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The application of historic sample-specific variables in evaluating the biodiversity patterns of the South African azooxanthellate scleractinians (Cnidaria: Anthozoa). --Manuscript Draft--

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Abstract:	A200xantheliate scleractinian corars, a group of species that lack a symbolic relationship with dinoflagellates, are influenced by environmental variables at various scales. As the global commitment to sustainably manage ocean ecosystems and resources rises, there is a growing need to describe biodiversity trends in previously unsampled areas. Benthic invertebrate research in South Africa is a developing field, and many taxa in deep water environments remain inadequately characterized. Recently, the South African azooxanthellate scleractinian fauna was taxonomically reviewed, but their distributional correlation with physical parameters have not been studied. Here we aim to understand the biodiversity gradients of the South African azooxanthellate coral fauna by analysing the environmental correlates of museum samples. The associated coordinate data were georeferenced and depth obtained from a national bathymetric dataset, prior to undertaking a multivariate analysis. Overall, our results confirmed two longitudinal groups (eastern margin [group A] vs southern and western margin [group B]) and 11 depths represented within two bathymetric zones (shallow [50-200 m] and deep [300-1000 m]). Both the longitudinal groups and depth zones partially explained coral distribution patterns, with depth highly correlated with species variation. Caryophylliids, flabellids, and dendrophylliids contributed the most towards distinguishing longitudinal and depth gradients. Data limitations within our data set resulted to unexplained variance, however, despite these limitations, the study demonstrates that historical museum samples provide a valuable data source that can fill research sampling gaps and improve our understanding of biodiversity patterns of	
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16	Abstract
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18	dinoflagellates, are influenced by environmental variables at various scales. As the global
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20	to describe biodiversity trends in previously unsampled areas. Benthic invertebrate research in
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23	reviewed, but their distributional correlation with physical parameters have not been studied. Here
24	we aim to understand the biodiversity gradients of the South African azooxanthellate coral fauna

25 by analysing the environmental correlates of museum samples. The associated coordinate data 26 were georeferenced and depth obtained from a national bathymetric dataset, prior to undertaking 27 a multivariate analysis. Overall, our results confirmed two longitudinal groups (eastern margin 28 [group A] vs southern and western margin [group B]) and 11 depths represented within two 29 bathymetric zones (shallow [50-200 m] and deep [300-1000 m]). Both the longitudinal groups and 30 depth zones partially explained coral distribution patterns, with depth highly correlated with 31 species variation. Caryophylliids, flabellids, and dendrophylliids contributed the most towards 32 distinguishing longitudinal and depth gradients. Data limitations within our data set resulted to 33 unexplained variance, however, despite these limitations, the study demonstrates that historical 34 museum samples provide a valuable data source that can fill research sampling gaps and improve 35 our understanding of biodiversity patterns of the coral fauna in under sampled marine ecosystems.

36

37 **Keywords:** stony Cold-water corals, longitude, depth, gradients, distribution.

38

1. Introduction

40 The distribution of azooxanthellate corals, a group of scleractinian species that lack a symbiotic 41 relationship with photosynthetic dinoflagellates, is influenced by environmental variables at 42 various scales (Guinotte et al., 2006; Hovland, 2008; Roberts et al., 2009; Davies and Guinotte, 43 2011; Angeletti et al., 2020). Physical and chemical oceanographic factors, as well as 44 geomorphologic settings affect food supply and, consequently, benthopelagic coupling (Roberts 45 et al., 2009). Overall, depth might be used as a variable linked to several oceanographic factors 46 that influence species distributions. For example, coral species have preferred thermal ranges 47 (Davies and Guinotte, 2011), and a global azooxanthellate coral richness trend has been 48 documented between 200 and 1000 m deep. This depth range often coincides with shelf and slope 49 features, which may provide suitable substrate for larval settlement and habitats for 50 azooxanthellate coral species to colonise (Cairns, 2007; Roberts et al., 2009). Furthermore, long-51 term environmental stability appears to also be important for the occurrence/distribution of deep 52 water stony coral species. In addition to the temporal and spatial stability of an environment, it is 53 well established that life history patterns, including reproduction strategies and relationship to 54 substrate, are of utmost importance for a species' distribution (Oakham, 2009). For instance, 55 attached deep water scleractinians require consolidated substrates to survive, whilst unattached 56 forms are found on or in unconsolidated substrates (Roberts et al., 2009; Hovland, 2008).

57

Given the difficulty of sampling in deep-water marine systems, the mapping and classification of 58 59 biodiversity into spatial units (which then act as surrogates for unmapped biodiversity) is a 60 common approach in spatial planning (Waters, 2008; Costello, 2009; Reygondeau and Dunn, 61 2018; Reygondeau 2019; Richter et al., 2022). Considering the growing concern regarding 62 declining ocean health, voluntary commitments to reach a national 30 % area protection by 2030, 63 and the United Nations call for better ocean governance (United Nations, 2018; 2019; 2020), such 64 spatial classifications are powerful tools to guide conservation and management strategies to support the achievement of the United Nations 14th Sustainable Development Goals (SDGs). 65 Although ocean basins have been mapped at broad scales (Sayre et al., 2017) and several global 66 and regional bioregionalisations exist (Ekman, 1953; Briggs, 1974; Spalding et al., 2007; Grant et 67 al., 2006; Cedras et al., 2020; McQuaid et al., 2023), few studies have used species data to describe 68 69 biological patterns in areas deeper than 200 m (Zezina, 1997; O'Hara et al., 2011; Watling et al., 70 2013; Cedras et al., 2020; Summers and Watling, 2021; Watling and Lapointe, 2022), particularly within benthic ecosystems (O'Hara et al., 2011; Summers and Watling, 2021; Watling and Lapointe, 2022). Developing nations, such as South Africa, particularly lack specialised resources to survey deeper waters, further constraining *in situ* research (Bell et al., 2022). Consequently, samples housed in scientific collections may be a valuable source of biological data to evaluate distribution patterns (Spalding et al., 2007; Thandar et al., 1989; Woolley et al., 2020) but should be interpreted with caution.

77

78 Despite the fact that the first marine collections along South Africa's shores dates back to the 79 1700s (Day, 1977; Griffiths et al., 2010), ocean resource management is still constrained by the 80 poor state of knowledge of key invertebrate species (Sink et al., 2019). Endeavouring to address 81 such species data gaps, local research advancements have recently been initiated by examining 82 natural history collections (Biccard 2012; Laird 2013; Filander, 2014; Olbers 2016; Boonzaaier, 83 2017; Landschoff 2011). Some of these studies have been integrated into the ecosystem map 84 developed by Sink et al. (2019; 2023) for the National Biodiversity Assessment (NBA). The NBA 85 used pelagic and benthic data, including biological available information (i.e., macrofauna, 86 epifauna, and fish) to produce an expert-driven ecosystem type map for national reporting 87 frameworks. It comprises four hierarchical levels that represent six ecoregions (two deep ocean 88 and four confined to the continental shelf), five depth zones (shore, shelf, slope, plateau, and 89 abyss), and different substrate types. Absent, however, from this national spatial classification map 90 is a holistic consideration of the South African azooxanthellate scleractinian fauna, given that this 91 taxonomic group was only recently reviewed (Filander et al., 2021), and its distribution patterns 92 had not yet been investigated. The NBA does however report on some distribution records of potential Vulnerable Marine Ecosystem indicator taxa, which includes records of two reef-building
azooxanthellate coral taxa (i.e., Dendrophylliidae Gray, 1847 and Caryophylliidae Dana, 1846).

95

96 Earlier international studies (Cairns, 2007) have grouped the available literature on 97 azooxanthellate Scleractinia into broad geographic regions. Although not a biodiversity analysis, 98 Cairns (2007) produced an output that served as a starting point for emerging taxonomists in the 99 field. Cairns and Keller (1993) did, nonetheless, summarise depth affiliations within the southwest 100 Indian Ocean, in which South African taxa reported off the eastern and southern margins were 101 represented. Apart from these two publications (Cairns and Keller, 1993; Cairns, 2007), the South 102 African azooxanthellate Scleractinia distribution pattern have not been investigated in light of its 103 relationship to physical variables. Therefore, this paper aims to examines the diversity measures of the South African azooxanthellate coral fauna, with respect to sample-specific environmental 104 105 gradients.

106

107 **2. Material and Methods**

108 Data considered for this study were based primarily on a subset of species distribution records for 109 the South African azooxanthellate scleractinian fauna recently reported by Filander et al. (2021). 110 The subset was compiled by including those occurrence data with coordinate information and a 111 sample number, resulting in 761 occurrence records (**Figure 1** and **Appendix A**: Occurrence data). 112 These coral occurrence data were predominately collected during six historical dredge surveys 113 undertaken between 1898 and 1990 (i.e., RV Anton Bruun, Benguela IV, RV Meiring Naude, RV 114 Pieter Faure, Sardinops, and University of Cape Town Ecological Surveys). The recent surveys undertaken in the 21st century are represented by two trawl (NANSEN and Department of 115

116	Environment, Forestry and Fisheries/South African Environmental Observation Network demersal			
117	surveys) and three dredge research surveys (ACEP: Deep-Secrets and IMIDA surveys,			
118	Department of Environment, Forestry and Fisheries) (see Filander et al., 2021).			
119				
120	Figure 1. Study domain and spatial coverage of the coral records forming the basis of the analysis.			
121				
122	It is worth noting that the historical data sets had a varying degree of reliability in terms of			
123	associated data and, therefore, required data validation. Consequently, all the occurrence records			
124	were first geo-referenced using ArcGIS 10.1. This step involved overlaying the coral point data on			
125	the National Biodiversity Assessment (Sink et al., 2019; 2023) marine ecosystem types map.			
120=	Records that were recovered on the coastline were moved to the closest polygon boundary of the			
127	ecosystem types with the near command in ArcGIS 10.1. This process was particularly beneficial			
128	for the Pieter Faure stations, which had positions in degrees magnetic North (not true North);			
129	whereby land bearings were used as a reference. In the next step, the spatial join tool was used to			
130	assign depth in relation to the most recent national bathymetric dataset (de Wet and Compton,			
131	2021) to each of the coral records, irrespective of whether depths were present in the coral archive			
132	data set or not. Depth contours started at 50 m and continued every 100 m isobath intervals to a			
133	maximum of 1000 m. The resulting data set consisted of 95 of the total 108 azooxanthellate			
134	scleractinian species known from South Africa (Filander et al., 2021).			

136 **2.1. Assumptions and sampling biases**

Over 80% of the resulting data is of historical origin, and therefore poses some limitations. One of these limitations is sampling coverage bias, given that past national marine surveys focused mainly on nearshore areas due to their accessibility, whilst sampling in areas beyond the continental shell

relied on international surveys (the Pieter Faure expeditions being an exception) (Griffiths et al., 140 141 2010). Nevertheless, these historical surveys represent decades of sampling effort but were not 142 systematic and provide presence-only data. Secondly, depth and co-ordinate information are the 143 only two variables commonly associated with such datasets, but may be unreliable in some 144 instances (i.e., the *Pieter Faure* collection). Thus, the use of the occurrence data in the multi-145 variate analysis required three assumptions that may not necessarily be a true reflection of the data 146 attributes. These include presenting the data as presence-absence, inferring the occurrence of the 147 historical records to the modern day, and applying longitude and depth as a proxy for ocean basin 148 and water mass properties; respectively. The above-mentioned two-part data geo-referencing 149 methodology was therefore also undertaken to standardize the associated collection specific 150 parameters for the application of longitude and depth data as abiotic variables.

151

152 Furthermore, the data preparation methodology does not follow the interpolation of the presenceabsence matrix (i.e., if species occur between two extreme points, then occurrence is assumed in 153 154 between) as conducted in preceding marine benthic invertebrate studies based on museum 155 specimens (Filander, 2014; Olbers, 2016; Boonzaaier, 2017). This approach would have yielded 156 unrealistic conclusions in the absence of fine-scale substrate data sources- as substrate type is one 157 of the primary drivers of coral settlement. Whereas the 2018 NBA (Sink et al., 2019) sub-divided 158 substrate into ecosystem types, the multibeam data represented less than 1% of the South African 159 seabed. Cawthra et al. (2021) review on existing South African core samples collected from areas 160 deeper that 130 m highlights the importance of high resolution hydroacoustic surveys to better 161 contextualize published core localities. It is for this reason that the substrate level was not 162 considered to support interpolation techniques.

Lastly, the average taxonomic distinctiveness (ATD) diversity measure was based on the 163 164 established phylogenetic clades (Kitahara et al., 2010; Stolarski et al., 2011). However, owing to 165 limited resolution regarding species relationships below family level, phylogenetic scores were 166 not assigned to taxonomic levels lower than family. Additionally, the existing phylogenetic 167 reconstructions still lack sufficient representation of azooxanthellate coral species occurring in 168 South Africa. For instance, less than 20% of South African coral species have been sequenced and 169 included in existing molecular trees (Kitahara et al., 2010; Stolarski et al., 2011). It is important to 170 note that ATD is a diversity calculation method that considers the distance between each species 171 and its closest relative outside the group. This calculation is then divided by the number of species 172 within the group being evaluated. The resulting ATD value provides an estimation of the group's 173 evolutionary uniqueness, with higher values indicating greater distinctiveness. Consequently, an 174 alphabetically arranged method would yield inaccurate results.

175

176 **2.2. Data analysis**

177 A presence-absence matrix (Appendix A: presence-absence) of the coral occurrence data was 178 compiled and all analyses were undertaken using the PRIMER 7 software package (Clarke and 179 Warwick, 2001; Clarke et al., 2014), with the PERMANOVA+ add on (Anderson et al., 2008). 180 The matrix, consisting of 488 columns (stations/samples) and 95 rows (species), was converted to 181 a resemblance matrix. The associated higher taxonomic classifications of these resulting species 182 identifications were thereafter extracted from the World Register of Marine Species batch match 183 online function (WoRMS Editorial Board, 2021) (Appendix A: Taxonomic attributes). Owing to 184 the patchy nature of the data set, in which 30 species were represented by only one sample and 22 185 species by less than ten samples (Appendix A: Number of records per species), the Gamma+

=

dissimilarity matrix was selected (Clark and Warwick, 1999; Clark et al., 2006). This measure
used the cophenetic distances derived from the phylogeny established in Kitahara et al. (2010) and
Stolarski et al. (2011) (e.g. "Basal", "Complex", and "Robust") (Appendix A: Taxonomic
attributes). Such a procedure allowed for biotic distances among samples to be quantified even
when they had zero or very few species in common.

191

192 The sample-specific data also required data preparation, which followed the biological data 193 assessment. Longitude and depth are the two sample-specific variables considered to determine 194 the environmental settings of the South African maritime domain (Appendix A: Sample-specific 195 abiotic data). For instance if a sample was recorded at a 31° longitude, then it was collected in the 196 Indian Ocean and influenced by the Agulhas Current. Additionally, water mass temperature (T) 197 and salinity (S) properties can be confirmed by depth, in which upper layer waters were identified 198 between the surface and 500 m deep, intermediate between 500 and 1500 m, and abyss greater that 199 1500 m (Emery, 2003). Each abiotic parameter was firstly classified accordingly (see below), prior 200 to running an one-way (unordered) similarity analysis (ANOSIM) test to evaluate variation of 201 corals species within the longitudinal and depth groups (Appendix A: Sample-specific abiotic 202 data). The ANOSIM test requires groupings and measures the mean of ranked dissimilarities 203 between groups to the mean of ranked dissimilarities within groups. A standard approach was 204 undertaken to investigate change in species attributes along the longitudinal and depth gradient, 205 whereby a SIMPER analysis was run to evaluate contributing taxa.

206

To classify the longitudinal data as a factor to be tested on the biological data, an auto select k-R cluster mean analysis was run on the normalized longitudinal data. A draftsman's plot was

210 present (Appendix B: Figure 2). The depth classification starts at a 50 m isobath and progresses 211 at 100 m increments. These depth readings were further classified according to zone, i.e., shallow 212 (50-200 m) vs deep (300-1000 m). The depth values and subsequent depth zones were also added 213 as a factor for testing. Similarity percentage (SIMPER) tests were run independently to distinguish 214 coral species contributing to the longitudinal groups identified by the k-R mean analysis and depth 215 groups. Sampling effort (i.e., denoted by N), species richness (denoted by S), Shannon index 216 (denoted by H'log^e), and taxonomic distinctiveness (denoted by delta+) across the longitudinal 217 and depth groups was quantified. The former was investigated by assigning coral records to 50x50218 km grids created with the fishnet ArcGis function, whereby the grid size was guided by the

produced to identify the number of longitudinal groups present and validate the cluster groups

219 boundary breaks of the k-R mean cluster groups.

220

209

221 Subsequently, a RELATE routine was undertaken to evaluate if the combined longitude and depth 222 spatial gradients correspond with those inferred from the coral species patterns. Here we used the 223 Gamma + matrix in relation to the associated depth and longitude information, which was 224 normalised into a Euclidean distance resemblance matrix. The RELATE technique calculated a 225 Spearman's p rank correlation coefficient between all elements of the coral assemblage and 226 environmental variable resemblance matrices, followed by a permutation test. Following this, a 227 biota and/or environment matching (BEST) test was conducted to confirm which variable 228 contributed the most to sample statistic given by the RELATE results. A species accumulation 229 model was lastly produced to assess how well the current observed azooxanthellate stony coral 230 data represents South Africa's predicted coral diversity.

232 **3. Results**

233 **3.1. Longitudinal gradient**

234 The k-R (non-hierarchal) cluster analysis yielded two longitudinal groups (R = 0.94), whereby 235 group A encompasses samples from the eastern margin of South Africa and group B are samples 236 from the southern and western margins (Figure 3). The SIMPER results suggested that biological 237 samples represented by each of the two longitudinal clusters had an overall low average similarity 238 in species composition (Appendix C: Long group SIMPER spp results). Communities in group 239 A, with an average similarity of 3.4 %, were characterised by a total of 11 species from five 240 families (i.e., Dendrophylliidae, Caryophylliidae, Micrabaciidae Vaughan, 1905, Turbinoliidae 241 Milne-Edwards & Haime, 1848, and Flabellidae Bourne, 1905). Group B had a total of eight 242 species contributing to the group identity with a slightly higher average similarity (4.6%), whereby 243 only two families (i.e., Dendrophylliidae and Caryophylliidae) were represented.

244

Figure 3. The 50x50 km gridded cells with samples in relation to the longitudinal groups defined by the k-R cluster analysis. Group A represents samples collected off the eastern margin and group B are samples collected off the southern and western margins.

248

Furthermore, there was a distinction between families contributing the most to the cluster identities. Three dendrophylliids (*Balanophyllia capensis* Verrill, 1865), *Pourtalopsammia togata* (van der Horst, 1927), and *Heteropsammia cochlea* (Spengler, 1781) contributed the most to the similarity within group A; and the caryophylliids (*Trochocyathus* sp. 2 sensu Filander et al. (2021), *Caryophyllia stellula* Cairns, 1998, and *Desmophyllum dianthus* (Esper, 1794)) defined group B. When comparing the two groups, a 98% average dissimilarity was observed when investigating species contributing to a minimum cut off of 70%, whereby different species influenced group

256 distinction. Heteropsammia cochlea, Letepsammia franki (Owens, 1994), Flabellum pavoninum 257 (Lesson, 1831), and Labyrinthocyathus delicatus (Marenzeller, 1904) were represented only in 258 group A. Among these species, L. delicatus is restricted to the Indian Ocean, whilst L. franki, F. 259 pavoninum and H. cochlea are widely distributed in the Indo-Pacific. Four species contributed the 260 most to group B, of which two are restricted to South African waters (Ednapsammia 261 columnapravia Filander, 2021 and Dendrophyllia sp. 1 sensu Filander et al. (2021)), one 262 considered cosmopolitan (Desmophyllum pertusum (Linnaeus, 1758)), and the remaining reported 263 from the Atlantic and Indian ocean basins (*Enallopsammia pusilla* (Alcock, 1902)) (Appendix C: 264 Long Group SIMPER spp results).

265

Overall, the number of samples between the two groups varied, whereby group A (eastern margin) had over two times more samples than group B (southern and eastern margin) (**Table 1**). Contrary to this, the related area (i.e., number of grids) representing these samples was larger in group B than in group A (**Table 1**). Diversity followed the same pattern of higher measures in group A as compared with group B.

271

272

Table 1. Summary of sampling effort in relation to longitudinal gradient

k-R cluster group	Number of samples	Number of 50x50 km grids	Species richness	Shannon's Index	Delta +
A (eastern margin)	569	37	86	3.964	90.907
B (southern & western margin)	192	43	37	3.249	89.289

273

274 **3.2. Depth gradients**

275 A direct relationship between the number of samples (N), species richness (S), and depth was 276 observed (Figure 4). The highest number of samples and observed species richness occurred 277 between depths of 50 and 200 m, with the greatest species richness and sample count recorded at 278 a depth of 50 m. The same two measures (S and N) fluctuated at the deep bathymetric zone (i.e., 279 300-1000 m). Within this deep zone, the highest coral diversity measures (S and N) were recorded 280 at 1000 m and the lowest at 800 m. The overall pattern suggests that species richness is influenced 281 by depth and sampling effort, but other factors may also play a role in shaping patterns of species 282 diversity across different depths. Average taxonomic distinctiveness (denoted by delta +), which 283 takes into account species phylogeny, did not show a clear pattern in coral diversity with depth 284 and species diversity was relatively constant from 50 to 200 m. However, according to this 285 measure, coral diversity was slightly higher at 1000 m despite the usage of a smaller number of 286 samples from this depth (42 samples compared to 269 samples at 50 m). In other words, eight 287 taxonomic families were recorded at 1000 m, while only seven were recorded at 50 m. In contrast, 288 however, the conservative Shannon diversity index mirrored the pattern of species richness with 289 depth (Figure 4).

290

Figure 4. The relationship between number of coral samples (N) and species richness (S) in conjunction with the average taxonomic distinctiveness (delta+) and Shannon diversity (H'log^e) index measures across depth gradients. The x-axis shows samples represented in depth values in metres and the y-axis shows values that represent diversity measures in arbitrary units. The inset shows the Shannon diversity (H'log^e) index repeated on a Y axis of 1-4.

295

The SIMPER results of the coral species data according to family suggested that the caryophylliids, dendrophylliids, and flabellids were the main contributing taxa towards both the shallow (50-200 m) and deep (300-1000m) stations. Whilst all three families collectively contributed towards the 299 zone comparison (i.e., shallow *vs* deep) at a 70% cut, the Caryophylliidae representatives were 300 more abundant in the deep stations compared with the Dendrophylliidae and the Flabellidae in the 301 shallow stations (**Appendix C**: Depth zones SIMPER family results).

302

When comparing the two bathymetric zones at lower taxonomical rankings, three species were exclusively recorded in the shallower zone (*Rhizopsammia compacta* Sheppard & Sheppard, 1991, *Truncatoflabellum inconstans* (Marenzeller, 1904), and *Dendrophyllia cornigera* (Lamarck, 1816)), whereas the deeper zones were characterized by three of the four species restricted to the longitudinal group B: *E. columnapravia, Dendrophyllia* sp. 1, and *E. pusilla*.

308

309 3.2. The correlation of sample-specific variables (longitude and depth groups) to coral 310 distribution patterns

311 The one-way ANOSIM results, which evaluated the rank differences in the coral pattern that may 312 be explained by the longitudinal k-R clusters and depth groupings independently, showed that the 313 two longitudinal clusters (R=0.05, p = 0.001), and the eleven depth groupings (R = 0.072, p =314 0.001) of the biological assemblages differed significantly from one another (Appendix D: Figure 315 5). Depth also showed significant differences (R = 0.105, p = 0.001) when grouped as shallow (50-316 200 m) and deep zones (300-500 m) (Appendix D: Figure 6). Nonetheless, even though some 317 overlap was observed in the coral structure when testing longitude and depth; the null hypothesis 318 (of no differences in rank groups) can be rejected.

320 The RELATE results showed a marginal correlation (Rho-value = 0.087) but a significant 321 difference (p-value = 0.001) when comparing the coral patterns modelled by the Gamma+

resemblance matrix to that of the Euclidean distance matrix (i.e., environmental variables longitude and depth; **Figure 7**). It is important to note that the null hypothesis in the RELATE function is that there is no correlation. Thus, although the correlation is closer to zero (unexplained variance), the p-value confirms that longitude and depth are good predictors for the coral distribution patterns. The BEST results further confirmed that depth had an independent correlation value of 0.094, whilst both environmental parameters (longitude and depth) accounted for a correlation value of 0.097.

329

Figure 7. Simulated distribution/histogram of the test statistic Rho under the null hypothesis that there is no

331 correlation between the modelled coral patterns and that of the environmental variables Rho = 0.041.

332

The majority of the species accumulation curves, which show how the number of species detected (i.e., observed or sampled) increases with increasing sampling effort (i.e., the number of individuals or samples collected), did not reach a plateau (**Figure 8**). All seven estimated curves, along with the observed or sampled species, started with a steep slope and indicated a rapid increase in the number of species observed with increasing sampling effort. Only two (MM and UGE) of the seven estimator curves followed the species observed pattern (Sobs), which appears to be levelling off as the sampling effort increases (**Figure 8**).

340

341 **Figure 8**. Species richness accumulation curve showing the species observed (Sobs= blue upright triangle) in

342 relation to five estimators (Chao 1= red downward triangle, Chao 2= green square, Jacknife 1= pink diamond,

Jacknife 2= blue circle, Bootstrap= grey cross). Two pairs of curves overlap, whereby the UGE estimator curve

follows the same pattern as the Sobs and the Chao1 has the same pattern as Chao 2.

345

346 **4. DISCUSSION**

347 The multivariate analyses suggests that the sample-specific associated data (e.g., longitude and 348 depth) are significant predictor variables for azooxanthellate Scleractinia coral diversity. 349 Nonetheless, unexplained variance exists. Diversity measures were assessed, in which the number 350 of samples showed an inversely proportional relationship with species richness. Contrary to this 351 observation, taxonomic distinctiveness (a diversity measure independent of the number of 352 samples) revealed an opposing pattern to that of the univariate Shannon index measure. Thus, 353 taxonomic distinctiveness accounted for the uneven species distribution across the South African 354 continental maritime domain.

355

356 An increasing species turnover along the west to east gradient was observed in our analysis. Such 357 distributional patterns have long been reported for other South African marine invertebrates (e.g., 358 Lang, 2012; Filander, 2014; Boonzaaier, 2017), suggesting that different oceanographic conditions 359 are influencing the South African marine fauna. The accompanying current regimes may also 360 govern these contrasting species profiles across the region. Thus, although the two longitudinal 361 boundaries (Group A = eastern margin vs Group B= western margin) established by the k-R mean 362 cluster analysis do not conform to the previously proposed oceanographic boundaries (Longhurst, 363 2007; Spalding et al., 2012;), the ANOSIM suggested a species pattern that may be explained by 364 the two longitudinal groups. These margins correspond to varying oceanographic variables and 365 currents, whereby the eastern margin (group A) is situated within the oligotrophic waters of the 366 Indian oceanic basin and influenced by the western boundary Agulhas current. Interestingly, group B encompasses the southern and western margins located in both the Indian and Atlantic basins 367 368 respectively. At the southern margin, the Agulhas current retroflects, moving away from the shelf,

and introduces Indo-Pacific waters into the Atlantic Ocean, the latter being regulated by thenorthward flowing Benguela current (Shannon, 1985).

371

372 The SIMPER results detailed a clear taxonomic/ family and species distinctions within these two 373 longitudinal groups. Dendrophylliids contributed the most to Group A samples and caryophylliids 374 to Group B. Additionally, the exclusivity in species found between Group A (H. cochlea, L. franki, 375 F. pavinonum, and L. delicatus) and Group B (E. columnapravia, Dendrophyllia sp. 1, D. 376 *pertusum*, and *E. pusilla*) (see longitudinal gradient results) corroborates with the proposal that 377 species have a temperature threshold (Roberts et al., 2009; Cairns, 2007). The physiological 378 characteristics of azooxanthellate coral species are indeed influenced by the properties of ambient 379 water temperature (Gori et al., 2016; Castellan et al., 2019). For example, an *ex-situ* experiment 380 undertaken on the reef-building corals D. pertusum and Madropora oculata revealed that they 381 respond differently when exposed to three temperatures (12, 9.0, and 6.0 °C; Naumann et al., 382 2014). The respiration response rates varied; M. oculata declined whereas D. pertusum was not 383 affected by temperatures being lowered. Two other physiological responses (i.e., calcification and 384 dissolved organic carbon) were measured, and neither showed a consistent trend when comparing 385 the two species. Thus, species belonging to different families or even congeners are expected to 386 exhibit varying thermal tolerance.

387

The recovered species longitudinal pattern of low sampling effort in Group A (eastern margin) but higher number of records and diversity observed herein, was particularly surprising as the western margin (which contributes to Group B) has a higher historic sampling effort (Griffiths et al., 2010). The greater presence of coral species in the eastern Agulhas region (Group A) may be explained

392 by the heterogenous seabed substrate types provided by the increased abundance of mesophotic 393 reefs, submarine canyons, and mosaic ecosystem types (Sink et al., 2019). Whilst the incising 394 submarine canyons along the eastern continental margin (Green et al., 2007; Green, 2008; 2009; 395 Green, 2011) may also give rise to a heterogenous environment, localised canyon substrate type 396 studies need to be undertaken to confirm such hypothesis (Filander et al., 2022). Even though the 397 Benguela Current in the South Atlantic (influencing the western passive margin) is unique in its 398 interactions with the western boundary Agulhas current (Longhurst, 2007), this region has 399 substrate predominately unconsolidated, resulting in a more homogenous environment 400 (Dingle, 1979; Cawthra et al., 2021; Filander et al., 2022). Additionally, dissolved oxygen levels 401 have been proposed to affect scleractinian growth (Hanz et al., 2019) and the Southern Benguela 402 Upwelling region does include a low-oxygen area/cell off St Helena Bay (Lamont et al., 2015). 403 Though nearshore, such cells are reported to show spatial variability and may modify offshore 404 upwelling water masses. These oxygen parameters superimposed with unconsolidated bottom 405 types and a slow current (< 3 m/s) may be a constraint for coral presence. The presence of coral is 406 however influenced by multiple factors operating at different scales, and it is crucial to consider 407 species-specific regional adaptation abilities to environmental gradients (i.e., dissolved oxygen) -408 even for cosmopolitan species (Orejas et al., 2021). Nonetheless, the prominence of anthropogenic 409 activities that interact with the seabed in the Southern Benguela Upwelling area (Atkinson et al., 410 2011; Majiedt et al., 2019) cannot be overlooked and may also influence the low number of species 411 records in the area.

412

413 The southern margin, which contributes to Group B, is a unique area that exhibits minimal 414 interaction with other landmasses and, as such, high endemism has been noted (Griffiths et al., 415 2010). In this region, the Agulhas current injects Indo-Pacific waters into the Atlantic, down to 416 depths of 2000 m in the form of anticyclonic rings (Beal et al., 2011), before retroflecting 417 eastwards towards the Southern Indian Ocean Gyre and the Antarctic circumpolar current 418 (Spalding et al., 2012). Schouten et al. (2000) noted that the location of the retroflection loop is 419 variable, but still within the southern region. Nonetheless, the Agulhas transport is estimated to increase from 65 Sv (1Sv - 10⁶ m ³s⁻¹) at 32°S to 95 Sv at the southern tip of South Africa, as it 420 421 breaks away from the shelf (Gordon et al., 1992; Duncombe Rae, 1991). Thus, the unpredictable 422 behaviour and velocity of the Agulhas current make this area challenging for sampling and, 423 therefore, the low number of records here may be attributed to limited sampling effort.

424

425 The analysis of depth gradients allowed patterns of species richness in relation to the depth to be 426 better understood. These results complement the longitudinal gradients whereby the univariant 427 biodiversity measures peaked at 50 m, which corresponds to the accessible eastern margin of the 428 South African maritime domain. In addition to the shelf being shallower (~ 50-150 m) and more 429 accessible, the western boundary Agulhas current (characteristic of this area) has been linked to 430 the highly diverse biological properties in the Southwest Indian Ocean, where eddies can trap and 431 transport material over long distances (Halo et al., 2014). These complex oceanographic eddies 432 can upwell deep nutrient-rich waters through surface divergence mechanisms (Halo et al., 2014), 433 creating environments that favour the continuous inflow of potential food sources. Thus, these 434 observations may provide grounds for a hypothesis to explain why azooxanthellate corals have a 435 higher presence within this area.

437 The multivariate taxonomic average distinctiveness measure (denoted by delta +) showed diversity 438 to be highest at 1000 m, in which eight of the eleven known South African coral families are 439 represented. This result aligns with the knowledge that the global azooxanthellate stony coral 440 pattern (Cairns, 2007) has overall higher species diversity between the 200 and 1000 m. 441 Irrespective, the SIMPER analysis distinguished three major families to contribute to bathymetric 442 zone delineation. The deeper depths (300-1000 m) were characterized by caryophylliids and 443 flabellids, and the shallow zone (50-200 m) by dendrophyllids. These results conforms with the 444 known depth affiliations of these families, in which Dendrophylliidae species occurrence is 445 reported to peak at shallower depths (50 to 300 m) (Cairns, 2001) and extant species of 446 Caryophylliidae and Flabellidae are more prominent in the deeper waters (more than 200 m) 447 (Kitahara, 2011).

448

449 The two sample-specific (i.e., depth and longitude) data sets were applied in combination to 450 extrapolate ocean basin properties (nutrient content, salinity, temperature, etc.), which characterise 451 the oceanographic settings influencing South African marine fauna (i.e., the colder Benguela 452 current along the western margin, and the warmer Agulhas current along the southern and eastern 453 margin). In this context, the permutation models (ANOSIM and RELATE) imply that longitude 454 and depth are good predictors for coral distribution patterns. However, the close to zero R-values 455 (R<0.5) suggests a non-linear relationship even though significant variability is evident in the 456 species composition within the factorial groups. Whilst depth is noted to be one of the main drivers 457 for coral distribution (as shown by BEST results), it is important to recognize that this parameter 458 encompasses several other properties, such as the Aragonite Saturation Horizon (ASH) that is the 459 depth below which calcium carbonate becomes unstable and tends to dissolve (Jiang et al., 2015;

460 Guinotte et al., 2006). Such a zone has been estimated at 700-1500 m depth range south of $\sim 20^{\circ}$ S 461 (Jiang et al., 2015). Eight of the eleven known South African coral families are recorded within 462 this depth range, suggesting these species are surviving within an aragonite saturation state. 463 Interestingly, coral species have been previously reported to withstand saturating conditions. For 464 example, a study undertaken in the Caribbean basin showed the depth of the aragonite saturation 465 horizon to be strongly related to coral assemblage variation, whereby *M. oculata* and *S. variabilis* 466 occur in patchy distributions at or above the saturation zone (Auscavitch et al., 2020). The response 467 of coral species to water properties, such as the ASH, are in no way consistent, highlighting the 468 need for further research to comprehend the underlying environmental drivers of coral distribution.

469

470 Although the azooxanthellate coral data reported herein represent an accumulation of samples over 471 30 years and are the best available representation of the South African fauna, all species richness 472 estimator models did not plateau, demonstrating that the area is still not well sampled and may be 473 much more diverse than currently known. The shape and slope of the curve typically provides 474 information on the species richness, evenness, and heterogeneity of the community being sampled, 475 as well as the adequacy of the sampling effort. In other words, if the species accumulation curve 476 keeps increasing with additional sampling effort then full extent of species diversity within the 477 study area has not yet been captured. Without a doubt, additional sampling coverage will provide 478 clearer conclusions on national coral diversity trends.

479

480 **4.** Conclusion and recommendations

This study examined the best available data for the South African azooxanthellate coral fauna and
presented a pre-processing methodology that caters for historical samples. Differences in

483 azooxanthellate coral species across South Africa's diverse and dynamic oceanographic conditions 484 were revealed, whereby species turnover increased on a west to east axis. A species depth gradient 485 was additionally observed, in which the multivariate diversity measure complemented the existing 486 knowledge on taxa trends. These patterns were evident despite the data limitations related to 487 museum samples. Whilst the lack of museum associated abiotic data still exists, the methodology 488 to standardize co-ordinate information and depth may be considered in other data sets with similar 489 attributes to inform further research to elucidate diversity patterns. In general, historical collections 490 (which represent years of sampling effort) provide a valuable biological data source but require 491 thorough validation. Despite the sparsity and unbalanced nature of the data, knowledge has been 492 advanced and sampling gaps identified. A purposeful application for this existing coral data set 493 will be its integration into multi-taxa biogeography analyses that will support ecosystem 494 description and delineation. The data set will also be valuable for spatial prioritisation and marine 495 spatial planning, particularly alongside taxa that share similar abiotic requirements.

496

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856	Supporting information
857	Appendix A: Excel
858	
859	Appendix B: Draftsman's plot.
860	Figure 2. Draftsman's plot showing the interrelations between depth and the longitudinal groups identified at by the
861	k-R cluster mean analysis.
862	
863	Appendix C: Excel
864	
865	Appendix D: Histograms
866	Figure 5. Simulated distribution/histogram of the test statistic R under the null hypothesis that there are "no
867	differences" in coral patterns within each condition: TOP. krLong groups with observed data at R=0.05 and p=
868	0.001; BOTTOM . z-level (depth) with observed data at $R = 0.072$ and $p = 0.001$.
869	
870	Figure 6: Simulated distribution/histogram of the test statistic R under the null hypothesis that there are "no
871	differences" in coral patterns within depth zones ($R = 0.105$, $p = 0.001$).













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