COVID-19 infections and deaths during the three phases, i.e., pre-lockdown (14 March 2020 to 24 March 2020) (r = 0.79), lockdown (25 March 2020 to 31 May 2020) (r = 0.87), and unlock (1 June 2020 to 3 May 2021) (r = -0.75), explaining the variability of about 20-30% in the lockdown period and 18-19% in the unlock period. NO₂ explained the maximum variability of 10% and 7% in the total confirmed cases and deaths among the air pollutants, respectively. A generalized linear model could explain 80% and 71% of the variability in confirmed cases and deaths during the lockdown and 82% and 81% variability in the unlock phase, respectively. These findings suggest that these factors may contribute to the transmission of the COVID-19 and its associated deaths. The study results would enhance the ongoing research related to the influence of environmental factors. They would be helpful for policymakers in managing the outbreak of COVID-19 in Delhi, India.

Keywords: air pollutants; COVID-19; Delhi; humidity; rainfall; temperature; wind speed

1. Introduction

The outbreak of the novel coronavirus (COVID-19), associated with Severe Acute Respiratory Syndrome Coronavirus 2 (SARS-CoV-2), began in December 2019 in Wuhan, China. The disease has been affirmed to have human-to-human transmissibility [1], which elevated colossal attention in China and worldwide [2]. Due to its devastating effects worldwide, COVID-19 was declared a global pandemic by the World Health Organization [3]. In India, the first case of COVID-19 was reported on 30 January 2020, in Kerala. The instances in India started increasing at a higher pace, more exponentially with each passing day. Until the beginning of May 2021, there were more than 20.66 million confirmed cases of COVID-19 and around 226,000 deaths in India, reported on the WHO official portal for



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Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). COVID-19 [4]. Delhi, the largest commercial city of North India, is one of the worst affected cities due to COVID-19 in India, with over 1.21 million cumulative infected cases and more than 17,000 casualties as of 3 May 2021 [5]. The first case of the COVID-19 pandemic in the national capital was reported on 2 March 2020, when an older person from East Delhi with a travel history to Italy tested positive for COVID-19 [6].

Delhi is the ninth most populated metro city globally, with a population of 20 million. Located in Northern India, the national capital territory covers an area of 1484 km², making it the largest city in terms of area in the country. Delhi possesses a dry winter and humid subtropical climate, bordering on a hot semi-arid climate. The average annual rainfall is approximately 886 mm, mostly falling out during the monsoon months of July through August. The maximum and minimum temperature in the city ranges between 2 °C and 47 °C, except for -2.2 °C and 48.4 °C, which are the lowest and highest temperatures, respectively [7]. It is one of the most polluted cities in the country, with the most increased volumes of airborne particulate PM_{2.5}, which is considered the most hazardous particulate to health, with 153 micrograms [8]. The rising levels of air pollution have significantly increased lung-related diseases, especially asthma and lung cancer, among children and women in Delhi [9,10]. COVID-19 is also a respiratory disease directly transmitted by close contact through respiratory droplets emitted from an infected person [11].

Recently, several studies from various parts of the world reported that meteorological/weather conditions have a crucial role in the spread of COVID-19 [1,12–14]. A study by Zoran et al. (2021) provides evidence that climate parameters, such as temperature, humidity, and wind speed, can trigger the transmission of COVID-19 in Spain [15]. Tosepu et al. (2020) reported that the average temperature is positively correlated with the COVID-19 pandemic in Jakarta, Indonesia [16]. Similarly, an investigation in Singapore revealed that temperature and dew point positively impact daily and cumulative COVID-19 cases [17]. Bolaño-Ortiz et al. (2020) reported enhanced airborne transmission by wind speed due to a correlational existence [18]. In another study, Şahin (2020) analyzed the impact of population and weather parameters on COVID-19 in Turkey [19], and reported a strong correlation among them. An association between the transmission of COVID-19 and environmental factors was also demonstrated by Muhammad et al. (2020) in New York, USA [20]. In India, studies from the states of Maharashtra and Punjab discussed the role of environmental factors in the spread of COVID-19 during different phases of the pandemic, reporting significant correlations between environmental variables and COVID-19 cases [21,22].

There is still an insufficiency of data in several COVID-19 hotspots in Delhi. Since the research related to environmental indicators in COVID-19 is still contradictory, this study will provide rigorous insight to understand this relationship effectively. More vigorous studies must understand these factors to improve forecasting models that can be effective for public health measures and examine the COVID-19 pandemic in Delhi, India, and elsewhere. Thus, the study was conducted in Delhi to analyze the impact of air pollutants (PM_{2.5}, PM₁₀, CO, NO, NO₂, and O₃) and meteorological parameters (temperature, humidity, wind speed, and rainfall) on new infections and mortality due to the COVID-19 outbreak.

2. Materials and Methods

2.1. Data Collection

The analysis was carried out in the national capital of India, Delhi, as displayed in Figure 1a. The data for environmental indicators were extracted from the Ministry of Environment, Forest, and Climate Change for the Government of India. It comprised concentrations of $PM_{2.5}$ (µg/m³), PM_{10} (µg/m³), NO (µg/m³), NO₂ (µg/m³), CO (mg/m³), ozone (µg/m³), temperature (°C), humidity (%), wind speed (m/s), and rainfall (mm) from 14 March 2020 to 3 May 2021. These parameters' data were collected for eight real-time air quality monitoring stations in Delhi, specifically under the Central Pollution Control Board (CPCB) [23]. The substations were selected so that they geographically became an effective representative of the national capital, covering North, South, East, West, and Central Delhi (Figure 1b). While analyzing these parameters, the average for the eight substations was considered for effective representation. The time series data on COVID-19 cumulative infections and deaths in Delhi were taken for the same period, i.e., 14 March 2020 to 3 May 2021, using a reliable, crowdsourced database repository [5]. The study period starting from 14 March 2020 was considered based on data availability and the need to incorporate the different phases of COVID-19 lockdown. Furthermore, the data of all the environmental and climate indicators, as well as COVID-19 occurrences, were classified into different phases of COVID-19 based on restrictive nationwide policies, such as the pre-lockdown phase (14–24 March 2020), lockdown phase (25 March–31 May 2020), and the unlock phase (1 June 2020–3 May 2021).



Figure 1. (a) Map showing the location of the study area of Delhi, India. (b) Map showing the outline of the study area and the location of eight air quality and weather monitoring stations in Delhi, India.

2.2. Spearman's Correlation Test

Due to the lack of normality in the dataset, we employed Spearman's rank correlation for studying the relationship between air pollutants, climate factors, and the impact of COVID-19 in Delhi during 14 March 2020 through 3 May 2021. A correlation matrix was calculated to describe the relationship between all the parameters and other components. The mathematical formula for Spearman's correlation coefficient is given by:

$$r_{\rm S} = 1 - 6 \; \frac{\sum_{i=1}^{n} d_i}{n(n^2 - 1)}$$

where n is the number of observations and d_i is the difference of the rank between two variables.

2.3. Generalized Linear Model

A generalized linear model was employed to analyze the extent of variability by the various air pollutants and the meteorological parameters in COVID-19 cases and deaths. In real life, the assumptions of normality and constant variance are not satisfied by the dataset, hence, a simple linear regression model is challenging to apply. The GLM is a unification of both linear and non-linear regression models that incorporates non-normal response distributions. Estimates with a *p*-value < 0.05 were observed to be significant during the analysis. R^2 values were used as a measure of variability that a

model explains. Furthermore, adjusted R² values were calculated for the complete models using the formula:

Adjusted
$$R^2 = 1 - (1 - R^2) * \frac{n - 1}{n - k - 1}$$

where n =sample size and k =number of independent variables.

The location maps in Figure 1a, b have been prepared using Tableau version 2020.4.5. The statistical analysis in the paper was performed using Microsoft Excel and R version 4.0.2.

3. Results

Daily COVID-19 infections and deceased cases are presented in Figure 2. This displays a speedy growth of COVID-19 in Delhi, starting from seven confirmed cases on 14 March 2020 until 24 March i.e., the beginning of lockdown, and then rapidly rising to 19,844 cumulative cases by the end of lockdown. Similarly, the death toll also increased rapidly. By the first week of May 2021, there were as high as 1.21 million cumulative infections and a total of 17,414 deaths.



Figure 2. Impact of COVID-19 in Delhi, India from 14 March 2020 to 3 May 2021.

The concentrations of PM_{2.5}, PM₁₀, NO, NO₂, CO, and O₃, taken from 14 March 2020 to 3 May 2021, are presented in Figure 3a,b. Additionally, the dependence between the air pollutants and COVID-19 incidence and mortality is examined using the Spearman rank correlation. Matrices for the three phases, i.e., pre-lockdown, lockdown, and unlock, are displayed in Tables 1–3. The correlation coefficients of very few pairs of parameters turned out to be significant before the lockdown (Table 1). The coefficient in the lockdown phase between the cumulative cases and PM_{2.5}, (r = 0.60), PM₁₀ (r = 0.62), NO₂ (r = 0.65), and CO (r = 0.53) share a high positive correlation with a *p*-value < 0.05. These parameters are also positively correlated with the cumulative deaths during the lockdown period (Table 2). The findings suggested a significant correlation between COVID-19 cases (and deaths) and the parameters PM_{2.5} (r = 0.57), PM₁₀ (r = 0.57), NO (r = 0.48), NO₂ (r = -0.25), CO (r = 0.11), and O₃ (r = -0.19) in the unlock phase (Table 3).

| Pre-Lockdown | $PM_{2.5}$ | PM ₁₀ | NO | NO_2 | CO | Ozone | Temp | RH | WS | RF | Confirmed Cases |
|-------------------|------------|------------------|--------|--------|---------|---------|--------|---------|-------|-------|-----------------|
| PM _{2.5} | 1 | | | | | | | | | | |
| PM_{10} | 0.95 * | 1 | | | | | | | | | |
| NO | 0.64 * | 0.75 * | 1 | | | | | | | | |
| NO ₂ | 0.55 | 0.69 * | 0.87 * | 1 | | | | | | | |
| СО | 0.52 | 0.41 | 0.42 | 0.26 | 1 | | | | | | |
| Ozone | 0.15 | 0.06 | 0.17 | 0.02 | 0.91 * | 1 | | | | | |
| Temp | 0.51 | 0.38 | -0.08 | -0.22 | 0.33 | 0.21 | 1 | | | | |
| RH | -0.15 | -0.05 | 0.39 | 0.44 | -0.39 | -0.45 | -0.53 | 1 | | | |
| WS | -0.16 | -0.21 | -0.28 | -0.06 | -0.64 * | -0.73 * | -0.28 | 0.25 | 1 | | |
| RF | -0.35 | -0.35 | -0.19 | -0.29 | -0.49 | -0.39 | 0.10 | 0.54 | 0.10 | 1 | |
| Confirmed Cases | 0.08 | -0.07 | -0.47 | -0.60 | 0.33 | 0.43 | 0.79 * | -0.85 * | -0.33 | -0.08 | 1 |

Table 1. Spearman's correlation coefficient during pre-lockdown i.e., between 14–24 March 2020.

* p < 0.05.

Table 2. Spearman's correlation coefficient during lockdown i.e., between 25 March–31 May 2020.

| Lockdown | PM _{2.5} | PM ₁₀ | NO | NO ₂ | СО | Ozone | Temp | RH | WS | RF | Confirmed Cases | Confirmed Deaths |
|-------------------|-------------------|------------------|---------|-----------------|---------|--------|---------|---------|--------|-------|------------------------|-------------------------|
| PM _{2.5} | 1 | | | | | | | | | | | |
| PM_{10} | 0.92 * | 1 | | | | | | | | | | |
| NO | 0.58 * | 0.48 * | 1 | | | | | | | | | |
| NO ₂ | 0.83 * | 0.74 * | 0.77 * | 1 | | | | | | | | |
| CO | 0.40 * | 0.43 * | -0.17 | 0.24 * | 1 | | | | | | | |
| Ozone | 0.27 * | 0.31 * | -0.27 * | 0.12 | 0.98 * | 1 | | | | | | |
| Temp | 0.64 * | 0.68 * | 0.40 * | 0.67 * | 0.42 * | 0.34 * | 1 | | | | | |
| RH | -0.61 * | -0.61 * | -0.43 * | -0.62 * | -0.28 * | -0.19 | -0.37 * | 1 | | | | |
| WS | -0.39 * | -0.31 * | -0.34 * | -0.36 * | 0.08 | 0.16 | 0.05 | 0.31 * | 1 | | | |
| RF | -0.35 * | -0.32 * | -0.05 | -0.25 * | -0.17 | -0.09 | -0.08 | 0.46 * | 0.35 * | 1 | | |
| Confirmed Cases | 0.60 * | 0.62 * | 0.31 * | 0.65 * | 0.53 * | 0.46 * | 0.87 * | -0.31 * | 0.20 | -0.02 | 1 | |
| Confirmed Deaths | 0.60 * | 0.62 * | 0.31 * | 0.65 * | 0.53 * | 0.46 * | 0.87 * | -0.31 * | 0.20 | -0.02 | 0.99 * | 1 |

* *p* < 0.05.

| Unlock | PM _{2.5} | PM ₁₀ | NO | NO ₂ | CO | Ozone | Temp | RH | WS | RF | Confirmed Cases | Confirmed Deaths |
|-------------------|-------------------|------------------|---------|-----------------|---------|---------|---------|---------|--------|---------|------------------------|-------------------------|
| PM _{2.5} | 1 | | | | | | | | | | | |
| PM_{10} | 0.98 * | 1 | | | | | | | | | | |
| NO | 0.81 * | 0.81 * | 1 | | | | | | | | | |
| NO ₂ | 0.86 * | 0.85 * | 0.94 * | 1 | | | | | | | | |
| CO | 0.38 * | 0.40 * | 0.18 * | 0.25 * | 1 | | | | | | | |
| Ozone | -0.04 | -0.03 | -0.22 * | -0.16 * | 0.84 * | 1 | | | | | | |
| Temp | -0.76 * | -0.71 * | -0.73 * | -0.80 * | -0.10 | 0.22 * | 1 | | | | | |
| RH | -0.27 * | -0.35 * | -0.10 | -0.18 * | -0.53 * | -0.42 * | 0.09 | 1 | | | | |
| WS | -0.46 * | -0.47 * | -0.54 * | -0.52 * | -0.26 * | -0.04 | 0.20 * | 0.00 | 1 | | | |
| RF | -0.54 * | 0.56 * | -0.39 * | -0.45 * | -0.39 * | -0.21 * | 0.41 * | 0.54 * | 0.31 * | 1 | | |
| Confirmed Cases | 0.57 * | 0.57 * | 0.48 * | -0.25 * | 0.11 * | -0.19 * | -0.65 * | -0.43 * | 0.03 | -0.44 * | 1 | |
| Confirmed Deaths | 0.57 * | 0.57 * | 0.48 * | 0.54 * | 0.11 * | -0.19 * | -0.65 * | -0.43 * | 0.03 | -0.44 * | 0.99 * | 1 |

Table 3. Spearman's correlation coefficient after lockdown i.e., between 1 June 2020 to 3 May 2021.

* p < 0.05.



Figure 3. (a) $PM_{2.5}$ and PM_{10} concentration ($\mu g/m^3$); (b) concentrations of O_3 , NO, NO₂ (ug/m^3), and CO (mg/m^3) from 14 March 2020 to 3 May 2021 in Delhi. The two dotted lines separate three lockdown periods: pre-lockdown, lockdown, and unlock, consecutively.

The variation in meteorological parameters during the pre-lockdown, lockdown, and unlock phases is shown in Figure 4a,b. In Delhi, summer starts in early April and peaks in late May or early June. Average temperatures near 38 °C are followed by monsoons that last until mid-September. Winter begins in November and peaks in January with an average temperature of around 6-7 °C and ends by the first week of March. Humidity is at its maximum during the monsoons in Delhi, and otherwise remains low to moderate [24]. Spearman correlation results in Table 1 show that only temperature and humidity were the significantly correlated variables with confirmed cases (r = 0.79; -0.85) in the pre-lockdown period. Temperature was strongly correlated during the lockdown (r = 0.87) and unlock period (r = -0.65) with the COVID-19 infections and deaths (Tables 2 and 3). Temperature was also significantly correlated with the air pollutants, such as $PM_{2.5}$, PM_{10} , NO, and NO₂, during the lockdown and unlock periods. The factors of wind speed and rainfall did not seem to have a good correlation with COVID-19 in this study. The correlation between relative humidity and COVID-19 cases and deaths was -0.85 in the pre-lockdown period, -0.31 in the lockdown period, and -0.43 in the unlock period, all significant for a p-value < 0.05. Figure 5 displays that the average concentrations of all the air pollutants showed the following trends: $PM_{2.5}$ in lockdown < $PM_{2.5}$ in pre-lockdown < $PM_{2.5}$ in unlock; PM₁₀ in lockdown < PM₁₀ in pre-lockdown < PM₁₀ in unlock; NO in lockdown < NO in pre-lockdown < NO in unlock; NO₂ in lockdown < NO₂ in pre-lockdown < NO₂ in unlock; CO in unlock < CO in pre-lockdown < CO in lockdown; O_3 in unlock < O_3 in pre-lockdown < O₃ in lockdown.



Figure 4. (a) Temperature (°C) and humidity (RH) and (b) wind speed (WS) (m/s) and rainfall (RF) (mm) from 14 March 2020 to 3 May 2021 in Delhi. The two dotted lines separate three lockdown periods: pre-lockdown, lockdown, and unlock, consecutively.



Figure 5. Average concentrations' change of different parameters (during pre-lockdown, lockdown, and post-lockdown period) in Delhi.

Furthermore, to understand the variability by different predictors, the generalized linear model was constructed for the lockdown and unlock periods. The results for these models are presented in Tables 4 and 5, respectively. The GLM model for pre-lockdown was omitted from the analysis because of an insufficient data quality and higher insignificance in correlations. For modeling, total confirmed cases and total deaths were dependent variables, with the other air pollutants and the climate parameters as independent variables. NO₂, among other pollutants, explained maximum variability in total cases (10%) and deaths (7%) for the lockdown period. Other air pollutants did not seem to contribute much

to the transmission of the virus. For instance, $PM_{2.5}$ and PM_{10} explained only 1% of the variability in the confirmed cases. At the same time, PM_{10} failed to explain any variation in total deaths during the lockdown. NO explained only 2% variation in the infections and deaths in the lockdown, and around 2% variability in the unlock period. Unlike the lockdown model, O₃ explained a total variability of 16% in total cases and deaths, followed by CO (13%; 12%), $PM_{2.5}$ (2%), and PM_{10} (1%) during the unlock phase. The complete models for confirmed cases in the lockdown showed an adjusted R² value of 80%, whereas, for the deaths, it was 71%.

| | Confirm | ed Cases (Loc | kdown) | Confirmed Deaths (Lockdown) | | | | | |
|-------------------|-------------------------|-----------------|----------------|---|-----------------|-----------------------|--|--|--|
| Parameter | R ² Adjusted | (Complete M | odel) = 0.80 | R ² Adjusted (Complete Model) = 0.71 | | | | | |
| | β | <i>p</i> -Value | R ² | β | <i>p</i> -Value | R ² | | | |
| PM _{2.5} | -90.48 | 0.07 | 0.01 | -2.10 | 0.09 | 0.02 | | | |
| PM_{10} | 23.14 | 0.25 | 0.01 | 0.12 | 0.80 | 0.00 | | | |
| NO | -781.93 | 0.01 * | 0.02 | -14.95 | 0.04 * | 0.02 | | | |
| NO ₂ | 806.80 | 0.00 * | 0.10 | 13.83 | 0.00 * | 0.07 | | | |
| CO | -6124.62 | 0.01 * | 0.02 | -87.26 | 0.16 | 0.01 | | | |
| O3 | 761.21 | 0.02 * | 0.02 | 8.93 | 0.26 | 0.01 | | | |
| TEMP | 1634.56 | 0.00 * | 0.23 | 39.01 | 0.00 * | 0.31 | | | |
| RH | 149.85 | 0.00 * | 0.04 | 2.50 | 0.02 * | 0.03 | | | |

Table 4. Variability in COVID-19 confirmed cases and deaths during the lockdown.

* *p* < 0.05.

Table 5. Variability in COVID-19 confirmed cases and deaths after the lockdown.

| | Confir | med Cases (U1 | nlock) | Confirmed Deaths (Unlock) | | | | | |
|-------------------|-------------------------|-----------------|----------------|---|-----------------|-----------------------|--|--|--|
| Parameter | R ² Adjusted | l (Complete M | odel) = 0.82 | R² Adjusted (Complete Model) = 0.81 | | | | | |
| | β | <i>p</i> -Value | R ² | β | <i>p</i> -Value | R ² | | | |
| PM _{2.5} | -2834.1 | 0.00 * | 0.02 | -49.00 | 0.00 * | 0.02 | | | |
| PM_{10} | 1519.2 | 0.00 * | 0.01 | 27.40 | 0.00 * | 0.01 | | | |
| NO | -5138.3 | 0.00 * | 0.02 | -75.45 | 0.00 * | 0.02 | | | |
| NO ₂ | -2613.3 | 0.01 * | 0.00 | -29.76 | 0.03 * | 0.00 | | | |
| CO | 347546.1 | 0.00 * | 0.13 | 4695.55 | 0.00 * | 0.12 | | | |
| O ₃ | -50344.1 | 0.00 * | 0.16 | -705.31 | 0.00 * | 0.16 | | | |
| TEMP | -35350.1 | 0.00 * | 0.18 | -496.73 | 0.00 * | 0.19 | | | |
| RH | -1577 | 0.00 * | 0.00 | -2.76 | 0.79 | 0.00 | | | |

* *p* < 0.05.

In contrast to this, the models in the unlock phase yielded an adjusted R² value of 82% and 81%, respectively, for total cases and deaths. Out of all the parameters, including air pollutants and meteorological factors, temperature significantly explained maximum variability for cumulative cases and deaths in both the lockdown (23%; 31%) and unlock periods (18%; 19%). Humidity, rainfall, and wind speed played no role in explaining the variability in COVID-19 transmission.

4. Discussion

4.1. Association of Air Pollutants with COVID-19 Cases and Deaths

Figure 3a,b show that most air pollutant levels were drastically reduced during the lockdown phase until the end of August. Although the air pollutants PM_{2.5}, PM₁₀, NO₂, and CO have shown a strong positive correlation with COVID-19 incidences, this may not be true on the ground, as, during the lockdown, all the pollutants were drastically reduced. The most significant reduction is seen in the concentrations of PM_{2.5}, PM₁₀, NO₂, NO, and CO. A similar observation was also reported from other megacities in India [25–27] and elsewhere [28,29].

PM_{2.5}, PM₁₀, NO₂, and O₃ concentrations increased drastically after the lockdown. Exposure to such air pollutants is harmful to the respiratory and cardiovascular systems in humans [30]. COVID-19 infection is related to the respiratory system. Vulnerability to such pollutants would increase the risk of deaths due to COVID-19. According to Yamada et al. (2020), an increase of 1% in long-term exposure to $PM_{2.5}$ results in a 5.7% increase in COVID-19 mortality [31]. Similar results suggested by Dales et al. (2020) show a significant association between increased PM_{2.5} and NO₂ levels and daily COVID-19 deaths [32]. However, in an ideal situation, the escalation in pollutant levels would have correlated COVID-19 infection and mortality and air pollutants more robustly than in the lockdown period. However, this was not the case on the ground. Even if the pollutant emission increased, it took a little while for the situation to normalize in terms of people's movement for work or other affairs, so the correlation levels with COVID-19 dipped instead of rose. Moreover, during the lockdown, the COVID-19 infections and deaths were positively correlated with $PM_{2.5}$ and PM_{10} (*p*-value < 0.05). However, this exposure is not necessarily related to COVID-19 conditions [31]. The increase in COVID-19 cases in Delhi might result from more underlying factors. For instance, a mass migration of people from city centers to hometown and rural areas was caused by excessive job loss and fear of lockdowns. Furthermore, the effect of unfavorable meteorological conditions needs more research [33,34].

The average concentration plot of different air pollutants (Figure 5) shows that O_3 and CO concentrations also increased in the lockdown period in Delhi. Zhao et al. (2020) reported an increase of 47% in O_3 concentrations during the lockdown period in mainland China [35]. Similar reporting in O_3 concentrations was observed in many European cities [36]. Ozone production is dependent on various factors. The anthropogenic emissions and volatile organic compounds (VOCs) are the primary precursors for O_3 generation. In addition to these pollutants, the meteorological parameters also play an essential role in the production of O_3 . Advection of warm and polluted air masses can also raise the near-surface O_3 concentrations [37–39].

4.2. Relationship with the Meteorological Variables

Climate parameters, such as temperature, humidity, and wind speed, are reported to be vital factors in the transmission of SARS-CoV2 [15]. Ma et al. (2020) reported a significantly positive relation between daily temperature and deaths due to COVID-19, and a negative correlation for relative humidity [14]. A similar result is obtained in this study. COVID-19 cases and deaths in the lockdown period positively correlate with temperature and negatively correlate with humidity throughout the study period. In low humidity, the moisture in the exhaled bioaerosols evaporates rapidly. It forms droplet nuclei that may remain in the air for a more extended period, facilitating the increased pathogen transmission [40].

Low humidity can reduce the airway cilia cells' ability to remove virus particles, thereby exposing the host to the virus [41]. These associations indicate that the human body is at a higher risk of infection by SARS-CoV-2 in high temperature and low humidity environments. However, various studies suggest that the weather variables (especially temperature) seem to have a more negligible effect on the transmission of COVID-19. A study reports no correlation between temperature and humidity with COVID-19 incidences [42]. At the same time, others suggest that there is no supporting evidence that the COVID-19 transmission will decline in warm temperatures [43].

4.3. Determining Factors of COVID-19 Cases and Deaths

The GLM model findings explain that the role of particulate matter ($PM_{2.5}$ and PM_{10}) or other pollutants, such as NO, CO, and O₃, in the transmission of the virus is quite negligible in the lockdown period. CO and ozone seem to have contributed quite well to the model with confirmed cases and deaths in the unlock period by explaining maximum variability among all the other air pollutants. The pessimistic estimate values for O₃ in Table 5 suggest that with an increase in O_3 concentration, the number of confirmed cases and deaths decreases. It may be due to the virucidal act of the ozone on the host defense. It has previously been reported that O_3 is hugely influential in disinfection and sterilization against many respiratory infections, like influenza and SARS-CoV-1 viruses [44]. The adjusted R^2 values for all four models vary in a noticeable pattern. The values for models with confirmed cases in the lockdown and unlock period give a better adjusted R^2 value. It can be inferred from this that the GLM models provide a better estimate for the confirmed cases than for the deaths. The estimates of temperature in the model for confirmed cases in lockdown hold a positive relationship. After the lockdown, it is negative. The lockdown period comprises March, April, and May, typically known as the summer season in India. As per the estimates of the GLM model (Table 4), COVID-19 incidences seem to increase with increasing temperature considering this period. The temperature in India after May begins to vary (Figure 4a) with the onset of monsoon season (June-September) slightly, followed by autumn (October–November), and finally winter (December–February) to summers starting in March. Therefore, the estimates from this period in the GLM model (Table 5) suggest that the total infection and deaths from COVID-19 decrease in the said temperature variation. The parameters of wind speed and rainfall have the most minuscule contribution in explaining any variability or being correlated with total cases or deaths.

4.4. Limitations

This study does not include certain factors, such as individual human behavior, recent mass gatherings, or new COVID-19 variants, that might influence the spread of COVID-19 and its associated mortality. There is also the unavailability of data for these measures at the regional level, especially in India. Given the data, more complex research can be done to incorporate these measures and understand the extent of COVID-19 in Delhi. Another limitation is that the study used aggregated data rather than for the individual. The prime focus was to study the impact of environmental indicators under different restriction phases imposed to control COVID-19 transmission. Therefore, the findings are ecological.

5. Conclusions

The present study favors the argument that the COVID-19 lockdown has significantly helped clean the air environment. A reduction in the levels of PM_{2.5}, PM₁₀, NO, and NO₂ was observed during the lockdown. This is because of the stringent conditions for the movement of vehicles and other kinds of restrictions. However, the concentration of O₃ increased during the lockdown, which is possible because it was enforced during the warmer months in Delhi i.e., April and May. In warm temperatures, ozone pollution is expected to increase. Besides this, the positive correlation between PM_{2.5}, PM₁₀, and CO concentrations with COVID-19 incidences needs more research to understand its mechanism. This research finds that increasing temperature and decreasing humidity may increase daily new infections and deaths due to the coronavirus. At the same time, other meteorological and air pollutants exhibit no significant relation with the COVID-19 pandemic. The GLM models suggest that the temperature is statistically a substantial contributor to the spread of the virus, but this could also be related to seasonal variations in the Indian capital. It is also found that the air pollutants and meteorological parameters in this study could correlate better with the confirmed infections than deaths in Delhi. This could be because the deaths due to a disease would depend more on the health infrastructure and affordable medical facilities, primarily not on these factors. Considering the current situation of the COVID-19 pandemic in Delhi, policy measures, such as imposing lockdown restrictions and reducing contact rates, are suggested to be helpful to control the spread. Therefore, the impact of these factors may be considered in policy development to control the COVID-19 pandemic.

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References

- Wang, C.; Horby, P.W.; Hayden, F.G.; Gao, G.F. A novel coronavirus outbreak of global health concern. *Lancet* 2020, 395, 470–473. [CrossRef]
- Shrestha, N.; Shad, M.Y.; Ulvi, O.; Khan, M.H.; Karamehic-Muratovic, A.; Nguyen, U.-S.D.; Baghbanzadeh, M.; Wardrup, R.; Aghamohammadi, N.; Cervantes, D.; et al. The impact of COVID-19 on globalization. *One Health* 2020, *11*, 100180. [CrossRef]
- 3. World Health Organisation. WHO Director-General's Opening Remarks at the Media Briefing on COVID-19—11 March 2020. Available online: https://www.who.int/director-general/speeches/detail/who-director-general-s-opening-remarks-at-the-media-briefing-on-covid-19---11-march-2020 (accessed on 3 May 2021).
- 4. World Health Organisation. WHO Coronavirus (COVID-19) Dashboard. Available online: https://covid19.who.int/region/searo/country/in (accessed on 3 May 2021).
- 5. COVID-19 India. Available online: https://www.covid19india.org/ (accessed on 3 May 2021).
- COVID-19 Pandemic in Delhi. Wikipedia. 2021. Available online: https://en.wikipedia.org/wiki/COVID-19_pandemic_in_Delhi (accessed on 31 July 2021).
- 7. Climate of Delhi. Wikipedia. 2021. Available online: https://en.wikipedia.org/wiki/Climate_of_Delhi (accessed on 31 July 2021).
- Park, M. Top 20 Most Polluted Cities in the World—CNN.com. Retrieved 5 February 2021. Available online: http://edition.cnn. com/2014/05/08/world/asia/india-pollution-who/ (accessed on 2 May 2021).
- 9. Choudhury, S.R. Children in Delhi Have Lungs of Chain-Smokers! 22 February 2014. Available online: https://www.indiatoday. in/india/north/story/pollution-in-delhi-cng-children-in-delhi-182151-2014-02-22 (accessed on 10 June 2021).
- 10. Rehman, A. Pollution Increasing Lung Cancer in Indian Women. *DNA*. 3 February 2014. Available online: https://www.dnaindia. com/health/report-pollution-increasing-lung-cancer-in-indian-women-1959054 (accessed on 5 June 2021).
- Edwards, E. How Does Coronavirus Spread? 24 January 2020. Available online: https://www.nbcnews.com/health/healthnews/how-does-new-coronavirus-spread-n1121856 (accessed on 5 June 2021).
- 12. Wang, M.; Jiang, A.; Gong, L.; Luo, L.; Guo, W.; Li, C.; Zheng, J.; Yang, B.; Zeng, J.; Chen, Y.; et al. Temperature Significant Change Covid-19 Transmission in 429 Cities. *MedRxiv* 2020. [CrossRef]
- 13. Yuan, J.; Wu, Y.; Jing, W.; Liu, J.; Du, M.; Wang, Y.; Liu, M. Association between meteorological factors and daily new cases of COVID-19 in 188 countries: A time series analysis. *Sci. Total Environ.* **2021**, *780*, 146538. [CrossRef]
- 14. Ma, Y.; Zhao, Y.; Liu, J.; He, X.; Wang, B.; Fu, S.; Yan, J.; Niu, J.; Zhou, J.; Luo, B. Effects of temperature variation and humidity on the death of COVID-19 in Wuhan, China. *Sci. Total Environ.* **2020**, *724*, 138226. [CrossRef]
- 15. Zoran, M.A.; Savastru, R.S.; Savastru, D.M.; Tautan, M.N.; Baschir, L.A.; Tenciu, D.V. Exploring the linkage between seasonality of environmental factors and COVID-19 waves in Madrid, Spain. *Process. Saf. Environ. Prot.* 2021, 152, 583–600. [CrossRef]
- Tosepu, R.; Gunawan, J.; Effendy, D.S.; Lestari, H.; Bahar, H.; Asfian, P. Correlation between weather and Covid-19 pandemic in Jakarta, Indonesia. *Sci. Total Environ.* 2020, 725, 138436. [CrossRef] [PubMed]
- 17. Pani, S.K.; Lin, N.-H.; Babu, R. Association of COVID-19 pandemic with meteorological parameters over Singapore. *Sci. Total Environ.* **2020**, 740, 140112. [CrossRef] [PubMed]
- Bolaño-Ortiz, T.R.; Caicedo, Y.C.; Puliafito, S.E.; Ruggeri, M.F.; Bolaño-Diaz, S.; Pascual-Flores, R.; Saturno, J.; Ibarra-Espinosa, S.; Mayol-Bracero, O.L.; Torres-Delgado, E.; et al. Spread of SARS-CoV-2 through Latin America and the Caribbean region: A look from its economic conditions, climate and air pollution indicators. *Environ. Res.* 2020, 191, 109938. [CrossRef]
- 19. Şahin, M. Impact of weather on COVID-19 pandemic in Turkey. Sci. Total Environ. 2020, 728, 138810. [CrossRef]
- 20. Shakil, M.H.; Munim, Z.H.; Tasnia, M.; Sarowar, S. COVID-19 and the environment: A critical review and research agenda. *Sci. Total Environ.* **2020**, 745, 141022. [CrossRef]
- Sahoo, P.K.; Mangla, S.; Pathak, A.K.; Salãmao, G.N.; Sarkar, D. Pre-to-post lockdown impact on air quality and the role of environmental factors in spreading the COVID-19 cases—A study from a worst-hit state of India. *Int. J. Biometeorol.* 2021, 65, 205–222. [CrossRef]
- 22. Sahoo, P.K.; Chauhan, A.K.; Mangla, S.; Pathak, A.K.; Garg, V.K. COVID-19 pandemic: An outlook on its impact on air quality and its association with environmental variables in major cities of Punjab and Chandigarh, India. *Environ. Forensics* **2021**, *22*, 143–154. [CrossRef]
- 23. CPCB. Central Control Room for Air Quality Management—All India. Available online: https://app.cpcbccr.com/ccr/#/caaqm-dashboard-all/caaqm-landing/caaqm-comparison-data (accessed on 3 May 2021).
- 24. Overview of Seasonal Distribution. 2021. Available online: https://en.wikipedia.org/wiki/Climate_of_Delhi#Overview_of_seasonal_distribution (accessed on 31 July 2021).

- 25. Sharma, S.; Zhang, M.; Anshika; Gao, J.; Zhang, H.; Kota, S.H. Effect of restricted emissions during COVID-19 on air quality in India. *Sci. Total Environ.* **2020**, *728*, 138878. [CrossRef] [PubMed]
- Mahato, S.; Pal, S.; Ghosh, K.G. Effect of lockdown amid COVID-19 pandemic on air quality of the megacity Delhi, India. Sci. Total Environ. 2020, 730, 139086. [CrossRef]
- 27. Jain, S.; Sharma, T. Social and Travel Lockdown Impact Considering Coronavirus Disease (COVID-19) on Air Quality in Megacities of India: Present Benefits, Future Challenges and Way Forward. *Aerosol Air Qual. Res.* 2020, 20, 1222–1236. [CrossRef]
- 28. Xu, K.; Cui, K.; Young, L.H.; Wang, Y.F.; Hsieh, Y.K.; Wan, S.; Zhang, J. Air Quality Index, Indicatory Air Pol-Lutants and Impact of Covid-19 Event on the Air Quality near Central China. *Aerosol. Air Qual. Res.* **2020**, *20*, 1204–1221. [CrossRef]
- Krecl, P.; Targino, A.C.; Oukawa, G.; Junior, R.P.C. Drop in urban air pollution from COVID-19 pandemic: Policy implications for the megacity of São Paulo. *Environ. Pollut.* 2020, 265, 114883. [CrossRef]
- 30. American Heart Association. Long-Term Exposure to Low Levels of Air Pollution Increases Risk of Heart and Lung Disease. 22 February 2021. Available online: https://www.sciencedaily.com/releases/2021/02/210222082622.htm (accessed on 5 June 2021).
- Yamada, T.; Yamada, H.; Mani, M. Long-Term Causal Effects of PM2.5 Exposure on COVID-19 in India. SSRN Electron. J. 2020. [CrossRef]
- 32. Dales, R.; Blanco-Vidal, C.; Romero-Meza, R.; Schoen, S.; Lukina, A.; Cakmak, S. The association between air pollution and COVID-19 related mortality in Santiago, Chile: A daily time series analysis. *Environ. Res.* **2021**, *198*, 111284. [CrossRef]
- 33. Lee, J.; Mahmud, M.; Morduch, J.; Ravindran, S.; Shonchoy, A. *Migration, Externalities, and the Diffusion of COVID-19 in South Asia;* European Economic Association: Tessenderlo, Belgium, 2020.
- 34. Wang, P.; Chen, K.; Zhu, S.; Wang, P.; Zhang, H. Severe air pollution events not avoided by reduced anthropogenic activities during COVID-19 outbreak. *Resour. Conserv. Recycl.* 2020, 158, 104814. [CrossRef]
- Zhao, Y.; Zhang, K.; Xu, X.; Shen, H.; Zhu, X.; Zhang, Y.; Hu, Y.; Shen, G. Substantial Changes in Nitrogen Dioxide and Ozone after Excluding Meteorological Impacts during the COVID-19 Outbreak in Mainland China. *Environ. Sci. Technol. Lett.* 2020, 7, 402–408. [CrossRef]
- 36. Grange, S.K.; Lee, J.D.; Drysdale, W.S.; Lewis, A.C.; Hueglin, C.; Emmenegger, L.; Carslaw, D.C. COVID-19 Lock-Downs Highlight a Risk of Increasing Ozone Pollution in European Urban Areas. *Atmos. Chem. Phys. Discuss.* **2020**, 1–25. [CrossRef]
- Sun, W.; Hess, P.; Liu, C. The impact of meteorological persistence on the distribution and extremes of ozone. *Geophys. Res. Lett.* 2017, 44, 1545–1553. [CrossRef]
- 38. Garrido-Perez, J.M.; Ordóñez, C.; García-Herrera, R.; Schnell, J.L. The differing impact of air stagnation on summer ozone across Europe. *Atmospheric Environ.* **2019**, 219, 117062. [CrossRef]
- 39. Ordóñez, C.; Garrido-Perez, J.M.; García-Herrera, R. Early spring near-surface ozone in Europe during the COVID-19 shutdown: Meteorological effects outweigh emission changes. *Sci. Total Environ.* **2020**, 747, 141322. [CrossRef] [PubMed]
- 40. Lowen, A.C.; Mubareka, S.; Steel, J.; Palese, P. Influenza Virus Transmission Is Dependent on Relative Humidity and Temperature. *PLoS Pathog.* 2007, *3*, e1470–e1476. [CrossRef] [PubMed]
- Sun, Z.; Thilakavathy, K.; Kumar, S.S.; He, G.; Liu, S.V. Potential Factors Influencing Repeated SARS Outbreaks in China. Int. J. Environ. Res. Public Health 2020, 17, 1633. [CrossRef]
- 42. Yao, Y.; Pan, J.; Liu, Z.; Meng, X.; Wang, W.; Kan, H.; Wang, W. No association of COVID-19 transmission with temperature or UV radiation in Chinese cities. *Eur. Respir. J.* 2020, *55*, 2000517. [CrossRef]
- Xie, J.; Zhu, Y. Association between ambient temperature and COVID-19 infection in 122 cities from China. *Sci. Total Environ.* 2020, 724, 138201. [CrossRef]
- 44. Elvis, A.M.; Ekta, J.S. Ozone therapy: A clinical review. J. Nat. Sci. Biol. Med. 2011, 2, 66–70. [CrossRef] [PubMed]