Supporting Information

Release of Volatile Per- and Polyfluoroalkyl Substances from Plant Fiber-based Food Packaging and Municipal Solid Waste to Gas Under Simulated Landfill Conditions

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2 **Materials and Methods**

3 **Particle Induced Gamma-ray Emission (PIGE) Spectroscopy.** Briefly, each sample was cut 4 into \approx 2.5 \times 2.5 cm strips with methanol-rinsed scissors. Two gamma-ray characteristics generated 5 at 110 keV and 197 keV due to the decay of the ¹⁹F nucleus under continuous proton bombarding 6 (3.4 MeV beam) were measured with a high-purity germanium detector to quantify the total 7 fluorine concentration of the exposed strip. A calibration curve was established using a set of 8 paper spiked with known concentrations of total fluorine. The final total fluorine of each sample 9 was adjusted based on the actual density of each sample (**Equation 1**): 10 Total fluorine (corrected, ng/g) = Total fluorine (measured, ng/g) ∗ r/800(1) 11 (Total fluorine is described as PIGE-F in the manuscript.) The density was determined by cutting 12 4×4 cm strips of each collected food packaging material. Multiple layers of strips of each sample 13 were packed together and the thickness was measured using Vernier calipers. The obtained 14 thickness of the multilayered sample was then divided by the number of layers to get the 15 thickness of a single-layer (d, m) . The food packaging material density $(\rho, kg/m^3)$ was calculated 16 by dividing the average weight of a 4×4 cm strip by the strip's volume.

17 **Inoculum Preparation.** The inoculum used to initiate anaerobic decomposition was 18 developed from a methane-producing consortia derived from decomposing MSW that has been 19 grown on ground residential MSW for over 25 years.¹ To eliminate PFAS contamination in the 20 inoculum, the culture was transferred to 85 mL biological growth medium containing PFAS-21 free Whatman #2 filter paper as the substrate in 160 mL serum bottles. After at least three 22 transfers, the culture was used to inoculate larger vessels so to ultimately prepare \sim 20 L to be 23 used to inoculate reactors. The presence of ionic PFAS in the methanogenic inoculum were 24 analyzed using EPA 537M (**Table S6**) by a commercial lab (SGS). A trip control was sent with 25 the samples to monitor any contamination during sample transport. No PFAS were detected in 26 the inoculum or trip control.

27 **Cellulose, Hemicellulose, and Klason lignin Analysis.** About 1 g sample was extracted with 28 140 mL of toluene/ethanol (2:1, v/v) and then dried.² A known weight (∼0.1- 0.3 g) of ground 29 sample was subjected to a two-stage acid hydrolysis. Sugars (arabinose, galactose, glucose, 30 mannose, and xylose) liberated from acid digestion were then analyzed by HPLC using an ICS 31 2500 pulsed electrochemical detector (Dionex, Sunnyvale, CA). Anhydro correction was used to

32 convert glucose to cellulose and the other sugars to hemicellulose. Klason lignin was measured

33 from the solids remaining after acid hydrolysis as the weight loss on ignition at 550 °C for 2 h.

34 In the Klason lignin method, any organic material that is not soluble in the initial toluene/ethanol

35 wash (which removes lipophilic extractives), and does not dissolve in sulfuric acid, will be

36 counted as lignin.

37 **Residual Solid Extraction for Rinsate Gas Chromatography-Mass Spectrometer (GC-MS)**

38 **Analysis.** Residual solids at the end of the reactor test in the reactors underwent an in-vial solid-

39 liquid extraction for 11 volatile PFAS targets (Table S3) using methodology developed by

40 Rewerts et al.³ Briefly, 10 mg (\pm 4 mg) of residual solids were added to an autosampler vial

41 with 1440 µL of methanol (99.8%, VWR, Radnor, PA). The vial was spiked with 60 µL of eight

42 isotopically labeled internal standards for a final concentration of 100 pg/ μ L. Each vial was

43 vortexed, sonicated for 10 minutes, and allowed to sit for at least 1 hour to settle solid

44 particulates. If the concentration of the solid residual was beyond the calibration curve, the mass

45 extracted was halved (5 mg (\pm 2 mg)). If the concentration was still above the calibration curve,

46 a 1:20 (v/v) dilution of the 5 mg extract into methanol was done in a 2 mL centrifuge tube

47 (VWR, Randor, PA) before being added to the autosampler vial to get the concentration of PFAS

48 targets within the range of the calibration curve. Rinsate of the reactor did not undergo a pre-

49 concentration step before analysis due to the potential loss of volatile PFAS. Therefore, to an

50 autosampler vial with an insert, 288 μ L of rinsate was added to 12 μ L of eight isotopically

51 labeled internal standards for a final concentration of 100 $pg/\mu L$ and vortexed. All analyses for

52 volatile PFAS were done using concurrent solvent recondensation-large volume injection GC-53 MS.³

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55 **GC-MS Analysis**. For volatile PFAS analysis, 10 µL of extract was injected in splitless mode 56 with an inlet temperature of 280 °C. A 4 mm i.d. single taper Topaz inlet liner with 15 mg 57 deactivated quartz wool (Restek, Bellefonte, PA) was used. Helium was used as the carrier gas in 58 a constant flow mode of 1 mL/min. Separations were performed using a deactivated, fused silica 59 tubing capillary column (Agilent, $5 \text{ m} \times 0.53 \text{ mm}$ i.d.) connected to an Rxi-624Sil MS capillary 60 column (Restek, $30 \text{ m} \times 0.25 \text{ mm}$ i.d., 1.40 μ m film thickness). The GC oven temperature 61 program was as follows: 50 °C for 2 min, ramped to 188 °C at a rate of 5 °C/min, then ramped to

62 300 °C at a rate of 15 °C/min for a total run time of 37.07 min. The Agilent 6890 GC was 63 connected to an Agilent 5973N MS (Santa Clara, CA) that was operated in positive chemical 64 ionization mode and in selected ion monitoring mode with methane as the reagent gas at a flow 65 rate of 1 mL/min. Analyte concentrations for FTOHs, sFTOHs, FOSAs, and FOSEs were 66 determined by a calibration curve with a minimum of 6 points with 1/x weighted linear 67 regression or quadratic regression. All standards were prepared in the range of $1-2000$ pg/ μ L. 68 Continuing calibration verification standards (10 and 100 pg/ μ L) were analyzed after every five 69 samples and concentrations were expected to fall within \pm 30%. Method blanks and solvent 70 blanks were analyzed to monitor potential carryover introduced during the experimental 71 procedure; however, none was observed.⁴

72

73 **Leachate Analytical Method.** A solid phase extraction method was used to isolate the analytes 74 from the leachate matrix before analysis. The frozen leachate samples were thawed and 75 centrifuged to remove solids and spiked with an IS to a concentration of 800 ng/L. In preliminary 76 work, a pilot study confirmed that the freezing and re-thawing steps did not negatively impact 77 the analyzed ionic PFAS compounds. The spiked leachate was then loaded onto weak ion 78 exchange cartridges (Waters Corp., 150 mg, 30 µm), which were pre-cleaned, pre-conditioned, 79 and equilibrated with 4 mL 0.03% NH4OH in methanol solution, 4 mL methanol, and 4 mL 80 water before use. The samples were eluted with methanol and 0.03% NH₄OH methanol solution. 81 Both eluents were collected for each sample. To ensure consistent sample composition for all 82 analyzed leachate samples, the extracted samples were dried in a SpeedVac without heat and at 83 the lowest pressure setting (0.1 Torr). The dried samples were stored in a -20°C freezer before 84 being reconstituted with 2 mL of 90:10 water: methanol (*v/v*). Methanol was added first to 85 reconstitute highly hydrophobic PFAS, then water was added to reconstitute the remaining PFAS 86 to minimize PFAS loss.⁵

87 The leachate samples were analyzed with an Agilent 1290 LC coupled to a 6495c Agilent QqQ 88 MS/MS using two separate methods: the first method consisted of a large panel of PFAS that has 89 been previously published^{5, 6} and the second method a FTCA only panel. An InfinityLab 90 Poroshell HPH-C18 delay column (3.0 x 50 mm, 4μm; Agilent, Santa Clara, CA) was used for 91 the large PFAS panel. Approximately 250 μL of leachate sample was injected into the instrument 92 for analysis of 52 ionic PFAS analytes (Table S5). Sample preparation controls consisted of a

S7

- 93 neat positive control (light and heavy calibration mixes spiked into distilled water [800 ng/L,
- 94 PFAS panel; 3ng/mL, FTCA panel]) which was used to monitor extraction efficiency, and a
- 95 negative neat control (spiked with only the heavy mix) to monitor whether contamination
- 96 occurred during the SPE step. The instrument controls were a matrix matched extracted positive
- 97 control [800 ng/L, PFAS panel; 3ng/mL, FTCA panel] to track instrument sensitivity over time
- 98 and a negative control (IS only into matched matrix) to track potential carryover. Those controls
- 99 were analyzed after the calibration curve and after every twenty unknown samples.
- 100
- 101 For the FTCA panel method, a 50 μL aliquot was injected onto a Kinetex F5 (2.1 x 100 mm, 100
- 102 Å; Phenomenex, Torrance, CA) analytical column at 50°C for separation. Aqueous (solvent A:
- 103 water with 5% ACN and 0.1% formic acid) and organic (solvent B: ACN with 5% water and
- 104 0.1% formic acid) solvents were run at 500 μL/min using the following gradient: 0 min: 1% B,
- 105 0.5 min 50% B, 4 min: 99% B, 6 min: 99% B, 6.01 min: 1% B, & 7 min: 1%. Additionally, an
- 106 InfinityLab Poroshell HPH-C18 delay column (3.0 x 50 mm, 4μm; Agilent, Santa Clara, CA)
- 107 was installed in the flow path of the LC to delay any potential PFAS contamination in the HPLC
- 108 solvent from interfering with the sample analysis.
- 109 The method detection limits were 2500 ng/L for 5:3 FTAC and 6:2 FTCA, and 1-2 ng/L for the
- 110 other ionic PFAS analytes. A higher calibration range was needed for the FTCA panel due to the
- 111 high background noise from the PFAS free matrix.
- 112

113 **Results and Discussion**

- 114 **PFAS Release to Leachate.**
- 116 Cumulative PFAS release to leachate is presented in **Figures S12** - **S16**, and in **Table S11**.
- 117

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118 **Release of PFAS to Leachate from the High F Materials**

- 119 High concentrations of 5:3 FTCA and 6:2 FTCA were detected in the leachate from most of the
- 120 high F materials (**Figure S16**B). The high F materials also had the highest volatile PFAS yields
- 121 (343-799 ng/g, dominated by 6:2 FTOH) (**Table 2**). Both 5:3 FTCA and 6:2 FTCA have been

122 reported as the principal biotransformation products of 6:2 FTOH in landfill leachate,⁷ biosolids,⁸ 123 and river sediments.⁹ While 6:2 FTOH was detected in the gas phase around day 10, FTCAs 124 were detected in the leachate as early as day 33 (Compostable bowls) and as late as day 254 125 (Bagasse containers). The ratio of both 5:3 FTCA plus 6:2 FTCA released in leachate to the 126 summed 6:2 FTOH released into the gas phase ranged from 42% to 198% in the Natural plates a-127 c, Compostable bowls, Biodegradable boxes, and Bagasse containers reactors (**Table S11**). This 128 is consistent with the observation that 6:2 FTOH is released early and that some is converted to 129 FTCAs. Of course, the FTOH released to the gas phase represents removal from a reactor and 130 presumably, some FTOH was converted to FTCAs prior to its release in the reactor gas. The 131 discontinuous presence of FTCAs in the Natural plates (**Figure S13**) could be due to the high 132 LOQ of 5:3 FTCA (2500 ng/L) and 6:2 FTCA (2500 ng/L).

133 Both 5:3 FTCA and 6:2 FTCA were released from all high F reactors, except for the Popcorn 134 bags and Eco-friendly plates. In the case of Popcorn bag 1, it is possible that monitoring was 135 terminated prior to FTCA generation. Another possibility is that it was not detected due to the 136 high LOQ. Surprisingly, FTCA was not detected in Popcorn bag 2 or Eco-friendly plates despite 137 the extended monitoring time and high FTOH release. Perfluorohexanoic acid (PFHxA) was the 138 dominant compound measured in the leachate from both Popcorn reactors as well as from the 139 Eco-friendly plates. PFHxA has been reported as the primary impurity, degradant, and 140 metabolite of the side-chain fluorinated polymers and fluorosurfactants due to the shift to short-141 chain fluorotelomer-based products.¹⁰⁻¹³

142

143 **Release of PFAS to Leachate from the Low F and MSW Materials**

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- 144 Leachate PFAS release from the low F materials is summarized in **Figure S16** and **Table S11**.
- 145 PFAS release is generally 2 orders of magnitude lower than that from the high F materials, which
- 146 is consistent with the trend observed for volatile PFAS release. In contrast to the high F reactors,
- 147 neither 6:2 FTCA or 5:3 FTCA were detected (**Table S11**). Instead, relatively low
- 148 concentrations of PFBA, PFPeA, 6:2 FTS, and GenX were released to the leachate. Similarly, in
- 149 the leachate from the three MSW reactors that exhibited typical anaerobic decomposition (MSW-
- 150 May a, MSW-August a, and MSW-August b), small amounts of PFBA, PFPeA, PFHxA, 6:2
- 151 FTS, and GenX were released (**Table S11**). No FTCAs were detected in the MSW leachate
- 152 although they are the dominant PFAS in landfill leachate (e.g., 5:3 FTCA, 7:3 FTCA).^{14, 15}
- 153 However, similar to landfill leachate, other short-chain PFAS that are significant in landfill
- 154 leachate were detected (e.g., PFBA, PFBS, PFPeA, PFHxA).¹⁵
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- 156

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204 **Table S1**. Fluorine content of food packaging materials as screened by PIGE and GC-MS for 205 total fluorine and 6:2 FTOH, respectively.

206 ^aFood packaging materials with 'L' designations were not analyzed using GC-MS.

207 ^b ^bIn the case of the 2022 popcorn bag samples, PIGE analyses were not available and the

208 classification as High or Low F was based on the methanol extraction data alone.

209 ^cFood packaging materials with 'H' designations are *italicized.*

210 ^dpopcorn bag e-2021 was classified as High F despite the relatively low 6:2 FTOH based on the

211 high F as measured by PIGE $(9.99 \times 10^8 \text{ ng/g})$ and the possibility that popcorn bag e-2021

212 contained other volatile PFAS.

213 ^eThe total F measured using PIGE does not account for the density differences between the

214 samples and the standard. Therefore, the total F values used to help identify the high (H) and low

215 (L) designations have been corrected to account for density differences (see eqn. 1).

 $216 \quad \angle$ LOD = less than limit of detection.

 217 NA = not analyzed.

218

219 **Table S2**. Organic content and biochemical methane potential of materials tested in landfill

220 simulation reactors.

Table S3. List of target volatile PFAS analytes for food packaging materials after methanol extraction and for gas sample analysis by TD-GC-MS.^a

^aThe 5:2-sFTOH and 7:2-sFTOH were targets in the liquid extracts because target standards were acquired after gas phase analyses were complete. However, they were suspects for gas phase analyses.

 $b.e$ All PFAS have " x " in the same row is included in this method.

Wellington: Wellington Laboratories (Guelph, ON), SynQuest: SynQuest Laboratories (Alachua, FL), Matrix: Matrix Scientific (Columbia, SC), a.m.u.: atomic mass unit.

Table S4. List of 17 suspect volatile PFAS analytes for gas sample analysis.

CAS#	Analyte	Abbreviation	Neutral Molecular Formula	Matching ISc	d LOQ (ng/L)	$%$ Recovery \pm SD ^e
			Per and Polyfluoroalkyl carboxylic acid (PFCA)			
375-22-4	Perfluorobutanoic acid	PFBA	$C_4HF_7O_2$	\times	10	96 ± 9
2706-90-3	Perfluoropentanoic acid	PFPeA	$C_5HF_9O_2$	\times	5	103 ± 12
307-24-4	Perfluorohexanoic acid	PFHxA	$C_6HF_{11}O_2$	\times	$\sqrt{2}$	88 ± 10
375-85-9	Perfluoroheptanoic acid	PFHpA	$C_7HF_{13}O_2$	\times	5	86 ± 11
335-67-1	Perfluorooctanoic acid	PFOA	$C_8HF_{15}O_2$	\times		64 ± 11
375-95-1	Perfluorononanoic acid	PFNA	$C_9HF_{17}O_2$	\times	100	88 ± 11
335-76-2	Perfluorodecanoic acid	PFDA	$C_{10}HF_{19}O_2$	\times	100	86 ± 10
2058-94-8	Perfluoroundecanoic acid	PFUnDA	$CF3(CF2)9CO2H$	\times		
$307 - 55 - 1$	Perfluorododecanoic acid	PFDoDA	$C_{12}HF_{23}O_2$	\times	100	91 ± 7
72629-94-8	Perfluorotridecanoic acid	PFTrDA	$C_{13}HF_{25}O_2$			
376-06-7	Perfluorotetradecanoic acid	PFTeDA	$CF_3(CF_2)_{12}CO_2H$	\times	100	94 ± 12
67905-19-5	Perfluorohexadecanoic acid	PFHxDA	$C_{16}HF_{31}O_2$	\times		
16517-11-6	Perfluorooctadecanoic acid	PFODA	$C_{18}HF_{35}O_2$			
			Per and Polyfluoroalkyl sulfonic acids (PFSA)			
375-73-5	Perfluorobutanesulfonic acid	PFBS	$C_4HF_9O_3S$	\times	10	95 ± 16
2706-91-4	Perfluoropentanesulfonic acid	PFPeS	$C_5HF_{11}O_3S$			
355-46-4	Perfluorohexanesulfonic acid	PFHxS	$C_6HF_{13}O_3S$	\times	10	87 ± 8
375-92-8	Perfluoroheptanesulfonic acid	PFHpS	$C_7HF_{15}O_3S$			
1763-23-1	Perfluorooctanesulfonic acid	PFOS	$C_8HF_{17}O_3S$	\times	10	92 ± 10
68259-12-1	Perfluorononanesulfonic acid	PFNS	$C_9HF_{19}O_3S$			
2806-15-7	Perfluorodecanesulfonic acid	PFDS	$C_{10}F_{21}NaO_3S$			
			Per and Polyfluoroether carboxylic acid (PFECA)			
$674 - 13 - 5$	Perfluoro-2-methoxyacetic acid	PFMOAA	$C_3HF_5O_3$			
267239-61-2	Perfluoro-2-ethoxypropanoic acid	PEPA	$C_5HF_9O_3$			
39492-89-2	Perfluoro-3,5,7-trioxaoctanoic acid	PFO3OA	$C_5HF_9O_5$			
13252-13-6	Perfluoro-2-propoxypropanoic acid	Gen-X	$C_6HF_{11}O_3$	\times	10	83 ± 10
39492-90-5	Perfluoro-3,5,7,9-butaoxadecanoic acid	PFO4DA	$C_5HF_9O_5$			
39492-91-6	Perfluoro-3,5,7,9,11- pentaoxadodecanoic acid	PFO5DoDA	$C_7HF_{13}O_7$			
773804-62-9	Perfluoroethoxsypropanoic acid	Hydro-EVE	$C_8H_2F_{14}O_4$			
919005-14-4	4,8-Dioxa-3H-perfluorononanoic acid	NaDONA	$C_7H_2F_{12}O_4$			
			Per and Polyfluoroether sulfonic acid (PFESA)			

Table S5. List of 52 target ionic PFAS analytes in leachate using liquid chromatography tandem-mass spectrometry (LC-MS/MS).^{a,b}

a Included in the Fluoryx Mix.

^bThe green highlighted compounds are within the range of 70 to 130%. Uncolored compounds did not pass the data quality control. 24 out of 52 analytes passed the data quality control, and average % recovery and standard deviation were added to these compounds. c All PFAS with " \times " in the same row have match IS in the method.

^dLOQ: Limit of quantitation.

eAverage of 5 standard additions.

Table S6. List of 21 target analytes in EPA 537M method used for ionic PFAS background check of the reactor system. Samples were analyzed using LC-MS/MS by SGS North America Inc.

^aRL: Reporting Limit

bMDL: Method Detection Limit

Table S7. Mass of materials and volume of inoculum added to each reactor and the extent of cellulose and hemicellulose decomposition in each material.^a

^aDecomposition is 1 minus the mass of cellulose or hemicellulose recovered at the end of the decomposition cycle divided by the initial mass.

PFAS yield (ng/g bag)	Popcorn bag 1 day 27	Popcorn bag 2-day 29	Popcorn bag 2-day 203		
4:2 FTOH	0.00213	θ	θ		
6:2 FTOH	14.56	52.8	99.5		
8:2 FTOH	0.0062	θ			
10:2 FTOH	0.025	$_{0}$			
MeFOSA	0.0045	θ			
EtFOSA	0.0079	θ			
6:2 FTAc	0.00044	θ			
8:2 FTAc	0.038	θ			
10:2 FTAc	0.0068	θ			
8:2 FTMAc	0.011	θ			
6:2 FTO	0.84	θ	0.0017		
8:2 FTO	0.0055	θ			
10:2 FTO	0.0016	θ			
12:2 FTO	0.003	θ	0.011		
PFHxI	4.7	0			
PFOI	0.081	θ			
$4:2$ FTI	0.045	θ			
6:2 FTI	0.0083	0			
$7:2$ - s FTOH	0.083	0			

Table S8. Volatile PFAS released in Popcorn Bags 1 reactor after 27 days and Popcorn Bags 2 reactor at day 29 and day 203 (final day).

Table S9. Measured 6:2 FTOH concentration ranges in the generated gas for each high F material.

	Reactors											
$PFAS$ (ng/g)		MSW-May b	MSW-August a		Low F							
	MSW-May a			MSW-August b	a	h						
$6:2$ FTOH	$<$ LOO	$<$ LOO	$<$ LOO	$<$ LOO	$<$ LOO	0.0066						
$6:2$ FTOH ^b		14000		8000	NA							
$8:2$ FTOH	$<$ LOO	$<$ LOQ	$<$ LOO	$<$ LOO	$<$ LOO	0.0033						
$10:2$ FTOH	$<$ LOO	$<$ LOQ	$<$ LOO	0.013	0.02	0.0045						
$7:2$ -sFTOH	0.034	0.62	0.1	0.135	0.07	0.4						
Σ PFAS	0.078	0.62	0.1	0.148	0.09	0.41						

Table S10. Cumulative volatile PFAS yield in Low F and MSW reactors (ng/g) at the end of the monitoring period.^a

aif a compound is LOD in every case, then it was not included.

b6:2 FTOH in ground MSW material prior to testing.

Reactor groups	PFBA	PFPeA	PFHxA	PFHpA	PFOA ^b	4:2 FTS	6:2 FTS	GenX	5:3 FTCA	6:2 FTCA	Summed FTCAs	Summed PFAS (Leachate)	Summed PFAS (gas)	Summed FTCAs/6:2 FTOH in gas phase
high F-Natural Plates (a,b,c)	2.64	0.906	9.83	0.245		0.034			115	220	335	340	799	0.42
	4.71	0.228	6.2						410	142	553	561	487	1.13
	10.1	0.314	11.4						342	130	472	445	556	0.85
high F-Compostable Bowls	2.77	1.3	14.3	0.07				0.01	225	154	379	279	343	1.11
high F-Biodegradable Boxes	2.29	0.173	4.7						176	93	269	276	484	0.57
high F-Bagasse Containers	3.07	1.45	11.9	1.13				0.045	182	58	240	257	125	1.98
high F-Eco-friendly Plates	4.98	1.47	18.3	1.1			0.098	0.001				25.6	62	
high F-Popcorn Bag 1	0.009	$\overline{0}$	0.757	0.203								0.96	20.4	
high F-Popcorn Bag 2	\sim LOD		0.862	0.648			0.029		$a <$ LOQ	$a <$ LOQ		15.9	99.7	
Low F-mixed paper	0.793	0.284										1.02	0.09	
plates, eco-friendly food trays, and polycoated freezer paper (A, B)	0.254	0.32					0.116	0.08				0.488	0.41	
MSW -May (a, b)	3.57	0.378	0.095		0.074		0.057	0.029				4.09	0.078	
	0.129	0.113	0.003				0.04					0.197	0.62	
	0.212	0.314										0.525	0.1	
MSW-August (a, b)		0.335						0.004				0.335	0.148	

Table S11. Cumulative PFAS yield (ng/g) in leachte for all reactors and the ratio of FTCA released in leachate to 6:2 FTOH released to gas phase.

^aLOQ is 2500 ng/L.

 $\frac{b_{\text{In}}}{c_{\text{In}}}$ the case of PFOA, concentrations were 22 to 57% above the highest concentration in a control reactor.

All results <LOD are left empty.

Table S12. Profile of volatile PFAS and total F for select samples pre- and post-digestion.

^a<LOD: below limit of detection ^bNA: not analyzed ^cNC: not calculated

^dThe average of the three low F substrates (22400) was used for calculations since equal masses of each were added to the duplicate low F reactors

The summed PFAS includes suspect compounds.

Table S13. Summed PFAS as F released in leachate and ratio of leachate F to gas phase F.

Figure S1. Schematic of the test reactor. **A.** 2L-glass reactor with flange connection part; **B.** drainage layer of pea gravel; **C.** #120 stainless steel woven wire mesh to reduce clogging of the reactor outlet; **D.** glass cap for the reactor with flange connection part; **E.** flange clamps to hold the two-pieces of glass together and keep the system sealed; a 245 EPDM gasket was placed between the two pieces of glass; **F.** stainless steel adjustable clamp; **G.** 10 mL Norm-Ject Luer lock polypropylene syringe; **H.** T-shape plastic 3-way barb fitting, polypropylene, 1/4" OD; **I.** Cole-Parmer female/male Luer hose barb adapter, 1/4" OD; **J.** McMaster quick-disconnect tube couplings, polysulfone, 1/4" OD; **K.** tubing flow control hose clamp, 3/4" OD 54; **L.** stainless steel quick-turn tube couplings, 1/4" OD; **M.** one-way Luer stopcock; **N.** tygon tube, 1/4" ID, flexible, clear, PVC; **O.** 10 L-foil gas bag (SKC. Inc.), **P.** 2L-Thermo Scientific™ Labtainer™ bioProcess container, with 3 Ports, MPC Insert, MPC Body, and Luer lock.

Figure S2. Gas sampling system. **A**. 0.5 L-foil gas bag; **B.** check valve; **C.** 160 mL-serum bottle; **D.** H₂SO₄ aqueous solution (pH <3.7), 1 g/L methyl orange was added as a pH indicator (3.1-4.4); **E.** 18-gauge needles; **F.** rubber stopper to seal serum bottle. **G.** straight adapters, tube pushon barbed hose, 1/4" ID; **H**. Markes Universal gas sampling tube 1/4" OD (IS spiked TD tube); **I**. One-way Luer stopcock.

Figure S3. Measured PFAS concentrations in the Whatman #2 filter paper Control reactors. **A**. Control 1 reactor and **B**. Control 2 reactor.

Figure S4. pH of all reactors: **A.** Control and Low F; **B.** MSW; **C.** High F (Natural Plates and Popcorn Bags), and **D.** High F (Compostable Bowls, Biodegradable Boxes, Bagasse Containers, and Eco-friendly Plates).

Figure S5. CH4 yield in the reactors of **A.** Whatman #2 filter paper control; **B.** MSW; **C.** Popcorn bags, and **D.** High F (Compostable Bowls, Biodegradable Boxes, Bagasse Containers, and Eco-friendly Plates).

Figure S6. CH₄ generation rate (mL/day-dry g) of all reactors: **A.** Control and Low F; **B.** MSW; **C.** High F (Natural Plates and Popcorn Bags), and **D.** High F (Compostable Bowls, Biodegradable Boxes, Bagasse Containers, and Eco-friendly Plates).

Figure S7. The total fluorine concentration of 46 tested materials (5 popcorn bags sampled in 2022 were not tested by PIGE).

Figure S8. Volatile PFAS released from Popcorn Bags 1 reactor during the 27-day monitoring period. PFAS data are presented as a yield. As such, a flat line indicates that no PFAS was detected in a sample. The first gas sampling and detection day was day 12, and concentrations before day 12 are extrapolated.

Figure S9. Time-dependent release of volatile PFAS to the gas phase in **A.** Natural Plates a and **B.** Natural Plates b. The summed PFAS and 6:2 FTOH lines overlap in both reactors.

Figure S10. Volatile PFAS released from High F reactors. **A**. Compostable Bowls; **B.** Biodegradable Boxes; **C.** Bagasse Containers; **D.** Eco-friendly plates. The summed PFAS and 6:2 FTOH lines overlap in all reactors.

Figure S11. Time-dependent release of volatile PFAS in **A**. Low F a, **B**. Low F b, **C**. MSW-August b, and **D**. MSW-May a reactor. The concentration of 7:2-sFTOH was first detected on day 74 for Low F a reactor and day 51 for Low F b reactor, 8:2 FTOH were first detected on day 43 in Low F b reactor, 10:2 FTOH were first detected on day 148 and 43 in Low F a and Low F b reactors; yields of these compounds before first detection were extrapolated. The summed PFAS and 7:2-sFTOH lines overlap in Figure S11C.

Figure S12. PFAS concentration and relative abundance in leachate samples over time for the High F reactors: **A**&**B**. Compostable Bowls, **C**&**D**. Biodegradable Boxes, **E**&**F**. Bagasse Containers, **G**&**H**. Eco-friendly Plates, **I**&**J**. Popcorn Bags 1, and **K**&**L**. Popcorn Bags 2.

Figure S13. PFAS concentration and relative abundance and CH₄ generation rate in the leachate samples over time for: **A**&**B**. High F Natural Plates a, **C**&**D**. High F Natural Plates b, and **E**&**F**. High F Natural Plates c. The discontinuous reporting of FTCAs concentrations in Natural Plates a could be due to the high LOQ of 5:3 FTCA (2500 ng/L) and 6:2 FTCA (2500 ng/L).

Figure S14. The PFAS concentration and relative abundance of each PFAS component in the leachate samples of **A**&**B.** Low F a **C**&**D.** Low F b.

Figure S15. The PFAS concentration in the leachate samples of **A.** MSW-May a, **B.** MSW-May b, **C.** MSW-August a, and **D.** MSW-August b. In the case of PFOA, concentrations were 22 to 57% above the highest concentration in a control reactor.

Figure S16 A. Cumulative summed and individual ionic PFAS released to leachate from each reactor. **B.** The fraction of each PFAS released to the leachate.

