# Denitrification, Acetylene Reduction, and Methane Metabolism in Lake Sediment Exposed to Acetylene

ROGER KNOWLES

Department of Microbiology, Macdonald Campus of McGill University, Ste. Anne de Bellevue, Quebec, H9X 1C0 Canada

# **Received for publication 10 June 1979**

Samples of sediment from Lake St. George, Ontario, Canada, were incubated in the laboratory under an initially aerobic gas phase and under anaerobic conditions. In the absence of added nitrate (NO<sub>3</sub><sup>-</sup>) there was O<sub>2</sub>-dependent production of nitrous oxide (N<sub>2</sub>O), which was inhibited by acetylene (C<sub>2</sub>H<sub>2</sub>) and by nitrapyrin, suggesting that coupled nitrification-denitrification was responsible. Denitrification of added NO<sub>3</sub><sup>-</sup> was almost as rapid under an aerobic gas phase as under anaerobic conditions. The N<sub>2</sub>O that accumulated persisted in the presence of 0.4 atm of C<sub>2</sub>H<sub>2</sub>, but was gradually reduced by some sediment samples at lower C<sub>2</sub>H<sub>2</sub> concentrations. Low rates of C<sub>2</sub>H<sub>2</sub> reduction were observed in the dark, were maximal at 0.2 atm of C<sub>2</sub>H<sub>2</sub>, and were decreased in the presence of O<sub>2</sub>, NO<sub>3</sub><sup>-</sup>, or both. High rates of light-dependent C<sub>2</sub>H<sub>2</sub> reduction occurred under anaerobic conditions. Predictably, methane (CH<sub>4</sub>) production, which occurred only under anaerobiosis, was delayed by added NO<sub>3</sub><sup>-</sup> and inhibited by C<sub>2</sub>H<sub>2</sub>. Consumption of added CH<sub>4</sub> occurred only under aerobic conditions and was inhibited by C<sub>2</sub>H<sub>2</sub>.

The acetylene ( $C_2H_2$ ) inhibition method for the measurement of denitrification (2, 12, 29) was used for soil (28) and was further validated by <sup>13</sup>N (21) and <sup>15</sup>N (19) experiments. It has also been applied to sediments (15, 22), and subsequently an in situ field method for sedimentwater systems was described in which denitrification and  $C_2H_2$  reduction (N<sub>2</sub> fixation) were measured simultaneously (7). However, there is little information on the effects of O<sub>2</sub>, nitrate (NO<sub>3</sub><sup>--</sup>), methane (CH<sub>4</sub>), and light on sediment processes during a "C<sub>2</sub>H<sub>2</sub> inhibition" assay.

The present paper describes the first laboratory studies of the production and reduction of nitrous oxide (N<sub>2</sub>O), reduction of C<sub>2</sub>H<sub>2</sub>, and production and consumption of CH<sub>4</sub> by samples of lake sediment in the presence and absence of C<sub>2</sub>H<sub>2</sub>, O<sub>2</sub>, NO<sub>3</sub><sup>-</sup>, CH<sub>4</sub>, and light.

### **MATERIALS AND METHODS**

Sediments were collected in June 1977 from 5- and 14-m depths in the eastern basin of Lake St. George, a small eutrophic lake in Richmond Hill County, north of Toronto, Ontario, Canada. The areas studied were free from macrophytes. From each of the areas sampled, 15 48-mm-diameter cores were taken using a Kajak-Brinkhurst core sampler. Sediment was gently extruded, and all the 0- to 5-cm and all the 5- to 10-cm depths were combined in two separate containers. For some experiments, 0- to 10-cm depths were combined. The sediments were gently mixed without introduction of air and stored at about 4°C. Physical characteristics of the sediments were as shown in Table 1. Ten-gram portions of fresh sediment were placed in 50-ml Erlenmeyer flasks. Solutions of NaNO<sub>3</sub> (0.5 ml to give 2  $\mu$ mol/g of sediment) and nitrapyrin (0.5 ml to give 20  $\mu$ g/g of sediment) were added as required. The flasks were closed with serum stoppers (Suba-Seal, England), and the atmospheres were either left airfilled (i.e., initially aerobic) or made anaerobic by evacuating and refilling to 1 atm three times with He. Appropriate amounts of C<sub>2</sub>H<sub>2</sub>, C<sub>2</sub>H<sub>4</sub>, CH<sub>4</sub>, and N<sub>2</sub>O were added by means of a syringe through the stoppers after removing an equivalent volume of the gas phase. Flasks were incubated statically in the dark or, if so indicated, under approximately 500 k from fluorescent lights at 20°C for up to 21 days.

At desired intervals, 0.2-ml samples of the gas phase were removed by means of a 1-ml syringe with Mininert valve (Precision Sampling Corporation, Baton Rouge, La.) and analyzed for CH<sub>4</sub>,  $C_2H_2$ ,  $C_2H_4$ ,  $CO_2$ , and N<sub>2</sub>O by single injection into a split-column gas chromatographic system with flame ionization and thermal conductivity detectors (17). After 8 days of incubation, the concentration of O<sub>2</sub> in initially aerobic flasks was determined by gas chromatography (6), and O<sub>2</sub> was added to replace that consumed. The O<sub>2</sub> was found to be rarely depleted below 0.1 atm. Subsequent reference to such conditions as aerobic or initially aerobic does not imply that the whole of the sediment sample was aerobic, merely that the sample was incubated under an oxygen-containing atmosphere.

Pure gases were obtained from Matheson (Canada) Ltd., and nitrapyrin (N-Serve) was obtained from Dow Chemical Co., Sarnia, Ontario. Data are the means of triplicate flasks and are expressed as micromoles per gram of sediment (fresh weight basis). Nitrous oxide data are corrected for solubility in the liquid phase Vol. 38, 1979

and for leakage as determined by using similarly treated flasks without sediment and supplemented with 10  $\mu$ mol of N<sub>2</sub>O per flask.

#### RESULTS

In the absence of added  $NO_3^-$ , no  $N_2O$  was produced except under an initially aerobic gas phase (henceforth referred to as aerobic) in the absence of  $C_2H_2$  (Fig. 1). This production of  $N_2O$ was inhibited by  $C_2H_2$ , and further information on this phenomenon is presented later. Patterns of  $N_2O$  production from added  $NO_3^-$  were similar under anaerobic and aerobic conditions, although rates were slightly lower in the latter case (Fig. 1). In the absence of  $C_2H_2$ ,  $N_2O$  peaked transiently at 2 to 3 days and then disappeared at 5 to 8 days. Acetylene (0.1 or 0.4 atm) caused rapid accumulation of  $N_2O$  with no subsequent reduction, except under anaerobic conditions with 0.1 atm  $C_2H_2$ , where some reduction occurred after 12 days. The second peak of  $N_2O$  accumulation under aerobic conditions at 16 days is not understood (Fig. 1).

Production of ethylene ( $C_2H_4$ ) did not occur in the absence of  $C_2H_2$  (Fig. 2). Furthermore, no

 
 TABLE 1. Dry weight, loss on ignition, and pH of the Lake St. George sediments used in this study

	Sedi-		Loss	pH"		
Water depth (m)	ment depth (cm)	Dry wt (% fresh wt)	on ig- nition (% dry wt)	Aerobic gas phase	Anaero- bic gas phase	
5	0–5	5.4	30	6.79	7.39	
	5-10	6.7	25	7.19	7.77	
14	0-5	8.6	16	6.61	7.41	
	5-10	11.2	18	6.68	7.30	

" Mean pH values of sediment after 21 days of incubation under an initially aerobic gas phase (aerobic) and under He (anaerobic).

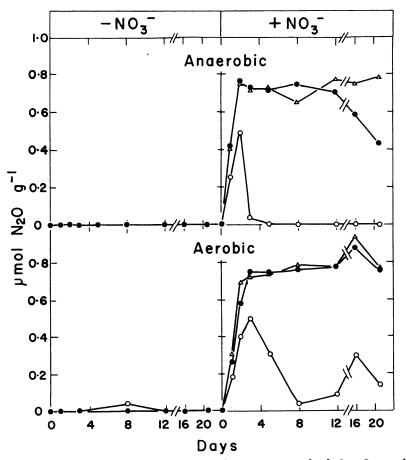


FIG. 1. Production of N<sub>2</sub>O by Lake St. George sediment from 14-m water depth, 0- to 5-cm sediment depth. Samples were incubated statically under He (upper) or air (lower), in the absence (left) and in the presence (right) of 2 µmol of NO<sub>3</sub><sup>-</sup> per g of sediment. The gas phase contained no ( $\bigcirc$ ), 0.1 atm ( $\bigcirc$ ), or 0.4 atm of C<sub>2</sub>H<sub>2</sub> ( $\triangle$ ).

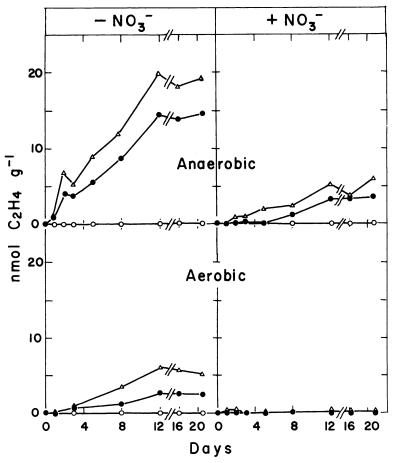


FIG. 2. Production of  $C_2H_4$  by Lake St. George sediment from 14-m water depth, 0- to 5-cm sediment depth. Conditions and symbols as in Fig. 1.

metabolism of an added 40 nmol of  $C_2H_4$  per g of sediment occurred under either anaerobic or aerobic conditions (data not presented). Most  $C_2H_4$  was produced under anaerobic conditions without  $NO_3^-$  supplement. No  $C_2H_4$  was produced under aerobic conditions with  $NO_3^-$ , and the other conditions gave intermediate rates (Fig. 2). Slightly greater rates of  $C_2H_4$  production were observed with 0.4 atm than with 0.1 atm  $C_2H_2$ , and further data on this aspect will be presented later.

As was expected, release of CH<sub>4</sub> into the gas phase did not occur under aerobic conditions, but did occur under anaerobiosis (Fig. 3). In the absence of  $NO_3^-$ , production of CH<sub>4</sub> occurred with no lag and at a constant rate; however, the addition of  $NO_3^-$  caused a lag of about 5 to 8 days before production began. Methane production was completely inhibited by C<sub>2</sub>H<sub>2</sub> (Fig. 3). The base-line levels of CH<sub>4</sub> observed in the presence of  $C_2H_2$  in aerobic flasks (Fig. 3) probably represent CH<sub>4</sub>, already present in the sediment, which equilibrated with the atmosphere, its metabolism being inhibited by  $C_2H_2$ . Such endogenous CH<sub>4</sub> would be removed by the evacuation involved in the creation of anaerobic conditions.

The production of N<sub>2</sub>O and of C<sub>2</sub>H<sub>4</sub> was affected by the concentration of C<sub>2</sub>H<sub>2</sub> to which the sediment was exposed (Fig. 4). The initial period of rapid production of N<sub>2</sub>O from NO<sub>3</sub><sup>-</sup> (e.g., at day 1) was not affected by C<sub>2</sub>H<sub>2</sub> concentration. The smaller amount of N<sub>2</sub>O present after 8 days at C<sub>2</sub>H<sub>2</sub> concentrations lower than 0.4 atm represents the somewhat less effective inhibition of subsequent reduction of C<sub>2</sub>H<sub>4</sub> between 1 and 8 days (Fig. 4) was maximal at 0.2 atm but somewhat submaximal at 0.1 atm of C<sub>2</sub>H<sub>2</sub>.

In similar studies of other samples of sedi-

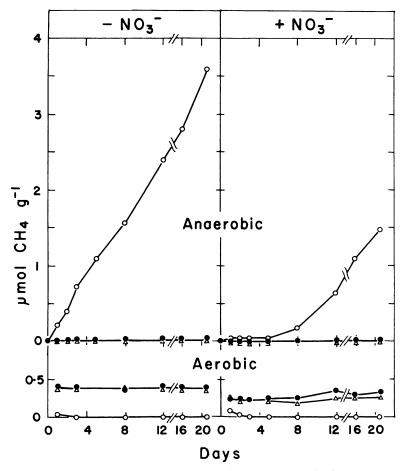


FIG. 3. Production of CH<sub>4</sub> by Lake St. George sediment from 14-m water depth, 0- to 5-cm sediment depth. Conditions and symbols as in Fig. 1.

ments, patterns of production of N<sub>2</sub>O, C<sub>2</sub>H<sub>4</sub>, and CH<sub>4</sub> were similar to those shown in Fig. 1 to 3. Rates calculated from the straight-line parts of such curves are summarized in Table 2. Denitrification potential (N<sub>2</sub>O production from NO<sub>3</sub><sup>-</sup> in the presence and absence of C<sub>2</sub>H<sub>2</sub>) was somewhat greater in the 5-m-depth sediments. The subsequent reduction of accumulated N<sub>2</sub>O in the presence of  $C_2H_2$  was negligible except in the 5m surface sediment, where it was appreciable. This is also reflected in the maximum amounts of  $N_2O$  observed in the presence of 0.1 atm of  $C_2H_2$  (Table 3). Least  $N_2O$  accumulated in the 5-m surface sediment and most in the 14-m subsurface samples. This relationship likely depends on the efficacy of the  $C_2H_2$  inhibition of  $N_2O$  reduction.

In the absence of a  $NO_3^-$  supplement,  $N_2O$  was produced only in aerobic conditions after a lag of 3 to 5 days (Table 2). Such production of

 $N_2O$  was completely inhibited by  $C_2H_2$ . This  $O_2$ dependent production of  $N_2O$  in the absence of added  $NO_3^-$  was investigated in several further experiments (Table 4). The data were rather variable, but  $N_2O$  was generally produced after a lag of 1 to 8 days in the absence (but not in the presence) of  $C_2H_2$ . The addition of the nitrification inhibitor, nitrapyrin, delayed the appearance of  $N_2O$ , but subsequently the rate of  $N_2O$ production was not very different from that in the controls.

Rates of production of  $C_2H_4$  from  $C_2H_2$  (Table 2) decreased as water depth and sediment depth increased and, as might be expected, were also reduced by aerobiosis and the addition of NO<sub>3</sub><sup>-</sup>. However, all such rates observed during dark incubation were very low.

Methane was produced only under anaerobic conditions, and addition of  $NO_3^-$  induced a lag of 4 to 10 days (Table 2). Rates were greater in

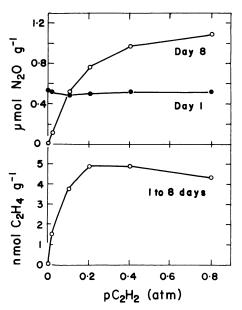


FIG. 4. Nitrous oxide present after 1 and 8 days of incubation and  $C_2H_4$  produced between 1 and 8 days of incubation by Lake St. George sediment from 5-m water depth, 0- to 10-cm sediment depth. Samples were supplemented at zero time with 2 µmol of  $NO_3^-$  per g of sediment and incubated under He with the indicated concentrations of  $C_2H_2$ .

the surface sediments and were negligible in the 14-m subsurface samples. In all cases,  $CH_4$  production was inhibited by 0.1 atm of  $C_2H_2$ .

Light markedly stimulated the rate of C<sub>2</sub>H<sub>2</sub> reduction in 5-m surface sediment (Fig. 5), but this effect was seen only under anaerobiosis. Microscopic examination of such enriched samples revealed cells which appeared to be unicellular cyanobacteria. No purple or green photosynthetic bacteria were seen, although the former occur in the water column of this lake during late stratification (D. R. S. Lean, personal communication). Dark C<sub>2</sub>H<sub>4</sub> production was inhibited by addition of CH4 under both anaerobic and aerobic conditions (Fig. 5), but this phenomenon was not further investigated. Predictably, added CH<sub>4</sub> (3.6  $\mu$ mol/g) was not metabolized under anaerobic conditions or in the presence of  $C_2H_2$ . It was rapidly oxidized under aerobiosis in the absence of C<sub>2</sub>H<sub>2</sub> and had completely disappeared within 3 days (Fig. 5).

# DISCUSSION

The lack of  $N_2O$  production by anaerobically incubated sediment in the absence of added  $NO_3^-$  indicates that there was no endogenous  $NO_3^-$  present at the time of the experiments. This is the first report of the production of  $N_2O$ by sediment under an aerobic gas phase in the

depth dept	Sedi-	ND (2	O <sub>3</sub> (2 Gas phase	N <sub>2</sub> O (nmol/g per day)				C₂H₄ produc- tion	CH₄ production (nmol/g per day) 0– 21 days	
	depth (cm)			Production 0-2 days		Reduction 2-12 days		(pmol/g per day) 0–3 days	$-C_2H_2$	$+C_2H_2$
				$-C_2H_2$	$+C_2H_2$	$-C_2H_2$	$+C_2H_2$	$(+C_2H_2)$		
5	0-5	-	Anaerobic	0	0	_		2,700	130	0
			Aerobic	0	0	_	_	400	0	0
		+	Anaerobic	315	530	320	96	400	244 (6)	0
			Aerobic	260	390	51	0	0	0	0
	5-10	-	Anaerobic	0	0		_	2,100	18 (4)	0
			Aerobic	9 (5)	0	0	—	70	0	0
		+	Anaerobic	460	500	-8	0	300	27 (10)	0
			Aerobic	250	340	0	0	0	0	0
14	14 0-5	_	Anaerobic	0	0		_	1,300	210	0
			Aerobic	8 (3)	0	100		400	0	0
		+	Anaerobic	220	350	430	4	200	105 (6)	0
			Aerobic	170	260	84	-1	0	0	0
	5-10	-	Anaerobic	0	0	_	—	700	3	0
			Aerobic	16 (3)	0	12		320	0	0
		+	Anaerobic	360	370	180	0	0	0	0
			Aerobic	240	280	107	0	0	0	0

TABLE 2. Rates of production and subsequent reduction of  $N_2O$  and of production of  $C_2H_4$  and  $CH_4$  in the presence and absence of 0.1 atm of  $C_2H_2$  by Lake St. George sediments<sup>a</sup>

<sup>a</sup> Data are reported on a sediment fresh weight basis. —, Not determined or not applicable. Figures in parentheses indicate the number of days of lag before initiation of the activity reported.

TABLE 3. Maximum amounts of  $N_2O$  observed up to 21 days after addition of 2 µmol of  $NO_3^-$  per g of sediment in the presence of 0.1 atm of  $C_2H_2$ 

Weter Janet	0.1	N <sub>2</sub> O (µmol/g)			
Water depth (m)	Sediment depth (cm)	Anaerobic gas phase	Aerobic gas phase		
5	0-5	0.59	0.68		
	5-10	0.65	0.69		
14	0-5	0.78	0.94		
	5-10	0.84	1.03		

TABLE 4. Rates of production of  $N_2O$  by Lake St. George sediments incubated in air in the presence and absence of nitrapyrin and  $C_2H_2^a$ 

	Sedi-	N <sub>2</sub> O (nmol/g per day)				
Water depth (m)	ment depth (cm)	Control	+Nitra- pyrin (20 μg/g)	+C <sub>2</sub> H <sub>2</sub> (0.1 atm)		
5	0-10	73 (8)	62 (12)	0		
		30 (1-7)	37 (12)	0		
	0-5	0	0	0		
		17 (5)		0		
	5-10	9 (5)	9 (8)	0		
14	0-5	8 (3)	41 (8)	0		
		0	_	0		
	5-10	16 (3)	62 (8)	0		

<sup>a</sup> Data are reported on a sediment fresh weight basis. No  $NO_3^-$  was added to any of these samples. -, Not determined. Figures in parentheses indicate the number of days of lag before initiation of the reported activity.

absence of added  $NO_3^-$ . It suggests that coupled nitrification and denitrification occurred due to the existence of an aerobic-anaerobic interface within the sediment. The fact that this N<sub>2</sub>O formation was inhibited by C<sub>2</sub>H<sub>2</sub> and delayed by nitrapyrin is consistent with this interpretation, since C<sub>2</sub>H<sub>2</sub> inhibits nitrification of NH<sub>4</sub><sup>+</sup> by Nitrosomonas europaea (13, 27), as does nitrapyrin (3). The latter compound, however, is difficult to apply uniformly in experimental work (5) and is reported to lose effect after about 7 days in sediments (26). This possibly explains the formation of N<sub>2</sub>O after at least 8 days with nitrapyrin under the present conditions.

The denitrification of added  $NO_3^-$  occurred almost as rapidly under an initially aerobic gas phase as under anaerobic conditions, in agreement with the report that  $O_2$  up to 10 mg/liter in the water column did not greatly inhibit the process in sediment (26). Thus, providing lake bottom water contains  $NO_3^-$ , sediment denitrification in anaerobic microenvironments and especially below the aerobic-anaerobic interface is likely to occur rapidly regardless of the dissolved  $O_2$  concentration in the overlying water. The marked but transient accumulation of N<sub>2</sub>O in the absence of  $C_2H_2$  indicates a high mole fraction of N<sub>2</sub>O in the denitrification products during the first 2 days and suggests that the N<sub>2</sub>O-reducing system here is quite sensitive to NO<sub>3</sub><sup>-</sup> concentrations of the order of  $1 \mu mol/g$  of fresh sediment, as has been shown for other systems (4, 18). In many of the present experiments (especially with shallow water and surface sediments), although the  $N_2O$  accumulation in the presence of 0.1 atm of C<sub>2</sub>H<sub>2</sub> probably reflected total denitrification during the first 2 or 3 days. it did not subsequently represent complete conversion of the added NO<sub>3</sub><sup>-</sup>. The data suggest that this was partly due to the incomplete inhibition of N<sub>2</sub>O reduction by C<sub>2</sub>H<sub>2</sub> at concentrations of the order of 0.1 to 0.2 atm. However, it may also reflect some dissimilatory or assimilatory reduction of  $NO_3^-$  to  $NH_4^+$ , which has been observed in some sediments (16, 23).

High rates of  $C_2H_2$  reduction were seen in anaerobically incubated shallow-water surface sediment in the light. This activity was attributed to unicellular cyanobacteria which would not be inhibited by the O<sub>2</sub> they produce at the low illumination employed (25). Acetylene reduction rates in the dark were highest in shallow water and in surface sediments under anaerobic conditions. As was expected, activity was inhibited somewhat by an aerobic atmosphere and by the addition of 2  $\mu$ mol of NO<sub>3</sub><sup>-</sup> per g. Partial to complete inhibition of sediment C<sub>2</sub>H<sub>2</sub> reduction was also reported for concentrations in the range 0.2 to 10 mM  $NO_3^{-}$  (14). Data showed that maximum rates of C<sub>2</sub>H<sub>4</sub> production were observed with 0.2 atm of  $C_2H_2$ . Much higher concentrations were reported to be necessary for maximum rates in other sediments (e.g., R. Sylvester-Bradley, Ph.D. thesis, University of Edinburgh, Edinburgh, Scotland, 1976).

The concentrations of  $C_2H_2$  required to inhibit N<sub>2</sub>O production completely and to saturate the nitrogen-fixing system appear to depend on the nature of the sediment. Concentrations in equilibrium with a gas phase containing 0.1 to 0.2 atm of  $C_2H_2$  seem to be adequate, particularly in short-term experiments. The high rates of light-dependent  $C_2H_2$  reduction observed would not significantly deplete the concentration of  $C_2H_2$  introduced for a denitrification assay.

As might be expected, methane was produced with little or no lag under anaerobic conditions by all except the deep-water subsurface sediments. The addition of  $NO_3^-$  imposed a 6- to 10day lag, as has been observed by others (1, 8). Added CH<sub>4</sub> was metabolized more rapidly under aerobic conditions than this gas was produced under anaerobic conditions. This indicates that

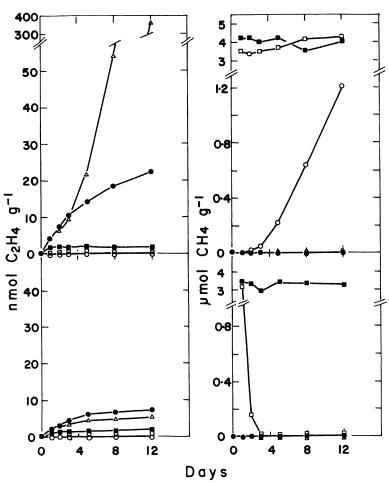


FIG. 5. Production of  $C_2H_4$  from  $C_2H_2$  (left) and metabolism of  $CH_4$  (right) by Lake St. George sediment from 5-m water depth, 0- to 5-cm sediment depth. Samples did not receive  $NO_3^-$  and were incubated under He (upper) or air (lower). Other conditions were: no  $C_2H_2$  ( $\bigcirc$ ), 0.1 atm of  $C_2H_2$  ( $\bigoplus$ ), 3.6 µmol of  $CH_4$  per g of sediment ( $\square$ ),  $C_2H_2 + CH_4$  ( $\blacksquare$ ), and  $C_2H_2 + light$  ( $\triangle$ ).

the lack of release of CH<sub>4</sub> by sediment under air could have been due to the greater potential for consumption than for production. Both production and consumption of CH<sub>4</sub> were completely inhibited by 0.1 atm of  $C_2H_2$ , as was previously reported (11, 20).

The use of  $C_2H_2$  in assays of denitrification and  $N_2$  fixation thus clearly inhibits activity of any organisms whose growth is supported by CH<sub>4</sub> (9, 11). Denitrifying bacteria supported by CH<sub>4</sub> (10) and by methanol (24) have been reported, and methanotrophic  $N_2$  fixers have been studied (9, 11). There is little information on the contributions of such organisms in sediments, and the present data do not permit an estimation of the possible methane-supported activity. The  $N_2$  fixation activity and high denitrification activity that were observed, however, occurred in the presence of  $C_2H_2$  and therefore could not be attributed to the activity of methanotrophic bacteria.

#### ACKNOWLEDGMENTS

This work was supported by grants from the Inland Waters Directorate of Environment Canada and the Ministère de l'Education, Québec.

Permission to use the Lake St. George facilities was given by the Canada Centre for Inland Waters through D. R. S. Lean. Daniel Dubreuil and Tat-Yee Tam provided technical assistance, and M. Olson-Parkinson typed the manuscript.

#### LITERATURE CITED

 Balderston, W. L., and W. J. Payne. 1976. Inhibition of methanogenesis in salt marsh sediments and whole-cell suspensions of methanogenic bacteria by nitrogen oxides. Appl. Environ. Microbiol. 32:264-269. Vol. 38, 1979

- Balderston, W. L., B. Sherr, and W. J. Payne. 1976. Blockage by acetylene of nitrous oxide reduction in *Pseudomonas perfectomarinus*. Appl. Environ. Microbiol. 31:504-508.
- Billen, G. 1976. Evaluation of nitrifying activity in sediments by dark <sup>14</sup>C-bicarbonate incorporation. Water Res. 10:51-57.
- Blackmer, A. M., and J. M. Bremner. 1978. Inhibitory effect of nitrate on reduction of N<sub>2</sub>O to N<sub>2</sub> by soil microorganisms. Soil Biol. Biochem. 10:187-191.
- Bremner, J. M., A. M. Blackmer, and L. E. Bundy. 1978. Problems in use of nitrapyrin (N-Serve) to inhibit nitrification in soils. Soil Biol. Biochem. 10:441-442.
- 6. Brouzes, R., C. I. Mayfield, and R. Knowles. 1971. Effect of oxygen partial pressure on nitrogen fixation and acetylene reduction in a sandy loam soil amended with glucose, p. 481-494. In T. A. Lie and E. G. Mulder (ed.), Biological nitrogen fixation in natural and agricultural habitats. Plant Soil, Special Vol. Martinus Nijhoff, The Hague.
- Chan, Y.-K., and R. Knowles. 1979. Measurement of denitrification in two freshwater sediments by an in situ acetylene inhibition method. Appl. Environ. Microbiol. 37:1067-1072.
- Chen, R. L., D. R. Keeney, J. G. Konrad, A. J. Holding, and D. A. Graetz. 1972. Gas production in sediments of Lake Mendota, Wisconsin. J. Environ. Qual. 1:155-157.
- Dalton, H., and R. Whittenbury, 1976. The acetylene reduction technique as an assay for nitrogenase activity in the methane-oxidizing bacterium *Methylococcus cap*sulatus strain Bath. Arch. Mikrobiol. 109:147-151.
- Davies, T. R. 1973. Isolation of bacteria capable of utilizing methane as a hydrogen donor in the process of denitrification. Water Res. 7:575-579.
- De Bont, J. A. M., and E. G. Mulder. 1974. Nitrogen fixation and co-oxidation of ethylene by a methaneutilizing bacterium. J. Gen. Microbiol. 83:113-121.
- Fedorova, R. I., E. I. Milekhina, and N. I. Il'yukhina. 1973. [Evaluation of the method of "gas metabolism" for detecting extraterrestrial life. Identification of nitrogen fixing microorganisms.] Izv. Akad. Nauk SSSR, Ser. Biol. 1973(6):797-806.
- Hynes, R. K., and R. Knowles. 1978. Inhibition by acetylene of ammonia oxidation in *Nitrosomonas europaea*. FEMS Microbiol. Lett. 4:319-321.
- Jäger, D., and D. Werner. 1976. Physiologie und Mikrobiologie der N<sub>2</sub>-Fixierung in Bodenproben des Harkortsees (Ruhrtal). Ber. Dtsch. Bot. Ges. 89:609-630.
- Knowles, R. 1978. Common intermediates of nitrification and denitrification, and the metabolism of nitrous oxide,

p. 367-371. In D. Schlessinger (ed.), Microbiology-1978. American Society for Microbiology, Washington, D.C.

- Koike, I., and A. Hattori. 1978. Denitrification and ammonia formation in anaerobic coastal sediments. Appl. Environ. Microbiol. 35:278-282.
- Nelson, L. M., and R. Knowles. 1978. Effect of oxygen and nitrate on nitrogen fixation and denitrification by *Azospirillum brasilense* grown in continuous culture. Can. J. Microbiol. 24:1395-1403.
- Nömmik, H. 1956. Investigation on denitrification in soil. Acta Agric. Scand. 6:195-228.
- Paul, E. A., and R. L. Victoria. 1978. Nitrogen transfer between the soil and the atmosphere, p. 525-541. *In* W. Krumbein (ed.), Proceedings of the 3rd Symposium on Environmental Biogeochemistry, Wolfenbeutel, Germany. Ann Arbor Science, Ann Arbor, Mich.
- Raimbault, M. 1975. Etude de l'influence inhibitrice de l'acétylene sur la formation biologique du méthane dans un sol de rizière. Ann. Microbiol. (Inst. Pasteur) 126A: 247-258.
- Smith, M. S., M. K. Firestone, and J. M. Tiedje. 1978. The acetylene inhibition method for short-term measurement of soil denitrification and its evaluation using nitrogen-13. Soil Sci. Soc. Am. J. 42:611-615.
- Sørensen, J. 1978. Denitrification rates in a marine sediment as measured by the acetylene inhibition technique. Appl. Environ. Microbiol. 36:139-143.
- Sorensen, J. 1978. Capacity for denitrification and reduction of nitrate to ammonia in a coastal marine sediment. Appl. Environ. Microbiol. 35:301-305.
- Sperl, G. T., and D. S. Hoare. 1971. Denitrification with methanol. Selective enrichment for *Hyphomicrobium* species. J. Bacteriol. 108:733-736.
- Stewart, W. D. P. 1973. Nitrogen fixation by photosynthetic microorganisms. Annu. Rev. Microbiol. 27:283– 316.
- Van Kessel, J. F. 1977. Factors affecting the denitrification rate in two water-sediment systems. Water Res. 11:259-267.
- Walter, H. M., D. R. Keeney, and I. R. Fillery. 1979. Inhibition of nitrification by acetylene. Soil Sci. Soc. Am. J. 43:195-196.
- Yoshinari, T., R. Hynes, and R. Knowles. 1977. Acetylene inhibition of nitrous oxide reduction and measurement of denitrification and nitrogen fixation in soil. Soil Biol. Biochem. 9:177-183.
- Yoshinari, T., and R. Knowles. 1976. Acetylene inhibition of nitrous oxide reduction by denitrifying bacteria. Biochem. Biophys. Res. Commun. 69:705-710.