# Bacterial Disproportionation of Elemental Sulfur Coupled to Chemical Reduction of Iron or Manganese

BO THAMDRUP,<sup>1\*</sup> KAI FINSTER,<sup>1</sup> JENS WÜRGLER HANSEN,<sup>1</sup> AND FRIEDHELM BAK<sup>2</sup>

Department of Microbial Ecology, Institute of Biological Sciences, University of Aarhus, Ny Munkegade, DK-8000 Aarhus C, Denmark,<sup>1</sup> and Max Planck Institut für terrestrische Mikrobiologie, W-3550 Marburg an der Lahn, Germany<sup>2</sup>

Received 27 July 1992/Accepted 6 October 1992

A new chemolithotrophic bacterial metabolism was discovered in anaerobic marine enrichment cultures. Cultures in defined medium with elemental sulfur (S<sup>0</sup>) and amorphous ferric hydroxide (FeOOH) as sole substrates showed intense formation of sulfate. Furthermore, precipitation of ferrous sulfide and pyrite was observed. The transformations were accompanied by growth of slightly curved, rod-shaped bacteria. The quantification of the products revealed that S<sup>0</sup> was microbially disproportionated to sulfate and sulfide, as follows:  $4S^0 + 4H_2O \rightarrow SO_4^{2-} + 3H_2S + 2H^+$ . Subsequent chemical reactions between the formed sulfide and the added FeOOH led to the observed precipitation of iron sulfides. Sulfate and iron sulfides were also produced when FeOOH was replaced by FeCO<sub>3</sub>. Further enrichment with manganese oxide, MnO<sub>2</sub>, instead of FeOOH yielded stable cultures which formed sulfate during concomitant reduction of MnO<sub>2</sub> to Mn<sup>2+</sup>. Growth of small rod-shaped bacteria was observed. When incubated without MnO<sub>2</sub>, the culture did not grow but produced small amounts of SO<sub>4</sub><sup>2-</sup> and H<sub>2</sub>S at a ratio of 1:3, indicating again a disproportionation of S<sup>0</sup>. The observed microbial disproportionation of S<sup>0</sup> only proceeds significantly in the presence of sulfide-scavenging agents such as iron and manganese compounds. The population density of bacteria capable of S<sup>0</sup> disproportionation in the presence of FeOOH or MnO<sub>2</sub> was high, >10<sup>4</sup> cm<sup>-3</sup> in coastal sediments. The metabolism offers an explanation for recent observations of anaerobic sulfide oxidation to sulfate in anoxic sediments.

In aquatic sediments sulfur undergoes bacterial redox transformations which, in their entirety, are known as the microbial sulfur cycle. The reduction of sulfate to hydrogen sulfide ( $H_2S$ ) is catalyzed by the strictly anaerobic, sulfate-reducing bacteria. The microbiology of sulfate reduction has been studied intensively (37, 48, 49), and quantification of the process by radiotracer techniques has proved it to be a major pathway of carbon mineralization in marine sediments with significance for the global carbon cycle (21, 23, 42).

In the oxic zone of sediments the formed  $H_2S$ , or other partially reduced sulfur compounds, is reoxidized to sulfate by chemolithotrophic bacteria such as *Beggiatoa* and *Thiobacillus* spp. (27, 34). In the absence of  $O_2$ , some of these bacteria may use nitrate as the terminal electron acceptor (e.g., see reference 29).

There are several indications, however, that sulfide is oxidized completely to sulfate in the absence of oxygen and nitrate in marine and limnic sediments.

(i) The depth in sediments at which  $H_2S$  accumulates is often several centimeters below the oxic surface layer (23). The separating suboxic zone is characterized by the accumulation of  $Mn^{2+}$  and  $Fe^{2+}$  from metal oxide reduction in the pore water (19, 22, 43). Sulfate reduction continuously produces  $H_2S$  in the suboxic zone, but the concentrations are kept low (<1  $\mu$ M), presumably by reactions involving iron and manganese oxides (4, 9, 43).

(ii) Addition of manganese oxide to anoxic salt marsh sediments caused the oxidation of solid-phase sulfides to sulfate with apparent biological catalysis (1, 28). Addition of ferric hydroxide only resulted in slight sulfate production. The depth distribution of manganese and FeS in Danish coastal sediments also suggested manganese oxide to be an important oxidant for FeS-derived sulfur (10). So far, however, bacteria capable of oxidizing reduced sulfur compounds with iron or manganese at circumneutral pH have not been described.

(iii) A strong indication of an anaerobic oxidation of  $H_2S$  to sulfate came from recent radiotracer studies. In anoxic, nitrate-free sediments, added radiolabeled  $H_2S$  was oxidized to sulfate (14, 16, 24–26). The mechanism was not revealed, but thiosulfate was the major immediate oxidation product. The most important process in the further transformation of thiosulfate was bacterial disproportionation (3).

The studies mentioned above inspired us to search for bacteria oxidizing reduced sulfur species with manganese or iron oxide. Enrichment for bacteria oxidizing  $H_2S$  with manganese or iron oxides is complicated by the rapid chemical reaction between these compounds, yielding  $S^0$  as the major product (8, 38, 39). To avoid this, we chose elemental sulfur as the sulfur source in our enrichments. The outcome of the enrichment studies is presented here.

### MATERIALS AND METHODS

The procedures used for preparation of culture media and cultivation of anaerobic bacteria were basically those used for sulfate-reducing bacteria described by Widdel and Bak (49).

**Medium.** The basal medium had the following composition (in grams per liter of distilled water): NaCl, 15; NaHCO<sub>3</sub>, 2.5; MgCl<sub>2</sub> · 6H<sub>2</sub>O, 2; CaCl<sub>2</sub> · 2H<sub>2</sub>O, 0.4; KCl, 0.5; NH<sub>4</sub>Cl, 0.25; and KH<sub>2</sub>PO<sub>4</sub>, 0.2. The medium was autoclaved and cooled under an atmosphere of O<sub>2</sub>-free N<sub>2</sub>-CO<sub>2</sub> (88:12, vol/vol), after which the following components were added aseptically from sterile stock solutions (per liter): 30 ml of 1 M NaHCO<sub>3</sub>, 2 ml of trace element solution without chelating agent, and 1 ml each of three different vitamin

<sup>\*</sup> Corresponding author.

solutions (49). The pH was adjusted to 7.3. The medium was dispensed aseptically into 50- or 250-ml airtight screw-cap bottles. Elemental sulfur and manganese or iron oxide were added directly before inoculation.

Flowers of sulfur were thoroughly ground in a mortar together with distilled water, and the suspension was autoclaved at 114°C for 30 min. Sulfur was added in excess (ca.  $0.3 \text{ mmol liter}^{-1}$ ) to the medium with a sterile spatula.

Red, amorphous ferric hydroxide (FeOOH) was prepared by the method of Lovley and Phillips (30), and its uncrystalline nature was confirmed by x-ray diffraction analysis.

Ferrous carbonate (FeCO<sub>3</sub>) was synthesized by mixing  $Na_2CO_3$  powder into an anoxic solution of an equimolar amount of FeCl<sub>2</sub> (0.5 M) with vigorous stirring under oxygen-free  $N_2$ . The pale grey FeCO<sub>3</sub> precipitate was washed in anoxic distilled water until the NaCl concentration of the suspension was approximately equal to that of the medium. Dissolution of the FeCO<sub>3</sub> in 0.5 N HCl (see below) showed <10% Fe(III).

Manganese oxide  $(MnO_2)$  was prepared by the method of Burdige (7) with the modification that the amount of NaOH was doubled. X-ray diffraction analysis showed no lines, indicating poor crystallinity of the MnO<sub>2</sub>. The average oxidation state of Mn determined by the oxalate method (33) was  $3.98 \pm 0.05$  (n = 13). The FeOOH and MnO<sub>2</sub> suspensions contained ca. 0.5 mol of Fe or Mn liter<sup>-1</sup> and were stored at 4°C and added to the culture medium under O<sub>2</sub>-free N<sub>2</sub>.

Enrichment and cultivation. Completely filled 50-ml screwcap bottles were used for enrichments. Medium supplemented with elemental sulfur and FeOOH (40 mmol of Fe liter<sup>-1</sup>) or  $MnO_2$  (30 mmol of Mn liter<sup>-1</sup>) was inoculated with marine sediment from the 1- to 5-cm depth interval. The sediment was collected at a 16-m-deep location in Aarhus Bay, Denmark. Inoculation and subsequent transfers were made with 5 ml of inoculum into 50-ml screw-cap bottles with no headspace. Incubation was stationary at 30°C in the dark.

**Determination of stoichiometries.** For determination of stoichiometries, the enrichments were subcultured into 250-ml screw-cap bottles. These were supplied with S<sup>0</sup> and different amounts of  $MnO_2$ , FeOOH, or FeCO<sub>3</sub> or with S<sup>0</sup> only. Uninoculated controls were made in parallel.

Samples were withdrawn with sterile glass pipettes after vigorous shaking. The gas phase was immediately flushed with N<sub>2</sub>-CO<sub>2</sub> (88:12), and after the first sampling, the screw-caps were replaced by black rubber stoppers. To avoid loss of accumulating free H<sub>2</sub>S because of flushing with N<sub>2</sub>-CO<sub>2</sub>, the volume of liquid removed from subcultures of the S<sup>0</sup>-MnO<sub>2</sub> enrichment with S<sup>0</sup> only was replaced by pieces of glass and the bottles were sealed again immediately.

**Separations.** For determination of  $SO_4^{2-}$ ,  $H_2S$ ,  $Fe^{2+}$ , and  $Mn^{2+}$  and of particulate Mn(II), an aliquot was filtered through a 0.45-µm cellulose acetate filter under N<sub>2</sub>. For analysis of  $SO_4^{2-}$  and  $H_2S$ , filtrate was collected in 2% ZnCl<sub>2</sub> to stabilize sulfide as ZnS. The rest of the filtrate was acidified with HCl for analysis of dissolved  $Fe^{2+}$  or  $Mn^{2+}$ .

The Mn(II) in the precipitate [adsorbed Mn(II), MnCO<sub>3</sub>] remaining on the filter was measured after dissolution. After the filter was washed with distilled water,  $Mn^{2+}$  was dissolved and washed out with 2 ml of H<sub>2</sub>SO<sub>4</sub> and 2 ml of distilled water. The acid and water from the last wash were mixed and analyzed for Mn<sup>2+</sup>.

Ferric iron was dissolved by 16-h extraction in 0.5 N HCl. This treatment dissolved amorphous FeOOH completely. Ferric iron was determined as the difference between total Fe and  $Fe^{2+}$  in the solution.

For determination of sulfide precipitates, 2 ml of culture was initially fixed in 1 ml of 20% zinc acetate. The sulfide precipitates were extracted by three slightly different sulfide distillation procedures.

Samples from the incubations with FeCO<sub>3</sub> were analyzed by a two-step distillation that separated acid-volatile (AVS;  $H_2S$  and FeS) and chromium(II)-reducible (CRS; S<sup>0</sup> and pyrite, FeS<sub>2</sub>) sulfides (15). Since S<sup>0</sup> caused a high background of sulfide during chromium reduction, iron sulfides distilled in this step were quantified indirectly as iron dissolved. The FeCO<sub>3</sub> dissolved completely in the acid distillation step.

Two-step distillation was also used with the final samples from the time course experiments with  $FeCO_3$  and FeOOHcultures. Here, however, S<sup>0</sup> in the samples was removed by extraction with CS<sub>2</sub> prior to chromium distillation. Thus, the CRS measured only represented iron sulfides such as pyrite. Control experiments confirmed that the CS<sub>2</sub> treatment dissolved all S<sup>0</sup> in the samples.

Samples from the incubations with  $MnO_2$  were analyzed by acid distillation only. To reduce oxidized manganese, which could interfere by oxidizing the dissolving sulfide (28), the samples were mixed with a solution of 1%  $NH_2OH$  HCl for 10 min prior to acidification. Using this treatment, no loss of sulfide could be detected during control distillations of a defined mixture of ZnS and  $MnO_2$ .

Analyses. Sulfate was measured by nonsuppressed anionexchange chromatography (pump model 510, IC-PAK anion column, conductivity detector model 430; Waters). In a few cases,  $SO_4^{2-}$  was measured by  $BaSO_4$  precipitation and turbidometry (12). Sulfide was quantified by the methylene blue method (11) with a detection limit of 1  $\mu$ M. Soluble and HCl-extracted Fe<sup>2+</sup> were measured spectrophotometrically by reaction with Ferrozine (44) (0.02% in 50 mM HEPES [N-2-hydroxyethylpiperazine-N'-2-ethanesulfonic acid], pH 7), and total HCl-extractable Fe was measured similarly with reducing Ferrozine (Ferrozine with 1% hydroxylamine hydrochloride) (36). Iron dissolved during chromium distillation and Mn<sup>2+</sup> were determined by flame atomic absorption spectroscopy (Perkin-Elmer 370 A spectrometer).

Cell counts. Since the Mn and Fe precipitates did not allow growth measurement by the monitoring of increases in optical density, cells were counted directly by staining with acridine orange. Culture samples were preserved with glutaraldehyde (2.5% final concentration). Before counting, manganese precipitates were removed with 1.5 M  $NH_2OH \cdot HCl$  in 0.25 N HCl while iron precipitates were dissolved by a dithionite solution (sodium dithionite, 50 g liter<sup>-1</sup> in 0.2 M sodium citrate–0.35 M acetic acid [9]). Then, 50 to 500  $\mu$ l of preserved sample was added to 2 ml of the extractants. After 5 min, the mixture was filtered through a 0.2-µm polycarbonate filter, which was subsequently stained with 0.01% acridine orange for 1 min. The staining solution was then filtered off and the filter was washed in distilled water. All solutions were sterilized by filtering (0.2-µm filter). Cells were counted by epifluorescence microscopy.

Viable counts in sediment samples. To enumerate bacteria metabolizing  $S^0$  in the presence of either FeOOH or  $MnO_2$  in natural sediments, we used most-probable-number (MPN) dilutions (2). The dilution series were prepared as described by Bak and Pfennig (4). The medium used was identical to the enrichment medium. The tubes for counting were incubated for 6 months at 20°C and checked for bacterial growth every second week. Grown cultures were tested for  $H_2S$  and

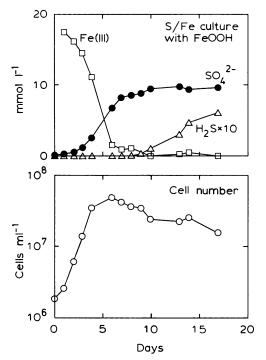


FIG. 1. Time course of  $SO_4^{2-}$  and  $H_2S$  production, Fe(III) reduction, and bacterial growth in the S/Fe enrichment culture incubated with amorphous FeOOH: average of two parallel cultures. (Top) Chemical analyses; note that H<sub>2</sub>S concentrations are multiplied by 10. (Bottom) Cell number; note logarithmic scale.

 $SO_4^{2-}$  formation. With  $S^0$  and  $MnO_2$ , counts were made with marine sediment from three locations around Denmark: Aarhus Bay at 16-m depth, a salt marsh on the west coast of Jutland, and manganese-rich sediment from the deepest part of Skagerrak at 695-m depth. With S<sup>0</sup> and FeOOH, enumerations were done in sediment from the salt marsh.

## RESULTS

With both FeOOH and MnO<sub>2</sub>, cultures that metabolized  $S^0$  with concomitant reduction of the metal oxide were obtained. The two enrichment cultures were designated S/Fe culture and S/Mn culture, respectively. The cultures were routinely transferred every other week when growth had ceased (see below). Neither growth nor reduction of metal oxides was detected within 2 months when the cultures were pasteurized (75°C, 20 min), not inoculated, or inoculated with an anoxically autoclaved enrichment culture. Pasteurization of growing cultures terminated all measured processes immediately. The determination of stoichiometries was carried out after more than 20 subcultures.

S/Fe culture. The enrichment culture with S<sup>0</sup> and amorphous ferric hydroxide grew and reduced the added FeOOH within 1 week, with concomitant  $SO_4^{2-}$  production (Fig. 1). A lag phase was not observed. The culture was dominated by rod-shaped, slightly curved cells, attached to the particulate phase. The growth rate during the exponential phase was about 1 day<sup>-1</sup>. During the iron reduction phase, the concentration of free Fe<sup>2+</sup> did not exceed 200  $\mu$ M. The red color of the FeOOH precipitate changed to black, and the yellow sulfur grains received a grey coating. Formation of black precipitates consisting of magnetite ( $Fe_3O_4$ ) has been described for cultures of dissimilatory Fe(III)-reducing bacteria (32). However, the black precipitate observed in our cultures was not magnetite but ferrous sulfides. The iron sulfides were recovered partly as AVS and partly as CRS (Table 1). In the precipitate of grown cultures, pyrite and S<sup>0</sup> were the only compounds detectable by x-ray diffraction analysis.

Free H<sub>2</sub>S was not detected in the growth phase, but after depletion of FeOOH, SO<sub>4</sub><sup>2-</sup> production continued briefly and H<sub>2</sub>S accumulated to a final concentration of 0.5 to 1 mM.

The pH of the culture medium decreased during growth. In the cultures sampled for the time course experiment, the pH declined to 6.5, while in grown cultures with larger amounts of Fe, values down to pH 4.5 were measured. Even when the cultures reached such low pH values, growth resumed shortly after transfer.

The molar ratio of Fe(III) added to  $SO_4^{2-}$  produced was close to 2 (Table 1). A minor  $SO_4^{2-}$  production was seen in the absence of FeOOH.

When the S/Fe culture was transferred to medium containing ferrous carbonate instead of ferric hydroxide, again bacterial growth and production of sulfate and iron sulfides occurred (Fig. 2 and Table 2). The rate of  $SO_4^{2-}$  production was lower than in the presence of FeOOH. Growth ceased after 5 days, while  $SO_4^{2-}$  production continued at a constant rate for more than 15 days. After this, H<sub>2</sub>S accumulated to ca. 1 mM.

When FeCO<sub>3</sub> was added in relatively low amounts (10 to 25 mmol liter<sup>-1</sup>), H<sub>2</sub>S accumulated to 0.5 to 1 mM after a few

Fe(III) reduced $(\text{mmol liter}^{-1})^a$	$SO_4^{2-}$ produced (mmol liter <sup>-1</sup> ) <sup>a</sup>	AVS produced $(\text{mmol liter}^{-1})^a$	CRS produced (mmol of S liter <sup>-1</sup> ) <sup><math>a,b</math></sup>	Fe(III)/SO <sub>4</sub> <sup>2-</sup> (mol:mol) <sup>a</sup>	Cells formed $(ml^{-1})^c$
17.4	9.1	9.0	9.7	1.91	$5.1 \times 10^{7}$
18.6	9.8	8.0	14.7	1.90	$4.1 \times 10^{8}$
42	19.7	$ND^{d}$	ND	2.13	ND
37.5	15.7			2.39	
14.6	6.7			2.18	
7.3	4.1			1.78	
0 <sup>e</sup>	0.2			0	
0 <sup>e</sup>	0.7			0	

TABLE 1. S/Fe cultures with amorphous ferric hydroxide: stoichiometry and bacterial growth

Grown cultures; all Fe(III) reduced to Fe(II).

<sup>b</sup> S<sup>0</sup> removed by CS<sub>2</sub> extraction.
<sup>c</sup> Difference between day 0 and maximum cell number.

<sup>d</sup> ND, not determined.

" No FeOOH added.

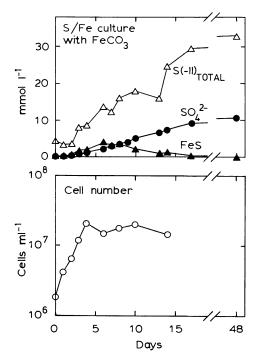


FIG. 2. Time course of  $SO_4^{2-}$  and sulfide production and bacterial growth in the S/Fe culture incubated with FeCO<sub>3</sub>. (Top) Chemical analyses. (Bottom) Cell number; note logarithmic scale. S(-II)<sub>TOTAL</sub> includes both sulfide bound in FeS and CRS.

days of incubation. At this time, the sum of AVS and CRS was about 2 mmol liter<sup>-1</sup>. Thus, most of the FeCO<sub>3</sub> remained unchanged and coexisted with  $H_2S$  in these cultures. In cultures with more FeCO<sub>3</sub> (37 mmol liter<sup>-1</sup>), the ratio of Fe(II) added to sulfate produced was close to 3 (Table 2).

S/Mn culture. In the culture grown with elemental sulfur and manganese oxide, all Mn(IV) was reduced to Mn(II) within 2 weeks (Fig. 3). A simultaneous production of sulfate was observed. The Mn reduction and  $SO_4^{2-}$  production typically started after a lag phase of 1 to 2 days and were accompanied by exponential bacterial growth. The culture was dominated by small, nonmotile, rod-shaped cells that were mainly attached to the particulate phase. The growth rate during the exponential phase was  $1 \text{ day}^{-1}$ , similar to that of the S/Fe culture. Reduction of MnO<sub>2</sub> changed the color of the precipitates from dark brown to whitish rosyred. Soluble Mn<sup>2+</sup> never exceeded 1.2 mM, indicating that most Mn(II) was associated with the solid phase. The mineral formed was presumably MnCO<sub>3</sub>, as reported from dissimilatory Mn-reducing cultures (31). No sulfide precipitates were detected by distillation. After depletion of the oxide, bacterial growth and the production of  $SO_4^{2-}$  ceased



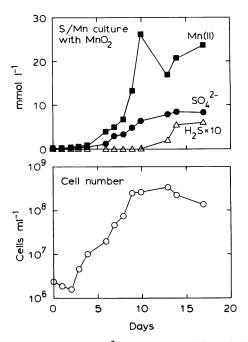


FIG. 3. Time course of  $SO_4^{2-}$ ,  $H_2S$ , and Mn(II) production and concomitant bacterial growth in the S/Mn enrichment culture incubated with MnO<sub>2</sub>: average of two parallel cultures. (Top) Chemical analyses; note that  $H_2S$  concentrations are multiplied by 10. (Bottom) Cell number; note logarithmic scale. Mn(II) includes both particulate and soluble forms. MnO<sub>2</sub> added: 23 mmol liter<sup>-1</sup>.

and a further release of about  $1 \text{ mM H}_2S$  (final concentration) was observed before all transformations stopped.

The pH of the medium increased. In the time course experiments it rose to 7.5, and in cultures with larger amounts of  $MnO_2$ , values of about 8.5 were measured. At this high pH, the medium turned yellow, presumably due to the presence of polysulfides, and some manganous sulfide might have formed, as indicated by the sulfide smell produced upon acidification of the precipitate.

The molar ratio of Mn(II) to  $SO_4^{2-}$  produced was 2.7 on average (Table 3). Control cultures without addition of MnO<sub>2</sub> concomitantly produced  $SO_4^{2-}$  and H<sub>2</sub>S (Fig. 4 and Table 3). The  $SO_4^{2-}$  production ceased when H<sub>2</sub>S reached a concentration of approximately 1.5 mM. The ratios of H<sub>2</sub>S to  $SO_4^{2-}$  produced were 3.61 and 3.68 in two parallel cultures. However, growth was insignificant in the absence of MnO<sub>2</sub>. The maximum cell number generally increased with increasing additions of MnO<sub>2</sub> (Table 3).

**Enumeration of bacteria in sediments.** Viable counts of bacteria metabolizing  $S^0$  in the presence of either FeOOH or MnO<sub>2</sub> were made by using the MPN method. The results are given in Table 4. The highest numbers were found with

TABLE 2. S/Fe cultures with ferrous carbonate: stoichiometry and bacterial production

$FeCO_3$ added (mmol liter <sup>-1</sup> )	$SO_4^{2-}$ produced (mmol liter <sup>-1</sup> ) <sup><i>a</i></sup>	AVS produced $(\text{mmol liter}^{-1})^a$	CRS produced (mmol of S liter <sup>-1</sup> ) <sup><math>a,b</math></sup>	FeCO <sub>3</sub> /SO <sub>4</sub> <sup>2-</sup> (mol:mol)	Cells formed $(ml^{-1})^c$
37.6	13.2	0.13	50.8	2.85	$9.3 \times 10^{6}$
37.6	11.2	0.11	43.4	3.36	$1.8 \times 10^{7}$

<sup>a</sup> 48 days of incubation.

<sup>b</sup> S<sup>0</sup> removed by CS<sub>2</sub> extraction.

<sup>c</sup> Difference between day 0 and maximum cell number.

TABLE 3. S/Mn cultures: stoichiometry and bacterial growth

Mn(II) produced (mmol liter <sup>-1</sup> ) <sup><math>a</math></sup>	$SO_4^{2-}$ produced (mmol liter <sup>-1</sup> ) <sup>a</sup>	Mn(II)/SO <sub>4</sub> <sup>2-</sup> (mol:mol) <sup>a</sup>	Cells formed (ml <sup>-1</sup> ) <sup>b</sup>
24.7	8.1	3.05	$4.8 \times 10^{8}$
22.6	8.5	2.66	$2.3 \times 10^{8}$
14.8	5.1	2.90	$8.2 \times 10^{7}$
8.0	3.1	2.22	$6.9 \times 10^{7}$
0°	1.6	0	$1.7 \times 10^{6}$
$0^c$	0.47	0	ND
0 <sup>c</sup>	0.43	0	

<sup>a</sup> Grown cultures; all Mn(IV) reduced to Mn(II).

<sup>b</sup> Difference between day 0 and maximum cell number. ND, not determined.

<sup>c</sup> No MnO<sub>2</sub> added.

FeOOH in the salt marsh sediment,  $10^5$  to  $10^6$  cells per cm<sup>3</sup>. With MnO<sub>2</sub>, the numbers obtained in the Aarhus Bay and salt marsh sediments were about  $10^4$  to  $10^5$  cells per cm<sup>3</sup>, while the numbers found in Skagerrak were significantly lower,  $<100 \text{ cm}^{-3}$ .

### DISCUSSION

Sulfur transformations in enrichment cultures. In marine enrichment cultures established in anoxic bicarbonate-buffered defined medium with elemental sulfur and either ferric hydroxide or manganese oxide, sulfur was transformed by biologically catalyzed processes, during which both sulfate and sulfide were formed (Fig. 1 and 3). The sulfur metabolism sustained bacterial growth which, since no organic compounds except vitamins were added, most likely involved fixation of bicarbonate.

The observed formation of sulfide, either in the form of ferrous sulfide minerals or as free H<sub>2</sub>S in late growth phases or in cultures without metal oxide additions (Fig. 1, 3, and 4), must be the result of a reduction of part of the added elemental sulfur. However, since no external electron donors, such as organic compounds or H<sub>2</sub>, were added, the measured sulfide cannot be the product of a dissimilatory microbial sulfur reduction. Thus, a microbial disproportionation of elemental sulfur, similar to the recently described disproportionation of thiosulfate and sulfite by sulfate-reducing bacteria (3), is the only plausible explanation for the formation of sulfide in our enrichment cultures. The formed H<sub>2</sub>S will react spontaneously with amorphous FeOOH, as expected by the following equation (e.g., see references 38 and 39):

$$3H_2S + 2FeOOH \rightarrow S^0 + 2FeS + 4H_2O \tag{1}$$

TABLE 4. Viable counts of bacteria in sediments disproportionating S<sup>0</sup> in the presence of MnO<sub>2</sub> or FeOOH<sup>a</sup>

Site depth	MnO <sub>2</sub> (cells per cm <sup>3</sup> )			FeOOH (salt marsh;	
(cm)	Aarhus Bay	Salt marsh	rsh Skagerrak	cells per cm <sup>3</sup> )	
0-1	$4.3 \times 10^{4}$	$2.4 \times 10^{5}$	93	$2.4 \times 10^{5}$	
1–2 2–3 <sup>b</sup> 4–6	$9.3 \times 10^4$ $1.5 \times 10^4$	$2.4 \times 10^{5}$ $4.6 \times 10^{4}$	4 23	$4.6 \times 10^{5}$ >1 × 10 <sup>6</sup> 1.1 × 10 <sup>6</sup>	

' The 95% confidence interval is 0.14 to 5.4 times the MPN or less.

<sup>b</sup> The site depth was 2 to 4 cm for FeOOH in this sampling.

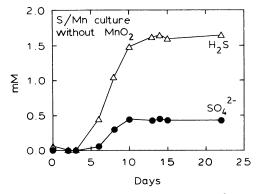


FIG. 4. Time course of concomitant  $H_2S$  and  $SO_4^{2-}$  production in the S/Mn enrichment culture in the absence of MnO<sub>2</sub>.

A complete spontaneous oxidation of  $H_2S$  to  $SO_4^{2-}$  as recently reported with hematite  $(\alpha - Fe_2O_3)$  (13) is not likely since this pathway was found to be dependent on undersaturation with respect to FeS, whereas FeS precipitated in our cultures.

A stoichiometric microbial disproportionation of S<sup>0</sup> to H<sub>2</sub>S and  $SO_4^{2-}$  is described by equation 2:

$$4S^{0} + 4H_{2}O \rightarrow 3H_{2}S + SO_{4}^{2-} + 2H^{+}$$
(2)

The combination of equations 1 and 2 yields the overall reaction for the disproportionation of S<sup>0</sup> in the presence of FeOOH:

$$3S^0 + 2FeOOH \rightarrow SO_4^{2-} + 2FeS + 2H^+$$
(3)

The determined ratio of Fe(III) to  $SO_4^{2-}$  of 2 (Table 1), the observed precipitation of iron sulfide, and the decrease of pH in the medium of the S/Fe culture are all in agreement with this equation, and we thus conclude that this is indeed the net reaction that occurred in this culture.

Accumulation of both AVS (H<sub>2</sub>S and FeS) and CRS was observed (Table 2). Since x-ray diffraction analysis indicated the presence of pyrite in the culture, it is likely that this compound mainly contributed to the CRS fraction when S<sup>0</sup> had been removed. The acid-resistant grey coatings observed on sulfur grains could be pyrite, similar to that described by Berner (5). Pyrite may form by reaction of FeS and  $S^0$  or by reaction of  $Fe^{2+}$  with polysulfides (5, 40). Since only half of the sulfur in pyrite is in the oxidation state of sulfide, S(-II), the amount of sulfide accumulated in the cultures can be calculated as the sum of AVS and ( $\frac{1}{2} \times CRS$ ). Thus, the ratio of S(-II) to  $SO_4^{2-}$  accumulated in the two cultures in Table 2 was 1.54 on average; i.e., 23% less S(-II) was recovered than theoretically expected (equation 3). This discrepancy is probably due to difficulties in sampling the rapidly settling sulfur grains on which iron sulfides were seen as coatings.

When the S/Fe culture was incubated with FeCO<sub>3</sub> as an alternative scavenger of  $H_2S$ , the calculated ratio of Fe(II) to  $SO_4^{2-}$  (Table 2) agreed well with that predicted from the combination of  $S^0$  disproportionation (equation 2) and FeS precipitation:

$$4S^0 + 3FeCO_3 + 4H_2O \rightarrow \qquad (4)$$

$$SO_4^{2-} + 3FeS + 3HCO_3^{-} + 5H^+$$

In the cultures which received the highest amounts of FeCO<sub>3</sub>, almost all sulfide was found as CRS. We have not identified the compound contributing to this fraction. Assuming it to be pyrite, the ratio of S(-II) to  $SO_4^{2-}$  was 1.92 on average for the two cultures presented in Table 2, 65% of that expected from equation 4. Particle discrimination in sampling might again explain part of the discrepancy. However, the measurements of iron released during chromium reduction (Fig. 2) indicate an iron sulfide of a composition other than pyrite. These iron concentrations should be equal to the amount of S(-II) bound in the CRS fraction. The ratio of iron released to sulfur recovered during the chromium reduction step at the end of the time course was 0.74, which corresponds to the Fe/S ratio of greigite (Fe<sub>3</sub>S<sub>4</sub>) rather than that of pyrite. If this was indeed the composition of the ferrous sulfide in the CRS fraction, the ratio of S(-II) to  $SO_4^{2-}$  is 2.91, which is close to 3, the expectation from equation 4.

The S/Mn culture produced  $Mn^{2+}$  and  $SO_4^{2-}$  with an average ratio of 2.7 (Fig. 3 and Table 3). This implies an oxidation of S<sup>0</sup> by MnO<sub>2</sub> by the reaction:

$$S^{0} + 3MnO_{2} + 2H_{2}O \rightarrow SO_{4}^{2-} + 3Mn^{2+} + 4OH^{-}$$
 (5)

The concomitant increase in pH supports this stoichiometry. In the culture,  $Mn^{2+}$  precipitated as  $MnCO_3$ . However, the simultaneous production of sulfate and sulfide observed when the culture was incubated without  $MnO_2$  (Fig. 2) again implies a disproportionation of S<sup>0</sup>. The measured ratio of sulfide to sulfate (3.7 on average) is in the range of that expected from equation 2.

In the presence of  $MnO_2$ ,  $H_2S$  is rapidly oxidized to  $S^0(8)$ :

$$H_2S + MnO_2 \rightarrow S^0 + Mn^{2+} + 2OH^-$$
 (6)

A combination of a biologically catalyzed disproportionation (equation 2) and a spontaneous sulfide oxidation (equation 6) would be stoichiometrically identical to a one-step  $S^0$  oxidation with MnO<sub>2</sub> (equation 5). For that reason, these two pathways are indistinguishable in our experimental setup. However, since sulfur is disproportionated in the S/Mn culture when MnO<sub>2</sub> is absent, it is most likely that  $S^0$  is also disproportionated in the presence of MnO<sub>2</sub>. The H<sub>2</sub>S concentration is kept below detection by the rapid reoxidation. The appearance of H<sub>2</sub>S after oxide depletion supports this interpretation (Fig. 3).

Abiological disproportionation of elemental sulfur to sulfide and sulfate in aqueous solution occurs at temperatures above  $80^{\circ}C$  (6, 20). Biologically catalyzed disproportionation of S<sup>0</sup> to sulfide and sulfate has, to our knowledge, only been reported as a photochemical process in CO<sub>2</sub>-free cultures of *Chlorobium limicola* (35). The authors suggested that *Chlorobium* sp. also disproportionates S<sup>0</sup> in the presence of CO<sub>2</sub> as a first step in the oxidation of S<sup>0</sup> to SO<sub>4</sub><sup>2-</sup>. In analogy to our S/Mn culture, the intermediate formation of H<sub>2</sub>S only becomes visible in the absence of the oxidant, here CO<sub>2</sub>.

**Thermodynamic considerations.** The observed microbial dismutation of elemental sulfur to sulfide and sulfate represents a new chemolithotrophic metabolism. The disproportionation (equation 2) is endergonic at pH 7 and under standard conditions,  $\Delta G^{o'} = 41$  kJ mol<sup>-1</sup> (45). The  $\Delta G$  is, however, strongly dependent on the activities of H<sub>2</sub>S and H<sup>+</sup>. Thus, at very low concentrations of H<sub>2</sub>S,  $\Delta G$  is sufficiently low that energy may be gained through the reaction. The process can therefore sustain bacterial growth as long as the H<sub>2</sub>S is removed, e.g., by reaction with MnO<sub>2</sub>, FeOOH, or FeCO<sub>3</sub>. With activities of H<sub>2</sub>S and SO<sub>4</sub><sup>2-</sup> of 10<sup>-7</sup> and 10<sup>-2</sup> M, respectively,  $\Delta G'$  is -92 kJ mol<sup>-1</sup>. This is sufficient for synthesis of at least 1 mol of ATP (45). The disproportion-

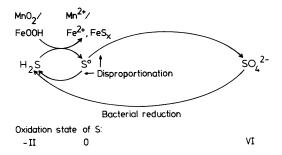


FIG. 5. Possible role of  $S^0$  in anaerobic oxidative pathways of the sulfur cycle. See text for details.

ation of  $S^0$  is thus dragged by the abiological removal of the  $H_2S$  produced.

**Ecological considerations.** Disproportionation of elemental sulfur as reported here is a novel pathway in the biogenic transformations of sulfur. Significant amounts of elemental sulfur are found in both marine and freshwater sediments (42, 46, 47). The concentration is often highest in the suboxic zone of the sediment where S<sup>0</sup> probably derives from abiological oxidation of H<sub>2</sub>S with Mn(IV) or Fe(III) (equations 3 and 5). Although H<sub>2</sub>S is produced at significant rates by sulfate reduction, the pore water concentration of H<sub>2</sub>S in the suboxic zone is often below 1  $\mu$ M (9). Thus, the suboxic environment is relatively similar to that of the growing enrichment cultures described herein, and this is the environment in which disproportionation of elemental sulfur would be expected to be of importance.

The sediment sections used for MPN counts (Table 4) were, except for the oxic surface layer, all from the suboxic zone of the sediments. Thus, the high numbers of  $S^{0}$ -disproportionating bacteria found in the two coastal sediments support the idea that the process could be of importance here. The low numbers in the manganese-rich Skagerrak sediment could indicate that the process is not ubiquitous or that  $S^{0}$ -disproportionating bacteria at this site require different conditions for growth.

The MPN counts for the coastal sediments (Table 4) are comparable to those of both  $S_2O_3^{2-}$ -disproportionating and  $SO_4^{2-}$ -reducing, hydrogen- or acetate-oxidizing bacteria found in Danish marine sediments (26). The authors suggested that the different sulfur transformations are carried out by largely the same groups of bacteria. This could also be true for S<sup>0</sup> disproportionation.

Elemental sulfur may have a role similar to that of  $S_2O_3^{2^-}$ as an intermediate in the anoxic transformations of sulfur in sediments (24). In analogy to the thiosulfate shunt, a small cycle may exist between  $H_2S$  and  $S^0$  on the basis of  $H_2S$ oxidation by iron or manganese oxides and bacterial  $S^0$ disproportionation (Fig. 5). For every 4 mol of  $H_2S$  oxidized to  $S^0$ , 1 mol of S is transferred to the  $SO_4^{2^-}$  pool (equation 2). This pathway could explain the observations of anoxic oxidation of <sup>35</sup>S-labeled  $H_2S$  to  $SO_4^{2^-}$  (e.g., see references 16 and 26). Even in the absence of an oxidant for  $H_2^{35}S$ , <sup>35</sup>S<sup>0</sup> may form rapidly after  $H_2^{35}S$  addition by isotope exchange reactions (17, 18). In the sediment studies,  $S_2O_3^{2^-}$  was found to be the major immediate oxidation product. We cannot exclude  $S_2O_3^{2^-}$  as a rapidly metabolized intermediate in our enrichment cultures.

A major difference between the effects of Fe and Mn on the pathways in Fig. 5 is the loss of sulfur through FeS and FeS<sub>2</sub> precipitation. When sulfur is limiting and FeS or FeS<sub>2</sub> precipitates, the  $SO_4^{2-}$  production is expected to be much smaller when coupled to FeOOH reduction rather than to  $MnO_2$  reduction. Working with salt marsh sediments, Aller and Rude (1) found  $SO_4^{2-}$  production by biological catalysis after  $MnO_2$  addition, while the  $SO_4^{2-}$  concentration only increased slightly after the addition of amorphous FeOOH. Bacterial disproportionation of S<sup>0</sup> might explain these observations.

In unmanipulated sediments, processes are often in a steady state, where substrates are supplied and metabolites are removed continuously. Thus, iron reduction might not lead to supersaturation with respect to FeS or  $FeS_2$  and precipitation of these. Therefore, the efficiency of ferric iron as an oxidant for sulfide, e.g., by the pathway suggested here, might be greater than indicated by batch-type experiments (1, 14, 28).

In conclusion, another twist has been added to the sulfur cycle. The disproportionation of  $S^0$  coupled to Mn or Fe reduction can explain at least some of the observations of anaerobic sulfide oxidation in sediments. The discovery raises new questions concerning both the ecological significance and the microbiology of the process.

### **ACKNOWLEDGMENTS**

We thank Bo Barker Jørgensen for support throughout this study and Nissemanden for laboratory assistance. S. E. Rasmussen kindly performed X-ray diffraction analyses.

This work was supported by the Danish Agency of Environmental Protection through Marine Research Programme 90 in Denmark. During part of the present study, F.B. received a research fellowship from the Deutsche Forschungsgemeinschaft.

#### REFERENCES

- Aller, R. C., and P. D. Rude. 1988. Complete oxidation of solid phase sulfides by manganese and bacteria in anoxic marine sediments. Geochim. Cosmochim. Acta 52:751–765.
- 2. American Public Health Association. 1969. Standard methods for the examination of water and wastewater, including bottom sediments and sludge, p. 604–609. American Public Health Association, Washington, D.C.
- Bak, F., and N. Pfennig. 1987. Chemolithotrophic growth of Desulfovibrio sulfodismutans sp. nov. by disproportionation of inorganic sulfur compounds. Arch. Microbiol. 147:184–189.
- Bak, F., and N. Pfennig. 1991. Microbial sulfate reduction in littoral sediment of Lake Constance. FEMS Microbiol. Ecol. 85:31-42.
- 5. Berner, R. A. 1970. Sedimentary pyrite formation. Am. J. Sci. 268:1-23.
- 6. Belkin, S., C. O. Wirsen, and H. W. Jannasch. 1985. Biological and abiological sulfur reduction at high temperatures. Appl. Environ. Microbiol. 49:1057–1061.
- 7. Burdige, D. J. 1983. The biogeochemistry of manganese redox reactions: rates and mechanisms. Ph.D. thesis. University of California, San Diego.
- Burdige, D. J., and K. H. Nealson. 1986. Chemical and microbiological studies of sulfide-mediated manganese reduction. Geomicrobiol. J. 4:361-387.
- 9. Canfield, D. E. 1989. Reactive iron in sediments. Geochim. Cosmochim. Acta 53:619-632.
- Canfield, D. E., B. Thamdrup, and J. W. Hansen. Submitted for publication.
- Cline, J. D. 1969. Spectrophotometric determination of hydrogen sulfide in natural waters. Limnol. Oceanogr. 14:454–458.
- Cypionka, H., and N. Pfennig. 1986. Growth yields of *Desulfo-tomaculum orientis* with hydrogen in chemostat culture. Arch. Microbiol. 143:396–399.
- dos Santos Afonso, M., and W. Stumm. 1992. The reductive dissolution of iron(III) (hydr) oxides by hydrogen sulfide. Langmuir 8:1671-1676.
- 14. Elsgaard, L., and B. B. Jørgensen. 1992. Anoxic transformations

of radiolabeled hydrogen sulfide in marine and freshwater sediments. Geochim. Cosmochim. Acta 56:2425-2436.

- 15. Fossing, H., and B. B. Jørgensen. 1989. Measurement of bacterial sulfate reduction in sediments: evaluation of a single-step chromium reduction method. Biogeochemistry 8:223-245.
- Fossing, H., and B. B. Jørgensen. 1990. Oxidation and reduction of radiolabeled inorganic sulfur compounds in an estuarine sydiment, Kysing Fjord, Denmark. Geochim. Cosmochim. Acta 54:2731-2742.
- 17. Fossing, H., and B. B. Jørgensen. 1990. Isotope exchange reactions with radiolabeled sulfur compounds in anoxic seawater. Biogeochemistry 9:223-245.
- Fossing, H., S. Thode-Andersen, and B. B. Jørgensen. 1992. Sulfur isotope exchange between <sup>35</sup>S-labeled inorganic sulfur compounds in anoxic sediments. Mar. Chem. 38:117-132.
- Froelich, P. N., G. P. Klinkhammer, M. L. Bender, N. A. Luedtke, G. R. Heath, D. Cullen, P. Dauphin, D. Hammond, B. Hartman, and V. Maynard. 1979. Early oxidation of organic matter in pelagic sediments of the eastern equatorial Atlantic: suboxic diagenesis. Geochim. Cosmochim. Acta 43:1075-1090.
- Giggenbach, W. F. 1974. Equilibria involving polysulfide ions in aqueous sulfide solutions up to 240°C. Inorg. Chem. 13:1724– 1730.
- Henrichs, S. M., and W. S. Reeburgh. 1987. Anaerobic mineralization of marine organic matter: rates and the role of anaerobic processes in the ocean carbon economy. Geomicrobiol. J. 5:191-237.
- 22. Howarth, R. W. 1984. The ecological significance of sulfur in the energy dynamics of salt marsh and coastal marine sediments. Biogeochemistry 1:5-27.
- 23. Jørgensen, B. B. 1983. Processes at the sediment-water interface, p. 477-509. In B. Bolin and R. B. Cook (ed.), The major biogeochemical cycles and their interactions. John Wiley & Sons, Inc., New York.
- Jørgensen, B. B. 1990. A thiosulfate shunt in the sulfur cycle of marine sediments. Science 249:152–154.
- 25. Jørgensen, B. B. 1990. The sulfur cycle of freshwater sediments: role of thiosulfate. Limnol. Oceanogr. 35:1329-1343.
- Jørgensen, B. B., and F. Bak. 1991. Pathways and microbiology of thiosulfate transformations and sulfate reduction in a marine sediment (Kattegat, Denmark). Appl. Environ. Microbiol. 57: 847-856.
- Kelly, D. P. 1988. Oxidation of sulfur compounds, p. 65–98. *In* J. A. Cole and S. J. Ferguson (ed.), The nitrogen and sulfur cycles. Cambridge University Press, Cambridge.
- King, G. M. 1990. Effects of added manganic and ferric oxides on sulfate reduction and sulfide oxidation in intertidal sediments. FEMS Microbiol. Ecol. 73:131-138.
- Kuenen, J. G., L. A. Robertson, and H. Van Gemerden. 1985. Microbial interactions among aerobic and anaerobic sulfuroxidizing bacteria, p. 1–59. *In* K. C. Marshall (ed.), Advances in microbial ecology, vol. 8. Plenum Publishing Corp., New York.
- Lovley, D. R., and E. J. P. Phillips. 1986. Organic matter mineralization with reduction of ferric iron in anaerobic sediments. Appl. Environ. Microbiol. 51:683-689.
- Lovley, D. R., and E. J. P. Phillips. 1988. Novel mode of microbial energy metabolism: organic carbon oxidation coupled to dissimilatory reduction of iron and manganese. Appl. Environ. Microbiol. 54:1472-1480.
- Lovley, D. R., J. F. Stolz, G. L. Nord, Jr., and E. J. P. Phillips. 1987. Anaerobic production of magnetite by a dissimilatory iron-reducing microorganism. Nature (London) 330:252–254.
- Murray, J. W., L. S. Balistrieri, and B. Paul. 1984. The oxidation state of manganese in marine sediments and ferromanganese nodules. Geochim. Cosmochim. Acta 48:1237–1247.
- Nelson, D. C., N. P. Revsbech, and B. B. Jørgensen. 1986. Microoxic-anoxic niche of *Beggiatoa* spp.: microelectrode survey of marine and freshwater strains. Appl. Environ. Microbiol. 52:161-168.
- Paschinger, H., J. Paschinger, and H. Gaffron. 1974. Photochemical disproportionation of sulfur into sulfide and sulfate by *Chlorobium limicola* forma *thiosulfatophilum*. Arch. Microbiol. 96:341-351.

- Phillips, E. J. P., and D. R. Lovley. 1987. Determination of Fe(III) and Fe(II) in oxalate extracts of sediments. Soil Sci. Soc. Am. J. 51:938-941.
- 37. Postgate, J. R. 1984. The sulfate reducing bacteria, 2nd ed. Cambridge University Press, Cambridge.
- Pyzik, A. J., and S. E. Sommer. 1981. Sedimentary iron monosulfides: kinetics and mechanism of formation. Geochim. Cosmochim. Acta 45:687-698.
- Rickard, D. T. 1974. Kinetics and the mechanism of sulfidation of goethite. Am. J. Sci. 274:941–952.
- 40. Rickard, D. T. 1975. Kinetics and mechanisms of pyrite formation at low temperatures. Am. J. Sci. 275:636-652.
- 41. Skyring, G. W. 1987. Sulfate reduction in coastal ecosystems. Geomicrobiol. J. 5:295-374.
- Smith, R. L., and M. J. Klug. 1981. Reduction of sulfur compounds in the sediment of a eutrophic lake basin. Appl. Environ. Microbiol. 41:1230-1237.
- Sørensen, J., and B. B. Jørgensen. 1987. Early diagenesis in sediments from Danish coastal waters: microbial activity and Mn-Fe-S geochemistry. Geochim. Cosmochim. Acta 51:1583–

1590.

- 44. Stookey, L. L. 1970. Ferrozine—a new spectrophotometric reagent for iron. Anal. Chem. 42:779–781.
- Thauer, R. K., K. Jungermann, and K. Decker. 1977. Energy conservation in chemotrophic anaerobic bacteria. Bacteriol. Rev. 41:100–180.
- 46. Thode-Andersen, S., and B. B. Jørgensen. 1989. Sulfate reduction and formation of <sup>35</sup>S-labeled FeS, FeS<sub>2</sub>, and S<sup>0</sup> in coastal marine sediments. Limnol. Oceanogr. 34:793–806.
- 47. Troelsen, H., and B. B. Jørgensen. 1982. Seasonal dynamics of elemental sulfur in two coastal sediments. Estuarine Coastal Shelf Sci. 15:255-266.
- Widdel, F. 1988. Microbiology and ecology of sulfate- and sulfur-reducing bacteria, p. 469–586. *In* A. Zehnder (ed.), Biology of anaerobic microorganisms. John Wiley & Sons, Inc., New York.
- Widdel, F., and F. Bak. 1991. Gram-negative mesophilic sulfatereducing bacteria, p. 3352–3378. In A. Balows et al. (ed.), The procaryotes, 2nd ed., vol 4. Springer-Verlag, New York.