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<b>Abstract:</b>	Millions of oceanic dolphins are exposed and presumably affected by military sonar annually; however empirically measuring its impact on the behavior of free-ranging dolphins has proven logistically challenging. Additionally, baseline variability or the frequency of vocal state-switching during undisturbed conditions is lacking among social oceanic delphinids, making it difficult to attribute changes in vocal behavior to anthropogenic disturbance. Using a network of drifting acoustic buoys and a three-phased controlled exposure experiment, we investigated the effects of mid-frequency (3-4 kHz) active sonar (MFAS) on whistle production in short-beaked ( <i>Delphinus delphis</i> ) and long-beaked common dolphins ( <i>D. bairdii</i> ) in southern California. Given the complexity of acoustic behavior exhibited by these group-living species, we conducted our response analysis over varying temporal windows (10 min – 5 sec) to describe both longer-term and instantaneous changes in sound production. We found that common dolphins exhibited an acute and dramatic change in acoustic behavior in the 5-seconds following exposure to experimental Navy sonar. In some cases, this response was sustained throughout repeated exposures to sonar in the experiment, suggesting that dolphins may not habituate to this kind of disturbance. These results suggest that common dolphins exhibit a brief, yet dramatic acoustic response to MFAS, and highlights how different sized temporal windows – tuned to key aspects of baseline vocal behavior as well as experimental parameters related to MFAS exposure – impacts the detection of behavioral responses. We suggest that future work with oceanic delphinid species explore baseline vocal rates a-priori and use information on the rate of change in vocal behavior to inform the analysis time-window over which behavioral responses are measured.
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All exported raw whistle counts from PAMGuard and associated buoy distances for each playback experiment will be made available via Dryad data repository <http://datadryad.org/>.



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2 **Short title: Dolphin acoustic response to sonar**

3

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18 **Abstract**

19 Millions of oceanic dolphins are exposed and presumably affected by military sonar annually;  
20 however empirically measuring its impact on the behavior of free-ranging dolphins has proven  
21 logistically challenging. Additionally, baseline variability or the frequency of vocal state-  
22 switching during undisturbed conditions is lacking among social oceanic delphinids, making it  
23 difficult to attribute changes in vocal behavior to anthropogenic disturbance. Using a network of  
24 drifting acoustic buoys and a three-phased controlled exposure experiment, we investigated the  
25 effects of mid-frequency (~~3-4 kHz~~) active sonar (MFAS) on whistle production in short-beaked  
26 (*Delphinus delphis*) and long-beaked common dolphins (*Delphinus bairdii*) in southern  
27 California. Given the complexity of acoustic behavior exhibited by these group-living species,  
28 we conducted our response analysis over varying temporal windows (10 min – 5 sec) to describe  
29 both longer-term and instantaneous changes in sound production. We found that common  
30 dolphins exhibited an acute and dramatic change in acoustic behavior in the 5-seconds following  
31 exposure to experimental Navy sonar. In some cases, this response was sustained throughout  
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35 aspects of baseline vocal behavior as well as experimental parameters related to MFAS exposure  
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37 delphinid species explore baseline vocal rates a-priori and use information on the rate of change  
38 in vocal behavior to inform the analysis time-window over which behavioral responses are  
39 measured.

40

41 **Introduction**

42 Sound production and reception play a critical role in the lives of cetaceans, aiding in important  
43 life-history events including maintenance of social relationships, coordination of group  
44 movement, foraging, and evasion of predators (1). Consequently, substantial effort has been  
45 directed toward describing cetacean acoustic behavior (2,3) and evaluating how it is impacted by  
46 human-generated disturbance (4-8). There are many sources of anthropogenic noise pollution  
47 (e.g., ~~vessel noise, oil and gas exploration, construction and facilities maintenance, fisheries and~~  
48 ~~aquaculture, military activity~~) which can have varying short and long-term impacts on marine  
49 mammal ~~health and behavior~~ (4, 9-11). Concentrated research efforts aimed at characterizing  
50 these impacts have led to the systematic development of acoustic exposure criteria, informing  
51 and improving effective management strategies for both regulators and industries (for ~~reviews on~~  
52 ~~auditory and behavioral~~ criteria, see 12). Such ~~assessments and criteria~~ have also highlighted  
53 species and noise exposure contexts for which information is sparse or unavailable.

54  
55 Among cetaceans, oceanic delphinids represent an important and logistically challenging group  
56 of species for which to evaluate the impact of anthropogenic noise on vocal behavior. These  
57 animals are highly soniferous, abundant, and often extremely gregarious (pods of > 500  
58 individuals are common for some species). Sound production has been demonstrated to play a  
59 particularly important role in the regulation of social interactions and cohesion among group  
60 members (e.g., 2,3). Oceanic delphinids are ubiquitous around some U.S. Navy operational areas  
61 where mid-frequency active sonar (MFAS; 1-10 kHz) is commonly used for submarine detection  
62 in training exercises ~~and warfare~~, resulting in associated large numbers of sonar exposures for  
63 these federally protected species. Some of the most powerful MFAS systems (e.g., ~~SQS-53C~~)

64 use repeated pings with fundamental frequencies in the 3-4 kHz range, ping lengths of  
65 approximately 1-3 seconds, and effective source levels as high as 235 dB re 1  $\mu$ Pa that may be  
66 transmitted for several minutes to hours at high duty cycles (13). Aside from the elevated  
67 background noise and potential disturbance that may result from these training exercises (*e.g.*,  
68 14), MFAS signals overlap with the frequencies that oceanic dolphins commonly rely on for  
69 social sound (whistle) production. Delphinid whistles are narrowband tonal sounds with most of  
70 the acoustic energy concentrated below 20 kHz (14-16).

71  
72 While MFAS has been linked to mass stranding events of cetaceans (13,17) and its effect on  
73 cetaceans has been experimentally evaluated in a handful of species (for a recent review, see 19)  
74 the effects of sonar on the acoustic behavior of oceanic delphinids have not been systematically  
75 explored. This is due in large part to the logistical challenges of applying previously developed  
76 methods used in other behavioral response studies of individual animals to large aggregations of  
77 dolphins. Much of the prior research on cetacean behavioral responses to noise has capitalized on  
78 the use of suction-cup attached motion-sensing and acoustic recording tags to characterize  
79 responses following controlled exposure to MFAS (21-23). Unfortunately, such tags are  
80 challenging to deploy and are easily shed by small dolphins due to the tag size relative to the  
81 smaller body surface of dolphins and high drag due to fast swimming speeds. Additionally,  
82 oceanic delphinids commonly occur in large groups that display remarkable coordination,  
83 making the collective vocal behavior of the group perhaps a more appropriate focus of analysis  
84 (20). Opportunistic passive acoustic studies relying on large, cabled hydrophone arrays have  
85 been used to quantify changes in vocal activity and thus as a proxy for presence of or absence of  
86 multiple animals before, during, and after exposure to MFAS (24,25). Such experiments are

87 valuable but require extensive, high-cost moored hydrophone arrays with restricted spatial  
88 coverage.

89

90 ~~Findings from previous studies of acoustic responses to Navy sonar in oceanic delphinids include~~  
91 ~~shifting specific frequency~~ components of whistle contours, increasing or decreasing calling rate,  
92 increasing call amplitude, and even mimicry of MFAS elements (14, 26-28). For example,  
93 tagged orcas (*Orcinus orca*) adjust the high-frequency component of their whistles during sonar  
94 exposure, and increase the number and amplitude of their calls following each ping (27). False  
95 killer whales (*Pseudorca crassidens*) appear to increase their whistle rate and produce more  
96 MFA-like whistles after exposure to simulated sonar (26).

97

98 One opportunistic study provided initial insights into the behavioral responses of some social  
99 oceanic delphinids to MFAS. Bottlenose dolphins (*Tursiops truncatus*), common dolphins  
100 (*Delphinus sp.*), Pacific white-sided dolphins (*Lagenorhynchus obliquidens*), and Risso's  
101 dolphins (*Grampus griseus*) ~~exposed to incidental~~ MFAS showed a cessation of vocalizations, an  
102 increase in the intensity of vocalizations, or a combination of both (14). Of all delphinid species,  
103 common dolphins displayed the widest range of responses, including changing their behavioral  
104 state or direction of travel when sonar stopped, increasing the intensity of vocalizations when  
105 sonar began, vocalizing very little or not at all during sonar, or a combination of these  
106 observations (14). These results are consistent with a more recent opportunistic evaluation of  
107 delphinid responses to an underwater ~~explosion~~, which showed that whistle rate, complexity, and  
108 frequency content all varied in response to the explosive event (30). Depending upon the  
109 frequency, intensity, and consistency of these kinds of noise exposures, such behavioral changes

110 could result in physiological consequences that impact overall population health  
111 (31). Unfortunately, information on baseline variability or the frequency of vocal state-switching  
112 during undisturbed conditions is lacking among social oceanic delphinids, making it challenging  
113 to interpret the responses observed. While opportunistic studies of delphinid acoustic responses  
114 to sonar are insightful, a detailed assessment under controlled experimental conditions is needed  
115 to understand the extent to which delphinids are impacted by this kind of disturbance.

116  
117 Quantifying vocal behavior in these taxa is complicated by the fact that dolphin acoustic  
118 behavior is dynamic, variable, and influenced by a myriad of social and environmental factors  
119 (32). Measurements of acoustic behavior and how it changes in response to disturbance must be  
120 made across some predetermined interval of time. In previous cetacean behavioral response  
121 studies using a conventional controlled exposure experimental (CEE) design, this temporal  
122 window was often dictated by logistical limitations of the technology being used (*e.g.*, battery  
123 power of tags, the feasibility of continuous behavioral observations, etc.), or designed to match  
124 the duration of anthropogenic noise source being evaluated. However, averaging vocal behavior  
125 over long time windows may result in missing instantaneous or shorter duration responses that  
126 occur at the onset of exposure or at scales that are more biologically meaningful to the  
127 individuals exposed. One way to address this is to examine acoustic metrics computed over a  
128 variety of time windows to determine if and/or when we can attribute a change in vocal behavior  
129 to a known, controlled disturbance.

130  
131 ~~Here, we apply and compare different broad and fine-scale analytical approaches to analyzing~~  
132 delphinid vocal behavior and effects of controlled MFAS on whistle production using

133 ~~experimental methods~~ in short-beaked (*Delphinus delphis*) and long-beaked common dolphins  
134 (*Delphinus bairdii*) in southern California. By assessing group-level vocal behavior across  
135 different time scales, we ~~aim to~~:

- 136 1. Describe the variability in baseline vocal behavior of common dolphin aggregations  
137 during control conditions.
- 138  
139 2. ~~Characterize the types of~~ vocal responses detected during a controlled exposure to MFAS  
140 ~~using broad and fine-scale temporal resolution.~~

141 Given the need to establish sampling regimes that can be applied and compared across studies,  
142 ~~our objective is~~ to provide an informative framework for assessing the complex acoustic  
143 behavior exhibited by group-living species. We highlight how using different sized temporal  
144 windows – tuned to key aspects of baseline vocal behavior as well as experimental parameters  
145 related to MFAS exposure – impacts the detection of behavioral responses.

146

## 147 **Methods**

### 148 **A. General Methods**

149 ~~Controlled exposure experiments (CEEs)~~ were conducted with two ~~species~~ of common dolphins  
150 as part of a broader effort aimed at quantifying group-level responsiveness of oceanic delphinids  
151 to military sonar using CEEs. Common dolphins occur as two separate species within our study  
152 area (33), around Santa Catalina Island located off the coast of southern California, USA. Since  
153 they regularly occur in mixed-~~species~~ groups, we ~~col~~lected data for the two species to describe  
154 baseline vocal data and included species as a potential explanatory variable in our models. The  
155 project ~~integrates~~ multiple different data streams, including shore-based tracking of dolphin

156 schools, passive acoustics to record vocal activity, and photogrammetry to measure fine-scale  
157 behavior (34). This work was conducted between 2017-2021. We chose our study area because it  
158 lies near the Southern California Offshore Range (SCORE) – a tactical training area for the U.S.  
159 Navy Pacific Fleet located off the west side of San Clemente Island – where animals regularly  
160 encounter the types of signals we used in our experiment.

161  
162 CEEs were composed of three discrete phases: pre-exposure (baseline), exposure using  
163 intermittent simulated MFAS signals, and post-exposure. In a subset of experiments, no MFAS  
164 signals were broadcast during the exposure phase, serving as controls. For details about the  
165 experimental source and sound source characteristics – including calculations of received levels  
166 – see Durban et al. 2022 (34). Each experimental phase was 10 minutes in duration. During  
167 exposure phases, MFAS ‘pings’ of 56 in duration consisting of three tonal and frequency  
168 modulated elements between 3.5-4 kHz were transmitted (see: Southall et al. 2012 (35) for  
169 additional details on the sound source and signal parameters). Pings were emitted at a broadband  
170 source level of 212 dB re 1  $\mu$ Pa RMS every 25 seconds, which is similar in repetition rate, duty  
171 cycle, and the absence of a ramped-up source level (as used in some previous MFAS CEEs) to  
172 some active Navy MFAS systems (e.g., helicopter-dipped sonar systems). Up to 24 total pings  
173 were emitted per 10-minute exposure phase, provided that no permit-mandated shutdowns  
174 occurred for animals occurring within 200 m of the active sound source (this occurred in only  
175 one playback). The sound source was positioned relative to focal animal groups using noise  
176 propagation modeling to ensure received levels at focal animals occurred at maximum levels of  
177 140-160 dB RMS.

178

179 For every CEE, species identity (based on differences in genetics, morphology and pigmentation)  
180 was determined using a combination of aerial images obtained from drone footage, genetic  
181 sequencing from biopsy samples, and visual observation. Additionally, group size was  
182 characterized by experienced shore-based observers using strong magnification binoculars or a  
183 binocular scope located at elevated locations (~70m) that enabled a broad overview of the  
184 research area (up to 20 km from shore). Shore-based tracking of animals could be conducted for  
185 groups up to 7 km from shore. Focal follows included estimation of low, best and high group  
186 size, the number of subgroups (defined as all individuals in closer proximity to each other than to  
187 other individuals in the area), the range of inter-individual spacings within subgroups, and  
188 distances between subgroups. These observations were taken continuously throughout the  
189 experiment at 2-minute intervals. For a full description of each of these methods see Visser et al.,  
190 2014 (36) and Durban et al., 2022 (34).

191

## 192 **B. Acoustic Data Collection and Processing**

### 193 *1. Passive Acoustic Monitoring*

194 Passive acoustic recordings were obtained from each target group of dolphins using three  
195 drifting, remote-deployed acoustic recording units. Up to three separate recording units were  
196 tactically positioned and recovered from a single small (~6 m) rigid-hull inflatable boat, with the  
197 objective of placing one recording unit within 500 m of the predicted trajectory of the dolphins  
198 during each CEE phase (Fig. 1). Each recording unit consisted of a surface buoy and flag with an  
199 underwater recorder suspended by 10 m of line. The recorder was either a SoundTrap ST300  
200 (Ocean Instruments NZ, Auckland, New Zealand) or a SNAP Recorder (Loggerhead  
201 Instruments, Sarasota, FL, USA). Both recorded via a single omnidirectional calibrated



202 hydrophone (SoundTrap: integrated hydrophone, frequency response 0.02–60 kHz  $\pm$  3 dB, end-  
203 to-end sensitivity -178 dB re 1 $\mu$ Pa/V; SNAP: HTI-96-MIN hydrophone, sensitivity -164 dB re  
204 1 $\mu$ Pa/V, frequency response 1.0-20 kHz  $\pm$  3 dB) which was suspended by a shock-mounted  
205 cable at a depth of 10 m. All recording units had a Global Positional System (GPS) tracking  
206 device (Trace, SPOT LLC, Chantilly, VA, USA) that recorded the location of the instrument  
207 once every minute (Fig. 1). Five-minute WAV files were continuously recorded at either 96 kHz  
208 sampling rate with 16-bit resolution (SoundTrap) or 44.1 kHz sampling rate with 16-bit  
209 resolution (SNAP).

210

211 **Figure 1. A schematic representation of the placement of acoustic recorders (see inset for**  
212 **floating acoustic recording unit) relative to the track of the focal group of dolphins.** The  
213 first buoy is placed within 500m of the animals during the pre-exposure, the second during the  
214 exposure, and the third during the post-exposure period. The sound source is located at  
215 approximately 1 km from the dolphins at the onset of the exposure period. The dashed line with  
216 associated arrows represents the movement path of the focal group. Note the presence of the  
217 shore station positioned on land.

218

219 To evaluate which PAM recording unit was closest to the focal group given the dolphins'  
220 frequently unpredictable course, the relative proximity of each hydrophone to the animals was  
221 determined *post-hoc*. The animals' location was known from an associated octocopter drone  
222 flight (APO-42, Aerial Imaging Solutions) centered over the focal group (see 34 for details). The  
223 relative distances (in meters) between the focal group (from the drone's GPS) and each recorder  
224 (from their flag-mounted GPS units) were estimated for every minute of the 30-minute

225 experiment using the Haversine formula and linear interpolation in a custom MATLAB script  
226 (Mathworks, Natick, MA, USA, Fig. 2). Recordings from the buoy closest to the focal group at  
227 1-minute intervals were used for all subsequent analyses. Any recorders that exceeded 1600  
228 meters from the focal group (even if they were the closest recorder deployed) were excluded.  
229 This threshold was selected based on a previous assessment of detection ranges of playbacks of  
230 odontocete whistles (10-20 kHz) by bottom hydrophones in southern California, which  
231 demonstrated a 95% probability of detection of a 135 dB re 1  $\mu$ Pa dolphin whistle at 1600 meters  
232 with an SNR of 2.2 dB (37). This assessment was supported by the drop-off in whistle amplitude  
233 observed in the spectrograms (*post-hoc*) when any buoy surpassed 1600-1800 m distance from  
234 the focal group (Fig. 2).

235

236 **Figure 2. Spectrograms from each of the three recorders strategically placed on the track-**  
237 **line of a moving group of dolphins.** The solid white line represents the distance between the  
238 recorder and the drone flight centered over the focal group (units on the right y-axis), the red  
239 dashed vertical lines denote the experimental period, and the horizontal white dashed line marks  
240 the 1600 m threshold. Estimates of the relative distance between the focal group and each of the  
241 recorders were produced every minute of the 30-minute experiment. The bottom panel shows the  
242 number of whistles detected on the closest recorder using the PAMGuard Whistle and Moan  
243 Detector. Times when the closest buoy switched is indicated by the blue dashed lines and  
244 associated blue numbers. MFAS pings denoted by the vertical lines on the spectrogram during  
245 the experimental period.

246

247 **2. Quantifying whistle production and variability**

248 While common dolphins are known to emit buzzes, echolocation clicks, and whistles, we  
249 focused our efforts on characterizing whistle production since they are the key signal for long  
250 distance communication and play a major role in group cohesion and coordination (2, 38).  
251 Extraction of whole whistle contours in high background noise with overlapping whistles is  
252 extremely challenging and results in high rates of missed detections, irrespective of  
253 methodology. To create a dataset in which error rates were kept constant across different phases  
254 of the experiment, we used the Whistle and Moan Detector (WMD) module in PAMGuard (v  
255 2.01.05, 39). The WMD deals with uncertainties by only detecting parts of whistles that clearly  
256 stand out above noise using standardized settings across extractions. It is important to note that  
257 this often leads to a fragmentation of whistles with one whistle being split into several  
258 independent sections. Thus, whistle counts reported here are not comparable to ones obtained  
259 with whole whistle extraction in other studies. However, for our assessment of changes in vocal  
260 activity between different experimental phases, it was more important to keep error rates  
261 constant to allow for relative comparisons.

262  
263 The WMD operates on the spectrogram output of the PAMGuard FFT Engine module. We  
264 optimized settings for the FFT Engine to provide comparable frequency and temporal resolution  
265 of the calculated spectrograms across the two recorders and sampling rates. For the SoundTrap  
266 recorders, which had a sampling rate of 96 kHz, the FFT Engine module calculated spectrograms  
267 with an FFT length 1024, hop size 512, and a Hann window providing a frequency resolution of  
268 93.75 Hz and time resolution of 10.67 ms. For the SNAP recorders, which had a sampling rate of  
269 44.1 kHz, spectrograms were calculated with a Hann window, FFT length 512 and hop size 256  
270 resulting in a frequency resolution of 86.13 and time resolution of 11.61 ms. The WMD was set

271 to detect whistles between 5 kHz and 20 kHz to exclude detection of the tonal sounds from the  
272 simulated mid-frequency sonar ~~playbacks~~ (below 5 kHz) and to standardize the upper detection  
273 limit across the two sampling rates and avoid any possible edge effects near the Nyquist  
274 frequency of the lower sampling rate. The detection threshold was set at 6.0 dB. Full WMD  
275 settings can be found in the supplementary materials (S1 Appendix). While the fundamental  
276 sonar tonals were excluded by the 5 kHz low-pass cut-off for detections, the high source level of  
277 the simulated MFAS resulted in the presence of harmonics in some of the recordings. All  
278 harmonics were manually annotated in PAMGuard Viewer using the Spectrogram Annotation  
279 module for later removal.

280

281 Whistles were exported from PAMGuard using the PAMGuard MATLAB tools  
282 (<https://github.com/PAMGuard/PAMGuardMatlab>) and R package `PAMPal` (40). MFAS  
283 harmonics were removed, and whistles were quantified at 1-second resolution. Because whistles  
284 are often longer than 1-second in duration, the total number of whistles starting within a 1-  
285 second bin was counted, providing a metric for whistle activity as detected whistles per second.  
286 For brevity, this is referred to as whistle count throughout the remainder of this manuscript.

287

## 288 **C. Characterization of baseline vocal behavior**

### 289 *1. Overall whistle count*

290 To assess common dolphin vocal behavior ~~in~~ control conditions, we calculated the mean and  
291 median whistle count per second for the full 30-minute experiment for each control CEE. These  
292 data were evaluated with reference to concurrent data collected ~~by experienced shore-based focal~~  
293 ~~follow observations~~ that quantified group size.

294

## 295 **2. Changepoint analysis**

296 To describe the natural variability in vocal behavior during control conditions, we applied a  
297 changepoint analytical approach to the control CEE data collected for both common dolphin  
298 species. Change point detection is used to pinpoint times when the probability distribution of a  
299 time series changes (*i.e.*, vocal state changes). The aim is to identify times at which either the  
300 mean or variance deviates from the expected trends in the dataset and estimate the number and  
301 position of all changepoints. Effectively, this approach detects points in time when a significant  
302 change in whistle count occurs. First, a 5-second smoothing window was applied to the raw 1-  
303 second whistle count data. Then, changepoints in both mean whistle count and whistle count  
304 variance were detected using the ‘changepoint’ package in R version 2.2.4 (41). The “BinSeg”  
305 (Binary Segmentation) algorithm was used. This provided the number and locations of all state  
306 changes in both the mean and the variance of whistle count over the total 30-minute sampling  
307 period.

308

### 309 **D. Assessing the impact of disturbance on vocal behavior**

310 We employed a hierarchical approach to characterizing the types of vocal responses that might  
311 be detected during controlled exposure to MFAS using broad and fine-scale time windows (Fig.  
312 3). All statistical analyses were carried out using R version 4.3.1 (R Core Team 2023).

313

314 **Figure 3. Flowchart of methods implemented to assess changes in common dolphin vocal**  
315 **behavior during controlled exposure to MFAS.** Methods include pre-processing of acoustic  
316 data to baseline vocal behavior analysis and hierarchical assessment of disturbance on vocal

317 behavior at four temporal scales. Alpha-numeric references corresponds to the sub-heading in the  
318 methods section.

319

### 320 ***1. Difference in changepoints by period across CEEs***

321 We conducted a changepoint analysis on all CEEs (both controls and MFAS) to evaluate whether  
322 common dolphins change the frequency of vocal state switching as a result of exposure to  
323 MFAS. We used the same general method as in Section C.2 but quantified the number of  
324 changepoints in the 10-minute pre- and 10-minute exposure periods separately. Changepoints  
325 were detected for both the mean and variance of the whistle count data. The difference in the  
326 number of changepoints between the two periods was calculated and an unpaired t-test was used  
327 to evaluate any significant differences between controls and MFAS experiments.

328

### 329 ***2. Characterize the impact of MFAS exposure on whistle count: 10-minute time scale***

330 To identify potential broad scale changes in whistle count in response to ~~a simulated MFAS~~  
331 ~~exposure we pooled all CEEs (both controls and MFAS exposures) and analyzed them~~ using a  
332 Generalized Linear Mixed Model (GLMM) approach, implemented using R package  
333 ``glmmTMB`` (42). We modeled the absolute difference in median whistle count between the 10-  
334 minute pre-exposure period and the 10-minute exposure period (periodDiff) as a function of CEE  
335 type (ceeType; either control or simulated MFAS), a random identity variable (ceeNum), dolphin  
336 ~~species~~ (species), the best estimate of total group size from the shore based observers  
337 (groupSize), and the mean distance between the focal group and the closest buoy for the full  
338 CEE (buoyDistance). Using the absolute value for difference in median whistle count enabled us

339 to explore the magnitude of a potential response. We modeled the relationship using a negative  
340 binomial distribution which fit the count-type data after the transformation. Our full model was:

341

342 
$$peridDiff \sim ceeType + ceeNum + species + groupSize + buoyDist$$

343

344 We used backward elimination and resulting AIC scores to select our final model and present the  
345 ~~results of the final model below.~~

346

### 347 **3. Characterize the impact of MFAS exposure on whistle count: 20-second time scale**

348 To characterize more instantaneous changes in whistle production in response to MFAS  
349 exposure, we compared dolphin whistle counts in the 20 seconds before and 20 seconds after  
350 each individual ping (n = 24 1-second pings per 10-minute experimental period, ~25 seconds  
351 between each ping) for both MFAS experiments and controls. We selected this time window to  
352 capture sustained variation in whistling behavior within a single ping cycle, without overlap  
353 between cycles. Differences between these two sequential time bins were calculated by  
354 subtracting the mean whistle count for the first bin from the mean whistle count of the second  
355 bin (Fig. 4B). The first ping started at time 0, the second ping at time 25 seconds, and so on.  
356 Because no actual pings were present in the control experiments, we calculated the change in  
357 whistle count surrounding ‘ghost pings’ ~~which were chosen to be timed~~ at the same time as when  
358 real pings would have occurred during a MFAS CEE.

359

360 **Figure 4. Example plots of (A) raw whistle counts over time, (B) changes in whistle count**  
361 **between 20 second duration sequential bins, and (C) changes in whistle count between 5**

362 **second duration sequential bins, for the pre-exposure, exposure, and post-exposure period**  
363 **within one CEE. The dashed vertical red line indicates the onset of exposure, and the**  
364 **sequential gray dashed lines represent each ping within the exposure period.** Pre-exposure  
365 and post-exposure periods were not included in modeling analysis but are presented here for  
366 reference.

367

368 Similar to the analysis at the 10-minute time scale, we used a ~~generalized linear mixed model~~  
369 approach (using R package `glmmTMB`) to identify potential significant differences  
370 immediately following pings (*pingChange*) in MFAS experiments compared to controls where  
371 no pings were present. In addition to the previous fixed effects included at the 10-minute scale  
372 (*ceeType*, *ceeNum*, *species*, *buoyDistance*, and *groupSize*), we also included median whistle  
373 count per second for the entire experimental period to account for the varied baseline whistling  
374 activity across CEEs (*medWhist*).

375

376 
$$pingChange \sim ceeType + ceeNum + species + groupSize + buoyDist + medWhist$$

377

378 Binned ping change data were generally normally distributed, but were zero inflated, so we used  
379 a Gaussian distribution for the primary model and additionally modeled the zero inflation as a  
380 function of median whistle count (*medWhist*). We used backward elimination,  $\Delta AIC$ , and  
381 analysis of variance (ANOVA) to select the best model.

382

383 ***4. Characterize the impact of MFAS exposure on whistle count: 5-second time scale***



384 We repeated the analysis conducted in section D3, but over a shorter 5 second time window to  
385 investigate potential shorter duration changes immediately following pings (Fig. 4C).

386  
387 Similar to the above analysis, we used a ~~generalized linear mixed model~~ approach to identify  
388 potential significant differences in whistle count changes in the 5 seconds following compared to  
389 the 5 seconds before each ping (*pingChange*) for MFAS experiments compared to controls. We  
390 used the same fixed effects implemented at the 20-second scale (*ceeType*, *ceeNum*, *species*,  
391 *buoyDistance*, and *groupSize*, *medWhist*) but also included an autocorrelation structure to this  
392 analysis *AR(1)* to account for clear temporal lag effects in exploratory plots.

393  
394 
$$pingChange \sim ceeType + ceeNum + species + groupSize + buoyDist + medWhist + AR(1)$$

395  
396 Like the 20-second scale, binned ping change data were normally distributed and zero inflated; a  
397 Gaussian distribution was used for the conditional model and zero-inflation was modeled as a  
398 function of median whistle count (*medWhist*). We used backward elimination,  $\Delta AIC$ , and  
399 ANOVA to select the final model.

400

#### 401 ***5. CEE-by-CEE analysis***

402 We assessed each playback individually at each time scale to better contextualize the severity,  
403 persistence, and directionality (*i.e.*, increase or decrease in whistle count) of responses, which  
404 were not measured by either the modeling approach or changepoint analysis. The methods and  
405 results for the CEE-by-CEE analysis can be found in the supplementary materials (S2  
406 Appendix).

407

## 408 **Ethics Statement**

409 Animal research was conducted under authorization of the United States National Marine  
410 Fisheries Service marine mammal research permits 19116 and 19091.

411

## 412 **Results**

### 413 **A. Characterization of baseline vocal behavior**

#### 414 *1. Basic Description*

415 This analysis includes nine control experiments, each conducted on separate days. Four of these  
416 experiments were conducted with *Delphinus delphis* and five were conducted with *Delphinus*  
417 *bairdii* (Table 1). This resulted in 270 minutes of baseline acoustic data for both species  
418 (pooled). The average group size across species was 190 individuals (range 45-300 animals). The  
419 dispersion of animals varied considerably within and between control experiments, including  
420 small to large groups (55-300 individuals) in tight to loose organization, joint (in a single group  
421 with no subgroups), or spread out over several subgroups (range: 2-6) at tens to several hundreds  
422 of meters apart (range: 10-800 m).

423

424 Whistles were successfully detected across all control experimental deployments. Mean (SD) and  
425 Median (IQR) whistle count per second varied between control CEEs (Table 1). In our  
426 assessment of the control experiments, we found the total number of whistles varied between  
427 0.3-4.6 whistles/second. Note, however, that the inherent uncertainty error in our group size  
428 assessment for large groups did not allow us to calculate accurate whistle rates/individual (see  
429 [table 1](#) for group size estimates and whistle counts).

CEE ID	species	estimated group size	type	RL (Max)	RL (Range)	whistles per second Median [IQR] (full 30 min period)	# changepoint s - mean - pre-exposure	# changepoint s - mean - exposure	# changepoint s - variance - pre-exposure	# changepoint s - variance - exposure
2019_01	Db	260	simMFAS	147 dB re 1µPa	140-147 dB re 1µPa	6.96 [6.39]	26	39	5	1
2019_02	Dd	350	control	n/a	n/a	4.80 [5.19]	37	29	0	5
2019_04	Db	200	control	n/a	n/a	0.25 [0.77]	2	0	4	2
2019_06	Db	45	control	n/a	n/a	0.32 [1.38]	5	0	4	9
2019_07	Db	300	simMFAS	154 dB re 1µPa	150-154 dB re 1µPa	2.88 [3.76]	36	20	3	3
2019_08	Db	250	simMFAS	142 dB re 1µPa	131-142 dB re 1µPa	1.50 [3.91]	30	12	4	4
2019_09	Dd	250	control	n/a	n/a	4.21 [5.38]	46	44	2	4
2019_10	Dd	30	simMFAS	149 dB re 1µPa	146-149 dB re 1µPa	0.091 [0.38]	0	0	3	7
2021_01	Db	150	control	n/a	n/a	3.14 [4.47]	3	51	1	3
2021_02	Db	200	control	n/a	n/a	4.00 [3.78]	18	29	0	4
2021_03	Dd	150	control	n/a	n/a	0.99 [2.01]	11	0	2	0
2021_04	Db	150	control	n/a	n/a	4.61 [4.15]	27	37	4	2
2021_05	Dd	250	control	n/a	n/a	0.36 [1.87]	0	5	8	4
2021_08	Db	30	simMFAS	153 dB re 1µPa	145-153 dB re 1µPa	14.04 [7.75]	50	49	2	4
2021_09	Db	200	simMFAS	157 dB re 1µPa	152-157 dB re 1µPa	1.17 [2.17]	3	6	2	2
2021_10	Db	300	simMFAS	159 dB re 1µPa	150-159 dB re 1µPa	14.62 [9.41]	42	61	3	2
2021_11	Db	10	simMFAS	153 dB re 1µPa	150-153 dB re 1µPa	0.0029 [0.063]	0	0	0	6
2021_12	Dd	150	simMFAS	152 dB re 1µPa	149-152 dB re 1µPa	2.85 [3.07]	7	24	4	6
2021_13	Dd	200	simMFAS	147 dB re 1µPa	139-147 dB re 1µPa	2.21 [4.73]	2	2	9	4

431

432 **Tab 1. Summary of each controlled exposure experiment including controls (no sound**

433 **emitted) and ‘simMFAS’ (playback of simulated mid-frequency active sonar). CEE-ID**

434 contains the year and the CEE number. Species abbreviations are Db for *D. bairdii* and Dd for *D.*

435 *delphis* and estimated group size is taken from shore observations. Received sound levels (RLs)

436 are reported from Durbin et al. 2022. Median and interquartile range (IQR, 25-75th percentiles)

437 are given for the raw whistle counts per second of the entire 30-minute CEE. Changepoints were

438 calculated on both the mean whistle count and the variance in whistle count, separately for the

439 pre-exposure and exposure periods. Note that “exposures” in controls were quiet periods for  
440 comparison to sound exposure in MFAS trials.

441

## 442 **2. Baseline Changepoint Analysis**

443 The baseline changepoint analysis revealed that mean detected whistle counts over the 30-  
444 minute ~~control~~ sampling periods changed once every minute, and variance in detected whistle  
445 count changed once every 3.5 minutes in control conditions.

446

### 447 **B. Assessing the impact of disturbance on whistle behavior**

448 A total of 10 MFAS CEEs were conducted - eight of which included *Delphinus bairdii*, and two  
449 of which included *Delphinus delphis*. The calculated average received level across all  
450 experiments was 151 dB re 1 $\mu$ Pa (range 122-159 dB re 1 $\mu$ Pa, Table 1). The average group size  
451 for MFAS CEEs was ~173 individuals (range 10-300, Table 1).

452


#### 453 **1. Number of changepoints in pre-exposure vs exposure**

454 Changepoint analysis was run for all control (9) and MFAS (10) experiments. The number of  
455 changepoints detected in both the mean and variance of whistle count during the pre-exposure  
456 and exposure periods are presented in Table 1 for both controls and MFAS experiments. There  
457 was no significant increase or decrease in the mean and variance of detected whistle counts  
458 following MFAS exposure when compared to the natural variance present during control  
459 conditions. The difference ( $\Delta$ ) in change points in variance in whistle count between the pre-  
460 exposure and exposure period did not differ significantly between controls and MFAS CEEs  
461 (controls:  $M = 3.1$ ,  $SD = 1.4$ ; MFAS:  $M = 2.5$ ,  $SD = 2.3$ ;  $t(17) = 0.69$ ,  $p = 0.5$ ). The same was

462 true when comparing the mean whistle count between both experimental types (controls: M =  
 463 11.3, SD = 14.2;MFAS, M = 8.7, SD = 8.5;  $t(17) = 0.5$ ,  $p = 0.63$ ).


464

465 **2. Impact of MFAS exposure on whistle count: 10-minute time scale**

466 Whistle counts did not change between the pre-exposure  exposure period during MFAS  
 467 experiments. At the 10-minute time scale, the preferred model was the simplest model with the  
 468 absolute value of the change in median whistle count as a function of only CEE type (either  
 469 MFAS or control). There was no significant effect of CEE type on the change in median whistle  
 470 count (negative binomial GLMM,  $n = 18$ ,  $p = 0.8$ ). The full model ( $\Delta AIC 5.4$ ) indicated that no  
 471 proposed predictor variables (CEE type, CEE number, species, buoy distance, or group size) had  
 472 a significant effect on the change in median whistle count between the pre-exposure and  
 473 exposure periods for both MFAS and control CEEs (negative binomial GLMM,  $n = 19$ :  $P > 0.05$   
 474 for all variables, table 2).

475

476 **Table 2. Overview of GLMMs used at three time-scales - 10 minutes, 20 seconds, and 5**  
 477 **seconds.**

conditional model 	zero-inflation model	distribution	$\Delta AIC$	degrees of freedom	dispersion ( $\sigma^2$ )
10-minute scale					
$abs(wrDiff) \sim ceeType$	n/a	nbinom2	0	3	5.44
$abs(wrDiff) \sim ceeType + (1   ceeNum)$	n/a	nbinom2	2	4	1.04
$abs(wrDiff) \sim ceeType + (1   ceeNum) + species + buoyDist + groupSize$	n/a	nbinom2	5.4	7	1.48
20-second scale					
$diff \sim ceeType + expMed$	~expMed	gaussian	0	6	7.1
$diff \sim ceeType$	~expMed	gaussian	5.1	5	7.23
$diff \sim ceeType + species + groupSize + buoyDist + expMed$	~expMed	gaussian	5.3	9	7.09
$diff \sim ceeType + (1   ceeNum)$	~expMed	gaussian	7.1	6	7.23
$diff \sim ceeType$	none	gaussian	28.3	3	6.71

5-second scale						
diff ~ ceeType + groupSize + expMed + ar1(times + 0   cee)	~expMed	gaussian	0	9	10.1	
diff ~ ceeType + (1   ceeNum) + species + groupSize + buoyDist + expMed + ar1(times + 0   cee)	~expMed	gaussian	2.6	12	10.1	
diff ~ ceeType + ar1(times + 0   cee)	~expMed	gaussian	13.8	7	9.83	
diff ~ ceeType + (1   ceeNum) + ar1(times + 0   cee)	~expMed	gaussian	15.8	8	9.83	
diff ~ ceeType	none	gaussian	207.8	3	10.9	

478

479




480 **3. Characterize the impact of MFAS exposure on whistle count: 20-second time scale**

481 Whistle counts did not differ significantly over the 20-second time windows surrounding each  
 482 ping. The preferred model at the 20 second scale included only predictor variables for CEE type  
 483 and median whistle count (table 2). The results of this model showed that ceeType did not have a  
 484 significant effect on changes in whistle count in the 20-seconds after each ping (GLMM, n=18,  
 485 slope = 0.47, SE = 0.27, P > 0.05, table 49, but that the baseline median whistle count for that  
 486 experimental period was a significant predictor for the change in whistle count following a ping  
 487 or ghost ping (GLMM, n = 19, slope = 0.76, SE = 0.028, p = 0.0075).


488

489 **4. Characterize the impact of MFAS exposure on whistle count: 5-second time scale**

490 The preferred model at the 5 second scale included the temporal autocorrelation structure and  
 491 three explanatory variables, CEE type, group size, and median whistle count, all of which had a  
 492 significant effect on the change in whistle count in the 5 seconds immediately following a ping  
 493 compared to the 5 seconds immediately before a ping. When accounting for all other variables,  
 494 the magnitude of the change in whistle count in the 5 seconds following a true MFAS ping was  
 495 1.4x greater than any change in whistle count following ghost pings in control CEEs (GLMM,  
 496 n=18, slope = 1.43, SE = 0.47, p = 0.002, table 50. The results of the CEE by CEE analysis (S2


497 Appendix) showed that in four of the  experiments in which sound was broadcasted,  
498 dolphins exhibited a significant elevation in whistle production in the 5-seconds after each ping  
499 throughout the entire 10-minute exposure period, while in three of the  MFAS exposures,  
500 animals showed an initial response to the first ping, with vocal responses slowly abating over the  
501 course of the experiment (plots of all raw whistles are provided in S3 Appendix). The effect was  
502 particularly pronounced (outside the 75<sup>th</sup> percentile; Appendix S2 Fig. S2.1) in the first ping of 6  
503 of the MFAS CEEs; whistle activity increased in the 5 seconds following the first MFAS ping up  
504 to 15 times the whistle count in the 5 seconds before the first MFAS (mean of all MFAS CEEs  
505 3.9, SD 5.2), compared to increases of only up to 1.4 times (mean 0.46, SD 0.88) at the first  
506 ghost ping of controls (Fig. 5, Appendix S2.1). Additionally, group size and median whistle  
507 count for the exposure period were found to be significant predictor variables. Larger groups  
508 showed larger changes in whistle count following pings and ghost pings (GLMM, n=18, slope =  
509 0.007, SE = 0.003, p = 0.008, table 2 ) and when the median background whistle count was  
510 higher, so too were the changes following pings or ghost pings (GLMM, n=18, slope = 0.25, SE  
511 = 0.048, p = < 0.005, table 2).

512

513 **Figure 5. (A) Spectrogram example of 5 seconds before and 5 seconds after the first ping**  
514 **for MFAS CEE 2021\_08, illustrating  the large increase in whistle count immediately**  
515 **following the cessation of the ping.** Focal group was comprised of approximately 30 long-  
516 beaked common dolphins. The MFAS signal can be seen between 3 and 4 kHz. (B) Boxplot of  
517 the change in whistle count from the 5 seconds before to the 5 seconds following each of the 24  
518 pings for CEE 2021\_08. Boxplot shows median, 25th, and 75th percentiles, with raw whistle  
519 count changes as open gray circles. The change following the first ping is shown as a red star.

520

521 **Discussion**

522 ~~We present a~~ hierarchical approach to quantifying the vocal response of large groups of common  
523 dolphins to Navy ~~Sonar and find~~ that the most pronounced acoustic response occurs within the 5-  
524 seconds following each ping during MFAS exposure. As in previous observational studies  
525 evaluating delphinid acoustic behavior (14), determining a singular behavioral response to  
526 MFAS presents several challenges. Multiple factors – including rapid changes in behavioral state  
527 over the course of the experiment and variation in group size and composition – make it difficult  
528 to ascribe changes in vocal behavior due to disturbance versus natural variability. To better  
529 understand typical acoustic variation among common dolphins, we assessed vocal behavior  
530 during control conditions and found that they exhibited natural vocal state changes (identified by  
531 the changepoint analysis) in whistle production approximately once every minute. This rapid  
532 acoustic state switching informed our analytical approach, which utilized a range of temporal  
533 windows to test for changes in whistle count (1  minutes - 5 seconds) before and after MFAS  
534 exposure. Across the ~~larger~~ time windows selected, we did not detect a change in whistle  
535 behavior that was attributable to MFAS. However, CEE type did have a significant effect on the  
536 change in whistle count in the 5-seconds immediately following a ping compared to the 5-  
537 seconds immediately before a ping.

538

539 The initial selection of the time-period over which behavior was assessed in response to Navy  
540 sonar (10-minutes) for this project was influenced by multiple factors, including the flight  
541 endurance of the drone used for calculating animal distance to our recording buoys (34), our  
542 ability to consistently track large groups of fast-moving dolphins, and other previous BRS




543 methods using MFAS (e.g., 43, 21). Many of the aforementioned constraints are imposed by the  
544 logistics of field work. An informed approach to identify behavioral responses to anthropogenic  
545 disturbances also requires some prior knowledge of the timing of behavioral state switching in  
546 the study species. For example, previous work with blue whales (*Balaenoptera musculus*)  
547 evaluated several behavioral metrics (e.g., maximum depth, dive time, ascent/descent rate) in  
548 responses to simulated MFAS exposure over a 30-minute time window (43). This exposure  
549 duration (which included a 30-minute pre-exposure period) was adequate to capture a behavioral  
550 change given the typical duration of their dive cycles (5-8 minutes, 44). In contrast, beaked  
551 whales are known to exhibit extremely long, deep foraging dives that often last over an hour  
552 followed by long periods of recovery (e.g., 45,46). Consequently, studies focused on direct  
553 measurements of behavioral response by Cuvier's beaked whales to MFAS extended their pre-  
554 exposure baseline period up to 9.3 hours to perform sound exposure during foraging dives and  
555 evaluated their response to sonar for up to 1.7 hours after the exposure period (47).

556

557 While the 10-minute exposure period seemed appropriate given the fast-paced lifestyle of  
558 common dolphins, neither the aggregate model nor the individual assessment of each CEE  
559 detected a change in acoustic behavior that could be ascribed to sonar exposure at this time scale.  
560 Even when we explored variation in whistle production during the 20-seconds surrounding each  
561 ping, the experimental period was not a significant predictor of changes in whistle count in either  
562 the MFAS or control experiments, as vocal state switching often occurs within a 40 second time  
563 window under natural conditions. While the impacts of sonar may be evident over the 10-minute  
564 exposure period for other behavioral metrics (i.e., changes in behavioral state, group

565 composition, diving behavior), our analysis reveals that changes in acoustic behavior are limited  
566 to an extremely narrow time-window in these species.

567

568 It was only at the 5-second time scale surrounding each ping that we observed dolphins  
569 exhibiting an acute acoustic response, which often included a rapid increase in whistle  
570 production relative to the 5 seconds immediately prior to sound exposure. On average, dolphins  
571 increased their whistle count 4 times the average count in the 5 seconds preceding the first ping  
572 of the exposure. In one MFAS experiment, dolphins increased their whistle production 15 times  
573 compared to the whistle count in the 5 seconds immediately before the first ping (Appendix  
574 S2.1). This elevated vocal response following the first ping of the exposure was seen in six of the  
575  MFAS exposures.

576

577 Elevations in whistle counts did not occur during the sonar signal ~~itself~~ – which lasted for 1.6  
578 seconds. Rather, the elevated vocal production occurred once the signal had been transmitted,  
579 often abating within ~ 10 seconds. The lack of whistle production occurring during the sonar  
580 transmission may be a tactic for reducing acoustic interference and masking which has been  
581 shown to impact the detection, discrimination, and localization of relevant signals (48). If the  
582 interfering signal is predictable (as is the case in our experiment), then animals should be able to  
583 adjust the timing of sound production to limit communication to periods in which noise is  
584 reduced (*e.g.*, 48-50). The ability of dolphins to learn the timing of intermittent noise has  
585 previously been demonstrated by Finneran et al., 2023 (51) who showed that individuals can  
586 modify their hearing sensitivity prior to the onset of predictably timed impulses, presumably to

587 mitigate ~~negative~~ auditory impacts. Surprisingly, little is known about their capacity to modify  
588 the timing of vocal production in response to interfering signals under ~~natural~~ conditions.

589

590 The sudden increase in vocal behavior following the first ping could be an example of the  
591 amplification of behavior of group members through recruitment or reinforcement (*i.e.*, positive  
592 feedback, 52). In this scenario, one dolphin may whistle in response to a surprising, salient  
593 stimuli and others follow suit. As this recruitment response continues, the number of dolphins  
594 producing whistles increases further and information is spread rapidly throughout the group (53).  
595 Alternatively, it could be an indicator that multiple animals are exchanging whistles to contact  
596 their closest social partners within the group in the presence of an unknown stimulus. This could  
597 be expected given the role of whistles in group cohesion and coordination (2,54) .

598

599 Whether animals continued to show an elevated acoustic response immediately following each  
600 ping varied between MFAS exposures. In four of the ~~nine~~ experiments in which sound was  
601 broadcasted, dolphins exhibited a significant elevation in whistle production in the 5-seconds  
602 after each ping throughout the entire 10-minute exposure period (S2 Appendix). This suggests  
603 that in these cases, dolphins did not habituate to successive pings (*i.e.*, show progressive decrease  
604 in the amplitude of a vocal behavioral response after repeated exposure). However, in three of  
605 the ~~nine~~ MFAS exposures, animals showed no initial response to the first ping, with vocal  
606 responses slowly abating over the course of the experiment (S2 Appendix). The population of  
607 dolphins tested in this study likely live a large portion of their lives in areas that are regularly  
608 ensonified by Navy sonar, and thus may exhibit an attenuated response when compared to  
609 individuals not regularly exposed to MFAS. Future work could compare the responsiveness of

610 animals in the Southern California Bight to nearby populations that occur in areas where Navy  
611 sonar is not regularly present (*e.g.*, Monterey Bay). Considering whether dolphins may be  
612 learning to modify their vocal behavior in response to repeated noise exposure may provide  
613 foundational evidence to support using vocal rates as a measure of sensitization or habituation to  
614 anthropogenic stimuli, as has been done in terrestrial species (55,56).

615  
616 Future management decisions mitigating the impact of sonar on oceanic delphinids should  
617 consider our reported results of clear responses during CEEs when analyses were conducted at  
618 the appropriate temporal resolution . With respect to the most recent methodology for assessing  
619 the relative response severity for free-ranging marine mammals to acoustic disturbance (57) –  
620 had the acoustic response of animals to MFAS been pooled across the 10-minute time window –  
621 common dolphins likely would have been assigned a behavioral response severity score of 0 (no  
622 response detected). However, when evaluating vocal behavior across a ~~shorter-time~~ 5-second  
623 time window, this species' response would be elevated to a category 3 severity which includes an  
624 increase in contact or alarm calls (57). Ultimately, continued work with this (and other closely  
625 related) species should also consider how observed behavioral responses vary with respect to  
626 other contextual parameters including behavioral state, group composition (*e.g.*, presence or  
627 absence of calves), seasonality, and environmental covariates. Concurrent efforts from this  
628 project aim to integrate passive acoustics with other remotely sensed datasets (*i.e.*, shore-based  
629 group tracking, aerial photogrammetry) to identify group-level behavioral changes and quantify  
630 exposure-response relationships (58). Paired with energetic modeling methods (*e.g.*, 31,59,60),  
631 these data can be used to link these observed short-term behavioral responses to long-term fitness  
632 outcomes in this species and inform the implementation of effective mitigation strategies.

633

634 Our study has some limitations that can be addressed in future work. Given the close  
635 phylogenetic relationship between short-beaked (~~*Delphinus delphis*~~) and long-beaked common  
636 dolphins (~~*Delphinus bairdii*~~), (61,62), the basic description of baseline vocal behavior (*i.e.*,  
637 whistle count data) was combined across species in our analysis. However, recent work by  
638 Oswald et al. discovered unique species-specific whistle frequency contours ~~in both *D. delphis*~~  
639 ~~and *D. bairdii*~~ and suggested that these distinctive acoustic signals could help facilitate  
640 recognition between these two *Delphinus* species (38). An increase in sample size for both long-  
641 beaked and short-beaked common dolphins would provide a more detailed understanding of their  
642 baseline vocal behavior and allow for the exploration of whether each species shows a  
643 differential or similar response in whistle type usage to MFAS. Additionally, photo-identification  
644 of individuals within these large, ephemeral groups is extremely difficult. Consequently, it is  
645 unknown whether individual dolphins around Catalina Island were exposed more than once to  
646 the experimental treatment. Future research could direct effort towards other oceanic delphinids  
647 with smaller group sizes where photo-identification is feasible and reliable (*e.g.*, bottlenose  
648 dolphins).



649

650 In any behavioral experiment, the response of the individual or group that is tested should be  
651 measured and interpreted over a time-window that is informed by their natural behavior. We  
652 suggest that future work with other oceanic delphinid species explore baseline vocal rates a-  
653 priori and use information on vocal state-switching to inform the analysis time-window over  
654 which behavioral responses are measured. Given these species' susceptibility to frequent MFAS  
655 exposure in Navy operational areas, it is of particular interest to evaluate how repeated exposure

656 influences responses. Future work should continue to explore key factors that are likely to  
657 influence the probability of response among these large groups, including their behavioral state  
658 and their proximity to the sound source and received level at the onset of exposure.

659

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672

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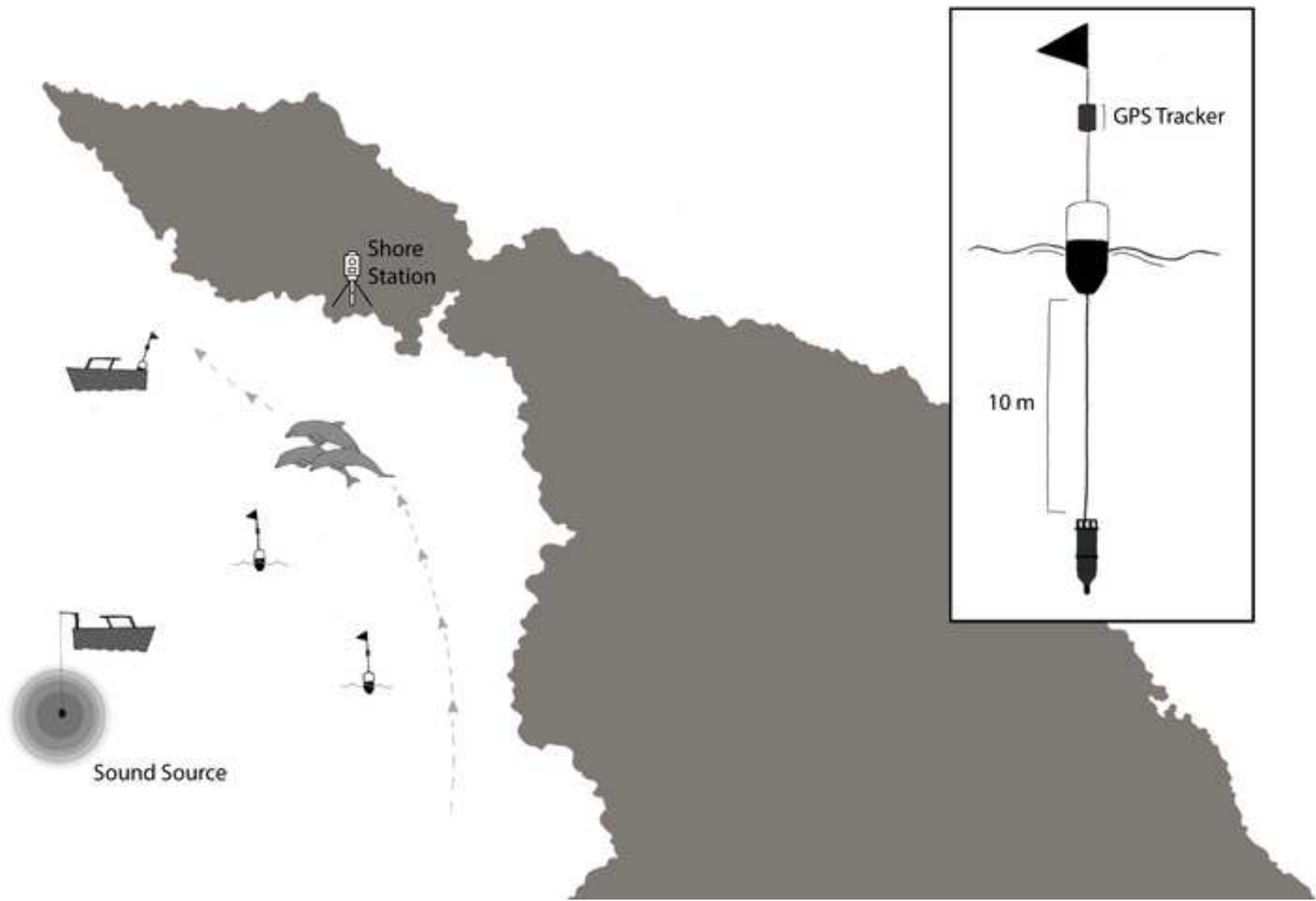
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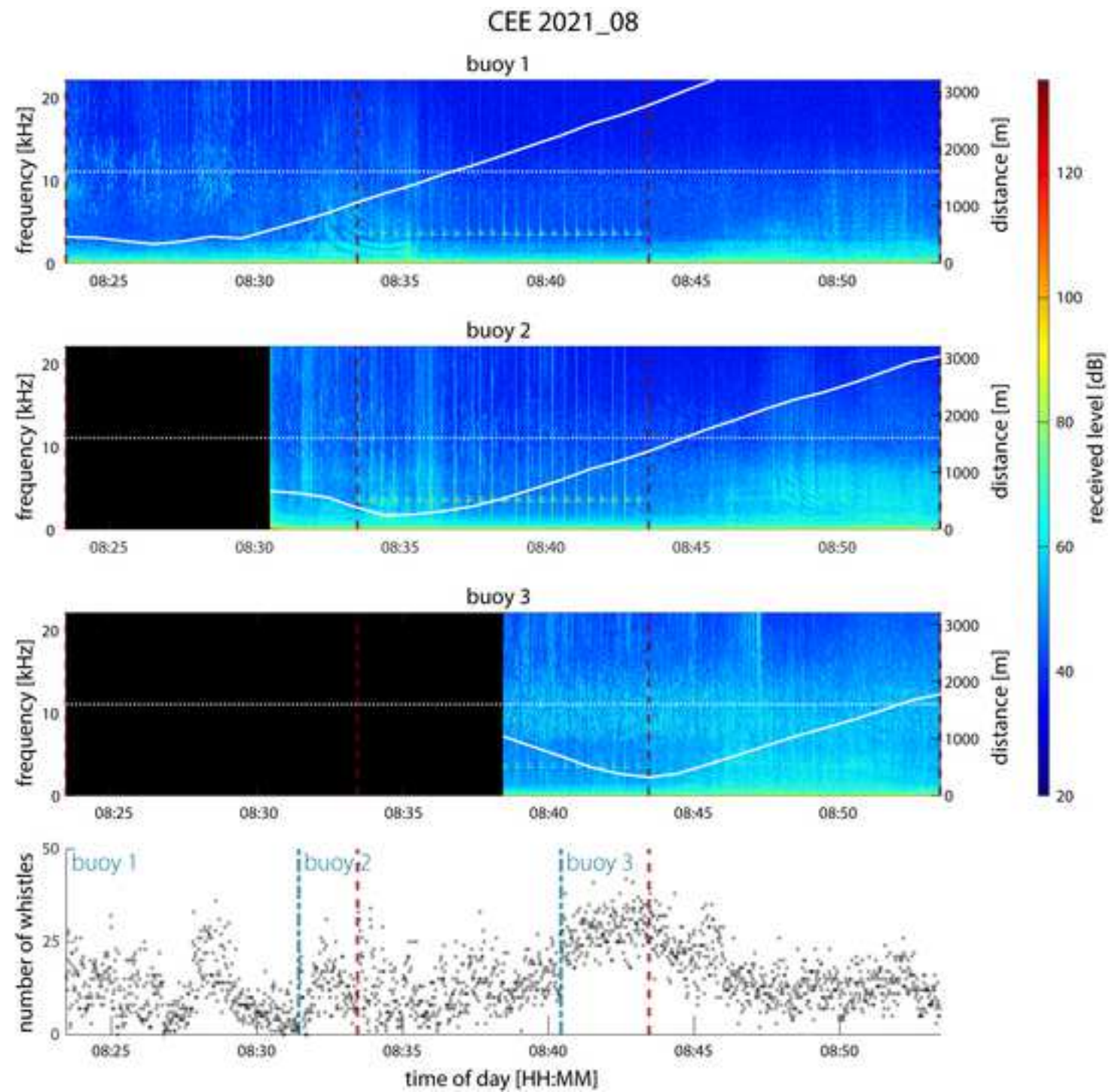
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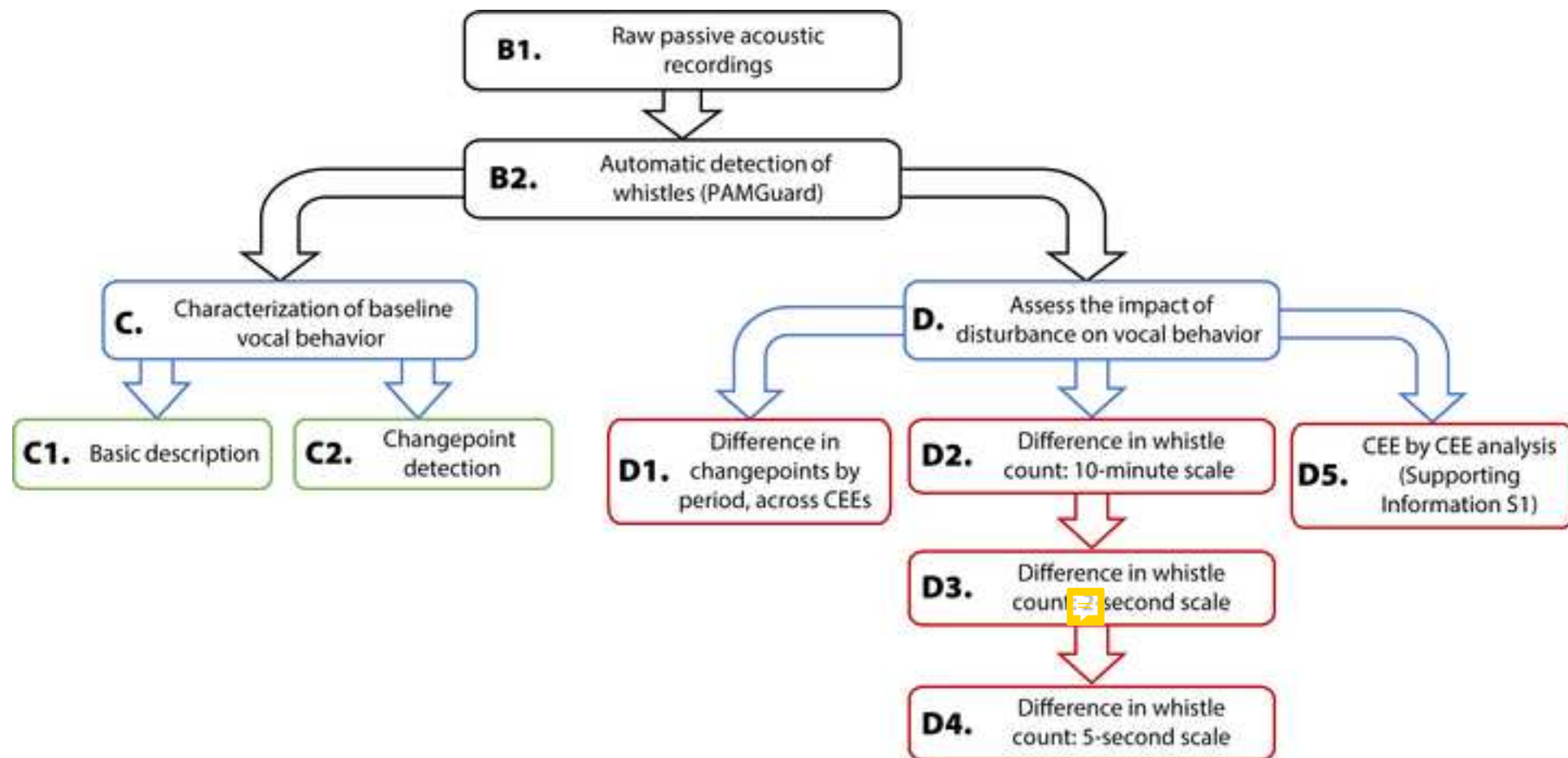
#### 849 **Author contributions**

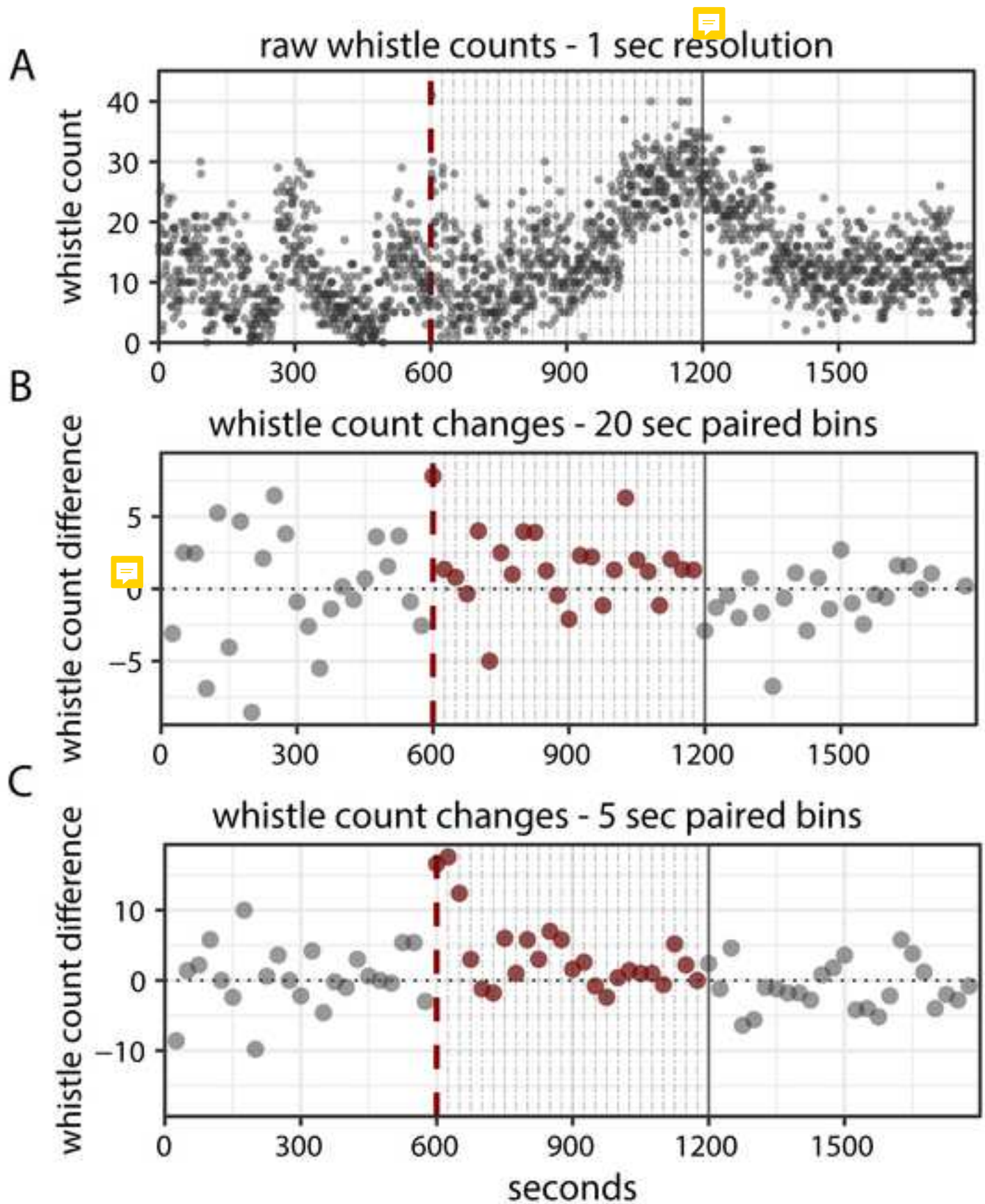
850 BLS acquired the funding for this project, and BLS, CC, and SF conceived, designed, and  
851 performed the experiment. CC and SF analyzed the data, while BLS, VMJ, and JNO provided  
852 analytical guidance. CC, SF, VMJ, JNO, and BLS wrote the manuscript.



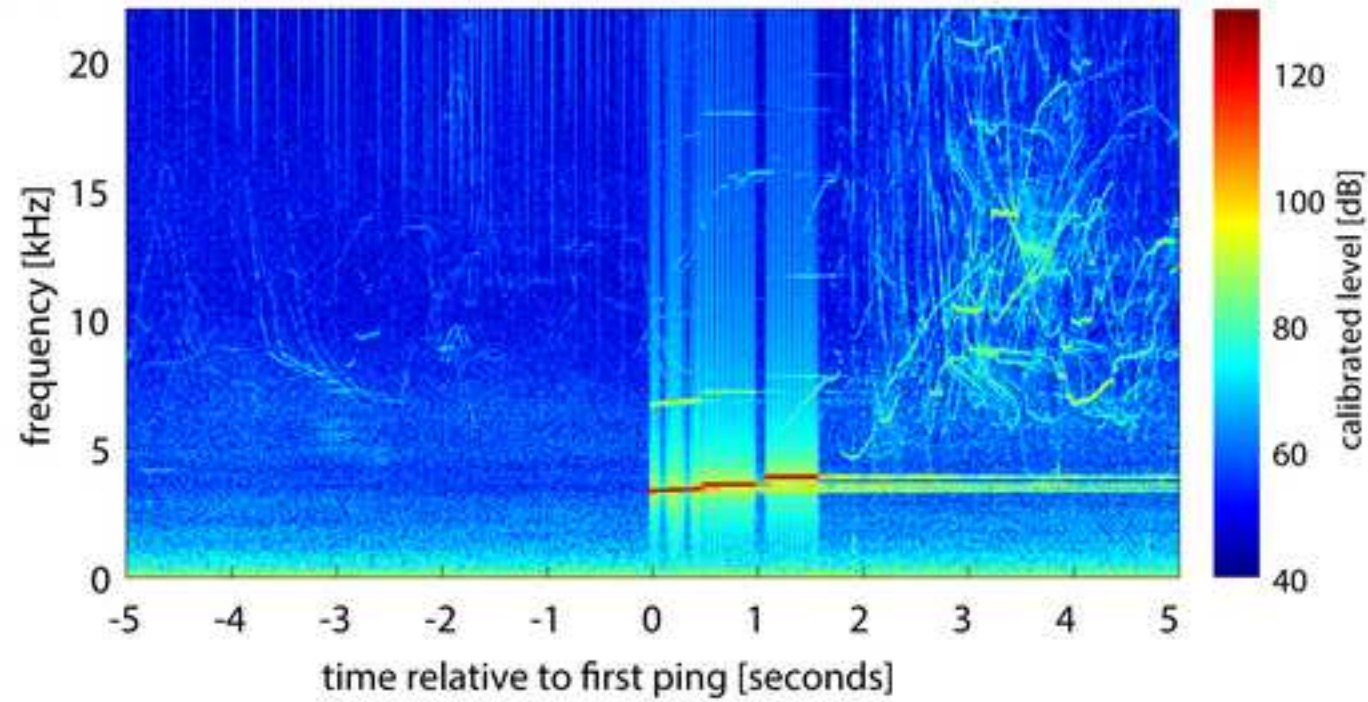




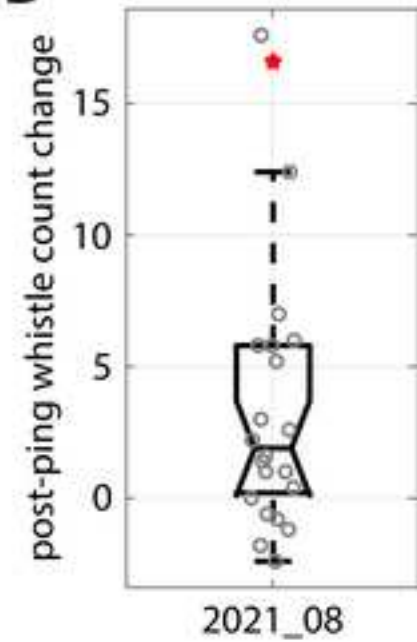


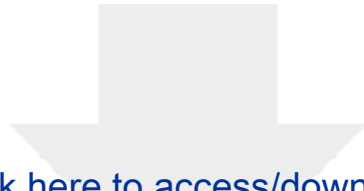


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