## **Supporting Information File 1 (SI1)**

## **Legacy and emerging plasticizers and stabilizers in PVC floorings and implications for recycling**

Environmental, Science & Technology

<https://doi.org/10.1021/acs.est.3c04851>

Helene Wiesinger<sup>1\*</sup>, Christophe Bleuler<sup>2</sup>, Verena Christen<sup>3</sup>, Philippe Favreau<sup>2</sup>, Stefanie Hellweg<sup>1,4</sup>, Miriam Langer<sup>3,5</sup>, Roxane Pasquettaz<sup>2</sup>, Andreas Schönborn<sup>6</sup>, Zhanyun Wang<sup>1,4,7\*</sup>

- <sup>1</sup> Chair of Ecological Systems Design, Institute of Environmental Engineering, ETH Zürich, 8093 Zürich, Switzerland
- $2$  Service de l'air, du bruit et des rayonnements non ionisants (SABRA), Geneva cantonal office for the environment, 1205 Geneva, Switzerland
- <sup>3</sup> Institute for Ecopreneurship, School of Life Sciences, University of Applied Sciences and Arts Northwestern Switzerland, FHNW, 4132 Muttenz, Switzerland
- <sup>4</sup> National Centre of Competence in Research (NCCR) Catalysis, Institute of Environmental Engineering, ETH Zürich, 8093 Zürich, Switzerland
- <sup>5</sup> Eawag Swiss Federal Institute of Aquatic Science and Technology, 8600 Dübendorf
- <sup>6</sup> Institute of Natural Resource Sciences, ZHAW Zurich University of Applied Science, 8820 Wädenswil, Switzerland
- $7$  Empa Swiss Federal Laboratories for Materials Science and Technology, 9014 St. Gallen, Switzerland

\* Contact information: Helene Wiesinger: [wiesinger@ifu.baug.ethz.ch;](mailto:wiesinger@ifu.baug.ethz.ch) Zhanyun Wang: [zhanyun.wang@ifu.baug.ethz.ch](mailto:zhanyun.wang@ifu.baug.ethz.ch)

## **TABLE OF CONTENTS**



## **TABLE OF TABLES**





## **TABLE OF FIGURES**



## **S1 BACKGROUND**

<span id="page-5-1"></span><span id="page-5-0"></span>

**Table S 1: Legal and industrial developments in the European Union relevant for** *ortho-***phthalates and metal(loids) in plastic products.**



## <span id="page-7-0"></span>**S2 METODS**

## <span id="page-7-1"></span>**S2.1 Samples**

An overview for all samples and the analysis results for each sample are provided on Sheet S1 in the supplementary information File 2 (SI2).



<span id="page-7-2"></span>**Table S 2 Pictures of samples used for the GC-MS analysis** 



**Sample characteristics.** Color of the top layer or decorative sheet, hardness, number of layers, presence of a grey layer, and presence of a glass-fiber layer were assigned to each sample (see Sheet S1 in SI2), based on one author's perception.

- Color of the top layer sometimes contained patterns or multiple colors (see [Table S 2](#page-7-2) for pictures) and was simplified to fit these categories based on the "main" color: *black, grey, wood, orange/beige/brown, red, blue/green, white/transparent*
- Hardness was determined by bending the samples: *hard* sample cannot be bent by hand, *medium –* sample can be bent, but with significant resistance*, soft –* sample can be easily bent, with little to no resistance
- Number of layers was determined by the number of different colored layers that can be seen without further magnification.
- Presence of a grey layer was determined based on all layers (including the top layer / coloring) and was used a proxy for recycled material (based on personal communication with a large PVC flooring retailer)
- Presence of a glass-fiber layer was determined based on a "cracking" sound, when bending the sample (based on personal communication with a large PVC flooring retailer)

### <span id="page-10-0"></span>**S2.2 Materials**

All chemical standards that were used in this study can be found in [Table S 4,](#page-12-0) their position on the chemical space plot ( $logKow - logKaw$ ) can be found in [Figure S 1.](#page-10-1) Further information on the substances, including other identifiers, physical-chemical properties and experimental properties are provided in the Sheet S2 – Substances in SI2.



<span id="page-10-1"></span>**Figure S 1: Chemical space of the substances in the suspect list. For the bottom plots, the iso-concentration curves are calculated for equal volumes of each compartment (i.e. water, air and octanol are exactly the same volume). For the bottom plot the Most suspects have a very high Kow and a low Kaw, meaning they are mainly found in octanol-like environments, this is especially striking for DEHP, DiNP, DiDP and the alternative plasticizers. Phthalates are more likely to vaporize than phosphate based plasticizers due to their higher Kaw.**

Information on the employed reference materials for the chemical analyses and materials for the bioassays can be found in [Table S 3.](#page-11-0)



<span id="page-11-0"></span>

<span id="page-12-0"></span>

Substance name	Abbr.	<b>CASRN</b>	MW [g/mol] Workflow		<b>Supplier</b>
ortho-Phthalates					
Dimethyl phthalate	<b>DMP</b>	$131 - 11 - 3$	194.06	q,s	Sigma-Aldrich: 41320 (Lot:BCBZ7340)
Diethyl phthalate	<b>DEP</b>	84-66-2	222.09	q,s	Sigma-Aldrich: 53008 (Lot:BCBV6074)
Diallyl phthalate	<b>DAP</b>	131-17-9	246.09	q,s	Sigma-Aldrich: 36925 (Lot:BCBS8034V)
Diisobutyl phthalate	<b>DiBP</b>	84-69-5	278.15	q,s	CHEM Service: N-11589-1G (Lot:7047300)
Di-n-butyl phthalate	<b>DBP</b>	84-74-2	278.15	q,s	Sigma-Aldrich: 43540 (Lot:BCBV9941)
Bis(-2-methoxyethyl) phthalate	<b>DMEP</b>	117-82-8	282.11	q,s	CHEM Service: N-11304-500MG (Lot:6923600)
Diisopentyl phthalate	<b>DiPP</b>	$605 - 50 - 5$	306.18	q,s	CHEM Service: N-11620-500mg (Lot:7060400)
Isopentylpentyl phthalate	nPiPP	776297-69-9	306.18	q,s	CHEM Service: N13811-1G (Lot:6777200)
Di-n-pentyl phthalate	<b>DPP</b>	$131 - 18 - 0$	306.18	q,s	synthonix: P59310 (Lot:994)
Benzyl butyl phthalate	<b>BBP</b>	85-68-7	312.14	q,s	CHEM Service: N11360-1G (Lot:6894600)
Dicyclohexyl phthalate	<b>DCHP</b>	$84-61-7$	330.18	q,s	Aldrich: 306150 (Lot:09019JD)
Dihexyl phthalate	<b>DHP</b>	$84 - 75 - 3$	334.21	q,s	CHEM Service: N-11596-1G (Lot:6748400)
Di(2-ethylhexyl) phthalate	<b>DEHP</b>	117-81-7	390.28	q,s	CHEM Service: N11226-1G (Lot:6962500)
Dioctyl phthalate	<b>DNOP</b>	$117 - 84 - 0$	390.28	q,s	Sigma-Aldrich: 88173 (Lot:BCBV7232)
Diisononyl phthalate	<b>DiNP</b>	68515-48-0	418.31	q,s	Aldrich: 376663 (Lot:STBH9661)
Diisodecyl phthalate	<b>DiDP</b>	68515-49-1	446.34	q,s	Sigma-Aldrich: 80135 (Lot:BCCB0561)
Deuterated ortho-Phthalates					
LGC phthalates mixture	LGC	n.a.	n.a.	q	LGC: DRE-A50000576DI (Lot: -)
Deuterated(d4) diisobutyl phthalate	$DiBP-d4$	358730-88-8	282.18	$\mathbf q$	CHIRON: 3123.16-100-IO (Lot:8282)
Deuterated(d4) di-n-butyl phthalate	$DBP-d4$	93952-11-5	282.18	$\mathbf q$	CHEM Service: N-FD68-C-0.25G (Lot:7108100)
Deuterated(d4) diisopentyl phthalate	$DiPP-d4$	1346597-80-5	310.21	$\mathbf q$	CHEM Cruz: SC-498746 (Lot:B0818)
Deuterated(d4) dipentyl phthalate	$DPP-d4$	358730-89-9	310.21	$\mathbf q$	CHIRON: 2893.18-100-IO (Lot:13203)
Deuterated(d4) benzyl butyl phthalate	BBP-d4	93951-88-3	316.16	$\mathbf q$	CHEM Service: S-FD67S-1.2ML (Lot:7108200)
Deuterated(d4) dihexyl phthalate	$DHP-d4$	1015854-55-3	338.24	$\mathbf q$	CHIRON: 9367.20-100-IO (Lot:11572)
Deuterated(d4) di(2-ethylhexyl) phthalate		DEHP-d4 93951-87-2	394.30	q	CHEM Service: N-FD66-C-0.25G (Lot:7109200)

<span id="page-12-1"></span>Table S 4: Overview over employed standards and in which workflow they were used (q = quantification of  $ortho$ -phthalates, s = suspect screening). The table is sorted based **on the substance group and the molecular weight.** 

#### **[Table S 4](#page-12-1) - continued**



## <span id="page-14-0"></span>**S2.3 Chemical analysis**

# **Dissolution & Precipitation** Cut & weig Add s

#### **Preperation**

- Dissolution of PVC w/ THF
- $\ddot{\phantom{0}}$ Reprecipitation of PVC w/ ACN
- Nylon filtration 0.45 um

orhto-Phthalates Suspects



#### **GC-MS**

- Phthalates accredited method Internal standard calibration
	- Quality checks: procedural & solvent blanks. reference material & solutions
- Suspects adjusted method
- Suspect list (antioxidants, plasticisers, flame retardants)
- GC-MS: DB5, slow rise to high final temp



#### Bioassays

- Concentration w/ Syncore
	- Different bioassays
	- Cytotoxicity MTT
	- ROS generation
	- Endocrine disruption (YES/YAS)
	- Genotoxicity (Ames/UmuC)

<span id="page-14-1"></span>**Figure S 2: Sample preparation and overview for GC-MS** *ortho-***phthalate quantifcaiton, GC-MS suspect screening and testing of biological activities.**

The conducted analyses use different types of sample processing which may impact the detected substances or effects. While XRF and FTIR are surface-specific techniques and thus need minimal processing, they but only yield information on the layer on top or bottom of the sample. By contrast, GC-MS and the bioassays required the extraction of compounds of interest from the polymer matrix but and yield results from the entire sample, furthermore bioassays required solvent evaporation removing any very volatile substances. For the results from surface-specific techniques and extraction techniques to be comparable, it has to be assumed that substances need to be equally dispersed in the product.

#### <span id="page-15-0"></span>S2.3.1 ATR-FTIR

An ATR-FTIR spectrum was recorded for each sample and each side using a Thermo Scientific Nicolet<sup>TM</sup> iS spectrometer with iD7 ATR accessory (settings in [Table S 5\)](#page-15-1). All recorded spectra can be found in SI6-Rawdata-ATR-FTIR. No sample pre-treatment was made, apart from cleaning the sample surface with ethanol where necessary.

<span id="page-15-1"></span>**Table S 5:** ATR-FTIR settings used to determine the polymer type and the presence of *ortho-*phthalates.

parameter	value
Spectral Range $[\text{cm}^1]$	$500 - 4000$
Number of scans per sample	15

**Polymer type determination:** The polymer type was determined using the ThermoFischer OMNIC Spectra Polymer Package and selected reference spectra.<sup>12,13</sup> Non-PVC samples (n=35) were not analyzed further.

*ortho-***Phthalates screening**: The presence of *ortho-*phthalates was determined using the characteristic *ortho-phthalate peaks at 1600*cm<sup>-1</sup> and 1580cm<sup>-1</sup>, with an approximate sensitivity according to the instrument manufacturer of 0.1weight% of *ortho-*phthalates.14,15 The quality of the screening was compared to the GC-MS quantification *ortho-*phthalates [\(Table S 6\)](#page-15-2). FTIR screening detected the majority of samples containing *ortho-*phthalates (sensitivity: 78.2%), and almost all samples containing more than 0.1wt% of *ortho-*phthalates (sensitivity: 97.2%), without many false positives (specificity: 85.4% resp. 80.9%).



<span id="page-15-2"></span>**Table S 6: Quality of ATR-FTIR screening for** *ortho-***phthalates, using confusion matrices, sensitivity and specificity.**

#### <span id="page-16-0"></span>S2.3.2 XRF elemental composition

The elemental composition of the samples was determined using a handheld XRF (Thermo Scientific<sup>™</sup> Niton<sup>™</sup> XL3 Gold Analyzer) with a plastic calibration. No specific sample pretreatment was made, apart from cleaning the sample surface with ethanol where necessary. Each side was measured for at least 30 seconds with each filter. Correct operation and equipment calibration was checked using a certified reference material, ERM-EC681m – Polyethylene (high level): the measured concentrations had to be within 20% of the certified levels. The XRF's limits of detection (LODs) are calculated according to the instrument manufacturer's protocol, as three times the minimum standard deviation of the analyte.<sup>16</sup> The calculated LODs and the LODs reported by the manufacturer are provided in [Table S 7.](#page-16-1) Concentrations and standard deviations were noted as determined by the instrument's plastic calibration, the original fluorescence spectra were not exported.

<span id="page-16-1"></span>



#### <span id="page-17-0"></span>S2.3.3 GC-MS quantification of phthalates

The official laboratory protocol was in French and was translated by DEEPL to English for better understanding of the reader. Both original and English version can be found in SI3.

*Sample preparation*. The samples were cut into smaller pieces and weighed exactly (~750mg), dissolved in a weighed amount of *tetrahydrofuran* (*THF*, CASRN: 109-99-9, ~4.5 mL) using an ultrasound bath for about two hours at room temperature. After adding a weighted amount of *acetonitrile* (*ACN*, CASRN: 75-05-8, ~9 mL), samples were left in the fridge (4°C) overnight for the polymers to re-precipitate, and subsequently filtered using 0.45um nylon filters (BGB SF2503-2). The resulting filtrate had a known concentration of PVC at ~55 mg/mL. Subsequently, the extracts were diluted using THF to two levels (40-fold and 1600-fold dilution) and spiked with the internal standards. Sample preparation was conducted in batches due to spatial and temporal constraints. For each batch, a procedural blank containing no PVC sample and a PVC reference material (SPEX CRM-PVC001) with certified levels of *ortho-*phthalate was prepared analogously to the samples.

*GC-MS analysis.* Seventeen *ortho-*phthalates were used as standards for the calibration curves [\(Table S 4,](#page-12-0) [Table S 8\)](#page-20-0), and seven deuterated *ortho-*phthalates were used as internal standards [\(Table](#page-12-0)  [S 4\)](#page-12-0). The calibration curve spanned points from 0.05 to 10 mg/mL for most standards (for *DiNP* and *DiDP*, it spanned 0.5 to 100 mg/mL). GC-MS analysis was conducted in batches to ensure proper operation. Besides calibration solutions and samples, each batch also contained a blank solution, a reference solution with a known concentration, and the solutions from the procedural blank and the reference material.

Briefly, all analyses were carried out on an Agilent GC-MS system (GC: Agilent 7890A, MS: Agilent 5975C) in single ion mode (SIM) with splitless injections with internal standard calibration. The compounds were separated on a DB 5MS column using a temperature gradient from 80°C to 320°C. The injection was performed in pulsed splitless mode (injection volume: 2 uL), to a woolfilled liner (Topaz, 4mm Single Taper w/Wool) to avoid build-up of dissolved short-chain PVC on the column. The compounds were separated on a DB-5MS column (length: 15 m, inner diameter: 0.25 mm, film thickness: 0.1 mm), using Helium as a carrier gas (constant flow rate: 1mL/min). The oven temperature was set from 80°C (initial hold: 2 min) to 320°C with a changing temperature gradient (20°C/min until 200°C, 8°C/min until 320°C). The interface temperature was set to 280°C.

Ionization was done by electron impact (Ionisation energy: 70 eV, Ion source temperature: 250°C). The MS was set to SIM mode with several retention time windows, with a quantification- and a control-ion for each calibration standard or internal standard eluting within a given window (see [Table S 8](#page-20-0) for retention time windows and the target ions for each standard). To preserve the detector, a solvent delay was set to 3.5 minutes and the 1600-fold dilutions were run first, and 40 fold dilutions were only run if no signal was recorded.

The chromatogram of the *ortho-*phthalate standards can be found below [\(Figure S 3\)](#page-18-0). The retention time and calibration curves are listed in [Table S 8.](#page-20-0)



orhto-Phthalates

<span id="page-18-0"></span>**Figure S 3: Chromatogram of all** *ortho-***phthalate standards at ~5 μg/mL (DiNP and DiDP at ~50 μg/mL) using the**  *ortho-***phthalate quantification workflow.**

*Data Analysis.* The recorded spectra were analyzed in an automatic quantitative workflow using Agilent Masshunter (the raw data are available as Agilent files in SI7-Rawdata-GCMS-Phthalates). Quadratic and weighted calibration curves using the relative signal of calibration standard to internal standard were used (Weight: 1/x). The automatic integration, the calibration curves, and the quantification of blanks, reference solution, and reference material, were double-checked manually.

*Quality assurance and control (OA/OC).* Quality assurance and control were implemented throughout the process. Specific quality management practices in the accredited laboratory were observed, including regular maintenance of the GC-MS and replacement of liners and septa, daily tune evaluation to ensure correct MS detection, analysis of blanks and references solutions to ensure correct GC-MS operation, analysis of procedural blanks and certified reference material to ensure correct extraction, and manual checks of the automatic data analysis workflow. For the quality control, measured concentrations of the reference solutions and certified reference material had to be within 20% of the certified levels

#### *S2.3.3.1 Target compounds*

**Table S 8:** *ortho-***Phthaltate standards used in the** *ortho-***phthalate quantification workflow. Calibration curves were fitted to a linear model (Relative area = a0 + a1 \*concentration) and a quadratic model (Relative area = a0 + a1\* concentration + a2\*concentration<sup>2</sup> ). The calibration was redone for each run (the displayed calibration curves were extracted from the run "220517\_Batch7-int"). The table is sorted based on retention time (RT). RT= retention time, Q-ion = Quantification ion, C-ion = Control ion, CASRN = Chemical Abstract Service Registry Number, MW = Molecular weight of isotope, Dyn. Range = Dynamic range.** 

<span id="page-20-0"></span>





#### <span id="page-23-0"></span>S2.3.4 GC-MS suspect screening

*Sample preparation.* The same extraction procedure and dilutions as above were used (section [S2.3.3\)](#page-17-0), but without adding internal standards.

*Suspect substances and custom library:* Common alternative plasticizers and some antioxidants were used as suspect substances, for which analytical standards were used [\(Table S 4\)](#page-12-0). For example, *DEHT* [*Bis(2-ethylhexyl) terephthalate*, CASRN: 6422-86-2], *DINCH* [*Di(isononyl) cyclohexane-1,2-dicarboxylate*, CASRN: 166412-78-8], *DEHA* [*Bis(2-ethylhexyl) adipate*, CASRN: 103-23-1], *TPhP* [*Triphenyl Phosphate*, CASRN: 115-86-6], *TCP* [*Tricresyl phosphate*, CASRN: 78-32-0], and *Octicizer* (*2-Ethylhexyl diphenyl phosphate*, CASRN: 1241-94-7) were used.

The suitability of the extraction procedure for the suspects was ensured (1) by doing a simple solubility check in relevant solvent systems (THF, 1:2 THF:ACN) and (2) by spiking a PVC sample and following the regular extraction procedure. Semi-quantification (based on a signal calibration curve) and approximate detection limits (based on the lowest concentration with correct identification) were determined using a dilution series for the suspect standards (see below).

An Agilent custom library was created from their measured mass spectra at 5 mg/L (SI4). The chromatogram of the suspect standards can be found below ( $Figure S$  3), their retention time and mass spectra are in [Table S 8.](#page-20-0)

A dilution series (different dilutions) of the investigated standards were run to determine (a) approximate detection limits and (b) approximate calibration curves for the semi-quantification. This semi-quantification is more uncertain compared to the phthalate quantification as:

- (1) no internal standard was used and the MS response of a standards may depend on various external factors other than the concentration,
- (2) fewer concentration-response data points were collected for most standards as the aim of this was not proper quantification
- (3) the dilutions of our samples did not always fall within the dynamic range of our approximate calibration.

Overall, most standards had an approximate calibration slope (Area/concentration in μg/L) of 7.4  $\pm$  9.8 x 10<sup>5</sup> (1.3 x 10<sup>2</sup> – 3.3 x 10<sup>6</sup>), the detector response for *ortho*-phthalates was generally higher than for other standards [\(Figure S 6\)](#page-27-0).

*GC-MS analysis:* All measurements were conducted on a low-resolution Agilent GC-MS system (GC: Agilent 7890A, MS: Agilent 5975C) in scan mode. The injection was performed in splitless mode (injection volume: 2uL, injection temp: 140°C), to a wool-filled liner (Topaz, 4mm Single Taper w/Wool) to avoid build-up of dissolved short-chain PVC on the column. The compounds were separated on a DB-5MS column (length: 15 m, inner diameter: 0.25 mm, film thickness: 0.1 mm), using Helium as a carrier gas (flow rate: 1mL/min). The oven temperature was set from 40°C (initial hold: 2 min) to 300 $^{\circ}$ C (final hold: 20 min) with a change of 8 $^{\circ}$ C/min. The interface temperature was set to 280°C. Ionization was done by electron impact (Ionisation energy: 70 eV, Ion source temperature: 250°C). The MS was set to scan mode with a range of 30 - 800 amu (scan speed: 1.2 scan/s). To preserve the detector, the solvent delay was set to 8 minutes and the 1'600fold dilutions were run first, and 40-fold dilutions were only run if a low signal was recorded.



#### orhto-Phthalates

<span id="page-24-0"></span>**Figure S 4: Chromatogram of all** *ortho-***phthalate standards (PHT solution) using the suspect screening workflow**



<span id="page-25-0"></span>**Figure S 5: Chromatogram of alternative plasticizer standards (Add solution) using the suspect screening workflow**

*Data analysis.* All recorded chromatograms and mass spectra (available as Agilent files in SI8- Rawdata-GCMS-Suspect) were analyzed for the presence and approximate concentration of the suspect compounds, and for unknown substances using library identification. A qualitative Agilent Masshunter workflow was used for compound discovery (either using chromatogram integration or molecular feature) and compound identification (using the custom library first, and the NIST 14 library second) with the final output exported as an Excel file. For compound discovery, both "Find by integration" (considering all Lorentzian chromatogram peaks with an area larger than 0.001% of the largest peak) and "Find by molecular feature" (limited to Lorentzian peaks with more than 500 counts and the largest 200 compounds) were used. For compound identification, (1) a manually created suspect library of the scanned suspect standards was searched first and then (2) the NIST 14 library was searched (this old library version was used to limit overfitting the data). The suspect library was constructed from measurements of the suspect standards at 5 mg/L. The minimum matching score for both libraries was set to 50, but was usually above 70; only *TMPP* [*Tri(3,4-dimethylphenyl) phosphate*, CASRN: 3862-11-1], *TCPP* [*Tris-(2-chloroisopropyl) phosphate*, CASRN: 13674-84-5] and *DINCH* scored slightly below 70. The assignment of suspects was partially manually double-checked, based on retention time. Substances that appeared several times under different identifiers in the library were manually harmonized (e.g., *DEHT* appears in the NIST library under the CASRN "6422-86-2" or under the name "*1,4- Benzenedicarboxylic acid, 1,4-bis(2-ethylhexyl) ester*").

Further data processing was done in Python (SI5) and included:

- 1. combining the individual Excel files
- 2. flagging compounds discovered in procedural blanks as "Blanks"
- 3. assigning identification confidence of compounds (confirmed with standards as "Level 1
- standard confirmed", others as "Level 2 library confirmed")
- 4. ranking substances based on their importance (total signal area, number of samples)

5. semi-quantifying suspects based on the calibration curves from the dilution series (calibration curves in [Figure S 6](#page-27-0) for all standards, in [Table S 10](#page-28-0) for individual standard).

Samples were run at two dilutions, the final concentration was selected based on which detections were in range (the detailed algorithm is portrayed in [Table S 9](#page-26-0) and the final selected concentrations for the semi-quantification can be found in Sheet S8 in SI2).

*QA/QC*. The aforementioned QA/QC were also applied here. Furthermore, the workflow for the suspect substances was thoroughly pre-tested, including, (1) testing the suitability of the extraction procedure, (2) optimizing GC-MS and data analysis parameters, and (3) determining approximate LODs for all suspects. Blank samples and suspect standards were included in regular intervals to ensure correct GC-MS operation, and correct suspect identification was ensured by employing a costume suspect library with a matching score above 70 (in most cases) and manual checks.



<span id="page-26-0"></span>**Table S 9: Selection of most suitable value based on detection situation. Selected value is in bold, comment in normal text, color signifies possible mistakes**



<span id="page-27-0"></span>**Figure S 6: Peak area vs concentration for different types of standards used in the suspect screening workflow.**

#### *S2.3.4.1 Suspect list*

**Table S 10: Analytical standards (including** *ortho-***phthalates, alternative plasticizer, phosphate plasticizers, brominated flame retardants, antioxidants and bisphenols) used in the suspect-screening workflow. Overview of massspectra and approximate calibration curves for GC-MS suspect-screening. Approximate calibration curves were fitted to a constrained linear model with the intercept forced through zero (Area = a1\*concentration) and a regular linear model (Area = a0 + a1 \*concentration). The table is sorted**  based on substance group and retention time (RT). RT = Retention time, CASRN = Chemical Abstract Service Registry Number, MW = Molecular weight of isotope, Dyn. **Range = Dynamic range.** 

<span id="page-28-0"></span>















 $\pm$ 

#### <span id="page-36-0"></span>**S2.4 Details Bioassays**

*Sample preparation.* The same extraction procedure as above was used (section [S2.3.3\)](#page-17-0) except that samples were not diluted after filtration but concentrated, since most bioassays have a low solvent tolerance (MTT/ROS: max 0.1 volume%). Using a Syncore system from Buchi to avoid losses of volatile substances, the solvent from 12 samples was evaporated in parallel, from approximately 6mL to 300 μL (pressure: 210mbar, temperature top of the flasks: 60°C, temperature bottom of the flask:  $10^{\circ}$ C). These samples were stored at  $-20^{\circ}$ C. However, during the inter-laboratory shipping (2–3 days), the temperature may have risen to 20°C. Due to the high volatility of *THF*, the sample volumes decreased during the storage and transport. Before applying to each assay, samples were taken out from –20°C, the volume of each sample was inspected and, if necessary, filled up to 300 μl with THF. Then, samples were stored overnight at  $4^{\circ}$ C prior to the testing.

*Extract selection:* The samples screened for cytotoxicity (MTT assay) and reactive oxygen species generation (ROS assay) were selected at random (n=85). The selected samples can be seen in Sheet S1 and Sheet S10 in SI2. The samples for the endocrine activity assays, AMES test and planarumuC bioassay were selected as to be maximally different regarding their *ortho-*phthalate content and their activity in the MTT assay [\(Table S11\)](#page-36-1).

Sample_id	MTT viability $[\%]$	o-phthalate content [wt%]	<b>YES/YAS</b>	$planar$ -umu $C$	Ames
g <sub>5</sub>	53.76	12.08		$\mathbf X$	
$d80-2$	57.21	16.73		X	X
$d1-2$	49.82	0	X	X	
$d31-1$	50.69		$\mathbf X$		
$dl-1$	57.74			X	
g <sub>4</sub>	70.81	33.02	$\mathbf X$	$\mathbf X$	
gl	82	20.61	X		
$d21-1$	88.03	47.14		X	X
$d20-2$	97.84	40.13	X	X	
garl	105.99	40.35	X		
g2	125.51	18.19	$\mathbf X$		
$d20-1$	108.43	35.02		$\mathbf X$	$\mathbf X$
g <sub>3</sub>	99.72	18.5		X	
$d42-2$	92.99	0	X		
$dl3-2$	93.29	0		$\mathbf X$	X
g <sub>7</sub>	128.66	0.01		X	X
$d75 - 2$	110.25	$\theta$		X	

<span id="page-36-1"></span>**Table S11: Selected extracts for further screening with YES/YAS, umuC and AMES bioassays based on MTT viability and**  *ortho-***phthalate content.**

**Cytot***oxicity and oxidative stress.* Randomly selected extracts (n=85) were screened for cytotoxicity using MTT assays and for oxidative stress using ROS assays. Both assays were conducted on human liver cells (Huh7), according to Christen et al. 2014.<sup>17</sup> Cells were grown in DMEM with GlutaMAX™ (LuBioScience, Lucerne, Switzerland) supplemented with 10% FBS (Sigma-Aldrich, Taufkirchen, Germany) in a humidified incubator with 5%  $CO<sub>2</sub>$  at 37 °C. Cells were usually split every 4 days and sub-cultured at split ratios of about 1:6. Then, Huh7 cells were plated at a density of 25 000 cells per well in 96-well plates. After 24 h, cells were treated either with the highest possible test concentration (1 μl extract/1 ml cell culture medium, as solvent concentration should not exceed 0.1 volume%), or for selected ones, with a serial dilution of the extracts (1:2 dilution steps). The samples were classified based on the cell viability in the MTT assay: "highly toxic" for below 30%, "moderately toxic" for 30–60%, "slightly toxic" for 60–90%, and "not toxic" for above 90%.

*Endocrine activity.* Eight selected extracts were screened for estrogenic, anti-estrogenic, androgenic, and anti-androgenic activities using XenoScreen YES/YAS assays from Xenometrix (Allschwil, Switzerland). Serial dilutions of selected extracts (highest test concentration: 1:150 dilution of pure extract) were tested according to the manufacturer's protocol.

*Mutagenicity*. Nine selected extracts were analyzed for potential mutagenic activity using Ames MPF 98/100 from Xenometrix (Allschwil, Switzerland) with *Salmonella typhimurium* strains TA98 (for detection of frameshift mutations) and TA100 (for detection of base substitution mutations), in accordance with the manufacturer's protocol.

*Genotoxicity.* Twelve selected extracts were analyzed for potential direct genotoxic activity using the planar-umuC bioassay protocol of planar4 GmbH (Stäfa, Switzerland). The planar-umuC was conducted on normal phase, silica gel Si 60 HPTLC plates (Merck, Germany), with the *Salmonella typhimurium* strain TA1535 pSK1002 (Xenometrix, Allschwil, Switzerland). The raw samples (300 μl) were first diluted to 800 μl ACN/THF to facilitate handling. All samples were then diluted 1:10, 1:100 and 1:1000, and applied to the HPTLC plates using an Automatic TLC samples (ATS4, Camag, Switzerland). A solvent blank (ACN/THF, for sample dilution), a second solvent blank (solvent of positive control) and three 4-NQO positive controls with a mass per band of 100, 200 and 800 pg were also applied. The HPTLC plates were developed with ACN:DCM (dichloromethane) (1:1) from 20 mm to 75 mm. A total of 8 runs were conducted. The genotoxicity after metabolic activation was not determined because the respective planar-umuC protocol was not available at the time of the experiment.

*QA/QC*. Procedural and solvent blank samples were tested to ensure effects were caused by substances present in the samples. MTT and ROS screening were performed in triplicate, whereas the other assays were repeated as often as recommended by the respective protocols.

#### <span id="page-39-0"></span>**S2.5 Data treatment**

Data treatment included (1) treatment using specialized software for the analysis method (e.g. NITON plastics calibration for XRF, Agilent Masshunter for GC-MS), which is described in the respective sections, and (2) further combined data analysis, which was conducted in python (relevant scripts are provided in SI5). Further data treatment included combining and aligning data, creating the graphs for this paper, conducting principal component analysis, and clustering the data.

Furthermore, the raw data produced in this campaign is provided for further analysis in the following formats in the SI4: (1) ATR-FTIR spectra for each sample and side as '.csv', (2) XRF based elemental concentrations for each sample and side as 'excel', (3) SIM GC-MS spectra for phthalate measurements of each sample as 'Agilent' and 'mzXML', and calculated concentrations as 'excel', (4) SCAN GC-MS spectra for suspect screening of each sample as 'Agilent' and 'mzXML', and detection, identification and semi-quantification results as 'excel', (5) bioassay readings as 'excel'.

## <span id="page-40-0"></span>**S3 RESULTS**

#### <span id="page-40-1"></span>**S3.1 Concentrations and presence of individual substances**

<span id="page-40-2"></span>**Table S12: Concentration and presence of individual elements based on XRF elemental analysis. Summary statistics (minimum, median, mean, sd, und maximum) are shown for the detected fraction only. The limits of detection (LODs) were calculated according to the instrument manufacturer's protocol, as three times the minimum standard deviation of the analyte. The table is sorted based on the detection frequency, if not detected by the abbreviation of the element. Abbr. = Abbreviation, CASRN = Chemical Abstract Service Registry Number, DF=Detection frequency, LOD = Limit of detection, SD = Standard deviation.**



<span id="page-41-0"></span>**Table S 13: Concentration and presence of individual** *ortho-***phthalates based on the phthalate GC-MS quantification workflow. Summary statistics (minimum, median, mean, sd, und maximum) are shown for the detected fraction only. The table is sorted based on the detection frequency, if not detected by the retention time of the standard. Abbr. = Abbreviation, CASRN = Chemical Abstract Service Registry Number, DF=Detection frequency, LOD = Limit of detection, SD = Standard deviation.** 



<span id="page-42-0"></span>**Table S14: Concentration and presence of individual substances based on GC-MS suspect screening workflow. The LODs reported here are based on a dilution series and only give an approximate measure for the limit of detection. Concentration estimates are based on semi-quantification, and may be above 1'000'000 mg/kg for samples outside the calibration range. Summary statistics (minimum, median, mean, sd, and maximum) are shown for the detected fraction only. The table is sorted based on the substance group and the detection frequency, if not detected by the retention time of the standard. Abbr. = Abbreviation, CASRN = Chemical Abstract Service Registry Number, DF=Detection frequency, LOD = Limit of detection, SD = Standard deviation.**



Aside from our suspect list, also other compounds were discovered by the library identification. All discovered substances, their corresponding samples and a prioritization of the substances (based on total area and number of relevant samples) is presented on Sheet S9 in SI2. Their chemical space plot can be seen in [Figure S 7,](#page-43-0) weighted by the peak area, and [.Figure S 8,](#page-44-0) weighted by the detection frequency.



<span id="page-43-0"></span>**Figure S 7: Chemical space of the substances detected in suspect list screening, marker size scaled to their total chromatogram area. For the bottom plots, the iso-concentration curves are calculated for equal volumes of each compartment (i.e. water, air and octanol are exactly the same volume). Most additional substances that were not standards have a logKow around zero, meaning they are dynamic and easily leach from octanol-like environments.** 



<span id="page-44-0"></span>**Figure S 8: Chemical space of the substances detected in suspect list screening, marker size scaled to their detection frequency. For the bottom plots, the iso-concentration curves are calculated for equal volumes of each compartment (i.e. water, air and octanol are exactly the same volume). Most additional substances that were not standards have a logKow around zero, meaning they are dynamic and easily leach from octanol-like environments.** 

## <span id="page-45-0"></span>**S3.2 Total plasticizer content**

Approximate plasticizer composition and total amount per sample are displayed in [Figure S9,](#page-45-1) individual values for each sample can be found in Sheet S1 in SI2. The values for semi-quantified substances are highly uncertain as many were outside the respective calibration curve range.



<span id="page-45-1"></span>**Figure S9: Plasticizer composition (left) and amount (right) by sample. Sorted by the major plasticizer per sample. Semi quantification of some plasticizer resulted in very high concentration estimates (some concentrations are even above 100wt%), which is mainly due to calibration curve uncertainty especially for signals above the calibration curve range.**



## <span id="page-46-0"></span>**S3.3 Correlation between Substances**

<span id="page-46-1"></span>**Figure S 10: Correlation matrix for detection and logarithmic concentration of all measured samples**



## <span id="page-47-0"></span>**S3.4 Bioassay results**

<span id="page-47-1"></span>**Figure S 11: Cell viability and induction of oxidative stress in Huh7 cells after exposure to plastic extracts. Huh7 cells were exposed to a serial dilution with a dilution-factor of 2 (d1: highest concentration, d8: lowest concentration) of the seven plastic extracts which induced more than 40% of cell mortality in the first screen.**



<span id="page-48-0"></span>**Figure S 12: Endocrine activity of selected plastic extracts. Estrogenic, anti-estrogenic, androgenic, and anti-androgenic activities were analysed in yeast cells after exposure to selected plastic extracts. Shown are the estrogenic, anti-estrogenic, androgenic, and anti-androgenic controls of the kit and data of extract d20.2. A serial dilution with a dilution factor of 2 from the highest possible concentration (d20.2) to the lowest test concentration (d7) was analysed. Red dotted lines point to the expected hormonal activities.**



<span id="page-48-1"></span>**Figure S 13: Induction of mutagenicity by extract d21.1.** *Salmonella typhimurium* **strains TA98 and TA100, each with and without S9 liver fractions, were exposed to a serial dilution of extract d21.1 and controls for 48h. Positive controls: 2 nitrofluorene (2-NF), 2-aminoanthracene (2-AA), and 4-nitroquinoline (4-NQO). Data is presented as number of reverted mutations per concentration with standard deviation from one experiment. Red line: 2-fold increase over baseline. Red star: binominal B≥ 0.99.**



<span id="page-49-0"></span>**Figure S 14 Genotoxicity of selected samples, measured with the planar-umuC bioassay (samples g3, g4, g5, g7, d1.1 and d1.2 (tracks 2 to 7, all at Rf 0.9). The dark bands of tracks 1-7 at Rf 0.7 indicate an inhibition of the planar-umuC test system by the THF-ACN (1:3) solvent. Control tracks of solvent in (track 8). Positive control 4-NQO (tracks 9-12) in increasing concentration.**

## <span id="page-50-0"></span>**S3.5 Linear regression models**

#### <span id="page-50-1"></span>S3.5.1 Toxic metals – presence and concentration

<span id="page-50-2"></span>**Table S 15:** Linear regression model for predicting the chance of any toxic metal(loids), i.e. Cd, Pb, Cr, Ni, Hg, As, being present (in %) based on sample properties (independent variables).



*\*\* significant contributions (p < 0.05) \* possibly significant contributions (p < 0.10)*

<span id="page-50-3"></span>**Table S 16:** Linear regression model for predicting the concentration in ppm of toxic metal(loids), i.e. Cd, Pb, Cr, Ni, Hg, As, based on sample properties (independent variables)



*\*\* significant contributions (p < 0.05)*

#### <span id="page-51-0"></span>S3.5.2 *ortho-*Phthalates – presence and concentration

#### *S3.5.2.1 Any* ortho-*phthalates*

<span id="page-51-1"></span>**Table S 17:** Linear regression model for predicting the chance of any of *ortho-*phthalates being present (in %) based on sample properties (independent variables).



*\*\* significant contributions (p < 0.05)*

*\* possibly significant contributions (p < 0.10)*

<span id="page-51-2"></span>**Table S 18:** Linear regression model for predicting the concentration (in wt%) of *ortho-*phthalates based on sample properties (independent variables).



*\*\* significant contributions (p < 0.05)*

#### *S3.5.2.2 Restricted* ortho-*phthalates*

<span id="page-52-0"></span>**Table S 19:** Linear regression model for predicting the chance of regulated *ortho-*phthalates being present (in %) based on sample properties (independent variables).



**Chance of presence of any**

*\*\* significant contributions (p < 0.05)*

*\* possibly significant contributions (p < 0.10)*

<span id="page-52-1"></span>



*\*\* significant contributions (p < 0.05)*

#### <span id="page-53-0"></span>S3.5.3 Alternative plasticizers – presence

<span id="page-53-2"></span>**Table S 21** Linear regression model for predicting the chance of alternative plasticizers being present (in %) based on sample properties (independent variables).

![](_page_53_Picture_480.jpeg)

*\*\* significant contributions (p < 0.05)*

*\* possibly significant contributions (p < 0.10)*

#### <span id="page-53-1"></span>S3.5.4 Bioassays

<span id="page-53-3"></span>**Table S 22:** Linear regression model for predicting the chance of activity in any of the bioassay (in %) based on sample properties (independent variables).

![](_page_53_Picture_481.jpeg)

*\*\* significant contributions (p < 0.05)*

## <span id="page-54-0"></span>**S3.6 Screening quality metrics**

<span id="page-54-1"></span>**Table S 23: Quality of different screening methods for determining samples of clear concern and those of any concern (possible + clear concern) using confusion matrices, sensitivity (sens) and specificity (spec).**

![](_page_54_Figure_2.jpeg)

![](_page_55_Figure_0.jpeg)

![](_page_56_Figure_0.jpeg)

![](_page_57_Figure_0.jpeg)

<span id="page-57-0"></span>**Figure S 15: Utility of different screening methods. Reverse specificity (as a proxy for unnecessary waste) is plotted against sensitivity (as a proxy for removed hazardous substances) for selected screening methods. Methods are differentiated by how difficult it is to implement them on industrial scale for waste sorting and by the fraction of samples tested in our study.**

## <span id="page-58-0"></span>**S4 DISCUSSION**

## <span id="page-58-1"></span>**S4.1 Chemical substances in PVC flooring**

<span id="page-58-3"></span>**Table S 24 Recent studies investigating plasticizers and other substances present in PVC flooring. The country was not specified for all studies, the location of the main authors are given in parenthesis if no details were mentioned.**

<b>Reference</b>	Country Year n			<b>Major plasticizers</b>	Conc. range $[wt\%]$	Other substances	Conc. range $[wt\%]$
Clausen, et. al $(2004)$ <sup>18</sup>	(DNK)	2004 1		<b>DEHP</b>	17	not analyzed	
Afshari, et. al $(2004)$ <sup>19</sup>	(DNK)	2004 4		<b>DEHP</b>	17-18.5	not analyzed	$\blacksquare$
Chino, et.al. $(2009)^{20}$	(IPN)	2009 1		<b>DEHP</b>	10	not analyzed	$\overline{\phantom{a}}$
Xu. et al. $(2012)^{21}$	<b>DNK</b>	2012 1		<b>DEHP</b>	15	not analyzed	
Kumari, et al. $(2014)$ <sup>22</sup>	<b>IND</b>	2014 1		not analyzed	$\overline{\phantom{0}}$	BDE47, BDE153, <b>BDE209</b>	$<$ LOD
Liang, et al. $(2015)^{23}$	<b>USA</b>	2015 16		DEHP, BBP, DEHI, DiNP, <b>DBP</b>	$0.03 - 26.5$	not analyzed	$\sim$
Shi, et al. $(2018)$ <sup>24</sup>	<b>CHN</b>	2018 2		DEHP, BBP, DnOP (only low MW ortho-phthalates analyzed)	$4 - 15$	not analyzed	$\overline{\phantom{a}}$
<b>Bohlin-</b> Nizzetto, et al. $(2021)^{25}$	<b>NOR</b>	2021	6	TPhP, TBEP (ortho-phthalates not analyzed)	$0.0002 - 0.07$	<b>BFRs</b>	$<$ LOD $-$ $7x10^{-8}$
Lowe, et al. $(2021)^{26}$	<b>USA</b>	2021 43		DEHA, DEP, TXIB, DBP, ATBC, BBP, others	not quant.	Hexadecanoic acid, Octadecanoic acid, 1-Dodecanol, others	not quant.
This study	<b>CHE</b>	2021 151		DEHT, DiNP, DEHA, DEHP, DiDP, Octicizer, others	$\langle$ LOD - 46	<b>UV326, BPA</b>	not quant.

## <span id="page-58-2"></span>**S4.2 Chemical substances in other PVC products**

<span id="page-58-4"></span>![](_page_58_Picture_513.jpeg)

![](_page_58_Picture_514.jpeg)

<span id="page-59-0"></span>**Table S 26: Recent studies investigating plasticizers and other organic substances present in other PVC products (not flooring). The country was not specified for all studies, the location of the main authors are given in parenthesis if no details were mentioned.**

Reference			<b>Country Year Product n</b>		<b>Major plasticizers</b>	Conc. range $[wt\%]$	<b>Other</b> substances	Conc. range $[wt\%]$
Wahl, et al.	<b>GER</b>		1999 Medical 6		DEHP, BEHP, DBP, DiBP, DEP,	not quant.	BHT,	not quant.
$(1999)^{32}$					DEHA, DMP	(DEHP	Styrene,	
						largest area)	others	
Wang, et al. $(2005)^{33}$	<b>DNK</b>		2005 Medical 3		DEHP, DCHP, DEHA	$0.06 - 30$	<b>BHT</b>	not quant.
Welle, et al. $(2005)^{34}$	(GER)		2005 Medical 6		DEHP, DINCH, TEHTM, ATBC	30-49	not analyzed	
Radaniel, et al. $(2014)$ <sup>35</sup>	<b>GER</b>		2014 Medical 5		DEHP, ATBC, DEHT, DiNCH, TEHTM (sampled tubing with known contetnt for method validation)	29-36	not analyzed	
Bernhard, et al. $(2015)^{36}$	<b>FRA</b>		2015 Medical 4		DEHP, DEHT, TEHTM, DiNCH	28-31	not analyzed	$\overline{a}$
<b>Bourdeaux et</b> al. $(2016)^{37}$	<b>FRA</b>		2016 Medical 32		TEHTM, DEHP, DiNCH, DiNP, ATBC, DEHA, DEHT	24-36	not analyzed	
Faessler, et al. $(2017)$ <sup>38</sup>	<b>CHE</b>		2017 Medical 7		DEHP, DiNCH, DEHT, TOTM, <b>ESBO</b>	22-44	not analyzed	÷,
Jeon, et al. $(2018)$ <sup>39</sup>	<b>KOR</b>		2018 Medical 3		DEHP, DiOP, TEHTM	not quant.	not analyzed	$\blacksquare$
Fernandez-	(FRA)		2018 Medical 1		TOTM, DEHP, DEHT, DEHA	$0.1 - 45$	not	
Canal, et al. $(2018)$ <sup>40</sup>							analyzed	
Den Braver- Sewradj, et al. $(2020)^{41}$			2020 Medical	$\sim$	Review (extensive use of DEHP, mail alternatives: TEHTP, DiNCH, DEHA, ATBC, DiNP)		not analyzed	
Rastogi, et al. $(1998)$ <sup>42</sup>	(DNK)	1998	<b>Toys</b>	$\overline{7}$	DEHP, DiNP, DiDP	$<$ LOD - 40	not analyzed	
<b>US-CPSC</b> $(2010)^{43}$	<b>USA</b>	2010	<b>Toys</b>	37	ATBC (60%), Tributyl aconitate (49%), DiNCH (38%), DEHT (35%), TXIB (32%), DEHP (3%), DiNP $(3%)$	14-42	not analyzed	
Al-Natsheh, et al. $(2015)$ <sup>44</sup>	<b>JOR</b>	2015	<b>Toys</b>	$\mathbf{1}$	<b>DEHP</b>	0.06	<b>Not</b> analyzed	$\overline{\phantom{a}}$
McCombie, et al. $(2017)^{45}$	<b>CHE</b>	2017	<b>Toys</b>		118 ESBO (81%), DEHT (55%), TXIB (49%), DiNCH (31%), ATBC (31%), DEHP (9%), others	$0.9 - 51$	not analyzed	$\overline{a}$
Ashworth, et 46 al. (2018)	<b>NZL</b>	2018	<b>Toys</b>	49	DEHP, DiNP, DiDP, DiBP, DBP, <b>DNOP</b>	$0.1 - 54$	not analyzed	

#### <span id="page-60-0"></span>**S4.3 Exposure to** *ortho-***Phthalates and alternative plasticizers**

Usually, the major exposure pathway for all *ortho-phthalates* is dietary intake, in the  $\mu$ g kg<sub>bw</sub><sup>-1</sup> d<sup>-1</sup> range, and together with indoor exposure, relevant health limit values (e.g., a reference dose for *DEHP*: 20  $\mu$ g kg<sub>bw</sub><sup>-1</sup> d<sup>-1</sup>) can be exceeded, especially for vulnerable and at-risk populations (e.g., toddlers).<sup>47</sup> Another noteworthy exposure pathway for specific individuals is from medical devices, which are still commonly plasticized with DEHP as allowed by a re-authorization process: exposure from intravenous administration of different solutions may reach up the mg  $kg_{bw}^{-1} d^{-1}$ range.<sup>41,48</sup> The most important indoor exposure pathways for (semi-volatile) plasticizers such as *ortho-*phthalates from indoor products are the ingestion of dust, inhalation of air-borne particles, and direct skin contact [\(Table S 28\)](#page-63-0). 18,49–52 While for higher-molecular weight *ortho-*phthalates, dust and dietary intake dominate the total exposure, for lower-molecular weight *ortho-*phthalates (e.g., DMP, DEP, DBP, DiBP), inhalation and dermal uptake (due to use in personal care products) are additionally important.47,53,54 Typically, steady-state air concentrations for *ortho-*phthalates have been found in the low  $\mu$ g/m<sup>3</sup> range in chamber experiments with PVC floorings (e.g., 0.8–1) μg/m<sup>3</sup> for DEHP),<sup>19,21,55</sup> and air measurements in residential buildings (e.g., 0.1–20 μg/m<sup>3</sup> for total phthalates) 56,57. This clearly demonstrates the releases of these substances from PVC floorings. Once released from the PVC matrix, partition to skin, dust, and air-borne particles is mainly governed by the octanol-air partition coefficients *K*OA, which is high for the major plasticizers in this study [\(Figure S 1,](#page-10-1) [Table S 4\)](#page-12-0).<sup>58–61</sup> A similar pattern to the original PVC flooring is, thus, expected in dust and skin wipes. Plasticizers have been measured in indoor dust samples in the  $\mu$ g/g to mg/g range, with strong correlation with the use of PVC floorings as can be expected.<sup>52,56,57,62–64</sup> The main plasticizers in dust vary by region, likely due to different flooring compositions across markets. For example, a recent Swedish study found mainly DiNP, DEHP, DiDP, DEHT, and DINCH (~100  $\mu$ g/g), which is in good agreement with our findings.<sup>62</sup> Another German study found DEHP, DiNP, and DiDP being the major ones, and DEHP reducing and the others rising over time.<sup>64</sup> A study in Canada found mainly DiNP, DEHP, and DBP  $(-14-200)$  $\mu$ g/g).<sup>57</sup> Studies from China, Republic of Korea and the US mainly reported DEHP, DBP, DiBP, and BBP (also  $\sim$ 100 μg/g).<sup>52,54,63,65</sup>. In general, exposure from dust ingestion has been estimated to be in the lower  $\mu$ g kg<sub>bw</sub><sup>-1</sup> d<sup>-1</sup>, while dermal absorption of dust is in the low ng kg<sub>bw</sub><sup>-1</sup> d<sup>-1</sup>.<sup>52,54,57</sup> Concentrations of *ortho-*phthalates on skin have typically been measured in the ng/cm<sup>2</sup> to μg/cm<sup>2</sup> range, resulting in estimated dermal exposure to air in the lower  $\mu$ g kg<sub>bw</sub><sup>-1</sup> d<sup>-1</sup> range. Dermal exposure is typically lower than exposure from dust ingestion.<sup>47,52,54,57,66</sup> Reported plasticizers on skin were again regionally dependent: DiNP, DNOP, DEHP, DBEP, and DMEP were the main plasticizers in China,<sup>66</sup> DEHP and DiNP were reported in the US and Canada,<sup>52,57</sup> and DiNP, DEHP, and DiDP were the main plasticizers reported in Norway (which shows a similar plasticizer profile as in this study). $47$ 

Alternatives are found in similar concentrations, albeit slightly lower than *ortho-*phthalates, in the different compartments [\(Table S 27\)](#page-62-0): DEHA, DINCH, DEHT and ATBC were found in the air around 10–100 ng/m<sup>3</sup> (an order of magnitude below *ortho-*phthalates),<sup>67,68</sup> DEHA, DINCH, DEHT, and ATBC were found in dust around  $10-100 \mu g/g$  (the same order of magnitude as *ortho-*phthalates).62,68,69

			ortho-phthalates				<b>Alternative</b> plasticizers		
		<b>Type of</b> study	<b>Restricted</b>		Other o-PHT		Detected in this study	Other	<b>Ref</b>
			<b>DEHP</b>	<b>BBP, DBP, DiBP</b>	DiNP, DiDP	<b>Others</b>	e.g. DEHT, DEHA, DINCH		
	logKoa		11.7	$8.2 -$	$11 -$	$5.7 -$	$10.8 -$	$6-$	70
				9.8	11.5	11.7	11.7	18	
Conc.	Flooring	M	$32 -$	$39-$	500-	$50-$	$80-$	n.a.	This
	[µg/g]		204'700	4'000	471'300	18'600	1'000'000*		study
	$Air [ \mu g/m^3 ]$	M	0.02	$0.0006 -$	$0.01 -$	$0.004 -$	$<$ LOD	$\overline{a}$	47,57,71,72
			$-3.69$	4.6	0.03	2.5			
		$\mathbf C$	$0.8 -$	$0.1 -$		$\blacksquare$	$\sim$	÷.	19,21,55
			1.0	0.2					
	Dust [ $\mu$ g/g]	M	100	$5.5-$	$29-$	$0.12 -$	$32.8 -$	$\sim$	47,57,62
			$-232$	15.2	282	6.3	34.5		
		$\mathbf C$	$0.5 -$	$\overline{a}$		$\blacksquare$	$\overline{\phantom{a}}$	$\blacksquare$	$\overline{55}$
			0.9						
	Dust [ $\mu$ g/m3]	M	$0.04 -$	$0.0003 -$		$0.0002 -$	$\sim$	÷.	72,73
			2.2	2.3		1.5			
		$\mathsf{C}$	$0.5 -$	$\overline{\phantom{a}}$		$\overline{\phantom{a}}$	$\overline{\phantom{a}}$	$\overline{\phantom{a}}$	$\overline{55}$
			0.9						
	Skin $\lceil \mu g/m2 \rceil$	M	$0.000001 -$	$0.0001 -$	$0.0001 -$	$0.0001 -$	$<$ LOD	$\overline{\phantom{0}}$	47,57
			55.7	2.7	56.9	3.6			
	Surface	M	$0.000001 -$	$0.0001 -$	$0.00001 -$	$0.00003 -$		$\overline{\phantom{a}}$	21,57
	[ $\mu$ g/m2]		1'241	0.004	0.037	0.0006			

**Table S 27: Measured and modelled indoor media concentrations of different plasticizers. M= measurements, C= chamber / model**

<span id="page-62-0"></span>\* alternative plasticizer concentrations in this study are highly uncertain, due to lack of internal standard and in-range calibration, which lead to partially implausible estimates

\*\* PVC flooring in a ventilated room

<span id="page-63-0"></span>![](_page_63_Picture_306.jpeg)

**Table S 28 Estimated exposure to different plasticizers.**

#### <span id="page-64-0"></span>**REFERENCES**

- (1) European Chemicals Agency (ECHA). Candidate List of substances of very high concern for Authorisation https://echa.europa.eu/candidate-list-table (accessed Apr 2, 2020).
- (2) European Chemicals Agency (ECHA). SCIP Database https://echa.europa.eu/scip-database (accessed Sep 25, 2023).
- (3) European Chemicals Agency (ECHA). Authorisation List https://echa.europa.eu/authorisation-list (accessed Sep 25, 2023).
- (4) European Chemicals Agency (ECHA). Restriction List https://echa.europa.eu/substancesrestricted-under-reach.
- (5) European Chemicals Agency (ECHA). Phthalates https://echa.europa.eu/hottopics/phthalates (accessed Sep 25, 2023).
- (6) European Parliament; Council of the European Union. Council Directive 2009/48/EC on the safety of toys http://data.europa.eu/eli/dir/2009/48/2019-11-18 (accessed Sep 25, 2023).
- (7) European Parliament; Council of the European Union. Council Directive 2005/84/EC on phthalates in toys and childcare articles https://eur-lex.europa.eu/eli/dir/2005/84/oj (accessed Sep 25, 2023).
- (8) European Parliament; Council of the European Union. Council Directive 2011/65/EU on the restriction of the use of certain hazardous substances (RoHS) in electrical and electronic equipment (EEE) http://data.europa.eu/eli/dir/2011/65/2021-04-01 (accessed Sep 25, 2023).
- (9) vinylPlus. The European PVC industry's experience in replacing lead and cadmium-based stabilisers https://www.stabilisers.eu/wp-content/uploads/2015/11/VinylPlus\_Contribution-Cefic\_Eu-Industry.pdf (accessed Sep 25, 2023).
- (10) European Stabilisers Producers Association (ESPA(). Stabilisers What's new ? https://www.stabilisers.eu/wp-content/uploads/2016/01/ESPA-stabilisers\_update\_January-2017.pdf (accessed Sep 25, 2023).
- (11) Everard, M. 5 PVC and Sustainability. In *PVC Additives*; Schiller, M., Ed.; Hanser, 2015; pp 369–410. https://doi.org/https://doi.org/10.3139/9781569905449.005.
- (12) Kuptsov, A. H.; Zhizhin, G. N. Spectra. In *Handbook of Fourier Transform Raman and Infrared Spectra of Polymers*; Kuptsov, A. H., Zhizhin, G. N. B. T.-P. S. D., Eds.; Elsevier, 1998; Vol. 45, pp 1–500. https://doi.org/10.1016/S0921-318X(98)80016-7.
- (13) Thermo Scientific. Fast , affordable solutions for polymers and plastics analysis https://assets.thermofisher.com/TFS-Assets/MSD/Flyers/FL52273-ftir-polymer-analysiskits.pdf.
- (14) Lowry, S.; Bradley, M.; Thermo Scientific. Using FT-IR Spectroscopy to Characterize Plastics and Other Materials. *Adv. Mater. Process.* **2011**, *169* (4), 22–25.
- (15) Agilent; Wang, Y. Polymer and Phthalate Analysis with FTIR Spectroscopy. Agilent 2018.
- (16) Thermo Fisher Scientific. XL3 Analyzer version 7.0.1 User Guide https://www.tttenviro.com/wp-content/uploads/Manual-XL3-Series-v7.0.11.pdf.
- (17) Christen, V.; Camenzind, M.; Fent, K. Silica Nanoparticles Induce Endoplasmic Reticulum Stress Response, Oxidative Stress and Activate the Mitogen-Activated Protein Kinase (MAPK) Signaling Pathway. *Toxicol. Reports* **2014**, *1*, 1143–1151. https://doi.org/10.1016/j.toxrep.2014.10.023.
- (18) Clausen, P. A.; Hansen, V.; Gunnarsen, L.; Afshari, A.; Wolkoff, P. Emission of Di-2- Ethylhexyl Phithalate from PVC Flooring into Air and Uptake in Dust: Emission and Sorption Experiments in FLEC and CLIMPAQ. *Environ. Sci. Technol.* **2004**, *38* (9), 2531– 2537. https://doi.org/10.1021/es0347944.
- (19) Afshari, A.; Gunnarsen, L.; Clausen, P. A.; Hansen, V. Emission of Phthalates from PVC and Other Materials. *Indoor Air* **2004**, *14* (2), 120–128. https://doi.org/10.1046/j.1600- 0668.2003.00220.x.
- (20) Chino, S.; Kato, S.; Seo, J.; Ataka, Y. Study on Emission of Decomposed Chemicals of Esters Contained in PVC Flooring and Adhesive. *Build. Environ.* **2009**, *44* (7), 1337–1342. https://doi.org/10.1016/j.buildenv.2008.07.003.
- (21) Xu, Y.; Liu, Z.; Park, J.; Clausen, P. A.; Benning, J. L.; Little, J. C. Measuring and Predicting the Emission Rate of Phthalate Plasticizer from Vinyl Flooring in a Specially-Designed Chamber. *Environ. Sci. Technol.* **2012**, *46* (22), 12534–12541. https://doi.org/10.1021/es302319m.
- (22) Kumari, K.; Sharma, J. K.; Kanade, G. S.; Kashyap, S. M.; Juwarkar, A. A.; Wate, S. R. Investigation of Polybrominated Diphenyl Ethers in Old Consumer Products in India. *Environ. Monit. Assess.* **2014**, *186* (5), 3001–3009. https://doi.org/10.1007/s10661-013- 3596-2.
- (23) Liang, Y.; Xu, Y. Emission of Phthalates and Phthalate Alternatives from Vinyl Flooring and Crib Mattress Covers: The Influence of Temperature. *Environ. Sci. Technol.* **2014**, *48* (24), 14228–14237. https://doi.org/10.1021/es504801x.
- (24) Shi, S.; Cao, J.; Zhang, Y.; Zhao, B. Emissions of Phthalates from Indoor Flat Materials in Chinese Residences. *Environ. Sci. Technol.* **2018**, *52* (22), 13166–13173. https://doi.org/10.1021/acs.est.8b03580.
- (25) Bohlin-Nizzetto, P. Content and migration of chemical additives from plastic products. (NILU report 9/2022). https://hdl.handle.net/11250/2992965 (accessed Sep 25, 2023).
- (26) Lowe, C. N.; Phillips, K. A.; Favela, K. A.; Yau, A. Y.; Wambaugh, J. F.; Sobus, J. R.; Williams, A. J.; Pfirrman, A. J.; Isaacs, K. K. Chemical Characterization of Recycled Consumer Products Using Suspect Screening Analysis. *Environ. Sci. Technol.* **2021**, *55* (16), 11375–11387. https://doi.org/10.1021/acs.est.1c01907.
- (27) Kumar, A.; Pastore, P. Lead and cadmium in soft plastic toys http://www.jstor.org/stable/24099126 (accessed Sep 25, 2023).
- (28) Ismail, S. N. S.; Mohamad, N. S.; Karuppiah, K.; Abidin, E. Z.; Rasdi, I.; Praveena, S. M. Heavy metals content in low-priced toys http://www.arpnjournals.org/jeas/research\_papers/rp\_2017/jeas\_0317\_5787.pdf (accessed Sep 25, 2023).
- (29) Oyeyiola, A. O.; Akinyemi, M. I.; Chiedu, I. E.; Fatunsin, O. T.; Olayinka, K. O. Statistical Analyses and Risk Assessment of Potentially Toxic Metals (PTMS) in Children's Toys. *J.*

*Taibah Univ. Sci.* **2017**, *11* (6), 842–849. https://doi.org/10.1016/j.jtusci.2017.02.005.

- (30) Meng, J.; Xu, B.; Liu, F.; Li, W.; Sy, N.; Zhou, X.; Yan, B. Effects of Chemical and Natural Ageing on the Release of Potentially Toxic Metal Additives in Commercial PVC Microplastics. *Chemosphere* **2021**, *283* (October 2020), 131274. https://doi.org/10.1016/j.chemosphere.2021.131274.
- (31) Turner, A.; Filella, M. Polyvinyl Chloride in Consumer and Environmental Plastics, with a Particular Focus on Metal-Based Additives. *Environ. Sci. Process. Impacts* **2021**, *23* (9), 1376–1384. https://doi.org/10.1039/d1em00213a.
- (32) Wahl, H. G.; Hoffmann, A.; Häring, H.-U.; Liebich, H. M. Identification of Plasticizers in Medical Products by a Combined Direct Thermodesorption–Cooled Injection System and Gas Chromatography–Mass Spectrometry. *J. Chromatogr. A* **1999**, *847* (1–2), 1–7. https://doi.org/10.1016/S0021-9673(99)00138-7.
- (33) Wang, Q.; Storm, B. K. Separation and Analysis of Low Molecular Weight Plasticizers in Poly(Vinyl Chloride) Tubes. *Polym. Test.* **2005**, *24* (3), 290–300. https://doi.org/10.1016/j.polymertesting.2004.12.002.
- (34) Welle, F.; Wolz, G.; Franz, R. Migration of plasticizers from PVC tubes into enteral feeding solutions http://pieweb.plasteurope.com/members/pdf/P204322b.PDF (accessed Sep 25, 2023).
- (35) Radaniel, T.; Genay, S.; Simon, N.; Feutry, F.; Quagliozzi, F.; Barthélémy, C.; Lecoeur, M.; Sautou, V.; Décaudin, B.; Odou, P.; Bernard, L.; Bourdeaux, D.; Chennell, P.; Richard, D.; Pereira, B.; Azaroual, N.; Christine Barthélémy; Décaudin, B.; Dine, T.; Feutry, F.; Genay, S.; Kambia, N.; Lecoeur, M.; Odou, P.; Simon, N.; Vaccher, C.; Cueff, R.; Feschet, E.; Breysse, C. Quantification of Five Plasticizers Used in PVC Tubing through High Performance Liquid Chromatographic-UV Detection. *J. Chromatogr. B Anal. Technol. Biomed. Life Sci.* **2014**, *965*, 158–163. https://doi.org/10.1016/j.jchromb.2014.06.027.
- (36) Bernard, L.; Cueff, R.; Breysse, C.; Décaudin, B.; Sautou, V. Migrability of PVC Plasticizers from Medical Devices into a Simulant of Infused Solutions. *Int. J. Pharm.* **2015**, *485* (1–2), 341–347. https://doi.org/10.1016/j.ijpharm.2015.03.030.
- (37) Bourdeaux, D.; Yessaad, M.; Chennell, P.; Larbre, V.; Eljezi, T.; Bernard, L.; Sautou, V.; Azaroual, N.; Barthelémy, C.; Décaudin, B.; Dine, T.; Feutry, F.; Genay, S.; Kambia, N. las; Lecoeur, M.; Masse, M.; Odou, P.; Simon, N.; Vaccher, C.; Daudet, X.; Richard, D.; Pereira, B.; Clauson, H.; Cueff, R.; Feschet, E.; Breysse, C. Analysis of PVC Plasticizers in Medical Devices and Infused Solutions by GC-MS. *J. Pharm. Biomed. Anal.* **2016**, *118*, 206–213. https://doi.org/10.1016/j.jpba.2015.10.034.
- (38) Faessler, D.; McCombie, G.; Biedermann, M.; Felder, F.; Subotic, U. Leaching of Plasticizers from Polyvinylchloride Perfusion Lines by Different Lipid Emulsions for Premature Infants under Clinical Conditions. *Int. J. Pharm.* **2017**, *520* (1–2), 119–125. https://doi.org/10.1016/j.ijpharm.2017.01.046.
- (39) Jeon, S. H.; Kim, Y. P.; Kho, Y.; Shin, J. H.; Ji, W. H.; Ahn, Y. G. Development and Validation of Gas Chromatography-Triple Quadrupole Mass Spectrometric Method for Quantitative Determination of Regulated Plasticizers in Medical Infusion Sets. *J. Anal. Methods Chem.* **2018**, *2018*. https://doi.org/10.1155/2018/9470254.
- (40) Fernandez-Canal, C.; Pinta, P. G.; Eljezi, T.; Larbre, V.; Kauffmann, S.; Camilleri, L.; Cosserant, B.; Bernard, L.; Pereira, B.; Constantin, J. M.; Grimandi, G.; Sautou, V. Patients' Exposure to PVC Plasticizers from ECMO Circuits. *Expert Rev. Med. Devices* **2018**, *15* (5), 377–383. https://doi.org/10.1080/17434440.2018.1462698.
- (41) Den Braver-Sewradj, S. P.; Piersma, A.; Hessel, E. V. S. An Update on the Hazard of and Exposure to Diethyl Hexyl Phthalate (DEHP) Alternatives Used in Medical Devices. *Crit. Rev. Toxicol.* **2020**, *50* (8), 650–672. https://doi.org/10.1080/10408444.2020.1816896.
- (42) Rastogi, S. C. Gas Chromatographic Analysis of Phthalate Esters in Plastic Toys. *Chromatographia* **1998**, *47* (11–12), 724–726. https://doi.org/10.1007/BF02467461.
- (43) United States Consumer Product Safety Comission (US CPSC). Phthalates and Phthalate Substitutes in Children's Toys https://www.cpsc.gov/s3fs-public/phthallab.pdf (accessed Sep 25, 2023).
- (44) Al-Natsheh, M.; Alawi, M.; Fayyad, M.; Tarawneh, I. Simultaneous GC-MS Determination of Eight Phthalates in Total and Migrated Portions of Plasticized Polymeric Toys and Childcare Articles. *J. Chromatogr. B Anal. Technol. Biomed. Life Sci.* **2015**, *985*, 103–109. https://doi.org/10.1016/j.jchromb.2015.01.010.
- (45) McCombie, G.; Biedermann, S.; Suter, G.; Biedermann, M. Survey on Plasticizers Currently Found in PVC Toys on the Swiss Market: Banned Phthalates Are Only a Minor Concern. *J. Environ. Sci. Heal. - Part A Toxic/Hazardous Subst. Environ. Eng.* **2017**, *52* (5), 491–496. https://doi.org/10.1080/10934529.2016.1274176.
- (46) Ashworth, M.; Chappell, A.; Ashmore, E.; Fowles, J. Analysis and Assessment of Exposure to Selected Phthalates Found in Children's Toys in Christchurch, New Zealand. *Int. J. Environ. Res. Public Health* **2018**, *15* (2), 200. https://doi.org/10.3390/ijerph15020200.
- (47) Giovanoulis, G.; Bui, T.; Xu, F.; Papadopoulou, E.; Padilla-Sanchez, J. A.; Covaci, A.; Haug, L. S.; Cousins, A. P.; Magnér, J.; Cousins, I. T.; de Wit, C. A. Multi-Pathway Human Exposure Assessment of Phthalate Esters and DINCH. *Environ. Int.* **2018**, *112* (April 2017), 115–126. https://doi.org/10.1016/j.envint.2017.12.016.
- (48) Schettler, T.; Skakkebæk, N. E.; De Kretser, D.; Leffers, H. Human Exposure to Phthalates via Consumer Products. *Int. J. Androl.* **2006**, *29* (1), 134–139. https://doi.org/10.1111/j.1365-2605.2005.00567.x.
- (49) Xu, Y.; Cohen Hubal, E. A.; Little, J. C. Predicting Residential Exposure to Phthalate Plasticizer Emitted from Vinyl Flooring: Sensitivity, Uncertainty, and Implications for Biomonitoring. *Environ. Health Perspect.* **2010**, *118* (2), 253–258. https://doi.org/10.1289/ehp.0900559.
- (50) Little, J. C.; Weschler, C. J.; Nazaroff, W. W.; Liu, Z.; Cohen Hubal, E. A. Rapid Methods to Estimate Potential Exposure to Semivolatile Organic Compounds in the Indoor Environment. *Environ. Sci. Technol.* **2012**, *46* (20), 11171–11178. https://doi.org/10.1021/es301088a.
- (51) Eichler, C. M. A.; Hubal, E. A. C.; Xu, Y.; Cao, J.; Bi, C.; Weschler, C. J.; Salthammer, T.; Morrison, G. C.; Koivisto, A. J.; Zhang, Y.; Mandin, C.; Wei, W.; Blondeau, P.; Poppendieck, D.; Liu, X.; Delmaar, C. J. E.; Fantke, P.; Jolliet, O.; Shin, H. M.; Diamond, M. L.; Shiraiwa, M.; Zuend, A.; Hopke, P. K.; Von Goetz, N.; Kulmala, M.; Little, J. C.

Assessing Human Exposure to SVOCs in Materials, Products, and Articles: A Modular Mechanistic Framework. *Environ. Sci. Technol.* **2021**, *55* (1), 25–43. https://doi.org/10.1021/acs.est.0c02329.

- (52) Kim, H.-H.; Yang, J.-Y.; Kim, S.-D.; Yang, S.-H.; Lee, C.-S.; Shin, D.-C.; Lim, Y.-W. Health Risks Assessment in Children for Phthalate Exposure Associated with Childcare Facilities and Indoor Playgrounds. *Environ. Health Toxicol.* **2011**, *26*, e2011008. https://doi.org/10.5620/eht.2011.26.e2011008.
- (53) Koch, H. M.; Lorber, M.; Christensen, K. L. Y.; Pälmke, C.; Koslitz, S.; Brüning, T. Identifying Sources of Phthalate Exposure with Human Biomonitoring: Results of a 48h Fasting Study with Urine Collection and Personal Activity Patterns. *Int. J. Hyg. Environ. Health* **2013**, *216* (6), 672–681. https://doi.org/10.1016/j.ijheh.2012.12.002.
- (54) Wang, W.; Wu, F. Y.; Huang, M. J.; Kang, Y.; Cheung, K. C.; Wong, M. H. Size Fraction Effect on Phthalate Esters Accumulation, Bioaccessibility and in Vitro Cytotoxicity of Indoor/Outdoor Dust, and Risk Assessment of Human Exposure. *J. Hazard. Mater.* **2013**, *261*, 753–762. https://doi.org/10.1016/j.jhazmat.2013.04.039.
- (55) Clausen, P. A.; Liu, Z.; Kofoed-Sørensen, V.; Little, J.; Wolkoff, P. Influence of Temperature on the Emission of Di-(2-Ethylhexyl)Phthalate (DEHP) from PVC Flooring in the Emission Cell FLEC. *Environ. Sci. Technol.* **2012**, *46* (2), 909–915. https://doi.org/10.1021/es2035625.
- (56) Lucattini, L.; Poma, G.; Covaci, A.; de Boer, J.; Lamoree, M. H.; Leonards, P. E. G. A Review of Semi-Volatile Organic Compounds (SVOCs) in the Indoor Environment: Occurrence in Consumer Products, Indoor Air and Dust. *Chemosphere* **2018**, *201*, 466–482. https://doi.org/10.1016/j.chemosphere.2018.02.161.
- (57) Yang, C.; Harris, S. A.; Jantunen, L. M.; Kvasnicka, J.; Nguyen, L. V.; Diamond, M. L. Phthalates: Relationships between Air, Dust, Electronic Devices, and Hands with Implications for Exposure. *Environ. Sci. Technol.* **2020**, *54* (13), 8186–8197. https://doi.org/10.1021/acs.est.0c00229.
- (58) Sukiene, V.; Gerecke, A. C.; Park, Y. M.; Zennegg, M.; Bakker, M. I.; Delmaar, C. J. E.; Hungerbühler, K.; Von Goetz, N. Tracking SVOCs' Transfer from Products to Indoor Air and Settled Dust with Deuterium-Labeled Substances. *Environ. Sci. Technol.* **2016**, *50* (8), 4296–4303. https://doi.org/10.1021/acs.est.5b05906.
- (59) Dodson, R. E.; Camann, D. E.; Morello-Frosch, R.; Brody, J. G.; Rudel, R. A. Semivolatile Organic Compounds in Homes: Strategies for Efficient and Systematic Exposure Measurement Based on Empirical and Theoretical Factors. *Environ. Sci. Technol.* **2015**, *49* (1), 113–122. https://doi.org/10.1021/es502988r.
- (60) Schossler, P.; Schripp, T.; Salthammer, T.; Bahadir, M. Beyond Phthalates: Gas Phase Concentrations and Modeled Gas/Particle Distribution of Modern Plasticizers. *Sci. Total Environ.* **2011**, *409* (19), 4031–4038. https://doi.org/10.1016/j.scitotenv.2011.06.012.
- (61) Garrido, J. A.; Parthasarathy, S.; Moschet, C.; Young, T. M.; McKone, T. E.; Bennett, D. H. Exposure Assessment for Air-To-Skin Uptake of Semivolatile Organic Compounds (SVOCs) Indoors. *Environ. Sci. Technol.* **2019**, *53* (3), 1608–1616. https://doi.org/10.1021/acs.est.8b05123.
- (62) Larsson, K.; Lindh, C. H.; Jönsson, B. A.; Giovanoulis, G.; Bibi, M.; Bottai, M.; Bergström, A.; Berglund, M. Phthalates, Non-Phthalate Plasticizers and Bisphenols in Swedish Preschool Dust in Relation to Children's Exposure. *Environ. Int.* **2017**, *102*, 114–124. https://doi.org/10.1016/j.envint.2017.02.006.
- (63) Zhang, Q.; Sun, Y.; Zhang, Q.; Hou, J.; Wang, P.; Kong, X.; Sundell, J. Phthalate Exposure in Chinese Homes and Its Association with Household Consumer Products. *Sci. Total Environ.* **2020**, *719*, 136965. https://doi.org/10.1016/j.scitotenv.2020.136965.
- (64) Nagorka, R.; Birmili, W.; Schulze, J.; Koschorreck, J. Diverging Trends of Plasticizers (Phthalates and Non-Phthalates) in Indoor and Freshwater Environments—Why? *Environ. Sci. Eur.* **2022**, *34* (1). https://doi.org/10.1186/s12302-022-00620-4.
- (65) Guo, Y.; Kannan, K. Comparative Assessment of Human Exposure to Phthalate Esters from House Dust in China and the United States. *Environ. Sci. Technol.* **2011**, *45* (8), 3788–3794. https://doi.org/10.1021/es2002106.
- (66) Zhao, A.; Wang, L.; Pang, X.; Liu, F. Phthalates in Skin Wipes: Distribution, Sources, and Exposure via Dermal Absorption. *Environ. Res.* **2022**, *204* (PB), 112041. https://doi.org/10.1016/j.envres.2021.112041.
- (67) Rudel, R. A.; Camann, D. E.; Spengler, J. D.; Korn, L. R.; Brody, J. G. Phthalates, Alkylphenols, Pesticides, Polybrominated Diphenyl Ethers, and Other Endocrine-Disrupting Compounds in Indoor Air and Dust. *Environ. Sci. Technol.* **2003**, *37* (20), 4543– 4553. https://doi.org/10.1021/es0264596.
- (68) Fromme, H.; Schütze, A.; Lahrz, T.; Kraft, M.; Fembacher, L.; Siewering, S.; Burkardt, R.; Dietrich, S.; Koch, H. M.; Völkel, W. Non-Phthalate Plasticizers in German Daycare Centers and Human Biomonitoring of DINCH Metabolites in Children Attending the Centers (LUPE 3). *Int. J. Hyg. Environ. Health* **2016**, *219* (1), 33–39. https://doi.org/10.1016/j.ijheh.2015.08.002.
- (69) Nagorka, R.; Conrad, A.; Scheller, C.; Süßenbach, B.; Moriske, H. J. Diisononyl 1,2- Cyclohexanedicarboxylic Acid (DINCH) and Di(2-Ethylhexyl) Terephthalate (DEHT) in Indoor Dust Samples: Concentration and Analytical Problems. *Int. J. Hyg. Environ. Health* **2011**, *214* (1), 26–35. https://doi.org/10.1016/j.ijheh.2010.08.005.
- (70) Mansouri, K.; Grulke, C. M.; Judson, R. S.; Williams, A. J. OPERA Models for Predicting Physicochemical Properties and Environmental Fate Endpoints. *J. Cheminform.* **2018**, *10* (1), 1–19. https://doi.org/10.1186/s13321-018-0263-1.
- (71) Wang, X.; Song, M.; Guo, M.; Chi, C.; Mo, F.; Shen, X. Pollution Levels and Characteristics of Phthalate Esters in Indoor Air in Hospitals. *J. Environ. Sci. (China)* **2015**, *37*, 67–74. https://doi.org/10.1016/j.jes.2015.02.016.
- (72) Zhang, L.; Wang, F.; Ji, Y.; Jiao, J.; Zou, D.; Liu, L.; Shan, C.; Bai, Z.; Sun, Z. Phthalate Esters (PAEs) in Indoor PM10/PM2.5 and Human Exposure to PAEs via Inhalation of Indoor Air in Tianjin, China. *Atmos. Environ.* **2014**, *85*, 139–146. https://doi.org/10.1016/j.atmosenv.2013.11.068.
- (73) Peeters, J. R.; Vanegas, P.; Kellens, K.; Wang, F.; Huisman, J.; Dewulf, W.; Duflou, J. R. Forecasting Waste Compositions: A Case Study on Plastic Waste of Electronic Display Housings. *Waste Manag.* **2015**, *46*, 28–39. https://doi.org/10.1016/j.wasman.2015.09.019.
- (74) Bui, T. T.; Giovanoulis, G.; Cousins, A. P.; Magnér, J.; Cousins, I. T.; de Wit, C. A. Human Exposure, Hazard and Risk of Alternative Plasticizers to Phthalate Esters. *Sci. Total Environ.* **2016**, *541*, 451–467. https://doi.org/10.1016/j.scitotenv.2015.09.036.
- (75) European Commission Directorate-General for Health and Food Safety. *The Safety of Medical Devices Containing DEHP Plasticized PVC or Other Plasticizers on Neonates and Other Groups Possibly at Risk (2015 Update)*; Brussels, Belgium, 2016; Vol. 76. https://doi.org/10.1016/j.yrtph.2016.01.013.
- (76) Panneel, L.; Cleys, P.; Breugelmans, C.; Christia, C.; Malarvannan, G.; Poma, G.; Jorens, P. G.; Mulder, A.; Covaci, A. Neonatal Exposure to Phthalate and Alternative Plasticizers via Parenteral Nutrition. *Int. J. Pharm.* **2023**, *631* (December 2022), 122472. https://doi.org/10.1016/j.ijpharm.2022.122472.